# Norton Sound Red King Crab Stock Assessment for the fishing year 2022 

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## Executive Summary

1. Stock. Red king crab, Paralithodes camtschaticus, in Norton Sound, Alaska.
2. Catches. This stock supports three important fisheries: summer commercial, winter commercial, and winter subsistence fisheries. Of those, the summer commercial fishery accounts for $85 \%$ of total harvest. The summer commercial fishery started in 1977. Catch peaked in the late 1970s with retained catch of over 2.9 million pounds. Since 1994, the Norton Sound Crab fishery operated as super exclusive. During the 2021 fishery season, 320 crab ( 922 lb .) were harvested in winter commercial and $2,892 \mathrm{crab}(5,784 \mathrm{lb})$ were harvested in the winter subsistence fishery. Summer commercial fishery opened in 2021, but 0 crab ( 0 lb ) were harvested. In total, $\mathbf{3 2 1 2} \mathrm{crab}(\mathbf{6 , 7 6 6} \mathrm{lb})$ were harvested during the 2021 season. This was below ABC of 0.35 million lb, and thus overfishing did not occur during the 2021 season.
3. Stock Biomass. Norton Sound red king crab is monitored not in biomass but in abundance. Abundance of the Norton Sound Red King Crab stock has been monitored by trawl surveys since 1976 by NOAA (1976-1991), NOAA NBS (2010-2019), and ADF\&G (1996-2020). Historical survey abundance of the Norton Sound red king crab of carapace length greater than $63 \mathrm{~mm}(\mathrm{CL}>63 \mathrm{~mm})$ ranged from 1.41 million to 5.9 million crab. In 2021 the survey abundance was 2.4 million crab (CV 0.6) by ADF\&G trawl survey and 2.37 million crab (CV 0.43) by NOAA NBS trawl survey.
4. Recruitment. Recruitment is not monitored directly. It is inferred by the assessment model. The model estimated recruitment since 1980s were around 0.7 million ranging from 0.2 to 1.6 million.
5. Management performance.

Status and catch specifications (million lb.)

| Year | MSST | Biomass <br> (MMB) | GHL | Retained <br> Catch | Total <br> Catch | OFL | ABC |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2018 | 2.41 | 4.08 | 0.30 | 0.31 | 0.34 | 0.43 | 0.35 |
| 2019 | 2.24 | 3.12 | 0.15 | 0.08 | 0.08 | 0.24 | 0.19 |
| 2020 | 2.28 | 3.67 | 0.17 | Conf. | Conf. | 0.29 | 0.21 |
| 2021 | 2.25 | 5.05 | 0.31 | 0.007 | 0.007 | 0.59 | 0.35 |
| $2022^{\mathrm{a}}$ | 2.08 | 5.33 |  |  |  | 0.89 | $0.53,0.40$ |
| $2022^{\mathrm{b}}$ | 2.15 | 4.79 |  |  |  | 0.96 | $0.58,0.43$ |

Notes:
MSST was calculated as $\mathrm{B}_{\mathrm{MSY}} / 2$
OFL-ABC 2018-2020 are retained only
2019, 2020: Total catch equals retained catch. Discards catch was estimated only for summer commercial fishery, but summer commercial fishery did not occur.
2022 MSST, MMB, OFL, and ABC are CPT adopted after Jan 2022 CPT meeting
$2022^{a}$ model $21.0,2022^{b}$ model 21.5
Status and catch specifications (1000t)

| Year | MSST | Biomass <br> (MMB) | GHL | Retained <br> Catch | Total <br> Catch | OFL | ABC |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2018 | 1.09 | 1.85 | 0.13 | 0.14 | 0.15 | 0.20 | 0.16 |
| 2019 | 1.03 | 1.41 | 0.07 | 0.04 | 0.04 | 0.11 | 0.09 |
| 2020 | 1.04 | 1.66 | 0.08 | Conf. | Conf. | 0.13 | 0.09 |
| 2021 | 1.02 | 2.29 | 0.14 | 0.003 | 0.003 | 0.20 | 0.16 |
| $2022^{\mathrm{a}}$ | 0.95 | 2.42 |  |  |  | 0.40 | $0.24,0.18$ |
| $2022^{\mathrm{b}}$ | 0.98 | 2.17 |  |  |  | 0.44 | $0.26,0.20$ |

Conversion to Metric ton: 1 Metric ton $(\mathrm{t})=2.2046 \times 1000 \mathrm{lb}$
$2022^{\mathrm{a}}$ model $21.0,2022^{\mathrm{b}}$ model 21.5

Biomass in millions of pounds

| Year | Tier | BMSY | Current <br> MMB | B/BMSY <br> (MMB) | Fofl | Years to <br> define <br> BMSY | M | ABC <br> Buffer | ABC |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2018 | 4 b | 4.82 | 4.08 | 0.9 | 0.15 | $1980-2018$ | 0.18 | 0.2 | 0.35 |
| 2019 | 4 b | 4.57 | 3.12 | 0.7 | 0.12 | $1980-2019$ | 0.18 | 0.2 | 0.19 |
| 2020 | 4 b | 4.56 | 3.66 | 0.8 | 0.14 | $1980-2020$ | 0.18 | 0.3 | 0.21 |
| 2021 | 4 a | 4.53 | 5.05 | 1.1 | 0.18 | $1980-2021$ | 0.18 | 0.4 | 0.35 |
| $2022^{\mathrm{a}}$ | 4 a | 4.17 | 5.33 | 1.3 | $0.18,0.61$ | $1980-2022$ | $0.18,0.61$ | $0.4,0.55$ | $0.53,0.40$ |
| $2022^{\mathrm{b}}$ | 4 a | 4.30 | 4.79 | 1.1 | $0.26,0.59$ | $1980-2022$ | $0.26,0.59$ | $0.4,0.55$ | $0.58,0.43$ |
| $2022^{\mathrm{a}}$ model 21.0,2022 model 21.5 |  |  |  |  |  |  |  |  |  |

Biomass in 1000t

| Year | Tier | Bmsy | Current <br> MMB | B/BMSy <br> (MMB) | Fofl | Years to <br> define | M | ABC <br> Buffer | ABC |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |


|  |  |  | BMSY |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :---: | :---: | :---: | :---: | :---: |
| 2018 | 4 b | 2.07 | 1.85 | 0.9 | 0.15 | $1980-2018$ | 0.18 | 0.2 | 0.16 |
| 2019 | 4 b | 2.06 | 1.41 | 0.7 | 0.12 | $1980-2019$ | 0.18 | 0.2 | 0.09 |
| 2020 | 4 b | 2.07 | 1.66 | 0.8 | 0.14 | $1980-2020$ | 0.18 | 0.3 | 0.09 |
| 2021 | 4 a | 2.05 | 2.29 | 1.1 | 0.18 | $1980-2021$ | 0.18 | 0.4 | 0.16 |
| $2022^{\mathrm{a}}$ | 4 a | 1.90 | 2.42 | 1.3 | $0.18,0.61$ | $1980-2022$ | $0.18,0.61$ | $0.4,0.55$ | $0.24,0.18$ |
| $2022^{\mathrm{b}}$ | 4 a | 1.95 | 2.17 | 1.1 | $0.26,0.59$ | $1980-2022$ | $0.26,0.59$ | $0.4,0.55$ | $0.26,0.20$ |

6. Probability Density Function of the OFL and memc estimates of model 21.0 and 21.5.

7. The basis for the ABC recommendation.

For Tier 4 stocks, the default maximum ABC is based on $\mathrm{P}^{*}=49 \%$ that is essentially identical to the OFL. Annual ABC buffer is determined by accounting for uncertainties in assessment and model results. However, criteria for determining the mount of ABC buffer are undefined. The SSC chose to use $90 \%$ OFL ( $10 \%$ Buffer) for the NARKC from 2011 to 2014. The buffer was increased to $20 \%$ ( $\mathrm{ABC}=80 \% \mathrm{OFL}$ ) in 20015, to $30 \%(\mathrm{ABC}=70 \% \mathrm{OFL})$ in 2020, and to $40 \%(\mathrm{ABC}=60 \% \mathrm{OFL})$ in 2021.

| Year | ABC Buffer |
| :--- | :--- |
| $2011-2014$ | $10 \%$ |
| $2015-2019$ | $20 \%$ |
| 2020 | $30 \%$ |


| 2021 | $40 \%$ |
| :--- | :--- |
| 2022 |  |

8. A summary of the results of any rebuilding analysis

NSRK is not overfished.

## A. Summary of Major Changes in 2022 assessment model

1. Changes to the management of the fishery.

None. Summer commercial fishery opened but no crabs were harvested.
2. Changes to the input data.

Input data update through 2021:
Winter subsistence, winter and summer commercial crab fishery harvest.
Trawl surveys: abundance, length-shell compositions: ADF\&G, NOAA NBS 2021
Standardized CPUE revision (See Appendix B)
3. Changes to the assessment methodology.

Seven assessment models are compared in this report based on the recommendations by the CPT and SSC:
a. Model 19.0e: with updated data,
b. Model 21.0: Model 19.0e + St. CPUE with 3 qs +2 summer commercial retention probabilities,
c. Model 21.1: Model 21.0 with $\mathrm{M}=0.18$ for all length size classes,
d. Model 21.2: Model 19.0e + St. CPUE data updated with 3 q ,
e. Model 21.3: Model 19.0e +2 summer commercial retention probabilities,
f. Model 21.4: Model 21.0 with M estimated equally for all length size classes
g. Model 21.5: Model 21.0 with M estimated for two length size classes ( $<124 \mathrm{~mm}$, $>123 \mathrm{~mm} \mathrm{CL}$ ).

These model scenarios focus on examination of $M$ values and time blocks of fishery catchability and commercial retained probability.
4. Changes to the assessment results.

Among the seven models, models 21.2 and 21.3 are incremental models between models 19.0 e and 21.0 that corrects standardized summer commercial fishery CPUE and time blocks of retention probability based on re-evaluation of fishery history. Changes in standardized CPUE slightly increased overall abundance, MMB, and biomass but slightly decreased the projected 2022 MMB and OFL. Separating retention probability values
before and after high grading reduced the retention probability of $94-104 \mathrm{~mm}$ size class during the latter period, thus slightly improving model fit to summer commercial retention size proportions. Overall, the changes in model results are very minor for these two models. Model 21.1 with $M=0.18$ for all size classes has the worst fit to the data among the seven models, followed by model 21.4. However, model 21.4 with a much higher estimated M fitted the data considerably better than the model 21.1. Model 21.5 had the best fit of data, slightly better than model 21.0. Both models have the same feature of $M$ : two $M$ values with one higher $M$ value for crab over 123mm CL ( $M=0.18$ \& 0.62 for model 21.0 and $M=0.26 \& 0.59$ for model 21.5). This suggests that the data are explained better by size-dependent $M$ rather than a single M for all lengths.
Models 21.0 and 21.5 have similar estimates of molting probability and selectivity profiles. Both models, similar to the other models, also underestimated size-proportions of larger and oldshell crab from trawl survey and overestimated the proportions of oldshell crab from summer commercial fishery retained data. Model 21.5 has slightly higher estimated mean total abundance and recruits ( 3.4 and 1.4 million) than model 21.0 ( 3.1 and 1.2 million), but the estimated legal abundances and MMB were similar. Mohn's rho values from retrospective analyses are slightly smaller for model 21.5 (0.191) than for model $21.0(0.209)$. Based on the model fits, we recommend model 21.0 or 21.5 for overfishing determination in 2022.

## B. Response to SSC and CPT Comments

## Following the request by SSC, CPT-SSC's requests/review (received in Jan 2021 and Oct 2021) and authors' responses were arranged by topic. Requests are italicized.

## I. NSRKC Biology-Ecology

## Natural Mortality

Revisit natural mortality assumptions. Both the assumed natural mortality for small crab and the larger natural mortality for crab greater than 123 mm CL should be better justified. The author noted that the maximum age observed in the tagging studies was 12 years, which is much lower than the assumed value of 25 years. Further, the " $1 \%$ method" used by the authors to calculate a natural mortality generally provides lower estimates of $M$ than empirical studies (see the tool at Barefoot Ecologist Toolbox for examples).

Authors reply:
Natural mortality $M$ was originally set to be 0.2 for Bering Sea red king crab stock (NPFMC 1998) and was changed to 0.18 with Amendment 24. Under this, $M$ of NSRKC assessment model was set to 0.18 from 0.3 of the initial assessment model. Since the inception of the crab

SAFE and adoption of NSRKC assessment model, the CPT has been requested to revisit $M$ assumptions for NSRKC. All those past attempts suggested that $M$ would be higher than 0.18 and more likely between 0.25 to 0.45 (NPFMC 2010, 2013, 2017). Under the Tier 4 harvest control, increasing $M$ will also increase OFL because default $\boldsymbol{F o F L}=\boldsymbol{M}$ (NPFMC 2010, 2013, 2017). Thus far, neither the CPT nor SSC recommended changing $M$ for NSRKC stock.

## Female clutch fullness

Future figures of clutch fullness should include confidence bounds.
Authors reply:
Clutch fullness and confidence bounds are listed in Table 3. Figure is not provided as females are not used for the assessment model.

## Growth

Revisit growth assumptions. Growth appears to be consistently overestimated in the assessment, producing too many large crab. The CPT looks forward to seeing the results from the laboratory studies on growth for NSRKC at the next meeting.

Authors reply:
In 2020, 36 (14 male) NSRKC were sent to Kodiak NOAA laboratory. Average molting growth was 13.1 mm for CL class $74-83 \mathrm{~mm}$ and 12.8 mm for CL class $84-93 \mathrm{~mm}$ that was smaller than observed growth from tag recovery data ( 15 mm for CL $74-83 \mathrm{~mm}$ and 16 mm for CL $84-93 \mathrm{~mm}$ ). However, the sample size was too small to evaluate statistical significance of the difference. To make more statistical comparison, > 30 crabs for each individual size classes needs to be captured and shipped to the lab.

As for overestimate of growth by the model, it should be reminded that the assessment model was not designed to fit observed growth increment, but to fit probability distributions of recaptured size classes based on the estimated transition matrix (Figure 13). Thus, the question should be not whether model estimated growth increments match to observed ones, but whether model estimated transition matrix predicted recaptured size proportion accurately.


Recovery after 2 years


Recovery after 3 years


The model fit between observed (bar) and predicted (Model 21.0 solid black, Model 21.1 dash red) generally matches. This suggest that transition size matrix derived from the tag-recovery data is unlikely overproducing the larger crab. We also included Appendix H that describes how the tag recovery data were assembled.

## Size at maturity

Investigations into size at maturity for this stock, referencing that of other red king crab stocks if useful.

Author reply:

As noted in previously (NPFMC 2018, 2019, 2020, 2021), size at maturity of Norton Sound male red king crab is highly uncertain. This is also true for other red king crab. First, maturity has two categories (biological and functional). Biological maturity indicate that male red king crab can biologically produce viable sperm, whereas functional maturity indicate that male red king crab is large enough to mate. The former can be determined using biological indicators, such as chela heights break point, whereas the latter is inferred by series of lab mating experiments. Few studies have been conducted to determine the size at functional maturity of male red king crab. The current NSRKC functional maturity size ( $>94 \mathrm{~mm}$ ) was inferred from Bristol Bay red king crab by incorporating the fact that Norton Sound red king crab is smaller. SSC suggested to investigate size at functional maturity of other stocks, such as of Barents Sea red king crab. However, it is unlikely that those metadata analyses would provide insights about size at maturity of Norton Sound red king crab because Norton Sound red king crab is the smallest among red king crab stocks. Authors were not able to find any other red king crab stocks that are comparable to the size of Norton Sound red king crab. We look forward progress in the laboratory studies. Although determining functional maturity size at maturity is important biologically, utility of this information for Tier 4 crab stock assessment is trivial. In Tier 4 stock assessment, size at maturity is used only for calculation of mature male biomass (MMB) and $\mathrm{B}_{\mathrm{MSY}}$ (average MMB). Harvest control (FOFL) is based on the ratio of projected MMB and $\mathrm{B}_{\mathrm{MSY}}$ (projected MMB/BMSY).
Tier 4 level and the OFL are determined by the $F_{M S Y}$ proxy, $B_{M S Y}$ proxy, and estimated legal male abundance and biomass:

| Level | Criteria | $F_{O F L}$ |
| :--- | :--- | :--- |
| A | $B / B_{M S Y^{p r o x}}>1$ | $F_{O F L}=\gamma M$ |
| B | $\beta<B / B_{M S Y \text { mox }} \leq 1$ | $F_{O F L}=\gamma M\left(B / B_{M S Y \text { pox }}-\alpha\right) /(1-\alpha)$ |
| C | $B / B_{M S Y^{p r o x}} \leq \beta$ | $F_{O F L}=$ bycatch mortality \& directed fishery $F=0$ |

where $B$ is a mature male biomass (MMB), $B_{M S Y}$ proxy is average mature male biomass over a specified time period, $M=0.18, \gamma=1, \alpha=0.1$, and $\beta=0.25$.

The $\mathrm{MMB} / \mathrm{B}_{\mathrm{MSY}}$ ratio is affected very little by changes of maturity size, unless the ratio is very close to 1.0 (Tier 4 a vs Tier 4 b borderline). To illustrate this, we present 2022 assessment model results with various minimum size at maturity cutoff, as follows.

Table: Effects of Tier 4 level by changing different maturity size.

| Maturity size | 94 mm <br> (default) | 74 mm | 84 mm | 104 mm | 114 mm | 124 mm | $>134 \mathrm{~mm}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| B MSY mil. lb $_{\text {M }}$ | 4.88 | 5.21 | 4.92 | 3.76 | 2.71 | 1.33 | 0.39 |
| MMB(2022) mil. lb | 5.21 | 5.91 | 5.61 | 4.42 | 2.86 | 1.03 | 0.27 |
| MMB/B MSY | 1.16 | 1.13 | 1.14 | 1.18 | 1.06 | 0.77 | 0.70 |
| Tier 4 level | a | a | a | a | a | b | b |
| FOFL | 0.18 | 0.18 | 0.18 | 0.18 | 0.18 | 0.13 | 0.12 |

As illustrated in the above table, changing minimum maturity size has little effects on $\mathrm{MMB} / \mathrm{B}_{\mathrm{MSY}}$ ratio and Tier 4 level designation. OFL and ABC are based on retained and unretained catch by size applied by Fofl.

However, this does not lessen biological importance of finding functional maturity size of NSRKC. The information would provide insights about productivity of the stock, as well as biological appropriateness of legal catch size.

## II. NSRKC Assessment Surveys and Data

## Discards Estimate

Further consider which of the methods to account for discards are most appropriate for NSRKC given probable future data availability. The CPT realizes that no method will be perfect, but an imperfect consideration of discards is better than ignoring them.

Authors reply:
As noted in Appendix C, we evaluated following 3 methods of estimating discards.

| Methods | Estimation <br> methodology | Assumption | Issue |
| :--- | :--- | :--- | :--- |
| LNR | Estimate total <br> discards from <br> observed discards <br> CPUE | Accurate observed <br> discards \& CPUE | Observer may not <br> know true discards. |
| Subtraction | Estimate total catch <br> from observed total <br> catch CPUE and then <br> subtract observed <br> retained | Accurate observed <br>  <br> CPUE | Discards can be < <br> 0, when total catch <br> CPUE is <br> underestimated. |
| Proportional | Estimate total <br> discards from <br> observed <br> discard/retained ratio. | Accurate <br> discards/retained <br> ratio. | Discards/retained <br> ratio may differ <br> greatly among <br> fishermen |

The major issue regarding NSRKC observer survey is that the observed fishermen are the most experienced and having larger boat, and thus their catch CPUE is higher than other fishermen. In fact, their catch CPUE during the observed periods were also higher than their post-season retained catch CPUE as well as post-season CPUE of other unobserved fishermen (except for 2012).

| Year | CPUEobs | CPUE $_{\text {FT. obs }}$ | CPUE $_{\text {FT.unobs }}$ |
| :--- | :--- | :--- | :--- |


| 2012 | 13.53 | 16.05 | 16.57 |
| ---: | ---: | ---: | ---: |
| 2013 | 10.88 | 8.67 | 7.47 |
| 2014 | 12.50 | 12.80 | 11.87 |
| 2015 | 24.29 | 17.26 | 15.62 |
| 2016 | 25.37 | 17.36 | 15.30 |
| 2017 | 19.76 | 14.33 | 13.33 |
| 2018 | 14.05 | 10.19 | 10.09 |
| 2019 | 5.07 | 4.58 | 4.56 |

LNR2 and Subtraction2 methods are intended to correct those by applying CPUE ratio between observed and Unobserved fishermen. CPT chose LNR2 discards observation method in 2021, despite that author recommended not to use discards estimates of any methods at all.

Given that discard estimation is required, authors propose using the Proportional methods for simplicity of assumptions. In Norton Sound commercial crab fishery observer survey, the number and length of discarded crabs are accurate because the observer also work as deckhand. However, representativeness of observed CPUE is highly uncertain, even adjustment is applied. On the other hand, proportion of discards can be more representable across all fishermen. Norton Sound commercial crab fishery pot configurations (and escapement mechanism) are largely standardized. Their fishery is also limited in geographically. Although red king crab distribution is patchy and spatial segregation among size classes and sex are possible, it is unlikely that the size dependent spatial segregation is occurring within the fishery grounds. The proportional method is also consistent with the estimation discards by the assessment model. Thus, for 2022 assessment model, we used discards data based on the proportional method.

It should be noted that ADF\&G terminated observer survey program in 2021, so that discards and total catch (retained + discards x handling mortality) will not be estimated. This also implies that management performance of Norton Sound red king crab (total catch OFL-ABC) cannot be evaluated.

## Discards Estimate in the absence of observer survey

## Bring forward methods to use historical data to estimate discard rates

## Authors reply:

As noted in the above and Appendix C, discard estimation method of the NSRKC is an ad hoc. Given that the NSRKC observer survey is terminated, developing a method for estimating discards biomass is also an ad hoc and highly speculative.

Hear I present an intended method for estimating discards mortality biomass when observer data are available. This method can also be used for estimation of discards when the data are not available.

When an observer survey is conducted and observer discards and the size distribution are available, discards abundance and biomass can be estimated as follows:

1. Estimate the number of discarded crab $\left(D_{n}\right)$ using the proportional method.

$$
D_{n}=\frac{n_{d i s}}{n_{r e t}} N_{r e t}=r_{d i s} \cdot N_{r e t}
$$

where $n_{\text {ret }}$ and $n_{\text {dis }}$ are the number of retained and discarded crab in the observer survey, and $N_{\text {ret }}$ is the number of retained crab from the commercial fisheries (tish ticket). $r_{d i s}$ is a discardsretained ratio.
2. Estimate biomass of discarded crab biomass $\left(D_{b}\right)$

$$
D_{b}=D_{n} \sum_{l} p_{d i s, l} \cdot w m_{l}=D_{n} \cdot w_{d i s}
$$

where $p_{\text {dis,l }}$ is the length $(l)$ proportions of observed discarded crab, $w m_{l}$ is the average weight of each length class $(l) . w_{\text {dis }}$ is a discard biomass unit per disgraced crab.

Combine the above two equations, discarded crab biomass is expressed as

$$
D_{b}=r_{d i s} \cdot w_{d i s} \cdot N_{r e t}
$$

Applying discards mortality of 0.2 , unretained catch biomass can be estimated as $0.2 \cdot D_{b}$, or $0.2 \cdot r_{d i s} \cdot w_{d i s} \cdot D_{n}, 0.2 \cdot r_{d i s} \cdot w_{d i s}$ is a discard mortality biomass unit per retained crab (Mort lb).

During the 2012-2019 periods, discarded crab size proportions, $r_{d i s}$, $w_{d i s}$, and Mort $l b$ are calculated as follows.

| Size <br> class | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | Average | Weight <br> wm (lb) |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 34 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.02 | 0.00 | 0.09 |
| 44 | 0.00 | 0.02 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.05 | 0.01 | 0.18 |
| 54 | 0.07 | 0.11 | 0.01 | 0.00 | 0.01 | 0.02 | 0.04 | 0.18 | 0.06 | 0.32 |
| 64 | 0.11 | 0.30 | 0.04 | 0.02 | 0.04 | 0.10 | 0.09 | 0.24 | 0.12 | 0.54 |
| 74 | 0.07 | 0.25 | 0.10 | 0.08 | 0.05 | 0.16 | 0.18 | 0.10 | 0.12 | 0.81 |
| 84 | 0.12 | 0.13 | 0.27 | 0.18 | 0.17 | 0.14 | 0.36 | 0.12 | 0.19 | 1.17 |
| 94 | 0.24 | 0.14 | 0.43 | 0.47 | 0.53 | 0.30 | 0.30 | 0.27 | 0.34 | 1.72 |
| 104 | 0.19 | 0.04 | 0.13 | 0.21 | 0.18 | 0.26 | 0.02 | 0.02 | 0.13 | 2.35 |
| 114 | 0.14 | 0.00 | 0.01 | 0.03 | 0.02 | 0.01 | 0.00 | 0.00 | 0.03 | 3.02 |
| 124 | 0.05 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.01 | 3.71 |
| 134 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 4.30 |
| $w_{\text {dis }}$ | 1.75 | 0.90 | 1.51 | 1.72 | 1.65 | 1.54 | 1.22 | 0.94 | 1.40 |  |
| $r_{\text {dia }}$ | 1.96 | 1.40 | 1.02 | 0.92 | 0.25 | 0.25 | 0.50 | 0.87 | 0.90 |  |
| $r_{\text {dis }}$ - dis | 3.43 | 1.27 | 1.53 | 1.58 | 0.42 | 0.39 | 0.61 | 0.82 | 1.26 |  |


| Mort lb | 0.786 | 0.254 | 0.306 | 0.316 | 0.084 | 0.078 | 0.122 | 0.164 | 0.252 |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

Mort lb ranged from 0.078 to 0.786 , with the mean of 0.252 . In the absence of observer data, unretained crab mortality can be estimated as $0.252 N_{\text {ret. }}$. However, this also indicates that applying the mean would overestimate unretained catch by 3.2 times $(0.252 / 0.078)$ or underestimate it by 0.32 times ( $0.252 / 0.786$ ).

Alternative methods:

## Alternative 1.

An alternative method is using the trawl survey length proportion data as proxy for true length proportion. The model estimated trawl survey selectivity is 1.0 for all lengths, which indicates that trawl survey length composition equals NSRKC length proportion. Discards length proportion $p_{\text {dis, } l}$ can be estimated by multiplying model estimated fishery selectivity $\left(S_{l}\right)$ and 1retention probability $\left(S_{\text {ret }, l}\right)$

$$
p_{d i s, l}=p_{t w l, l} \cdot S_{l} \cdot\left(1-S_{\text {ret } . l}\right)
$$

The discards-retained ratio ( $r_{d i s}$ ) is

$$
r_{d i s}=\frac{\sum_{l} p_{t w l, l} \cdot S_{l} \cdot\left(1-S_{r e t, l}\right)}{\sum_{l} p_{t w l, l} \cdot S_{l} \cdot S_{r e t, l}}
$$

The discard biomass unit ( $w_{d i s}$ ) is

$$
w_{d i s}=\frac{\sum_{l} p_{t w l, l} \cdot S_{l} \cdot\left(1-S_{r e t, l}\right) \cdot w m_{l}}{\sum_{l} p_{t w l, l} \cdot S_{l} \cdot\left(1-S_{r e t, l}\right)}
$$

During the 2012-2019 periods, trawl survey occurred in 2014, 2017, 2018, and 2019. The table below shows trawl survey length proportion, and model estimated selectivity and retention probability from the 2021 assessment model.

| Sizeclass | 2014 | 2017 | 2018 | 2019 | Selectivity <br> $(2021)$ | Retention <br> $(2021)$ | $l \mathrm{~b}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | :--- |
| 34 | 0.01 | 0.11 | 0.02 | 0.00 | 0.00 | 0.00 | 0.09 |
| 44 | 0.00 | 0.02 | 0.33 | 0.00 | 0.01 | 0.00 | 0.18 |
| 54 | 0.01 | 0.01 | 0.42 | 0.02 | 0.04 | 0.00 | 0.32 |


| 64 | 0.01 | 0.06 | 0.08 | 0.13 | 0.12 | 0.00 | 0.54 |
| :--- | :---: | :---: | :---: | :---: | :---: | ---: | :--- |
| 74 | 0.07 | 0.12 | 0.05 | 0.47 | 0.33 | 0.00 | 0.81 |
| 84 | 0.14 | 0.11 | 0.02 | 0.26 | 0.64 | 0.00 | 1.17 |
| 94 | 0.25 | 0.06 | 0.02 | 0.04 | 0.86 | 0.07 | 1.72 |
| 104 | 0.27 | 0.09 | 0.01 | 0.02 | 0.96 | 0.88 | 2.35 |
| 114 | 0.14 | 0.13 | 0.01 | 0.01 | 0.99 | 1.00 | 3.02 |
| 124 | 0.06 | 0.23 | 0.01 | 0.02 | 1.00 | 1.00 | 3.71 |
| 134 | 0.02 | 0.07 | 0.02 | 0.03 | 1.00 | 1.00 | 4.30 |
| wdis $^{r_{\text {dia }}}$ | 1.56 | 1.28 | 0.92 | 1.04 |  |  |  |
| $r_{\text {dis }}$ dis | 0.75 | 0.35 | 1.53 | 4.72 |  |  |  |
| Mort lb | 1.18 | 0.45 | 1.41 | 4.92 |  |  |  |
| \% Deviation | 0.236 | 0.090 | 0.282 | 0.984 |  |  |  |

Among the 4 years, the model estimated Mort lb multiplier ranged from 0.090 to 0.984 . Compared those with observed, the model deviation ranged from $-22 \%$ to $+500 \%$. The deviation was greater in 2018 and 2019.

Alternative 2.

The NSRKC assessment model estimates directly from the observed retained catch using the alternative model 1 approach except that the model uses predicted length composition. Use the adopted model (baseline model) with updated data to estimate discards biomass.

In all cases, the major difficulty is inferencing the amount and length composition of unobserved crabs that are highly variable and cannot be directly estimated from retained crab. Applying average Mort lb ignores the annual variations. Alternative models attempts to estimate annual variations with model estimated selectivity and retention probability; however, this also generate highly variable estimates.

## Pot loss

Reporting on pot loss, especially in regard to potential pot losses at the end of the season as noted in public testimony.

Authors reply:
Pot loss is inferred from "additional" pot permit requested by fishermen during the season (summer) and post-season self-reporting (winter commercial and subsistence). Although ADF\&G staff routinely ask reasons for additional permits, fishermen are NOT required to provide reason. Fishermen are not required to report pot loss to ADF\&G. Changes of regulations will be required to obtain accurate pot loss.

## VAST

Explore having Jon Richar work on a VAST model for Norton Sound trawl surveys.
Authors reply:
Jon Richar received an approval from his supervisor to work on a VAST model for the Norton Sound trawl surveys. We look forward to work with him when he is available.

## Standardized CPUE

Please explain how the SD was determined for the CPUE as it is the same from 2000-2019. Is this a fixed SD? If so shouldn't the CV be fixed rather than the SD?

Authors reply:
SD is a glm model estimate sigma of lognormal CPUE, exponent back to normal space (Appendix B, NPFMC 2013). For detailed description of the standardized CPUE calculation method and SD, please refer NPFMC (2013) for the original report and Appendix B for a brief data and model update.

## III. NSRKC Assessment model

## Data weighting

Continue exploration of data-weighting assumptions. Provide clarification and justification for the current data weighting scheme utilized in the model.

Authors reply:
Data-weighting is aimed to achieve a balance among various data sets. The current model data weighting schemes, although arbitrary were deemed appropriate by the CPT-SSC (NPFMC 2011, 2012). As illustrated in the figure below, increasing weight of size composition data (input sample size: from minimum) would lower model fit to the trawl survey abundance data.

Trawl survey crab abundance


Figure: Model 21.0 default input sample size (20: trawl, 10: others) (black) vs. increased input sample size $(200,100)$ (blue dash line), and reduce input size $(10,5)$ (red dash line).

Thus far, there is no objective criteria for determining the balance (i.e., how much a model should fit observed trawl abundance data vs. size composition data?). We have tried alternative weighting schemes (NPFMC 2019, 2020, 2021) and found current ones be appropriate. We welcome CPT and SSC's suggestions for alternative data weighting scheme.

In the Analytic approach, more descriptive text should be included in the sections describing the model and its assumptions, to reduce referring to Appendix A.

Authors reply:

Part of Appendix A are now in model description sections.

Furthermore, a thorough description of the model selection and evaluation criteria, and most particularly, the results of the author's recommended models (and the base model, if they differ) is a basic requirement for a complete assessment document. A list of figures and tables is not an acceptable description of results.

Authors reply:
Implemented

Variant of Model 21.0 with estimated natural mortality
Authors reply:
Models 21.4 and 21.5 estimates natural mortality.

## IV. NSRKC Management

## Legal sized crab

Explore and document the reasons for the changes in the relationship between carapace length and carapace width. Document which data sources are excluded or included and for what reason.

Authors reply:
In NSRKC, legal size is defined as carapace width greater than 4.75 inches that was conventionally equated as greater than 104mm carapace length. Since 1996 ADF\&G has started noting legal size crab based on carapace width in trawl, commercial fishery observer, and other miscellaneous surveys to complement the carapace length measurement. Originally, the proportion was based solely from the trawl survey. As more data are collected from commercial observer surveys, recent proportions are based on more observer data.

Proportion of legal (CW $>4.75$ inch $)$ crab in Trawl survey

| size class | 64 | 74 | 84 | 94 | 104 | 114 | 124 | 134 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{1 9 9 6}$ | 0.00 | 0.00 | 0.00 | 0.18 | 0.93 | 1.00 | 1.00 | 1.00 |
| $\mathbf{1 9 9 9}$ | 0.00 | 0.00 | 0.00 | 0.40 | 0.98 | 0.98 | 1.00 | 1.00 |
| $\mathbf{2 0 0 2}$ | 0.00 | 0.00 | 0.00 | 0.28 | 0.97 | 1.00 | 1.00 | 1.00 |
| $\mathbf{2 0 0 6}$ | 0.00 | 0.00 | 0.00 | 0.18 | 1.00 | 1.00 | 1.00 | 1.00 |
| $\mathbf{2 0 0 8}$ | 0.00 | 0.00 | 0.00 | 0.19 | 0.96 | 1.00 | 1.00 | 1.00 |
| $\mathbf{2 0 1 1}$ | 0.00 | 0.00 | 0.00 | 0.24 | 0.99 | 1.00 | 1.00 | 1.00 |
| $\mathbf{2 0 1 4}$ | 0.00 | 0.00 | 0.00 | 0.21 | 0.98 | 1.00 | 1.00 | 1.00 |
| $\mathbf{2 0 1 7}$ | 0.00 | 0.00 | 0.00 | 0.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| $\mathbf{2 0 1 8}$ | 0.00 | 0.00 | 0.00 | 0.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| $\mathbf{2 0 1 9}$ | 0.00 | 0.00 | 0.00 | 0.33 | 1.00 | 1.00 | 1.00 | 1.00 |
| $\mathbf{2 0 2 0}$ | 0.00 | 0.00 | 0.00 | 0.22 | 1.00 | 1.00 | 1.00 | 1.00 |
| Total | $\mathbf{0 . 0 0}$ | $\mathbf{0 . 0 0}$ | $\mathbf{0 . 0 0}$ | $\mathbf{0 . 2 5}$ | $\mathbf{0 . 9 8}$ | $\mathbf{1 . 0 0}$ | $\mathbf{1 . 0 0}$ | $\mathbf{1 . 0 0}$ |

Proportion of legal ( $\mathrm{CW}>4.75 \mathrm{inch}$ ) crab in Observer survey

| size class | 64 | 74 | 84 | 94 | 104 | 114 | 124 | 134 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{2 0 1 2}$ | 0.00 | 0.01 | 0.02 | 0.22 | 0.90 | 1.00 | 1.00 | 1.00 |
| $\mathbf{2 0 1 3}$ | 0.00 | 0.00 | 0.00 | 0.44 | 0.98 | 1.00 | 1.00 | 1.00 |
| $\mathbf{2 0 1 4}$ | 0.00 | 0.00 | 0.00 | 0.22 | 0.91 | 1.00 | 1.00 | 1.00 |
| $\mathbf{2 0 1 5}$ | 0.00 | 0.00 | 0.00 | 0.38 | 0.98 | 1.00 | 1.00 | 1.00 |
| $\mathbf{2 0 1 6}$ | 0.00 | 0.00 | 0.00 | 0.46 | 1.00 | 1.00 | 1.00 | 1.00 |
| $\mathbf{2 0 1 7}$ | 0.00 | 0.00 | 0.00 | 0.13 | 0.91 | 1.00 | 1.00 | 1.00 |
| $\mathbf{2 0 1 8}$ | 0.00 | 0.00 | 0.00 | 0.16 | 0.95 | 0.99 | 1.00 | 1.00 |
| $\mathbf{2 0 1 9}$ | 0.00 | 0.00 | 0.00 | 0.18 | 0.93 | 1.00 | 1.00 | 1.00 |
| Total | $\mathbf{0 . 0 0}$ | $\mathbf{0 . 0 0}$ | $\mathbf{0 . 0 0}$ | $\mathbf{0 . 3}$ | $\mathbf{0 . 9 5}$ | $\mathbf{1 . 0 0}$ | $\mathbf{1 . 0 0}$ | $\mathbf{1 . 0 0}$ |

The proportion of legal crab used in the assessment model is an average proportion based on observer survey data. In the assessment model, this proportion is used to estimate the number of retained crab in Winter and summer commercial fishery prior of 2008. During this time, all legal sized crab was assumed to be retained. Since 2009 commercially retained crab size is CW $>5.0$ and retention probability is estimated from the observer survey.

Plot the legal biomass over time using the different proportions of legal size crab to better understand the magnitude of the impact of the change.

## Author reply:

Norton Sound red king assessment model is based on abundance. Time series of legal crab is plotted in Figure 4. The plot is based on carapace length ( $\mathrm{CL}>104 \mathrm{~mm}$ ) not carapace width ( CW
$>4.75$ inch) that is the definition of legal crab. Legal crab biomass NOT used for calculation of the OFL. Thus, there is no meaning to plot the legal biomass time series by different proportion of legal crab.

The table below shows the proportion of legal vs. retained crab during the 2012-2019 observer survey, in response to request by the public.

| Year |  | 64 | 74 | 84 | 94 | 104 | 114 | 124 | 134 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{2 0 1 2}$ | Legal | 0 | 0.01 | 0.02 | 0.22 | 0.9 | 1 | 1 | 1 |
|  | Retained | $\mathbf{0}$ | $\mathbf{0}$ | $\mathbf{0}$ | $\mathbf{0 . 0 5}$ | $\mathbf{0 . 4 6}$ | $\mathbf{0 . 6 3}$ | $\mathbf{0 . 6 4}$ | $\mathbf{0 . 8 5}$ |
| $\mathbf{2 0 1 3}$ | Legal | 0 | 0 | 0 | 0.44 | 0.98 | 1 | 1 | 1 |
|  | Retained | $\mathbf{0}$ | $\mathbf{0}$ | $\mathbf{0}$ | $\mathbf{0 . 1 4}$ | $\mathbf{0 . 8 6}$ | $\mathbf{0 . 9 9}$ | $\mathbf{1}$ | $\mathbf{1}$ |
| $\mathbf{2 0 1 4}$ | Legal | 0 | 0 | 0 | 0.22 | 0.91 | 1 | 1 | 1 |
|  | Retained | $\mathbf{0}$ | $\mathbf{0}$ | $\mathbf{0}$ | $\mathbf{0 . 0 4}$ | $\mathbf{0 . 7 4}$ | $\mathbf{0 . 9 7}$ | $\mathbf{0 . 9 9}$ | $\mathbf{1}$ |
| $\mathbf{2 0 1 5}$ | Legal | 0 | 0 | 0 | 0.38 | 0.98 | 1 | 1 | 1 |
|  | Retained | $\mathbf{0}$ | $\mathbf{0}$ | $\mathbf{0}$ | $\mathbf{0 . 1 1}$ | $\mathbf{0 . 7 4}$ | $\mathbf{0 . 9 1}$ | $\mathbf{0 . 9 4}$ | $\mathbf{0 . 8 9}$ |
| $\mathbf{2 0 1 6}$ | Legal | 0 | 0 | 0 | 0.46 | 1 | 1 | 1 | 1 |
|  | Retained | $\mathbf{0}$ | $\mathbf{0}$ | $\mathbf{0}$ | $\mathbf{0 . 1 3}$ | $\mathbf{0 . 8 9}$ | $\mathbf{0 . 9 9}$ |  | $\mathbf{1}$ |
| $\mathbf{2 0 1 7}$ | Legal | 0 | 0 | 0 | 0.12 | 0.91 | 1 | 1 | $\mathbf{1}$ |
|  | Retained | $\mathbf{0}$ | $\mathbf{0}$ | $\mathbf{0}$ | $\mathbf{0 . 0 2}$ | $\mathbf{0 . 7 5}$ | $\mathbf{0 . 9 9}$ | 1 |  |
| $\mathbf{2 0 1 8}$ | Legal | 0 | 0 | 0 | 0.16 | 0.95 | 0.99 | $\mathbf{1}$ | 1 |
|  | Retained | $\mathbf{0}$ | $\mathbf{0}$ | $\mathbf{0}$ | $\mathbf{0 . 1 4}$ | $\mathbf{0 . 9 2}$ | $\mathbf{0 . 9 9}$ | $\mathbf{1}$ |  |
| $\mathbf{2 0 1 9}$ | Legal | 0 | 0 | 0 | 0.18 | 0.93 | 1 | $\mathbf{0}$ | 1 |
|  | Retained | $\mathbf{0}$ | $\mathbf{0}$ | $\mathbf{0}$ | $\mathbf{0 . 1 5}$ | $\mathbf{0 . 9 3}$ | $\mathbf{1}$ |  | $\mathbf{1}$ |

The proportion of legal sized crab retained from observer survey 2012-2019

| Year | 64 | 74 | 84 | 94 | 104 | 114 | 124 | 134 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{2 0 1 2}$ | 0 | 0 | 0 | 0.23 | 0.51 | 0.63 | 0.64 | 0.85 |
| $\mathbf{2 0 1 3}$ | 0 | 0 | 0 | 0.31 | 0.88 | 0.99 | $\mathbf{1}$ | 1 |
| $\mathbf{2 0 1 4}$ | 0 | 0 | 0 | 0.19 | 0.82 | 0.97 | 0.99 | 1 |
| $\mathbf{2 0 1 5}$ | 0 | 0 | 0 | 0.28 | 0.76 | 0.91 | 0.94 | 0.89 |
| $\mathbf{2 0 1 6}$ | 0 | 0 | 0 | 0.28 | 0.89 | 0.99 | 1 | 1 |
| $\mathbf{2 0 1 7}$ | 0 | 0 | 0 | 0.14 | 0.82 | 0.99 | 1 | 1 |
| $\mathbf{2 0 1 8}$ | 0 | 0 | 0 | 0.87 | 0.98 | 1 | 1 | 0.99 |
| $\mathbf{2 0 1 9}$ | 0 | 0 | 0 | 0.86 | 1 | 1 | 1 | 1 |

OFL

The OFL should be specified based on total catch including retained catch and non-surviving discard. Specifying the OFL based on legal crab would result in higher OFLs than if based on retained crab. This would then translate to higher exploitation rates on the exploitable crab than
the target rates and increased discard mortality on non-preferred size crab that must be sorted through to achieve the OFL.

Authors reply:
Corrected. Note that observer survey was terminated in 2021. Thus, even though total OFL and ABC are specified, total catch (retained and discarded x discard mortality) will not be directly calculated.

## LKTKS

The inclusion of local, traditional and subsistence knowledge (LKTKS) information in the assessment, an effort the SSC understands cannot be fully pursued until appropriate protocols are developed and pandemic conditions ease.

Authors reply:
We look for the Taskforce's progress in Norton Sound red king crab case study writeup that is projected to be finished in April 2022.

## C. Introduction

## 1. Species:

red king crab (Paralithodes camtschaticus) in Norton Sound, Alaska.

## 2. General Distribution:

Norton Sound red king crab (NSRKC) is one of the northernmost red king crab populations that can support a commercial fishery (Powell et al. 1983). It is distributed throughout Norton Sound with a westward limit of $167-168^{\circ} \mathrm{W}$. longitude, depths less than 30 m , and summer bottom temperatures above $4^{\circ} \mathrm{C}$. The Norton Sound red king crab management area consists of two units: Norton Sound Section (Q3) and Kotzebue Section (Q4) (Menard et al. 2011). The Norton Sound Section (Q3) consists of all waters in Registration Area Q north of the latitude of Cape Romanzof, east of the International Dateline, and south of $66^{\circ} \mathrm{N}$ latitude (Figure 1). The Kotzebue Section (Q4) lies immediately north of the Norton Sound Section and includes Kotzebue Sound. Commercial fisheries have not occurred regularly in the Kotzebue Section. This report deals with the Norton Sound Section of the NSRKC management area.

## 3. Evidence of stock structure:

Based on variability at 15 SNP loci and in mtDNA sequences (COI, 665 bp ), NSRK stock belongs to the Okhotsk Sea-Norton Sound-Aleutian Islands evolutionary lineage (SNPs, FCT = 0.054 ; mtDNA FCT $=0.222$ ) (Grant and Chen 2012). However, this does not indicate that NSRKC is a single stock. The study indicates that the study was incapable of detecting possible
evolutionary stock difference within NSRKC stock. No studies have investigated possible stock separation within Norton Sound management area (Figure 1).

## 4. Life history characteristics relevant to management:

Life history of NSRKC has not been well studied. One of the unique life-history traits of NSRKC is that they spend their entire lives in shallow water since Norton Sound is generally less than 40 m in depth. Distribution and migration patterns of NSRKC have not been well studied. Based on the 1976-2021 trawl surveys, NSRKC are found in areas with a mean depth range of 19 $\pm 6$ (SD) m and bottom temperatures of $7.4 \pm 2.5$ (SD) ${ }^{\circ} \mathrm{C}$ during summer. NSRKC are consistently abundant offshore of Nome.

NSRKC migrate between deep offshore and shallow inshore waters within Norton Sound. Timing of the inshore mating migration is unknown but is assumed to be during late fall to winter (Powell et al. 1983). Offshore migration occurs in late May - July (Jenefer Bell, ADF\&G, personal communication). The results from a study funded by North Pacific Research Board (NPRB) during 2012-2014 suggest that older/large crab (> 104mm CL) may stay offshore in the winter, based on findings that large crab are not found nearshore during spring offshore migration periods (Jenefer Bell, ADF\&G, personal communication). Molting occurs in fall: late August - November for male and Jan-March for female based on laboratory observation (Leah Zacher and Jennifer Gardner NOAA-AFSC personal comm). Trawl surveys show that crab distributions are patchy and dynamic.

## 5. Brief management history:

NSRKC fisheries consist of commercial and subsistence fisheries. The commercial red king crab fishery started in 1977 and occurs in summer (June - August) and winter (December - May). The majority of NSRKC is harvested during the summer commercial fishery, whereas the winter commercial and subsistence fisheries occur nearshore through ice.

## Summer Commercial Fishery

A large-vessel summer commercial crab fishery started in 1977 in the Norton Sound Section (Table 1) and continued from 1977 through 1990. No summer commercial fishery occurred in 1991 because there were no staff to manage the fishery. In March 1993, the Alaska Board of Fisheries (BOF) limited participation in the fishery to small boats. Then on June 27, 1994, a super-exclusive designation went into effect for the fishery. This designation stated that a vessel registered for the Norton Sound crab fishery may not be used to take king crabs in any other registration areas during that registration year. A vessel moratorium was put into place before the 1996 season. This was intended to precede a license limitation program. In 1998, Community Development Quota (CDQ) groups were allocated a portion of the summer harvest; however, no CDQ harvest occurred until the 2000 season. On January 1, 2000 the North Pacific License Limitation Program (LLP) went into effect for the Norton Sound crab fishery. The program dictates that a vessel which exceeds 32 feet in length overall must hold a valid crab license issued under the LLP by the National Marine Fisheries Service. Changes in regulations and the location of buyers resulted in eastward movement of the harvest
distribution in Norton Sound in mid-1990s. In Norton Sound, a legal crab is defined as $\geq 4$ 3/4 inch carapace width (CW, Menard et al. 2011), which is approximately equivalent to $\geq$ 104 mm carapace length (CL). In 2005 and 2006, commercial buyers (NSEDC: Norton Sound Economic Development Corporation) accepted only legal crab of $\geq 5$ inch CW. This preference became permanent since 2008.

Portions of Norton Sound area are closed to commercial fishing for red king crab. Since the beginning of the commercial fisheries in 1977, waters approximately 5-10 miles offshore of southern Seward Peninsula from Port Clarence to St. Michael have been closed to protect crab nursery grounds during the summer commercial crab fishery (Figure 2). The spatial extent of closed waters has varied historically. In 2020 the Board of Fisheries closed Norton Sound area east of 167 degrees W. longitude for commercial summer crab fishery. In 2021 NSEDC stopped purchasing NSRKC.

## CDQ Fishery

The Norton Sound and Lower Yukon CDQ groups divide the CDQ allocation. Only fishers designated by the Norton Sound and Lower Yukon CDQ groups are allowed to participate in this portion of the king crab fishery. Fishers are required to have a CDQ fishing permit from the Commercial Fisheries Entry Commission (CFEC) and register their vessel with the Alaska Department of Fish and Game (ADF\&G) before begin fishing. Fishers operate under the authority of each CDQ group. CDQ harvest share is $7.5 \%$ of total projected harvest, which can be prosecuted in both summer and winter fisheries season.

## Winter Commercial Fishery

The winter commercial crab fishery is a small fishery using hand lines and pots through the nearshore ice. On average 10 permit holders harvested 2,500 crab during 1978-2009. From 2007 to 2015 the winter commercial catch increased from 3,000 crab to over 40,000 (Table 2). In 2015 the winter commercial catch reached $20 \%$ of total crab catch. The BOF responded in May 2015 by amending regulations to allocate $8 \%$ of the total commercial guideline harvest level (GHL) to the winter commercial fishery, which became in effect since the 2017 season. The winter red king crab commercial fishing season was also set from January 15 to April 30, unless changed by an emergency order. The new regulation became in effect stating with the 2016 season.

## Subsistence Fishery

The winter subsistence fishery has a long history; however, harvest information is available only since the 1977/78 season. The majority of subsistence crab fishery occurs in winter using hand lines and pots through nearshore ice. Average annual winter subsistence harvest is 5,400 crab (1977-2010). Subsistence harvesters need to obtain a permit before fishing and record daily effort and catch. There are no size or sex specific harvest limits; however, the majority of retained catches are males of near legal size.

Summer subsistence crab fishery harvest has been monitored since 2004 with an average harvest of 712 crab per year. The summer subsistence fishery was not included in the assessment model.

## Harvest of both winter commercial and subsistence winter fisheries is influenced largely by availability of stable ice condition. Low harvest can occur due to poor ice condition, regardless of crab abundance.

1. Brief description of the annual ADF\&G harvest strategy

Since 1997 NSRKC has been managed based on a guideline harvest level (GHL). From 1999 to 2011 the GHL for the summer commercial fishery was determined by a prediction model and the model estimated predicted biomass: (1) $0 \%$ harvest rate of legal crab when estimated legal biomass $<1.5$ million lb ; $(2) \leq 5 \%$ of legal male abundance when the estimated legal biomass falls within the range $1.5-2.5$ million lb ; and ( 3 ) $\leq 10 \%$ of legal male when estimated legal biomass $>2.5$ million lb . In 2012 the summer commercial fishery GHL was revised to (1) $0 \%$ harvest rate of legal crab when estimated legal biomass $<1.25$ million lb ; $(2) \leq 7 \%$ of legal male abundance when the estimated legal biomass falls within the range 1.25-2.0 million lb ; $(3) \leq 13 \%$ of legal male abundance when the estimated legal biomass falls within the range $2.0-3.0$ million lb ; and $(3) \leq 15 \%$ of legal male biomass when estimated legal biomass $>3.0$ million lb .

In 2015 the BOF passed the following regulations regarding the winter commercial fisheries:

1) Revise GHL to include summer and winter commercial fisheries.
2) Set GHL for the winter commercial fishery $\left(\mathrm{GHL}_{\mathrm{w}}\right)$ at $8 \%$ of the total GHL
3) Dates of the winter red king crab commercial fishing season are from January 15 to April 30.

In practice, GHL is set to be below ABC retained that was derived from retained the OFL. Since 2021 the OFL and ABC of NSRKC is a total OFL that includes mortality of both retained and unretained crab. The historical management changes are summarized in the following table.

| Year | Notable historical management changes |
| :--- | :--- |
| 1976 | The abundance survey started |
| 1977 | Large vessel commercial fisheries began. Legal size was set to $\geq \mathbf{5}$ inch CW |
| 1978 | Legal size was changed to $\geq \mathbf{4 . 7 5}$ inch CW |
| 1991 | Fishery closed due to staff constraints |
| 1994 | Super exclusive designation went into effect. The end of large vessel commercial fishery <br> operation. |
| 1998 | Community Development Quota (CDQ) allocation went into effect |
| 1999 | Guideline Harvest Level (GHL) went into effect |
| 2000 | North Pacific License Limitation Program (LLP) went into effect. |
| 2002 | Change in closed water boundaries (Figure 2) |
| 2006 | The Statistical area Q3 section expanded (Figure 1) |
| 2008 | Start date of the open access fishery changed from July 1 to after June 15 by emergency order. <br> Pot configuration requirement: at least 4 escape rings $\mathbf{~}>\mathbf{4 . 5}$ inch diameter) per pot located <br> within one mesh of the bottom of the pot, or at least $1 / 2$ of the vertical surface of a square pot <br> or sloping side-wall surface of a conical or pyramid pot with mesh size $>\mathbf{6} .5$ inches. |
| 2008 | Commercially accepted legal crab size was changed to $\geq \mathbf{5}$ inch CW |
| 2012 | The BOF adopted a revised GHL for summer fishery. |
| 2016 | Winter GHL for commercial fisheries was established and modified winter fishing season dates <br> were implemented. |
| 2020 | The BOF closed summer commercial fishery E of 167 longitude |
| 2021 | NSEDC stopped purchasing NSRKC |

2. Summary of the history of the $B_{\mathrm{MSY}}$.

NSRKC is a Tier 4 crab stock. Direct estimation of the $B_{\text {MSY }}$ is not possible. The $B_{\text {MSY }}$ proxy is calculated as mean model estimated mature male biomass (MMB) from 1980 to present. Choice of this period was based on a hypothesized shift in stock productivity a due to a climatic regime shift indexed by the Pacific Decadal Oscillation (PDO) in 1976-77.

## D. Data

1. Summary of new information:

Winter commercial and subsistence fisheries:
The winter commercial fishery catch in 2021 was 320 crab ( 911 lb .). Subsistence retained crab catch was 2,892 and unretained was 1,763 crab or $38 \%$ of total catch (Table 2).

Summer commercial fishery:
The summer commercial fishery opened on $6 / 25 / 2021$ and closed on $9 / 03 / 2021$. Total of 0 crab ( 0 lb .) were harvested (Table 1).

Standardized CPUE

## Standardized CPUE for the years of 1993, 2020, 2021 were not calculated because commercial fishery did not occur (1993) or no crab was harvested (2020, 2021).

## Recalculate standardized CPUE:

3 periods:
1977-1993: Large Scale commercial fishery
1994-2007: Small boat commercial fishery
2008-2019: Small boat commercial fishery with high grading.
NOTE: Time periods revised.
Discards
Estimates of discards is based on author preferred proportional method, instead of LNR2 method that CPT selected in 2020.

## Summer Trawl Survey

Annual ADF\&G summer trawl survey was conducted in $7 / 19-8 / 3$ 2021. Because of unfavorable weather condition, 39 out of usual 60 stations were surveyed. Total male crab abundance estimate ( $\mathrm{CL}>63 \mathrm{~mm}$ ) is $2,400,000$ with a CV 0.60 .

Norton Sound portion of the NOAA NBS trawl survey was conducted in 7/29-8/7 2021. Total male crab abundance estimate (CL $>63 \mathrm{~mm}$ ) is $2,370,000$ with a CV 0.43 .
2. Available survey, catch, and tagging data

Available NSRKC data consist of followings: trawl survey that informs abundance and size composition, catch that informs size composition and standardized CPUE that informs an index of abundance, and tag recovery that informs growth-transition.

Trawl survey
Trawl survey consists of 3 surveys: NMFS triennial survey: 1976-1992, ADF\&G triennialannual survey: 1996-2019, and NOAA biannual survey: 2010, 2017-2019.

NMFS triennial survey:
Norton Sound trawl survey was initiated by NMFS in 1976 to assess stock status of crab and ground fish in Norton Sound and Kotezbue Sound. The survey established 10 nautical mile $(\mathrm{nm})$ grid survey stations throughout the entire Norton Sound and 15 nm grids outer Norton Sound area. The grid survey stations became the standard stations since 1976. Th survey was conducted from mid-late August to September-October, except for 1979. The survey used 83-112 Eastern Otter trawl gear, with tow distance of $1.3-1.7 \mathrm{~nm}$ ( 30 minutes tow). The survey was terminated in 1992.

ADF\&G triennial -annual survey:
After the termination of the NMFS trawl survey, ADF\&G began trawl survey in 1996 using the same survey stations, but smaller boat and survey coverage. The survey started as triennially but became annual survey in 2017. The survey usually occurs in late July - mid August, using 400 Eastern Otter trawl gear with tow distance of 1.0 nm . The survey used to have a re-tow protocol: when the first tow caught more than 7 red king crab, the station was re-towed. This protocol was dropped in 2012 in favor of more coverage.

NOAA biennial NBS survey:
NOAA NBS trawl survey started in 2010, and biennially since 2017. The survey occurs in late July-mid August, similar time as ADF\&G survey. The survey has 20 nm grid using 83112 Eastern Otter trawl gear, with tow distance of $1.3-2.5 \mathrm{~nm}$ ( 30 min tow).

Abundance estimation method
Methods of estimating abundance differed among the three surveys and throughout time periods. Abundance estimates have been revised many times.
Abundance and CV of the NMFS 1976-1991 surveys were provided by NOAA (Jon Richer NOAA personal communication). The abundance was estimated by averaging catch CPUE (\#/ $\mathrm{nm}^{2}$ ) of all stations (including survey stations out of Norton Sound) that was multiplied by standard Norton Sound Area ( $7600 \mathrm{~nm}^{2}$ ) (i.e., $\mathrm{N}=7600 *$ mean CPUE). On the other hand, ADF\&G survey abundance is calculated at each station CPUE* $\left(100 \mathrm{~mm}^{2}\right)$ and summed across all surveyed stations (i.e., $\mathrm{N}=$ sum of $100 *$ CPUEs) (Bell and Hamazaki 2019). Extent of the ADF\&G survey coverage differed among years due to survey conditions, and survey abundance was not standardized. NOAA NBS survey abundance is estimated by the author with the data limited only Norton Sound survey area that overlaps ADF\&G survey area ( $5841 \mathrm{~nm}^{2}$ ).


Figure ADF\&G trawl survey coverage (yellow shade) and NOAA NBS trawl survey coverage where abundance estimates were made (red hashed line)

Catchability of NSRKC appears to differ between ADF\&G and NMFS-NOAA NBS trawl surveys. ADF\&G trawl survey abundance tend to be higher than NMFS-NOAA NBS trawl survey even though NMFS-NOAA NBS survey coverages are greater than ADF\&G. The assessment model assumes (recommendation by CPT-SSC) that survey q of ADF\&G trawl survey be 1.0 , which resulted in $\mathrm{q}=0.7-0.8$ for NMFS and $0.7-0.96$ for NOAA NBS survey.

## Standardized CPUE

Standardized summer commercial fishery CPUE is included in the NSRKC assessment model as an index of NSRKC abundance that could supplement triennial trawl survey. CPUE standardization model was developed by Gretchen Bishop (ADF\&G) (NPFMC 2013). Since then, the same model has been applied with updated data (Appendix C).

## Tagging-recovery data

Norton Sound tagging study was initially conducted in 1980 as a part of mark-recapture abundance survey (Brannian 1987). The study was conducted in 1980-1982 and 1985. From 1986 to 2012 crabs were tagged during the winter pot survey. The winter surveys tagged more smaller (sublegal) crabs; however, very few were recovered. Tagging study was
resumed in 2012-2015 for spring migration movement survey. In all the above studies, tagged crabs were recovered by commercial fisheries.

## Length-Shell proportion data

Length-shell data have been collected in every research and harvest monitoring surveys. Of those, summer commercial harvest sampling, winter pot survey (terminated in 2012), and trawl survey have been consistent.

Time series of the data used for the NSRKC assessment model are summarized in the following figure and table.


|  | Years | Data Types | Tables |
| :--- | :--- | :--- | :--- |
| Summer trawl survey | $76,79,82,85,88,91,96,99$, | Abundance | 3 |
| Winter pot survey | $02,06,08,10,11,14,17-21$ | Length-shell comp | 6 |
| Summer commercial fishery | $71-87,89-91,93,95-00,02-12$ | Length-shell comp | 7 |
|  |  | Retained catch | 1 |
| Summer Com total catch | $12-19$ | Standardized CPUE, | 1, Appendix B |
| Summer Com Discards | $87-90,92,94$ | Length-shell comp | 4 |
| Winter subsistence fishery | $76-21$ | Length-shell comp | 9 |
| Winter commercial fishery | $78-21$ | Length-shell comp | 8 |
|  | $15-18$ | Total \& Retained catch | 2 |
| Tag recovery | $80-19$ | Retained catch | 2 |

Data available but not used for assessment

| Data | Years | Data Types | Reason for not used |
| :--- | :--- | :--- | :--- |
| Summer pot survey | $80-82,85$ | Abundance <br> Length proportion <br> Summer preseason survey | Uncertainties on how estimates <br> were made. |
| Summer subsistence | $2005-2019$ | retained catch | Just one year of data <br> fishery |
| Too few catches, ignored. |  |  |  |
| Winter Pot survey | $87,89-91,93,95-$ <br> $00,02-12$ | CPUE | CPUE data unreliable. |
| Preseason Spring pot <br> survey <br> Postseason Fall pot survey | $2011-15$ | CPUE, | Years of data too short |

Catches in other fisheries
In Norton Sound, the directed Pacific cod pot fishery was issued in 2018 under the CDQ permit. In 2018 and 2019 fishery seasons, a total of 8 and 13 kg (mortality applied) of NSRKC were taken from the groundfish fisheries (CPT 2020). However, all of bycatch occurred in the west of 168.0 longitude where none of NSRKC survey has been conducted. Norton Sound Fishery management area (Area Q3) extends to St. Lawrence Island and US-Russia border (Figure 1). In the absence of survey abundance extended to those area, it is questionable whether those bycatch mortalities be induced in the NSRKC population.

|  | Fishery | Data availability |
| :--- | :---: | :---: |
| Other crab fisheries | Does not exist | NA |
| Groundfish pot | Pacific cod | Y |


| Groundfish trawl | Does not exist | NA |
| :--- | :--- | :--- |
| Scallop fishery | Does not exist | NA |

Other miscellaneous data:
Tagging-recovery locations (2012-2019)
Satellite tag migration tracking (NOAA 2016, ADF\&G 2020-21)
Spring offshore migration distance and direction (2012-2015)
Monthly blood hormone level (indication of molting timing) (2014-2015)
Growth increment, molting, and mating of captured crab (2021)

Data aggregated:
Length data were aggregated by 10 mm range, starting from $64-73 \mathrm{~mm}$. Crab length greater than 133 mm were aggregated in $>133 \mathrm{~mm}$ class.
Shell condition data were aggregated to from very new, new, old, very old, very very old to simple newshell and oldshell.
Tag-recovery data were aggregated regardless tagging years.

Data estimated outside the model:
Summer commercial catch standardized CPUE (Table 1, Appendix B)
Proportions of legal size crab, estimated from trawl survey and observer data. (Table 13)
Average weight of crab by length class (Table 13)

## E. Analytic Approach

## 1. History of the modeling approach and issues:

The Norton Sound red king crab stock was assessed using a length-based synthesis model (Zheng et al. 1998). Since adoption of the model, the model had following model mismatches:

1. Model projects higher abundance-proportions of large size class (> 123 mm CL ) of crab than observed. This problem was further exasperated when natural mortality $M$ was set to 0.18 from previous $M=0.3$ in 2011 (NPFMC 2011).
2. Poor model fit to trawl survey abundance. This was further exasperated when M $=0.18$ for all length.

Those issues resulted in model overestimating projected abundance. Following describes historical model adjustments attempted.

1. Model projects higher abundance-proportions of large size class (> 123 mm CL ) of crab than observed.

This issue has been solved by assuming (3-4 times) higher $M$ for the large crab (i.e., $M=0.18$ for length classes $\leq 123 \mathrm{~mm}$, and higher M for $>123 \mathrm{~mm}$ ) (NPFMC 2012, 2013, 2014, 2015, 2016, 2017, 2018). However, because this solution is biologically unusual, several alternative assumptions have been considered in the past.
a. Large crabs move out of the survey and fishing area

In modeling, this was dealt with by setting dome shaped survey and commercial catch selectivity. This modeling configuration resulted in estimating MMB twice higher than the default model (NPFMC 2017). The NOAA NBS surveys (2010, 2017 , 2019) did not find high red king crab population in outer Norton Sound area. The large crab could also be near coastal area where commercial fishery is closed, and trawl survey is not conducted due to rocky bottom.
b. Higher natural mortality $(M)$ than assumed $M=0.18$

Profile analyses and estimating $M$ across all length classes resulted in higher $M$ (0.3-0.45) than default $M=0.18$ (NPFMC 2013, 2017). However, the model fit is slightly worse than the default model. In Tier 4, a higher M also results in higher OFL.
c. Higher natural mortality $(\mathrm{M}>0.18)$ for small and large crab having higher mortality than small crab.

This model configuration had the best fit to data (NPFMC 2017). However, the CPT and SSC rejected the model.
d. Model overestimating molting and growth probability (transition matrix)

The transition matrix was estimated from outside to inside of the assessment model (NPFMC 2014). However, the estimated transition matrix was similar to that estimated outside of the model. Individual length specific molting probability estimates were also similar to default inverse logistic molting function (NPFMC 2016). Time variant molting function (random walk) process did not improve model fit.
e. Gradual size dependent natural mortality.

The default assessment model assumes abrupt M increase at size CL 124 mm or greater. An alternative model suggested that $M$ gradually increasing from size as low as CL 94 mm ; however, the overall model fit did not greatly improve from the default model (NPFMC 2017).
2. Poor model fit to trawl survey abundance, especially NMFS survey (1976-1992) data

In addition to triennial trawl survey, standardized summer commercial catch CPUE was included in the assessment model (NPFMC 2013). Additional variance was also included in standardize CPUE model cv (NPFMC 2014).

In addition, time variant CPUE and trawl survey catchability $(q)$ were included. Thr CPUE q has two values: pre and post 1993, reflecting changes in fishery practices. Trawl survey $q$ was included for NMFS (1976-1992) and NBS (2010-2019), but trawl survey $q$ for ADF\&G trawl survey was assumed to be 1.0. Assuming the NMFS and NBS survey $q$ to be 1.0 resulted in ADF\&G trawl survey $q$ greater than 1.0 (i.e. trawl survey overestimates abundance), even though ADF\&G trawl survey area is generally smaller than NMFS and NBS survey areas.

When $M=0.18$ is assumed for all length classes, the model appears to ignore trawl survey data and assume low abundance. Survey $q$ values for NMFS and NBS surveys increased to 1.65 and 1.28 respectively (Model 21.1). This indicates that survey the trawl surveys overestimated NSRKC abundance.

This model fitting issue was also influenced by input sample sizes for size-shell compositions. Increasing the input sample size would also resulted in the model estimating lower abundance. Reducing the input sample sizes improved model fit to the trawl survey data but caused lower fit to size-shell composition data (NPFMC 2012, 2013, 2015). Alternative model weighting methods (e.g. Francis 2012) have been tried, but those did not improve model fit.

Historical Model configuration progression:
2011 (NPFMC 2011)
1). $M=0.18$.
2). $M$ of the last length class $=0.288$.
3). Include summer commercial discards mortality $=0.2$.
4). Weight of fishing effort $=20$.
5). The maximum effective sample size for commercial catch and winter surveys $=100$.

2012 (NPFMC 2012)

1) $M$ of the last length class $=3.6 \times M$.
2) The maximum effective sample size for commercial catch and winter surveys $=50$.
3) Weight of fishing effort $=50$.

2013 (NPFMC 2013)
4) Standardize commercial catch cpue and replace likelihood of commercial catch efforts to standardized commercial catch cpue with weight $=1.0$.
5) Eliminate summer pot survey data from likelihood.
6) Estimate survey $q$ of 1976-1991 NMFS survey with maximum of 1.0.
7) The maximum effective sample size for commercial catch and winter surveys $=20$.

2014 (NPFMC 2014)

1) Modify functional form of selectivity and molting probability to improve parameter estimates ( 2 parameter logistic to 1 parameter logistic).
2) Include additional variance for the standardized cpue.
3) Include winter pot survey cpue (But was removed from the final model due to lack of fit).
4) Estimate growth transition matrix from tagged recovery data.

2015 (NPFMC 2015)

1) Winter pot survey selectivity is an inverse logistic, estimating selectivity of the smallest length group independently.
2) Reduce weight of tag-recovery: $W=0.5$.
3) Model parsimony: one trawl survey selectivity and one commercial pot selectivity.

2016 (NPFMC 2016)

1) Length range extended from $74 \mathrm{~mm}-124 \mathrm{~mm}$ above to $64 \mathrm{~mm}-134 \mathrm{~mm}$ above.
2) Estimate multiplier for the largest ( $>123 \mathrm{~mm}$ ) length classes.

2017 (NPFMC 2017)

1) Change molting probability function from 1 to 2 parameter logistic. Assume molting probability not reaching 1 for the smallest length class.

2018 (NPFMC 2017) CPT-SSC suggested no model alternatives
2019 (NPFMC 2019)

1) Fit total catch length composition and estimate retention probability for summer and winter commercial fishery.
2) Include winter commercial retained length data.

2020 (NPFMC 2020) The CPT and SSC suggested no model alternatives
2021 (NPFMC 2021) Included discards data. The CPT and SSC requested

1) Models that bridge from the Model 19.0e to 21.0
2) Model 21.0 with natural mortality estimated by model.

## 2. Model Description

a. Description of overall modeling approach:

The model is a male-only size structured model and based on abundance that combines multiple sources of surveys, fishery catches and discards, and markrecovery data using a maximum likelihood modeling framework to estimate population dynamics under fisheries. The model is an extension of the length-based model developed by Zheng et al. (1998) for NSRKC. The detailed description of the model is in Appendix A.

The model estimates abundances of male crab with $\mathrm{CL} \geq 64 \mathrm{~mm}$ and with $10-\mathrm{mm}$ length intervals ( 8 length classes, $\geq 134 \mathrm{~mm}$ ) because few crab measuring less than 64 mm CL were caught during surveys or fisheries.

The model assumes newshell crab as molted and oldshell crab as unmolted.
One critical characteristic of the model is that the model does not estimate fishing mortality $(F)$. Observed harvests were considered accurate and thus directly subtracted from the model estimated abundance.

The modeling scheme and data is described in the following figure.


Figure: Norton Sound Red king crab model and data scheme. Bold type indicate data that were fitted to the model. Boxes in dotted line indicate model estimated parameters and quantities.

Natural mortality, M was set to 0.18 except for CL greater than 123 mm that was estimated in the model.

Timeline of calendar events and crab modeling events:

- Model year starts February $1^{\text {st }}$ to January $\mathbf{3 1}^{\text {st }}$ of the following year.
- Initial Population Date: February $1^{\text {st }}, 1976$, consisting only newshell crab.
- Instantaneous fishing mortality: winter (February $1^{\text {st }}$ ) and summer (July $1^{\text {st }}$ ) fisheries
- Instantaneous molting and recruitment occur on July $1^{\text {st }}$
- Critical model assumptions

NSRKC Crab Biology

1. Instantaneous annual natural mortality $(M)$ is 0.18 and increase at the size greater than CL $123 \mathrm{~mm} . M$ is constant over time.
2. Male crab size at maturity is 94 mm CL.
3. Molting occurs in right after the summer fishery.
4. Recruitment occurs in fall at the same time of molting.

In NSRKC assessment modeling, recruitment is not a function of mature males, but estimated model parameters entering to the immature length classes $64 \mathrm{~mm}-$ 93 mm . In modeling, this adjustment is done at the same time of molting-growth.
5. Molting probability is an inverse logistic function. Molted crab become newshell and unmolted crab become oldshell crab.
6. Growth increment is a function of length, constant over time. Molted crab does not shrink.

NSRKC Survey

1. $\mathrm{ADF} \& \mathrm{G}$ trawl survey abundance has the same scale as the population (i.e. catchability $\mathrm{q}=1.0$ ). Abundances by historical NMFS (1976-1992) and NOAA NBS (2010-present) survey are biased low (i.e. $q<1.0$ ).

Lower survey q for NOAA survey was recommended in 2013 assessment (NPFMC 2013). Model estimated survey q for ADF\&G trawl survey was greater than 1.0 (NPFMC 2013, 2019). The CPT and SSC recommended fixing the survey q of the ADF\&G survey to 1.0.
2. Size selectivity is an asymptotic one parameter logistic function of 1.0 at the length class CL 134 mm and the same across years and survey agencies.

$$
S_{l}=\frac{1}{1+e^{\left(\alpha\left(L_{\max }-L\right)+\ln (1 / 0.999-1)\right)}}
$$

This logistic function form was adopted during the crab workshop in 2005 as an way to reduce model parameters and parameter estimation stability.

Although the surveys differ among NOAA (1976-1991), ADF\&G (1996-present), and NOAA NBS (2010-2019) in terms of survey vessel and trawl net structure, selectivity of all surveys were assumed to be identical. Model fits separating and combining the surveys were examined in 2015; however, selectivity was essentially identical (1.0 across all size classes) (NPFMC 2015). For model parsimony, the SSC recommended that using only one selectivity.
3. Winter pot survey selectivity is a dome shaped function: a combination of a reverse logistic function of starting from length class CL 84 mm and model estimate for $\mathrm{CL}<84 \mathrm{~mm}$ length classes. The selectivity is constant over time.

Winter pot survey selectivity is a dome shaped function: a combination of a reverse logistic function of starting from length class CL 84 mm and model estimate for $\mathrm{CL}<84 \mathrm{~mm}$ length classes. Selectivity is constant over time.

$$
S_{w, l}=\frac{1}{1+e^{\alpha(L-\beta)}}
$$

This assumption is based on the fact that a low proportion of large crab are caught in the nearshore area where winter surveys occur. This does not necessarily imply that the crab pots are less selective to large crabs. Alternatively, this may imply that fewer large crab migrate into nearshore waters in winter.

NSRKC Fisheries

1. Fisheries occur twice on July 01 and Feb 01 and instantaneous.
2. Summer commercial fishery size selectivity is an asymptotic one parameter logistic function of 1.0 at the length class CL 134 mm . Selectivity is constant overtime.

$$
S_{l}=\frac{1}{1+e^{\left(\alpha\left(L_{\max }-L\right)+\ln (1 / 0.999-1)\right)}}
$$

This logistic function form was adopted during the crab workshop in 2005 as a way to reduce model parameters and parameter estimation stability. Although summer commercial fishery changed greatly among the periods (1977-1992, 1993-present) in terms of fishing vessel composition, pot configuration the selectivity of each period was assumed to be identical. Model fits of separating and combining the two periods were examined in 2015 and showed no difference between the two (NPFMC 2015). For model parsimony, the SSC recommended using only one selectivity.
3. Winter commercial pot selectivity is the same as the selectivity of the winter pot survey.

This assumption is based on the fact that the survey pot was similar to the one used for subsistence, and that many commercial fishermen are also catch for subsistence. However, by regulation winter commercial king crab pots can be any dimension (5AAC 34.925(d)) and recent popularity of winter commercial fishery may have deviated this assumption.
4. Winter subsistence fishery retains crab size greater than 94 mm .
5. Subsistence fishery does not have retainable size limit, so that we assumed that it retains crab smaller than legal sized crab ( $\sim 104 \mathrm{~mm}$ CL)
6. Discards handling mortality rate for all fisheries is $20 \%$.

Discards mortality was specified by CPT. No empirical estimates are available.

## Data quality

All size-shell composition data are collected accurately without systematic bias.
Annual retained catch is accurate without error.

In Norton Sound, almost all crabs are sold to NSEDC. This ensures accuracy of harvest.

Model data weighting

Survey data Input sample size

| Summer commercial, winter pot, <br> and summer observer | minimum of $0.1 \times$ actual sample size or 10 |
| :--- | :--- |
| Summer trawl and pot survey | minimum of $0.5 \times$ actual sample size or 20 |
| Tag recovery | $0.5 \times$ actual sample size |

Recruitment SD: 0.5.
Discards CV: 0.3
"Implied" effective sample sizes were calculated as

$$
n=\sum_{l} \hat{P}_{y, l}\left(1-\hat{P}_{y, l}\right) / \sum_{l}\left(P_{y, l}-\hat{P}_{y, l}\right)^{2}
$$

Where ${ }_{P_{y, l}}$ and $\hat{P}_{y, l}$ are observed and estimated length compositions in year $y$ and length group $l$, respectively. Estimated implied effective sample sizes vary greatly over time.

Changes of assumptions since last assessment:
None

## 3. Model Selection and Evaluation

a. Description of alternative model configurations.

For the 2021 final assessment, the CPT and SSC adopted model 19.0b (discard abundance estimated by the LNR2 method). For the 2022 draft assessment, we proposed model 21.0. For the 2022 final assessment, the CPT and SSC recommended following alternative models:

Model 19.0e: with updated data
Model 21.0: Model 19.0e+ St CPUE with 3qs +2 summer commercial retention probabilities.
Model 21.1: Model 21.0 with $\mathrm{M}=0.18$ for all length size classes.
Model 21.2: Model 19.0e + St CPUE data updated with 3 qs.
Model 21.3: Model 19.0e+2 summer commercial retention probability Model 21.4: Model 21.0 with M estimated equally for all length size classes
Model 21.5: Model 21.0 with M estimated for two length size classes $(<124 \mathrm{~mm}$, $>123 \mathrm{~mm} \mathrm{CL}$ ).

Models 21.2 and 21.3 are "bridging analyses" from model 19.0e to 21.0 requested by the CPT. Models 21.4 and 21.5 examine appropriateness of the $M=0.18$ assumptions. Model 21.4 assumes single $M$ for all size classes, and model 21.5 assumes size dependent $M$ same as the model 21.0

| Model | Discards <br> Est | St.CPUE <br> Q | Retention <br> probability | M | Parameters |
| :--- | :--- | :--- | :--- | :--- | :--- |
| 19.0 e | Prop | 1 | 1 | $0.18+$ est (L) |  |
| 21.0 | Prop | 3 | 2 | $0.18+$ est (L) | +4 |
| 21.1 | Prop | 3 | 2 | 0.18 | +3 |
| 21.2 | Prop | 1 | 2 | $0.18+$ est (L) | +2 |
| 21.3 | Prop | 3 | 1 | $0.18+$ est (L) | +2 |
| 21.4 | Prop | 3 | 2 | Est | +4 |
| 21.5 | Prop | 3 | 2 | Est (S,L) | +5 |

The updates are mostly due to changes in data inputs, which resulted in an increase of the number of parameters.

1. Revision of fishery time period and revised the start year of high grading fishery period from 2005 to 2008, after review of management documents
2. Updates in standardized CPUE from a single model with 3 block periods (19771993, 1994-2007, 2008-2019) to 3 models for each period. This increased the model fishery q parameter from 1 to 3 . In the 2019 model, the 3 periods were included as block effect of a single glm that was suggested by the CIE review. However, this was deemed to be incorrect.
3. Increase of summer and winter commercial retention probabilities from 1 model for each fishery to 2 model for each fishery, indicating changes in retention probability before and after high grading fishery. Model 19.0: indicating pre and post high grading in commercial fishery.
4. Models 21.2, 21.3 are intermediate transition from model 19.0e to 21.0 , requested by the CPT and SSC.
5. Comparisons among Models $21.0,21.1,21.4$, and 21.5 are proposed to examine the effects of $M$ assumptions, requested by the SSC.
b. Evaluation of negative log-likelihood values.

Table L. Negative log likelihood for alternative models.

|  |  |  |  |  |  |  |  | Sept <br> $\mathbf{2 0 2 1}$ | Sept <br> $\mathbf{2 0 2 1}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |


| Model | 19.0e | 21.0 | 21.1 | 21.2 | 21.3 | 21.4 | 21.5 | 21.0 | 21.1 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Additional Parameters |  | +4 | +3 | +2 | +2 | +4 | +5 | +4 | +3 |
| Total | 346.67 | 340.05 | 396.6 | 341.58 | 345.08 | 350.20 | 337.31 | 320.39 | 379.73 |
| TSA | 11.46 | 11.78 | 34.9 | 11.71 | 11.48 | 10.97 | 11.03 | 11.33 | 32.70 |
| St.CPUE | -17.21 | -23.57 | -19.0 | -23.73 | -17.25 | -24.19 | -24.45 | -36.35 | -28.04 |
| TLP | 129.81 | 129.67 | 133.30 | 129.94 | 129.66 | 132.03 | 128.34 | 124.89 | 128.07 |
| WLP | 39.30 | 38.88 | 40.75 | 38.91 | 39.28 | 39.78 | 38.87 | 38.51 | 40.04 |
| CLP | 50.00 | 48.28 | 62.55 | 49.60 | 48.64 | 51.49 | 47.79 | 48.39 | 64.28 |
| OBS | 24.11 | 24.36 | 25.47 | 24.23 | 24.20 | 27.93 | 25.10 | 24.49 | 25.06 |
| REC | 2.77 | 2.77 | 2.56 | 2.77 | 2.78 | 2.14 | 2.59 | 2.81 | 2.54 |
| WN | 18.95 | 20.36 | 20.14 | 20.40 | 18.95 | 20.78 | 21.07 | 20.15 | 19.22 |
| DIS | 3.40 | 3.59 | 3.89 | 3.68 | 3.30 | 3.07 | 3.20 | 3.61 | 4.03 |
| TAG | 84.13 | 83.98 | 92.04 | 84.14 | 84.03 | 86.21 | 83.77 | 82.55 | 91.39 |
| RMSE Trawl | 0.35 | 0.35 | 0.56 | 0.35 | 0.35 | 0.33 | 0.34 |  |  |
| RMSE CPUE | 0.45 | 0.40 | 0.43 | 0.39 | 0.46 | 0.39 | 0.40 |  |  |
| BMSY(mil.lb) | 4.21 | 4.17 | 2.40 | 4.20 | 4.17 | 4.21 | 4.30 | 4.48 | 2.55 |
| $\begin{array}{r} \hline \text { MMB } 2022 \\ \text { (mill.lb) } \end{array}$ | 5.17 | 5.33 | 4.90 | 5.17 | 5.35 | 4.70 | 4.79 | 5.22 | 4.85 |
| Retainable Crab (mil.lb) | 4.11 | 4.27 | 3.93 | 4.11 | 4.28 | 3.56 | 3.76 |  |  |
| Discards Crab (mil.lb) | 1.22 | 1.24 | 1.03 | 1.21 | 1,24 | 1.11 | 1.12 |  |  |
| $\begin{array}{r} \text { Total } \\ \text { OFL(mil.lb) } \end{array}$ | 0.92 | 0.89 | 0.64 | 0.89 | 0.88 | 1.12 | 0.96 |  |  |
| M | $\begin{aligned} & \hline 0.18 \\ & 0.62 \end{aligned}$ | $\begin{aligned} & \hline 0.18 \\ & 0.61 \end{aligned}$ | 0.18 | $\begin{aligned} & \hline 0.18 \\ & 0.62 \end{aligned}$ | $\begin{aligned} & \hline 0.18 \\ & 0.61 \end{aligned}$ | 0.42 | $\begin{aligned} & \hline 0.26 \\ & 0.59 \end{aligned}$ | $\begin{aligned} & \hline 0.18 \\ & 0.62 \end{aligned}$ | 0.18 |

TSA: Trawl Survey Abundance
St. CPUE: Summer commercial catch standardized CPUE
TLP: Trawl survey length composition:
WLP: Winter pot survey length composition
CLP: Summer commercial retention catch length composition
REC: Recruitment deviation
OBS: Summer commercial catch observer discards (Baseline) or total catch (Alternative models) length composition
TAG: Tagging recovery data composition
WN: Winter commercial length-shell composition
DIS: Summer commercial discards abundance

## 4. Results

Models 21.2 and 21.3 are bridge analyses from model 19.0 e to 21.0 that corrected standardized CPUE (Appendix C) and periods and types of retention probability based on reevaluation of fishery history. Those updates are considered as data correction, and the bridge model were intended to show how those data corrections would change NSRKC population
status. Changes in st.CPUE (19.0e vs. 21.2) increased overall abundance, MMB, and biomass (Figure 4, Table 14). This can also be manifested in decline of the NMFS survey q (q.1) from 0.805 (model 19.0e) to 0.772 (model 21.2) (Table 11). Simultaneously, the revised data also lowered projected 2022 MMB , which resulted in lower OFL (Table L). Separating retention probability before and after high grading (Model 21.0 vs. Model 21.2, Model 19.0 e vs. Model 21.3) showed different retention probability patterns in which the latter period had lower retention probability of $94-104 \mathrm{~mm}$ crab from 0.09 to 0.01 (Figure 3, Table 13). This also slightly improved model fit to summer commercial retention size proportion (CLP) (Table L). However, the dynamics of population and MMB for the two models were almost identical (Figure 4, 5).

Comparisons of models among 21.0, 21.4 and 21.5, center around exploration of alternative NSRKC life-history assumptions in the absence of studies verifying validity of natural mortality assumptions. Model 21.0 assumes $M=0.18$ and higher $M$ for $>123 \mathrm{~mm}$ CL crab, whereas Model 21.4 assumes single $M$ for all size classes and is estimated from the model. Model 21.5 allows the model to estimate two $M$ s for below and above 123 mm CL. As expected, models 21.4 and 21.5 estimated higher $M$ (Table L). Estimated $M$ of model 21.4 was 0.42 that was 2.3 times higher than default $M=0.18$. Estimated $M$ of the model 21.5 was 0.26 for under 124 mm CL and 0.59 for over 123 mm CL. Those values were closer to the model 21.0 of 0.18 and 0.62 (Table L). Those suggest that the data are explained by sizedependent $M$ rather than a single $M$ for all lengths. The shapes of estimated molting probability, fishery selectivity, and retention probability also differed greatly among the models (Figure 3, Table 13). Model 21.4 estimated slightly lower molting probability for larger crab, asymptotic selectivity for trawl survey, and lower selectivity for small crab in winter pot, and summer commercial fisheries (Figure 3, Table 13). Molting probability and selectivity profiles of model 21.5 were closer to those of the model 21.0 (Figure 3, Table 13). Regarding to fit to trawl survey abundance and standardized CPUE, the models 21.4 and 21.5 were slightly better than 21.0 (Figure 7, 8, 16). RMSE of the models 21.4 and 21.5 were 0.01-0.02 point smaller than the model 21.0 (Table L). As for fit to size-shell data, estimating single $M$ (models 21.1 and 21.4) showed lower fit to trawl and commercial catch shell-length compositions and tag-recovery size composition than size dependent M (models 21.0 and 21.5) (Table L). However, the differences among the models were visually very slight (Figures $8,9,10,11,12$, and 13). For trawl survey data, all models underestimated sizeproportion of larger and oldshell crab (Figure 9). For summer commercial retained data, all models overestimated the proportion of oldshell crab (Figure 8). Interestingly, model fit to total catch shell-size compositions appear to be better than fits to trawl and retained shell-size compositions (Figure 11).

Aside from the model fits, the great differences among models were estimates of abundance and MMB (Figure 4, 5). Model 21.4 had the highest mean abundance for total ( 4.1 million) and recruits ( 2.0 million), followed by model 21.5 ( 3.4 million and 1.4 million) and model 21.0 ( 3.1 million and 1.2 million). On the other hand, abundance of legal sized crab and MMB were similar among the 3 models (Figure 4, Table L). Among these three models, model 21.5 had the lowest negative log likelihood values, followed by model 21.0 (340.03) and model 21.4 (350.20) (Table L). This indicates that model 21.5 had the best model fit to the observed data, although difference of the fits among the models are small.

Retrospective analyses showed that Mohn's rho is the smallest for the model 21.4 (0.182), followed by model $21.5(0.191)$ and model 18.0 (0.209) (Figure 17). Hurtado-Ferro et al. (2015) provided a guideline that Mohn's rho exceeding the range of ( -0.15 to 0.2 ) for longer living species and ( -0.22 to 0.30 ) for shorter living species should cause for concern. It is unknown whether NSRKC can be considered having a shorter or longer living. If it's considered longer living, model 21.0 should cause for concern.

Similar model fits among the 3 models $(21.0,21.4,21.5)$ suggests that the observed data can be explained by several alternative life-history assumptions: high size independent mortality (model 21.4) size dependent mortality (models 21.0, and 21.5), or high migration of large crabs that explored previously (NPFMC 2017). For other NSRKC life-history traits that influence model fitting, such as time-invariant natural mortality, molt and recruitment timing, newshell-oldshell transition scheme (i.e., newshell crab $=$ molted, oldshell crab $=$ unmolted crab), only molt timing was confirmed outside the model.

Since inception of the model, the greatest discrepancies are overestimating the larger and oldshell crabs in trawl and summer commercial retained (Figures 8, 9). Some of those misfits could be due to uncertainties about shell conditions as well as selective discards of oldshell by fishermen.

Tag-recovery data (2012-2016) had 125 crabs that had no growth ( $+/-3 \mathrm{~mm}$ ) in one year of liberty. Of those 100 crabs were released as newshell and 25 crabs were released as oldshell. If no growth is considered unmolted, all those crabs should be recaptured as oldshell. However, $29 \%$ of crabs released as newshell were recaptured as newshell crab and $48 \%$ of crabs released as oldshell were recaptured as newshell.

Table: The number of tagged NSRKC released and recovered in 1 year with no ( $+/-3 \mathrm{~mm}$ ) changes in size (2012-2016)

| Released $\backslash$ Recovered | Newshell | Oldshell |
| :--- | :--- | :--- |
| Newshell | 29 | 71 |
| Oldshell | 12 | 13 |

Those crabs may be molted but did not grow, or did not molt but were incorrectly identified as newshell.

In the assessment model, retention probability is based on size; however, fishermen is more favor or retaining newshell than oldshell. Observer data showed that $21 \%$ of retained large crabs ( $>110 \mathrm{~mm}$ ) were oldshell, whereas $51 \%$ of discarded large crabs were oldshell crab.

Table: Distribution of $>110 \mathrm{~mm}$ CL crabs in observer survey

|  | Newshell | Oldshell |
| :--- | :--- | :--- |
| Retained | 6037 | 1618 |
| Discarded | 361 | 381 |

For the final 2022 assessment model, the authors recommend model 21.0 or model 21.5. The major difference between the two models is whether to specify $M=0.18$ of size classes under 124 mm CL. Between the two models, model 21.5 is better in retrospective analyses.

## F. Calculation of the OFL

1. Specification of the Tier level and stock status.

The NSRKC stock is placed in Tier 4. It is not possible to estimate the spawner-recruit relationship, but some abundance and harvest estimates are available to build a computer simulation model that captures the essential population dynamics. Tier 4 stocks are assumed to have reliable estimates of current survey biomass and instantaneous $M$; however, the estimates of M for the SRKC stock are uncertain.

At the Tier 4 level the OFL are determined by the $F_{M S Y}$ proxy, $B_{M S Y}$ proxy, and estimated legal male abundance and biomass:

| Level | Criteria | $F_{O F L}$ |
| :--- | :--- | :--- |
| A | $B / B_{M S Y^{p r o x}}>1$ | $F_{O F L}=\gamma M$ |
| B | $\beta<B / B_{M S Y^{p m a x}} \leq 1$ | $F_{O F L}=\gamma M\left(B / B_{M S Y^{\text {prax }}}-\alpha\right) /(1-\alpha)$ |
| C | $B / B_{M S Y^{\text {pox }}} \leq \beta$ | $F_{O F L}=$ bycatch mortality \& directed fishery $F=0$ |

where $B$ is a mature male biomass (MMB), $B_{M S Y}$ proxy is average mature male biomass over a specified time period, $M=0.18, \gamma=1, \alpha=0.1$, and $\beta=0.25$.

For NSRKC, MMB is defined as the biomass of males $>94 \mathrm{~mm}$ CL on February 01 (Appendix A). $B_{M S Y}$ proxy is
$B_{M S Y}$ proxy $=$ average model estimated MMB from 1980-2022.

## Estimated $B_{M S Y}$ proxy : $\quad$ Model 21.0: 4.17 million lb or 1.90 k t.

Model 21.5: 4.30 million lb or 1.95 k t.

Predicted mature male biomass in 2022 on February 01

Mature male biomass: $\quad$ Model 21.0: $\mathbf{5 . 3 3}$ million lb or 2.42 k t.
Model 21.5: 4.79 million lb or 2.17 kt .

Since the projected MMB is above $B_{M S Y}$ proxy,

The NSRKS red status is Tier 4a,

And $F_{O F L}$ for calculation of the OFL is $F_{O F L}=\gamma M$
2. Calculation formula of NSRKC OFL.

OFL of NSRKC is total OFL $\left(O F L_{T}\right)$ that is a sum of the retained and unretained OFL $\left(O F L_{r}\right.$, $O F L_{n r}$ ).

$$
O F L_{T}=O F L_{r}+O F L_{u r}
$$

where

$$
O F L_{r}=\text { retained }_{-} B \cdot F_{O F L} \text { and } O F L_{n r}=\text { unretained }_{-} B \cdot F_{O F L} \cdot h m
$$

retained $B$ is a biomass of crab subject to fisheries that is a sum of the products of crab abundance $\left(N_{w, l}+O_{w, l}\right)$, fishery selectivity $\left(S_{s, l}\right)$, retention probability ( $S_{r, l}$ ), and average weight lb (wmi) by length class ( $l$ ).
retained_B $=\sum_{l}\left(N_{w, l,}+O_{w, l,}\right) S_{s, l} S_{r, l} w m_{l}$
uretained_B is a biomass of crab subject to fisheries and is a sum of the products of crab abundance ( $N_{w, l}+O_{w, l}$ ), fishery selectivity $\left(S_{s, l}\right), 1$ minus retention probability ( $S_{r, l}$ ), and average weight lb ( $w m_{l}$ ) by length class ( $l$ ).
unretained_B $=\sum_{l}\left(N_{w, l,}+O_{w, l,}\right) S_{s, l}\left(1-S_{r, l}\right) w m_{l}$
$h m$ is handling mortality of default 0.2

The NSRKC fishery consists of two distinct fisheries: winter and summer. The two fisheries are discontinuous with 5 months ( 0.42 year) between the two fisheries during which natural mortality occurs. To estimate the OFL for the two fisheries, the CPT in 2016 recommended the following formula that sum of winter and summer catch $(H w, H s)$ equals to total OFL (OFL $=H w+H s)$ and that winter catch is a fraction ( p ) of total OFL: $H_{w}=p$. OFL, where $\boldsymbol{p}$ is predetermined fraction of the winter fishery to total fishery.
where

$$
\begin{aligned}
& H_{w}=B_{w}\left(1-\exp \left(-x \cdot F_{O F L}\right)\right), \\
& H_{s}=B_{s}\left(1-\exp \left((1-x) \cdot F_{O F L}\right)\right), \text { and } \\
& B_{s}=\left(B_{w}-H w\right) e^{-0.42 \cdot M}
\end{aligned}
$$

$B_{w}$ is the winter NSRKC biomass, $B_{s}$ is the summer NSRKC biomass, and $x$ is a fraction parameter,

Solving $x$ of the above (see Appendix A for derivation), retained and unretained OFL is calculated as:

$$
\text { OFL }_{r}=\text { retained }_{-} B_{w}\left(1-e^{-\left(F_{\text {orl }}+0.42 M\right)}-\left(1-e^{-0.42 M}\right)\left(\frac{1-p \cdot\left(1-e^{-\left(F_{\text {OFL }}+0.42 M\right)}\right)}{1-p \cdot\left(1-e^{-0.42 M}\right)}\right)\right)
$$

and

$$
O F L_{n r}=\text { unretained } B_{w}\left(1-e^{-\left(F_{\text {ofi }}+0.42 M\right)}-\left(1-e^{-0.42 M}\right)\left(\frac{1-p \cdot\left(1-e^{-\left(F_{\text {OFL }}+0.42 M\right)}\right)}{1-p \cdot\left(1-e^{-0.42 M}\right)}\right)\right) \cdot h m
$$

Because $M$ of NSRKC is length-dependent, so that proper calculation of NSRKC OFL should account for length-dependent $M$ as:

$$
\text { OFL } L_{r}=\sum_{l}\left[\text { retained } B_{w, l}\left(1-e^{-\left(F_{\text {oF }, l}+0.42 M_{l}\right)}-\left(1-e^{-0.42 M_{l}}\right)\left(\frac{1-p \cdot\left(1-e^{-\left(F_{\text {orl }, l}+0.42 M_{l}\right)}\right)}{1-p \cdot\left(1-e^{-0.42 M_{l}}\right)}\right)\right)\right]
$$

and

$$
O F L_{u r}=\sum_{l}\left[\text { unretained } B_{w, l}\left(1-e^{-\left(F_{\text {ofLl } l}+0.42 M_{l}\right)}-\left(1-e^{-0.42 M_{l}}\right)\left(\frac{1-p \cdot\left(1-e^{-\left(F_{\text {oFLl }}+0.42 M_{l}\right)}\right)}{1-p \cdot\left(1-e^{-0.42 M_{l}}\right)}\right)\right)\right] \cdot h m
$$

where $M_{l}$ is a size specific natural mortality,

The SSC recommended the length-dependent $M$ and $F_{\text {OFL }}$ for calculation of NSRKC OFL, but reversed to applying length-independent $F_{\text {OFL }}$ because of uncertainties about applying lengthdependent $F_{\text {OFL }}$ for Tire 4 stock, but more over to the fact that the length-dependent $F_{\text {OFL }}$ increased the value of OFL greatly from the previous year (NPFMC 2017, 2018). In 2021 the SSC (2021 Oct) said that "The rationale that it may result in a higher OFL should not prevent exploring a higher value for M if that may be the best description of the dynamics." (Appendix D). We welcome the SSC's revised opinion on this issue and recommend length-dependent $F_{\text {OFL }}$ for calculation of NSRKC OFL. This is a logical extension of the revised position should NSRKC assessment model estimates length-dependent $M$ (e.g., model 21.5). Length-dependent $F_{\text {OFL }}$ applied to NSRKC are $0.18(\mathrm{CL}<124 \mathrm{~mm})$ and $0.61(\mathrm{CL}>123 \mathrm{~mm})$ for model 21.0 and $0.26(\mathrm{CL}<124 \mathrm{~mm})$ and $0.59(\mathrm{CL}>123 \mathrm{~mm})$ for model 21.5.
3. Determination of NSRKC OFL for the 2022 fishery season.

Projected legal male biomass catchable to fishery and discards in 2022 are
Model 21.0: 4.27 and 1.24 million lb or 1.94 and 0.56 kt .
Model 21.5: 3.75 and 1.12 million lb or 1.70 and 0.51 k t.

With specified $p=0.16$. Total OFL of NSRKC for 2022 fishery is
OFL $=$
Model 21.0: 0.89 million lb or 0.40 kt .
Model 21.5: 0.96 million lb or 0.44 kt .

## G. Calculation of the $A B C$

1. Specification of the probability distribution of the OFL.

ABC is calculated as (1-ABC buffer).OFL
In 2015 ABC buffer of Norton Sound Red King Crab was set to $20 \%$, which was increased to $30 \%$ in 2020 and to $40 \%$ in 2021.

Applying the $40 \%$ buffer, NSRKC ABC for the 2022 fishery is

$$
\begin{aligned}
& \mathrm{ABC}= \\
& \quad \text { Model 21.0: } 0.534 \text { million lb or } 0.24 \mathrm{kt} .
\end{aligned}
$$

Model 21.5: 0.576 million lb or 0.26 k t.

Taking further account of uncertainty of length-dependent $M$ and $F_{O F L}$, the ABC buffer can be increased to $55 \%$ as:
$\mathrm{ABC}=$
$\quad$ Model 21.0: 0.400 million lb or 0.18 k ton.
Model 21.5: $\mathbf{0 . 4 3 2}$ million lb or 0.20 k ton.

Incidentally, the $55 \%$ buffer of the model 21.0 corresponds to ABC $40 \%$ buffer, if OFL were calculated with default length-independent $F_{\text {OFL }}$. OFL of the model 21.0 with lengthindependent $F_{\text {OFL }}$ is $\mathbf{0 . 6 7}$ million $\mathbf{l b}$ or $\mathbf{0 . 3 0} \mathbf{k} \mathbf{t}$, and ABC with $40 \%$ buffer is $\mathbf{0 . 4 0 2}$ million $\mathbf{l b}$ or 0.18 k t.

## H. Rebuilding Analyses

Not applicable

## I. Data Gaps and Research Priorities

The major data gap of NSRKC is understandings of its biology, including natural mortality, size at maturity, spatial and temporal distribution and abundance, molting frequency and growth, as well as female abundance, fecundity, size at maturity, mating timing, spatial-temporal distribution and abundance. Specifically, the model assumes size dependent natural mortality (i.e. high natural mortality of $>123 \mathrm{~mm}$ ). Further missing is analyses of LK/TK and socioeconomic impacts of NSRKC fisheries that could bring further insights about NSRKC biology and could be greatly significant in determination of management matrix such as ABC buffer.

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Table 1. Historical summer commercial red king crab fishery economic performance, Norton Sound Section, eastern Bering Sea. Bold type shows data that are used for the assessment model.

| Year | Guideline Harvest Level (lb) ${ }^{\text {b }}$ | Commercial Harvest (lb) ${ }^{\text {a, }}$ b |  | Number Harvest | Total Number (Open Access) |  |  | Total Pots |  | ST CPUE |  | Season Length |  | Midday from July |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Open |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  | Access | CDQ |  | Vessels | Permits | Landings | Registered | Pulls | CPUE | SD | Days | Dates |  |
| 1977 | c | 517.787 |  | 195,877 | 7 | 7 | 13 |  | 5,457 | 3.29 | 0.68 | 60 | c | 0.049 |
| 1978 | 3,000.000 | 2,091.961 |  | 660,829 | 8 | 8 | 54 |  | 10,817 | 4.68 | 0.65 | 60 | 6/07-8/15 | 0.142 |
| 1979 | 3,000.000 | 2,931.672 |  | 970,962 | 34 | 34 | 76 |  | 34,773 | 2.87 | 0.64 | 16 | 7/15-7/31 | 0.088 |
| 1980 | 1,000.000 | 1,186.596 |  | 329,778 | 9 | 9 | 50 |  | 11,199 | 3.07 | 0.65 | 16 | 7/15-7/31 | 0.066 |
| 1981 | 2,500.000 | 1,379.014 |  | 376,313 | 36 | 36 | 108 |  | 33,745 | 0.86 | 0.64 | 38 | 7/15-8/22 | 0.096 |
| 1982 | 500.000 | 228.921 |  | 63,949 | 11 | 11 | 33 |  | 11,230 | 0.2 | 0.62 | 23 | 8/09-9/01 | 0.151 |
| 1983 | 300.000 | 368.032 |  | 132,205 | 23 | 23 | 26 | 3,583 | 11,195 | 0.9 | 0.65 | 3.8 | 8/01-8/05 | 0.096 |
| $\backslash 1984$ | 400.000 | 387.427 |  | 139,759 | 8 | 8 | 21 | 1,245 | 9,706 | 1.59 | 0.65 | 13.6 | 8/01-8/15 | 0.110 |
| 1985 | 450.000 | 427.011 |  | 146,669 | 6 | 6 | 72 | 1,116 | 13,209 | 0.5 | 0.66 | 21.7 | 8/01-8/23 | 0.118 |
| 1986 | 420.000 | 479.463 |  | 162,438 | 3 | 3 |  | 578 | 4,284 | 1.74 | 0.7 | 13 | 8/01-8/25 | 0.153 |
| 1987 | 400.000 | 327.121 |  | 103,338 | 9 | 9 |  | 1,430 | 10,258 | 0.61 | 0.64 | 11 | 8/01-8/12 | 0.107 |
| 1988 | 200.000 | 236.688 |  | 76,148 | 2 | 2 |  | 360 | 2,350 | 2.36 | 0.86 | 9.9 | 8/01-8/11 | 0.110 |
| 1989 | 200.000 | 246.487 |  | 79,116 | 10 | 10 |  | 2,555 | 5,149 | 1.21 | 0.61 | 3 | 8/01-8/04 | 0.096 |
| 1990 | 200.000 | 192.831 |  | 59,132 | 4 | 4 |  | 1,388 | 3,172 | 1.08 | 0.68 | 4 | 8/01-8/05 | 0.099 |
| 1991 | 340.000 |  |  | 0 |  | ummer | hery |  |  |  |  |  |  |  |
| 1992 | 340.000 | 74.029 |  | 24,902 | 27 | 27 |  | 2,635 | 5,746 | 0.17 | 0.6 | 2 | 8/01-8/03 | 0.093 |
| 1993 | 340.000 | 335.790 |  | 115,913 | 14 | 20 | 208 | 560 | 7,063 | 0.9 | 0.35 | 52 | 7/01-8/28 | 0.093 |
| 1994 | 340.000 | 327.858 |  | 108,824 | 34 | 52 | 407 | 1,360 | 11,729 | 0.81 | 0.34 | 31 | 7/01-7/31 | 0.044 |
| 1995 | 340.000 | 322.676 |  | 105,967 | 48 | 81 | 665 | 1,900 | 18,782 | 0.42 | 0.34 | 67 | 7/01-9/05 | 0.093 |
| 1996 | 340.000 | 224.231 |  | 74,752 | 41 | 50 | 264 | 1,640 | 10,453 | 0.51 | 0.34 | 57 | 7/01-9/03 | 0.101 |
| 1997 | 80.000 | 92.988 |  | 32,606 | 13 | 15 | 100 | 520 | 2,982 | 0.84 | 0.35 | 44 | 7/01-8/13 | 0.074 |
| 1998 | 80.000 | 29.684 | 0.00 | 10,661 | 8 | 11 | 50 | 360 | 1,639 | 0.79 | 0.36 | 65 | 7/01-9/03 | 0.110 |
| 1999 | 80.000 | 23.553 | 0.00 | 8,734 | 10 | 9 | 53 | 360 | 1,630 | 0.92 | 0.36 | 66 | 7/01-9/04 | 0.104 |
| 2000 | 336.000 | 297.654 | 14.87 | 111,728 | 15 | 22 | 201 | 560 | 6,345 | 1.24 | 0.34 | 91 | 7/01-9/29 | 0.126 |
| 2001 | 303.000 | 288.199 | 0 | 98,321 | 30 | 37 | 319 | 1,200 | 11,918 | 0.64 | 0.34 | 97 | 7/01-9/09 | 0.104 |
| 2002 | 248.000 | 244.376 | 15.226 | 86,666 | 32 | 49 | 201 | 1,120 | 6,491 | 1.23 | 0.34 | 77 | 6/15-9/03 | 0.060 |
| 2003 | 253.000 | 253.284 | 13.923 | 93,638 | 25 | 43 | 236 | 960 | 8,494 | 0.85 | 0.34 | 68 | 6/15-8/24 | 0.058 |
| 2004 | 326.500 | 314.472 | 26.274 | 120,289 | 26 | 39 | 227 | 1,120 | 8,066 | 1.27 | 0.34 | 51 | 6/15-8/08 | 0.033 |
| 2005 | 370.000 | 370.744 | 30.06 | 138,926 | 31 | 42 | 255 | 1,320 | 8,867 | 1.19 | 0.34 | 73 | 6/15-8/27 | 0.058 |
| 2006 | 454.000 | 419.191 | 32.557 | 150,358 | 28 | 40 | 249 | 1,120 | 8,867 | 1.31 | 0.34 | 68 | 6/15-8/22 | 0.052 |
| 2007 | 315.000 | 289.264 | 23.611 | 110,344 | 38 | 30 | 251 | 1,200 | 9,118 | 1.02 | 0.34 | 52 | 6/15-8/17 | 0.036 |
| 2008 | 412.000 | 364.235 | 30.9 | 143,337 | 23 | 30 | 248 | 920 | 8,721 | 1.32 | 0.34 | 73 | 6/23-9/03 | 0.079 |
| 2009 | 375.000 | 369.462 | 28.125 | 143,485 | 22 | 27 | 359 | 920 | 11,934 | 0.84 | 0.34 | 98 | 6/15-9/20 | 0.090 |
| 2010 | 400.000 | 387.304 | 30 | 149,822 | 23 | 32 | 286 | 1,040 | 9,698 | 1.22 | 0.34 | 58 | 6/28-8/24 | 0.074 |
| 2011 | 358.000 | 373.990 | 26.851 | 141,626 | 24 | 25 | 173 | 1,040 | 6,808 | 1.58 | 0.34 | 33 | 6/28-7/30 | 0.038 |
| 2012 | 465.450 | 441.080 | 34.91 | 161,113 | 40 | 29 | 312 | 1,200 | 10,041 | 1.29 | 0.34 | 72 | 6/29-9/08 | 0.093 |
| 2013 | 495.600 | 373.278 | 18.585 | 130,603 | 37 | 33 | 460 | 1,420 | 15,058 | 0.67 | 0.33 | 74 | 7/3-9/14 | 0.110 |
| 2014 | 382.800 | 360.860 | 28.148 | 129,657 | 52 | 33 | 309 | 1,560 | 10,127 | 1.12 | 0.34 | 52 | 6/25-8/15 | 0.052 |
| 2015 | 394.600 | 371.520 | 29.595 | 144,255 | 42 | 36 | 251 | 1,480 | 8,356 | 1.45 | 0.34 | 26 | 6/29-7/24 | 0.033 |
| 2016 | 517.200 | 416.576 | 3,583 | 138,997 | 36 | 37 | 220 | 1,520 | 8,009 | 1.27 | 0.34 | 25 | 6/27-7/21 | 0.025 |
| 2017 | 496,800 | 411,736 | 0 | 135,322 | 36 | 36 | 270 | 1,640 | 9,401 | 1.1 | 0.34 | 30 | 6/26-7/25 | 0.027 |
| 2018 | 319,400 | 298,396 | 0 | 89,613 | 34 | 34 | 256 | 1,400 | 8,797 | 0.64 | 0.34 | 35 | 6/24-7/29 | 0.030 |
| 2019 | 150,600 | 73,784 | 1,239 | 24,506 | 24 | 26 | 146 | 1,096 | 5,438 | 0.26 | 0.34 | 62 | 6/25-9/03 | 0.068 |
| 2020 | 170,000 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | NA | NA | 0 | 6/25-9/03 | NA |
| 2021 | 290.000 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | NA | NA | 0 | 6/15-9/03 | NA |

[^0]Table 2. Historical winter commercial and subsistence red king crab fisheries, Norton Sound Section, eastern Bering Sea. Bold typed data are used for the assessment model.

|  |  | Commercial |  |  | Subsistence |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Model | Year ${ }^{\text {a }}$ | \# of | \# of Crab |  |  |  |  | Tota | Crab |
| Year | Year | Fishers | Harvested | Winter ${ }^{\text {b }}$ | Issued | Returned | Fished | Caught ${ }^{\text {c }}$ | Retained ${ }^{\text {d }}$ |
| 1978 | 1978 | 37 | 9,625 | 1977/78 | 290 | 206 | 149 | NA | 12,506 |
| 1979 | 1979 | $1^{\text {f }}$ | $221{ }^{\text {f }}$ | 1978/79 | 48 | 43 | 38 | NA | 224 |
| 1980 | 1980 | $1^{\text {f }}$ | $22^{\text {f }}$ | 1979/80 | 22 | 14 | 9 | NA | 213 |
| 1981 | 1981 | 0 | 0 | 1980/81 | 51 | 39 | 23 | NA | 360 |
| 1982 | 1982 | $1^{\text {f }}$ | $17^{\text {f }}$ | 1981/82 | 101 | 76 | 54 | NA | 1,288 |
| 1983 | 1983 | 5 | 549 | 1982/83 | 172 | 106 | 85 | NA | 10,432 |
| 1984 | 1984 | 8 | 856 | 1983/84 | 222 | 183 | 143 | 15,923 | 11,220 |
| 1985 | 1985 | 9 | 1,168 | 1984/85 | 203 | 166 | 132 | 10,757 | 8,377 |
| 1986 | 1985/86 | 5 | 2,168 | 1985/86 | 136 | 133 | 107 | 10,751 | 7,052 |
| 1987 | 1986/87 | 7 | 1,040 | 1986/87 | 138 | 134 | 98 | 7,406 | 5,772 |
| 1988 | 1987/88 | 10 | 425 | 1987/88 | 71 | 58 | 40 | 3,573 | 2,724 |
| 1989 | 1988/89 | 5 | 403 | 1988/89 | 139 | 115 | 94 | 7,945 | 6,126 |
| 1990 | 1989/90 | 13 | 3,626 | 1989/90 | 136 | 118 | 107 | 16,635 | 12,152 |
| 1991 | 1990/91 | 11 | 3,800 | 1990/91 | 119 | 104 | 79 | 9,295 | 7,366 |
| 1992 | 1991/92 | 13 | 7,478 | 1991/92 | 158 | 105 | 105 | 15,051 | 11,736 |
| 1993 | 1992/93 | 8 | 1,788 | 1992/93 | 88 | 79 | 37 | 1,193 | 1,097 |
| 1994 | 1993/94 | 25 | 5,753 | 1993/94 | 118 | 95 | 71 | 4,894 | 4,113 |
| 1995 | 1994/95 | 42 | 7,538 | 1994/95 | 166 | 131 | 97 | 7,777 | 5,426 |
| 1996 | 1995/96 | 9 | 1,778 | 1995/96 | 84 | 44 | 35 | 2,936 | 1,679 |
| 1997 | 1996/97 | $2^{\text {f }}$ | $83^{\text {f }}$ | 1996/97 | 38 | 22 | 13 | 1,617 | 745 |
| 1998 | 1997/98 | 5 | 984 | 1997/98 | 94 | 73 | 64 | 20,327 | 8,622 |
| 1999 | 1998/99 | 5 | 2,714 | 1998/99 | 95 | 80 | 71 | 10,651 | 7,533 |
| 2000 | 1999/00 | 10 | 3,045 | 1999/00 | 98 | 64 | 52 | 9,816 | 5,723 |
| 2001 | 2000/01 | 3 | 1,098 | 2000/01 | 50 | 27 | 12 | 366 | 256 |
| 2002 | 2001/02 | 11 | 2,591 | 2001/02 | 114 | 61 | 45 | 5,119 | 2,177 |
| 2003 | 2002/03 | 13 | 6,853 | 2002/03 | 107 | 70 | 61 | 9,052 | 4,140 |
| 2004 | 2003/04 | $2^{\text {f }}$ | $522{ }^{\text {f }}$ | 2003/04 ${ }^{\text {g }}$ | 96 | 77 | 41 | 1,775 | 1,181 |
| 2005 | 2004/05 | 4 | 2,091 | 2004/05 | 170 | 98 | 58 | 6,484 | 3,973 |
| 2006 | 2005/06 | $1^{\text {f }}$ | $75^{\text {f }}$ | 2005/06 | 98 | 97 | 67 | 2,083 | 1,239 |
| 2007 | 2006/07 | 8 | 3,313 | 2006/07 | 129 | 127 | 116 | 21,444 | 10,690 |
| 2008 | 2007/08 | 9 | 5,796 | 2007/08 | 139 | 137 | 108 | 18,621 | 9,485 |
| 2009 | 2008/09 | 7 | 4,951 | 2008/09 | 105 | 105 | 70 | 6,971 | 4,752 |
| 2010 | 2009/10 | 10 | 4,834 | 2009/10 | 125 | 123 | 85 | 9,004 | 7,044 |
| 2011 | 2010/11 | 5 | 3,365 | 2010/11 | 148 | 148 | 95 | 9,183 | 6,640 |
| 2012 | 2011/12 | 35 | 9,157 | 2011/12 | 204 | 204 | 138 | 11,341 | 7,311 |
| 2013 | 2012/13 | 26 | 22,639 | 2012/13 | 149 | 148 | 104 | 21,524 | 7,622 |
| 2014 | 2013/14 | 21 | 14,986 | 2013/14 | 103 | 103 | 75 | 5,421 | 3,252 |
| 2015 | 2014/15 | 44 | 41,062 | 2014/15 | 155 | 153 | 107 | 9,840 | 7,651 |
| 2016 | 2015/16 | 25 | 29,792 | 2015/16 | 139 | 97 | 64 | 6,468 | 5,340 |
| 2017 | 2017 | 43 | 26,008 | 2017 | 163 | 163 | 109 | 7,185 | 6,039 |
| 2018 | 2018 | 28 | 9,180 | 2018 | 123 | 120 | 82 | 5,767 | 4,424 |
| 2019 | 2019 | 6 | 1,050 | 2019 | 101 | 101 | 60 | 2,080 | 1,545 |
| 2020 | 2020 | 1 | conf | 2020 | 79 | 79 | 50 | 813 | 548 |
| 2021 | 2021 | 5 | 320 | 2021 | 103 | 103 | 76 | 4,655 | 2,892 |

a Prior to 1985 the winter commercial fishery occurred from January 1 - April 30. As of March 1985, fishing may occur from
November 15 - May 15.
b The winter subsistence fishery occurs during months of two calendar years (as early as December, through May).
c The number of crab actually caught, including females
d The number of crab retained is the number of crab caught and kept, including females
f Confidentiality was waived by the fishers.
h Prior to 2005, permits were only given out of the Nome ADF\&G office. Starting with the 2004-5 season, permits were given out in
Elim, Golovin, Shaktoolik, and White Mountain.

Table 3. Summary of NSRKC trawl survey abundance estimates ( $x \mathbf{1 0 0 0}$ ) ( $C L \geq 64 \mathrm{~mm}$ ). NMFS and ADF\&G trawl survey abundance estimate is based on $10 \times 10 \mathbf{n m}^{2}$ grids, and NBS trawl survey is based on $\mathbf{2 0} \times \mathbf{2 0} \mathbf{~ n m}^{2}$ girds. Bold typed data are used for the assessment model.

| Year | Dates | Survey Agency | Survey method | Abundance $\geq 64 \mathrm{~mm}$ |  | N | Female |  | \% clutch full $95 \%$ CI |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  | CV |  | $\begin{gathered} \% \\ \text { barren } \end{gathered}$ |  |  |
| 1976 | 9/02-9/25 | NMFS | Trawl | 4301.8 | 0.31 | 181 | 2.5 | 66.7 | 62.4-71.0 |
| 1979 | 7/26-8/05 | NMFS | Trawl | 1457.4 | 0.22 | 42 | 25.0 | 79.9 | 64.8-94.8 |
| 1980 | 7/04-7/14 | ADF\&G | Pots | 2092.3 | N/A |  |  |  |  |
| 1981 | 6/28-7/14 | ADF\&G | Pots | 2153.4 | N/A |  |  |  |  |
| 1982 | 7/06-7/20 | ADF\&G | Pots | 1140.5 | N/A |  |  |  |  |
| 1982 | 9/05-9/11 | NMFS | Trawl | 3548.9 | 0.25 | 269 | 0 | 84.3 | 81.5-87.2 |
| 1985 | 7/01-7/14 | ADF\&G | Pots | 2320.4 | 0.083 |  |  |  |  |
| 1985 | 9/16-10/01 | NMFS | Trawl | 2424.9 | 0.26 | 151 | 0 | 87.5 | NA |
| 1988 | 8/16-8/30 | NMFS | Trawl | 2702.3 | 0.29 | 219 | 1.0 | 80.7 | 77.3-84.2 |
| 1991 | 8/22-8/30 | NMFS | Trawl | 3132.5 | 0.43 | 105 | 0 | 69.3 | 57.7-80.8 |
| 1996 | 8/07-8/18 | ADF\&G | Trawl | 1283.0 | 0.25 | 168 | 30.8 | 71.9 | 65.9-77.9 |
| 1999 | 7/28-8/07 | ADF\&G | Trawl | 2608.0 | 0.24 | 81 | 4.7 | 80.4 | 76.0-84.7 |
| 2002 | 7/27-8/06 | ADF\&G | Trawl | 2056.0 | 0.36 | 168 | 4.7 | 76.8 | 73.4-80.2 |
| 2006 | 7/25-8/08 | ADF\&G | Trawl | 3336.0 | 0.39 | 194 | 3.6 | 67.3 | 63.2-71.5 |
| 2008 | 7/24-8/11 | ADF\&G | Trawl | 2894.2 | 0.31 | 28 | 3.3 | 56.1 | 48.5-61.7 |
| 2010 | 7/27-8/09 | NBS | Trawl | 1980.1 | 0.44 | 116 | 0 | 70.2 | 63.8-78.5 |
| 2011 | 7/18-8/15 | ADF\&G | Trawl | 3209.3 | 0.29 | 135 | 9.8 | 67.2 | 61.7-72.6 |
| 2014 | 7/18-7/30 | ADF\&G | Trawl | 5934.6 | 0.47 | 60 | 0 | 60.4 | 54.3-66.6 |
| 2017 | 7/28-8/08 | ADF\&G | Trawl | 1762.1 | 0.22 | 43 | 21.4 | 71.6 | 60.0-82.7 |
| 2017 | 8/18-8/29 | NBS | Trawl | 1035.8 | 0.40 | 58 | 0 | 80.0 | 72.5-87.5 |
| 2018 | 7/22-7/29 | ADF\&G | Trawl | 1108.9 | 0.25 | 424 | 15.8 | 76.3 | 59.7-83.5 |
| 2019 | 7/17-7/29 | ADF\&G | Trawl | 4660.8 | 0.60 | 386 | 47.8 | 50.6 | 43.1-56.4 |
| 2019 | 8/04-8/07 | NBS | Trawl | 2532.4 | 0.26 | 94 | 17.6 | 47.9 | 36.8-58.9 |
| 2020 | 7/31-8/14 | ADF\&G | Trawl | 1716.5 | 0.27 | 186 | 4.5 | 66.2 | 61.6-70.8 |
| 2021 | 7/19-8/03 | ADF\&G | Trawl | 2400.0 | 0.60 | 90 | 3.4 | 59.8 | 54.9-64.6 |
| 2021 | 7/29-8/07 | NBS | Trawl | 2370.0 | 0.43 | 138 | 2.6 | 61.1 | 58.8-63.4 |

Abundance of NMFS survey was estimated by NMFS, by multiplying the mean CPUE (\# NRKC/nm²) across all hauls (including re-tows) to a standard survey area ( $7600 \mathrm{~nm}^{2}$ ). Abundance of ADF\&G and NBS survey was estimated by ADF\&G by multiplying CPUE (\# NRKC/ $\mathrm{nm}^{2}$ ) of each station to the grid represented by the station and summing across all surveyed station (ADF\&G: $4700-5200 \mathrm{~nm}^{2}$. NBS $5841 \mathrm{~nm}^{2}$ ).
\%barren is calculated by dividing the number of mature females with no eggs by a total number of mature females.

Mean and $95 \%$ CI of \%fullness is calculated among non-barren mature females. Clutch fullness of each non-barren females was assigned by fullness index that was converted to percentage in the table below.

Clutch fullness index of both NMFS-NBS and ADF\&G were converted as follows

| NMFS <br> and NBS <br> Code | NMFS and <br> NBS <br> Fullness | Assigned <br> $\%$ | ADF\&G <br> code | ADF\&G <br> Fullness | Assigned <br> $\%$ |
| :--- | :--- | :--- | :--- | :--- | :--- |
| 2 | $0-1 / 8$ | 6.25 | 3 | $1-29 \%$ | 15 |


| 3 | $1 / 8-1 / 4$ | 18.75 | 4 | $30-59 \%$ | 45 |
| :--- | :--- | :--- | :--- | :--- | :--- |
| 4 | $1 / 4-1 / 2$ | 27.5 | 5 | $60-89 \%$ | 75 |
| 5 | $1 / 2-3 / 4$ | 62.5 | 6 | $90-100 \%$ | 95 |
| 6 | $3 / 4-1$ | 87.5 |  |  |  |
| 7 | $>1$ | 100 |  |  |  |

Table 4. Summer commercial retained catch length-shell compositions.

|  |  | New Shell |  |  |  |  |  |  |  | Old Shell |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Sample | $\begin{aligned} & 64- \\ & 73 \end{aligned}$ | 74-83 | 84-93 | $\begin{aligned} & \hline 94- \\ & 103 \end{aligned}$ | $\begin{aligned} & \hline 104- \\ & 113 \end{aligned}$ | $\begin{gathered} \hline 114- \\ 123 \end{gathered}$ | $\begin{gathered} \hline 124- \\ 133 \end{gathered}$ | 134+ |  | $\begin{aligned} & 4-74- \\ & 3 \quad 83 \end{aligned}$ | $\begin{gathered} 84- \\ \hline 93 \end{gathered}$ | $\begin{array}{ll} 4- & 94- \\ 3 & 103 \end{array}$ | $\begin{aligned} & \hline 104- \\ & 113 \end{aligned}$ | $\begin{gathered} 114- \\ 123 \end{gathered}$ | $\begin{aligned} & \hline 124- \\ & 133 \end{aligned}$ | 134+ |
| 1977 | 1549 | 0 | 0 | 0 | 0.00 | 0.42 | 0.34 | 0.08 | 0.05 | 0 | 0 |  | 00.00 | 0.06 | 0.04 | 0.01 | 0.00 |
| 1978 | 389 | 0 | 0 | 0 | 0.01 | 0.19 | 0.47 | 0.26 | 0.04 | 0 | 0 |  | 00.00 | 0.01 | 0.01 | 0.01 | 0.00 |
| 1979 | 1660 | 0 | 0 | 0 | 0.03 | 0.23 | 0.38 | 0.26 | 0.07 | 0 | 0 |  | 00.00 | 0.03 | 0.00 | 0.00 | 0.01 |
| 1980 | 1068 | 0 | 0 | 0 | 0.00 | 0.10 | 0.31 | 0.37 | 0.18 | 0 | 0 |  | 00.00 | 0.00 | 0.01 | 0.02 | 0.01 |
| 1981 | 1784 | 0 | 0 | 0 | 0.00 | 0.07 | 0.15 | 0.28 | 0.23 | 0 | 0 |  | 00.00 | 0.00 | 0.05 | 0.12 | 0.09 |
| 1982 | 1093 | 0 | 0 | 0 | 0.04 | 0.19 | 0.16 | 0.22 | 0.29 | 0 | 0 |  | 00.00 | 0.01 | 0.02 | 0.03 | 0.03 |
| 1983 | 802 | 0 | 0 | 0 | 0.04 | 0.41 | 0.36 | 0.06 | 0.03 | 0 | 0 |  | 00.00 | 0.04 | 0.01 | 0.02 | 0.02 |
| 1984 | 963 | 0 | 0 | 0 | 0.10 | 0.42 | 0.28 | 0.06 | 0.01 | 0 | 0 |  | 00.01 | 0.07 | 0.05 | 0.01 | 0.00 |
| 1985 | 2691 | 0 | 0 | 0.00 | 0.06 | 0.31 | 0.37 | 0.15 | 0.02 | 0 | 0 |  | 00.00 | 0.03 | 0.03 | 0.01 | 0.00 |
| 1986 | 1138 | 0 | 0 | 0 | 0.03 | 0.36 | 0.39 | 0.12 | 0.02 | 0 | 0 |  | 00.00 | 0.02 | 0.04 | 0.02 | 0.00 |
| 1987 | 1985 | 0 | 0 | 0 | 0.02 | 0.18 | 0.29 | 0.27 | 0.11 |  | 0 |  | 00.00 | 0.03 | 0.06 | 0.03 | 0.01 |
| 1988 | 1522 | 0 | 0.00 | 0 | 0.02 | 0.20 | 0.30 | 0.18 | 0.04 | 0 | 0 |  | 00.01 | 0.06 | 0.10 | 0.07 | 0.02 |
| 1989 | 2595 | 0 | 0 | 0 | 0.01 | 0.16 | 0.32 | 0.17 | 0.05 | 0 | 0 |  | 00.00 | 0.06 | 0.12 | 0.09 | 0.02 |
| 1990 | 1289 | 0 | 0 | 0 | 0.01 | 0.14 | 0.35 | 0.26 | 0.07 | 0 | 0 |  | 00.00 | 0.04 | 0.07 | 0.05 | 0.01 |
| 1991 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1992 | 2566 | 0 | 0 | 0 | 0.02 | 0.20 | 0.27 | 0.14 | 0.09 | 0 | 0 |  | 00.00 | 0.08 | 0.13 | 0.06 | 0.02 |
| 1993 | 17804 | 0 | 0 |  | 0.01 | 0.23 | 0.39 | 0.23 | 0.03 | 0 | 0 |  | 00.00 | 0.02 | 0.04 | 0.03 | 0.01 |
| 1994 | 404 | 0 | 0 | 0 | 0.02 | 0.09 | 0.08 | 0.07 | 0.02 | 0 | 0 |  | 00.02 | 0.19 | 0.25 | 0.20 | 0.05 |
| 1995 | 1167 | 0 | 0 | 0 | 0.04 | 0.26 | 0.29 | 0.15 | 0.05 | 0 | 0 |  | 00.01 | 0.05 | 0.07 | 0.06 | 0.01 |
| 1996 | 787 | 0 | 0 | 0 | 0.03 | 0.22 | 0.24 | 0.09 | 0.05 | 0 | 0 |  | 00.01 | 0.12 | 0.14 | 0.08 | 0.02 |
| 1997 | 1198 | 0 | 0 | 0 | 0.03 | 0.37 | 0.34 | 0.10 | 0.03 | 0 | 0 |  | 00.00 | 0.06 | 0.04 | 0.03 | 0.01 |
| 1998 | 1055 | 0 | 0 | 0 | 0.03 | 0.23 | 0.24 | 0.08 | 0.03 | 0 | 0 |  | 00.02 | 0.11 | 0.14 | 0.08 | 0.03 |
| 1999 | 562 | 0 | 0 |  | 0.06 | 0.29 | 0.24 | 0.18 | 0.09 | 0 | 0 |  | 00.00 | 0.02 | 0.05 | 0.04 | 0.00 |
| 2000 | 17213 | 0 | 0 | 0 | 0.02 | 0.30 | 0.39 | 0.11 | 0.02 | 0 | 0 |  | 00.00 | 0.05 | 0.07 | 0.04 | 0.01 |
| 2001 | 20030 | 0 | 0 | 0 | 0.02 | 0.22 | 0.37 | 0.21 | 0.07 | 0 | 0 |  | 00.00 | 0.02 | 0.05 | 0.02 | 0.01 |
| 2002 | 5219 | 0 | 0 | 0 | 0.04 | 0.23 | 0.28 | 0.25 | 0.07 | 0 | 0 |  | 00.00 | 0.03 | 0.04 | 0.03 | 0.01 |
| 2003 | 5226 | 0 | 0 | 0 | 0.02 | 0.37 | 0.32 | 0.12 | 0.03 | 0 | 0 |  | 00.00 | 0.02 | 0.05 | 0.05 | 0.01 |
| 2004 | 9606 | 0 | 0 | 0 | 0.01 | 0.38 | 0.39 | 0.11 | 0.03 | 0 | 0 |  | 00.00 | 0.03 | 0.03 | 0.01 | 0.01 |
| 2005 | 5360 | 0 | 0 | 0 | 0.00 | 0.25 | 0.47 | 0.16 | 0.02 | 0 | 0 |  | 00.00 | 0.02 | 0.05 | 0.02 | 0.01 |
| 2006 | 6707 | 0 | 0 | 0 | 0.00 | 0.18 | 0.35 | 0.17 | 0.02 | 0 | 0 |  | 00.00 | 0.05 | 0.14 | 0.07 | 0.01 |
| 2007 | 6125 | 0 | 0 | 0 | 0.01 | 0.36 | 0.34 | 0.14 | 0.03 | 0 | 0 |  | 00.00 | 0.02 | 0.06 | 0.03 | 0.01 |
| 2008 | 5766 | 0 | 0 | 0 | 0.00 | 0.35 | 0.35 | 0.06 | 0.01 | 0 | 0 |  | 00.00 | 0.09 | 0.09 | 0.04 | 0.01 |
| 2009 | 6026 | 0 | 0 | 0 | 0.01 | 0.34 | 0.33 | 0.11 | 0.02 | 0 | 0 |  | 00.00 | 0.08 | 0.08 | 0.02 | 0.01 |
| 2010 | 5902 | 0 | 0 | 0 | 0.01 | 0.39 | 0.36 | 0.10 | 0.01 | 0 | 0 |  | 00.00 | 0.05 | 0.05 | 0.02 | 0.00 |
| 2011 | 2552 | 0 | 0 |  | 0.00 | 0.32 | 0.40 | 0.12 | 0.02 | 0 | 0 |  | 00.00 | 0.06 | 0.06 | 0.02 | 0.00 |
| 2012 | 5056 | 0 | 0 | 0 | 0.00 | 0.24 | 0.46 | 0.18 | 0.02 | 0 | 0 |  | 00.00 | 0.03 | 0.04 | 0.02 | 0.00 |
| 2013 | 6072 | 0 | 0 | 0 | 0.00 | 0.24 | 0.37 | 0.24 | 0.06 | 0 | 0 |  | 00.00 | 0.01 | 0.04 | 0.02 | 0.00 |
| 2014 | 4682 | 0 | 0 | 0 | 0.01 | 0.28 | 0.24 | 0.18 | 0.07 | 0 | 0 |  | 00.00 | 0.04 | 0.09 | 0.07 | 0.02 |
| 2015 | 4173 | 0 | 0 | 0 | 0.01 | 0.48 | 0.28 | 0.10 | 0.03 | 0 | 0 |  | 00.00 | 0.02 | 0.03 | 0.03 | 0.01 |
| 2016 | 1543 | 0 | 0 | 0 | 0.00 | 0.25 | 0.47 | 0.16 | 0.03 | 0 | 0 |  | 00.00 | 0.02 | 0.02 | 0.03 | 0.01 |
| 2017 | 3412 | 0 | 0 | 0 | 0.00 | 0.18 | 0.39 | 0.21 | 0.03 | 0 | 0 |  | 00.01 | 0.03 | 0.12 | 0.05 | 0.01 |
| 2018 | 2609 | 0 | 0 | 0 | 0.00 | 0.11 | 0.32 | 0.32 | 0.08 | 0 | 0 |  | 00 | 0.01 | 0.08 | 0.08 | 0.02 |
| 2019 | 1136 | 0 | 0 | 0 | 0.01 | 0.32 | 0.23 | 0.13 | 0.03 | 0 | 0 |  | 00 | 0.02 | 0.10 | 0.14 | 0.03 |

Table 5. Winter commercial catch length-shell compositions.

|  |  | New Shell |  |  |  |  |  |  |  | Old Shell |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Sample | $\begin{gathered} 64- \\ 73 \end{gathered}$ | 74-83 | 84-93 | $\begin{aligned} & \hline 94- \\ & 103 \end{aligned}$ | $\begin{aligned} & \hline 104- \\ & 113 \end{aligned}$ | $114-$ | $\begin{gathered} \hline 124- \\ 133 \end{gathered}$ | 134+ |  |  | $\begin{gathered} \hline 84- \\ 93 \end{gathered}$ | $\begin{aligned} & \hline 94- \\ & 103 \end{aligned}$ | $\begin{gathered} \hline 104- \\ 113 \end{gathered}$ | $\begin{aligned} & \hline 114- \\ & 123 \end{aligned}$ | $\begin{aligned} & \hline 124- \\ & 133 \end{aligned}$ | 134+ |
| 2015 | 576 | 0 | 0 | 0 | 0.07 | 0.50 | 024 | 0.06 | 0.01 | 0 | 0 | 0 | 0.01 | 0.04 | 0.03 | 0.03 | 0.01 |
| 2016 | 1016 | 0 | 0 | 0 | 0.03 | 0.45 | 0.31 | 0.03 | 0.00 | 0 | 0 | 0 | 0.01 | 0.09 | 0.04 | 0.02 | 0.01 |
| 2017 | 540 | 0 | 0 | 0 | 0.00 | 0.20 | 0.30 | 0.13 | 0.02 | 0 | 0 | 0 | 0.00 | 0.08 | 0.19 | 0.06 | 0.02 |
| 2018 | 401 | 0 | 0 | 0 | 0.00 | 0.11 | 0.25 | 0.27 | 0.05 | 0 | 0 | 0 | 0 | 0.04 | 0.16 | 0.10 | 0.02 |

Table 6. Summer Trawl Survey length-shell compositions.

|  | New Shell |  |  |  |  |  |  | Old Shell |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Survey | Sample ${ }^{64-}$ | 4-83 | $\begin{aligned} & \hline 84- \\ & 93 \end{aligned}$ | $\begin{aligned} & \hline 94- \\ & 103 \end{aligned}$ | $\begin{aligned} & \hline 104- \\ & 113 \end{aligned}$ | $\begin{gathered} \hline 114- \\ 123 \end{gathered}$ | $\begin{array}{ll} \hline 124- \\ 133 & 134+ \\ \hline \end{array}$ | $\begin{aligned} & 64- \\ & 73 \end{aligned}$ | $74-$ | $\begin{aligned} & \hline 84- \\ & 93 \end{aligned}$ | $\begin{aligned} & \hline 94- \\ & 103 \end{aligned}$ | $\begin{aligned} & 104- \\ & 113 \end{aligned}$ | $\begin{gathered} \hline 114- \\ 123 \end{gathered}$ | $\begin{array}{cc} \hline 124- \\ 133 & 134+ \\ \hline \end{array}$ |
| 76 NMFS | 13260.01 | 0.02 | 0.10 | 0.19 | 0.34 | 0.18 | 0.020 .00 | 0.0 | 0.00 | 0.01 | 0.02 | 0.03 | 0.04 | 0.01 |
| 1979 NMFS | 2200.01 | 0.01 | 0.00 | 0.02 | 0.05 | 0.05 | 0.030 .01 | 0.01 | 0.00 | 0.01 | 0.04 | 0.14 | 0.40 | 0.190 .0 |
| 1982 NMFS | 3270.22 | 0.0 | . 16 | 0.23 | 0.17 | 0.03 | 0.000 .00 | 0.0 | 0.00 | 0.01 | 0.02 | 0.03 | 0.02 | 0.020 .03 |
| 1985 | 3500.11 |  | 19 | 0.17 | 0.16 | 0.06 | 0.010 .00 |  | . 00 | . 00 | 0.02 | 0.05 | 0.08 | 0.050 .0 |
| 88 N | 3660.16 | 0.19 | 0.12 | 0.13 | 0.11 | 0.06 | 0.030 .00 | 0.0 | 0.00 | 0.01 | 0.01 | 0.03 | 0.0 | 0.0 |
| 91 NMFS | 3400.18 | 0.08 | . 22 | 0.03 | 0.06 | 0.03 | 0.010 .01 | 0.0 | 0.06 | 0.02 | 0.08 | 0.16 | 0.1 | 0.090 .0 |
| 1996 ADF\&G | 2690.29 | 0.2 | 0.13 | 0.09 | 0.05 | 0.00 | 0.000 .01 | 0.00 | 0.00 | 0.03 | 0.03 | 0.04 | 0.0 | 0.040 .0 |
| 999 | 2830.03 | 0. |  | 0.29 | 0.26 | 0.13 | 0.030 .01 | 0.00 | 0.0 | 0.00 | 0.03 | 0.05 | 0.0 | 0.020 .0 |
| 02 ADF\& | 2440.09 |  |  | 0.1 | 0.02 | 0.03 | 0.020 .01 |  | . 03 | . 07 | 0.10 | 0.09 | 0.0 | 0.050 .02 |
| 2006 ADF\& | 3730.18 | 0.26 |  | 0.1 | 0.06 | 0.04 | 0.020 .00 |  | 0.00 | 0.00 | 0.02 | 0.0 | 0.0 | 0.010 .00 |
| 08 ADF\& | 2750.12 | 0.1 |  | 0.1 | 0.10 | 0.03 | 0.020 .01 | 0.0 | 0.01 | 0.04 | . 06 | 0.08 | 0.01 | 0.040 .00 |
| 2010 NMF | 690.01 | 0.0 |  | 0.1 | 0.06 | 0.03 | 0.000 .00 | 0.0 | 0.03 | 0.09 | 0.20 | 0.19 | 0.0 | . 03 |
| 2011 A | 3150.13 | 0.1 |  | 0.1 | 0.18 | 0.14 | 0.030 .0 | 0.0 | 0.00 | 0.01 | 0.02 | 0.09 | 0.0 | . 03 |
| 14 | 3870.08 | 0.15 |  | 0.18 | 0.09 | 0.02 | 0.010 .01 | 0.0 | 0.00 | 0.0 | 0.10 | 0.05 | 0.0 | 0.010 .00 |
| 17 ADF\& | 1160.14 | 0. |  | 09 | 0.10 | 0.04 | 0.000 .00 | 0.01 | . 22 | 0.02 | 0.02 | 0.07 | 8 | 0.040 .00 |
| 17 N | 580.09 | 0.10 |  | 0.05 | 0.05 | 0.05 | 0.050 .03 | 0.0 |  | 0.03 | 0.05 | 0.03 | 0.1 | 0.050 .03 |
| 2018 ADF\&G | 730.37 | 0.10 |  | 0.03 | 0.01 | 0.03 | 0.040 .01 |  | 0.07 | 0.01 | 0.04 | 0.0 | 0.0 | 0.10 |
| 19 ADF\&G | 3070.55 | 0. | 0.03 | 0 | 0.00 | 0.00 | 0.00 | 0.0 | 0.00 | 0.01 | . 02 | 0.0 | 0.02 | 0.030 .0 |
| 2019 NMFS | 1350.36 | 0.30 | 08 | 0.04 | 0.01 | 0 | 0.010 .01 | 0.0 | , 1 | 0.0 | . 02 | 0.0 | 0.01 | 0.040 .01 |
| 2020 ADF\&G | 1110.13 | 0.22 | 0.30 | 0.06 | 0.05 | 0.01 | $0 \quad 0$ | 0.03 | 0.08 | 0.05 | 0.02 | 0.02 | 0.02 | 00.01 |
| 2021 ADF\&G | 1580.0 | 0.17 | 0.22 | 0.22 | 0.22 | 0.04 | 0.010 .01 | 0 | 0 | 0.01 | 0 | 0.02 | 0.01 | 0.010 .01 |
| 021 NMFS | 820.05 | 0.16 | 0.2 | 0.16 | 0.10 | 0.02 | $0 \quad 0$ | 0.01 | 0.05 | 0.11 | 0.06 | 0.06 | 0.01 | 0 |

Table 7. Winter pot survey length-shell compositions.

|  |  |  | New Shell |  |  |  |  |  | Old Shell |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | CPU | ampl | $\begin{aligned} & \hline 64- \\ & 73 \end{aligned}$ | $\begin{gathered} \hline 74- \\ 83 \end{gathered}$ | $\begin{array}{ccc} \hline 84- & 94- & 104- \\ 93 & 103 & 113 \end{array}$ | $\begin{aligned} & 114- \\ & 123 \end{aligned}$ | $\begin{aligned} & 124- \\ & 133 \end{aligned}$ | 134+ | $\begin{array}{ll} 64- & 74- \\ 73 & 83 \end{array}$ | $\begin{aligned} & \hline 84- \\ & 93 \end{aligned}$ | $\begin{aligned} & 94- \\ & 103 \end{aligned}$ | $\begin{gathered} 104- \\ 113 \end{gathered}$ | $\begin{gathered} 114- \\ 123 \end{gathered}$ | $\begin{aligned} & 124- \\ & 133 \end{aligned}$ | 34+ |
| 1981/82 | NA | 719 | 0.00 | 0.10 | 0.230 .210 .07 | 0.02 | 0.02 | 0.00 | 0.000 .05 | 0.11 | 0.11 | 0.04 | 0.02 | 0.02 | 0.00 |
| 1982/83 | 24.2 | 2583 | 0.03 | 0.08 | 0.280 .280 .21 | 0.07 | 0.01 | 0.00 | 0.000 .00 | 0.00 | 0.00 | 0.02 | 0.01 | 0.01 | 0.01 |
| 1983/84 | 24.0 | 1677 | 0.01 | 0.16 | 0.260 .230 .15 | 0.06 | 0.01 | 0.00 | 0.000 .00 | 0.00 | 0.02 | 0.06 | 0.03 | 0.01 | 0.01 |
| 1984/85 | 24.5 | 789 | 0.02 | 0.09 | 0.250 .350 .16 | 0.06 | 0.01 | 0.00 | 0.000 .00 | 0.00 | 0.01 | 0.03 | 0.02 | 0.00 | 0.00 |
| 1985/86 | 19.2 | 594 | 0.04 | 0.12 | 0.170 .240 .19 | 0.08 | 0.01 | 0.00 | 0.00 | 0.00 | 0.01 | 0.06 | 0.04 | 0.01 | 0.00 |
| 1986/87 | 5.8 | 144 | 0.00 | 0.06 | 0.150 .190 .07 | 0.04 | 0.00 | 0.00 | 0.000 .00 | 0.01 | 0.04 | 0.30 | 0.11 | 0.03 | 0.00 |
| 1987/88 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1988/89 | 13.0 | 500 | 0.02 | 0.13 | 0.19 | 0. | 0.03 | 0.0 | 0.000 .00 | 0.00 | 0.0 | 0.0 | 0.0 | 0.03 | 0.00 |
| 1989/90 | 21.0 | 2076 | 0.00 | 0.05 | 0.210 .260 .18 | 0.12 | 0.06 | 0.01 | 0.00 | . 00 | 0.00 | 0.03 | 0.06 | 0.02 | . 00 |
| 1990/91 | 22.9 | 1283 | 0.00 | 0.01 | 0.090 .290 .27 | 0.10 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.03 | 0.12 | 0.07 | . 02 |
| 1992/93 | 5.5 | 181 | 0.00 | 0.01 | 0.030 .060 .13 | 0.12 | 0.03 | 0.00 | 0.000 .00 | 0.00 | 0.02 | 0.19 | 0.27 | 0.10 | . 05 |
| 1993/94 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1994/95 | 6. | 858 | 0.01 | 0.0 | 0.080 .100 .26 | 0.23 | 0.07 | 0.01 | 0.0 | 0.00 | 0.00 | 0.0 | 0.0 | . 06 | 0.02 |
| 1995/96 | 9.9 | 1580 | 0.06 | 0.1 | 0.200 .190 .11 | 0.07 | 0.03 | 0.00 | 0.000 .00 | 0.00 | 0.01 | 0.06 | 0.07 | 0.03 | 0.01 |
| 1996/97 | 2.9 | 398 | 0.07 | 0.2 | 0.220 .110 .15 | 0.11 | 0.05 | 0.01 | 0.000 .00 | 0.00 | 0.00 | 0.02 | 0.03 | 0.01 | 0.01 |
| 1997/98 | 10.9 | 881 | 0.00 | 0.1 | 0.410 .270 .05 | 0.02 | 0.00 | 0.00 | 0.000 .00 | 0.01 | 0.02 | 0.03 | 0.02 | 0.02 | 0.01 |
| 1998/99 | 10.7 | 1307 | 0.00 | 0.02 | 0.120 .360 .36 | 0.08 | 0.01 | 0.00 | 0.000 .00 | 0.00 | 0.01 | 0.02 | 0.0 | 0.0 | . 00 |
| 1999/00 | 6.2 | 575 | 0.02 | 0.0 | 0.100 .160 .33 | 0.18 | 0.03 | 0.00 | 0.0 | 0.00 | 0.00 | 0.0 | 0.02 | 0.0 | 0.00 |
| 2000/01 | 3.1 | 44 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2001/02 | 13.0 | 828 | . 05 | 0.29 | 0.260 .170 .06 | 0.06 | 0.04 | 0.01 | 0.010 .00 | 0.01 | 0.01 | 0.00 | 0.01 | 0.00 | 0.00 |
| 2002/03 | 9.6 | 24 | 0.02 | 0.10 | 0.220 .280 .18 | 0.06 | 0.02 | 0.00 | 0.000 .01 | 0.01 | 0.02 | 0.02 | 0.03 | 0.02 | 0.01 |
| 2003/04 | 3.7 | 296 | 0.00 | 0.02 | 60.260 .32 | 0.14 | 0.01 | 0.00 | 0.000 .00 | 0.01 | 0.02 | 0.02 | 0.01 | 0.02 | 0.01 |
| 2004/05 | 4.4 | 405 | 0.00 | 0.07 | 0.140 .180 .22 | 0.19 | 0.07 | 0.00 | 0.000 .00 | 0.00 | 0.00 | 0.04 | 0.06 | 0.01 | 0.00 |
| 2005/06 | 6.0 | 512 | 0.00 | 0.14 | 0.230 .210 .16 | 0.05 | 0.02 | 0.00 | 0.000 .01 | 0.01 | 0.02 | 0.04 | 0.07 | 0.03 | 0.01 |
| 2006/07 | 7.3 | 159 | 0.07 | 0.14 | 0.190 .350 .13 | 0.04 | 0.00 | 0.00 | 0.000 .00 | 0.01 | 0.01 | 0.02 | 0.04 | 0.00 | 0.00 |
| 2007/08 | 25.0 | 3552 | 0.01 | 0.14 | 0.250 .170 .14 | 0.07 | 0.01 | 0.00 | 0.010 .04 | 0.07 | 0.03 | 0.03 | 0.01 | 0.01 | 0.00 |
| 2008/09 | 21.9 | 525 | 0.00 | 0.07 | 0.130 .350 .20 | 0.08 | 0.01 | 0.00 | 0.000 .00 | 0.00 | 0.00 | 0.04 | 0.10 | 0.00 | 0.00 |
| 2009/10 | 25.3 | 578 | 0.01 | 0.05 | 0.130 .210 .24 | 0.11 | 0.02 | 0.00 | 0.000 .00 | 0.01 | 0.06 | 0.10 | 0.05 | 0.01 | 0.00 |
| 2010/11 | 22.1 | 596 | 0.02 | 0.08 | 0.130 .200 .17 | 0.13 | 0.05 | 0.00 | 0.000 .00 | 0.01 | 0.03 | 0.11 | 0.05 | 0.01 | 0.00 |
| 2011/12 | 29.4 | 675 | 0.03 | 0.11 | 0.230 .190 .12 | 0.13 | 0.04 | 0.00 | 0.000 .00 | 0.00 | 0.01 | 0.05 | 0.05 | 0.03 | 0.00 |

Table 8. Summer commercial 1987-1994 observer discards length-shell compositions.

|  |  | New Shell |  |  |  |  |  |  |  | Old Shell |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | $\begin{aligned} & \hline 64- \\ & 73 \end{aligned}$ | $\begin{aligned} & \hline 74- \\ & 83 \end{aligned}$ | $\begin{aligned} & \hline 84- \\ & 93 \end{aligned}$ | $\begin{aligned} & \hline 94- \\ & 103 \end{aligned}$ | $\begin{gathered} \hline 104- \\ 113 \end{gathered}$ | $\begin{aligned} & \hline 114- \\ & 123 \end{aligned}$ | $\begin{aligned} & \hline 124- \\ & 133 \end{aligned}$ | 34+ | $\begin{array}{\|c} \hline 64- \\ 73 \end{array}$ | $\begin{aligned} & \hline 74- \\ & 83 \end{aligned}$ | $\begin{gathered} 84- \\ 93 \end{gathered}$ | $\begin{aligned} & \hline 94- \\ & 103 \end{aligned}$ | $\begin{aligned} & \hline 104- \\ & 113 \end{aligned}$ | $\begin{aligned} & \hline 114- \\ & 123 \end{aligned}$ | $\begin{aligned} & 124- \\ & 133 \end{aligned}$ |  |
| 1987 | 114 | 0.06 | 0.19 | 0.32 | 0.33 | 0.03 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.02 | 0.04 | 0.00 | 0.00 | 0.00 | 0.00 |
| 1988 | 722 | 0.01 | 0.04 | 0.15 | 0.48 | 0.14 | 0.00 | 0.00 | 0.00 | 0.00 | 0.01 | 0.03 | 0.10 | 0.04 | 0.00 | 0.00 | 0.00 |
| 1989 | 1000 | 0.07 | 0.19 | 0.24 | 0.22 | 0.03 | 0.00 | 0.00 | 0.00 | 0.02 | 0.03 | 0.07 | 0.11 | 0.03 | 0.00 | 0.00 | 0.00 |
| 1990 | 50 | 0.08 | 0.23 | 0.27 | 0.27 | 0.04 | 0.00 | 0.00 | 0.00 | 0.02 | 0.02 | 0.02 | 0.05 | 0.01 | 0.00 | 0.00 | 0.00 |
| 1992 | 580 | 0.11 | 0.17 | 0.30 | 0.29 | 0.03 | 0.00 | 0.00 | 0.00 | 0.01 | 0.02 | 0.02 | 0.04 | 0.01 | 0.00 | 0.00 | 0.00 |
| 1994 | 850 | 0.07 | 0.06 | 0.11 | 0.15 | 0.02 | 0.00 | 0.00 | 0.00 | 0.07 | 0.07 | 0.15 | 0.24 | 0.05 | 0.00 | 0.00 | . 0 |

Table 9. Summer commercial observer total catch length-shell compositions.

|  |  | New Shell |  |  |  |  |  |  |  | Old Shell |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | ample | $\begin{aligned} & \hline 64- \\ & 73 \end{aligned}$ | $\begin{aligned} & 74- \\ & 83 \end{aligned}$ | $\begin{aligned} & 84- \\ & 93 \end{aligned}$ | $\begin{aligned} & 94- \\ & 103 \end{aligned}$ | $\begin{gathered} 104- \\ 113 \end{gathered}$ | $\begin{aligned} & 114- \\ & 123 \end{aligned}$ |  | + | $\begin{array}{\|c} 64- \\ 73 \end{array}$ | $\begin{aligned} & 74- \\ & 83 \end{aligned}$ | $\begin{aligned} & 84- \\ & 93 \end{aligned}$ | $\begin{aligned} & 94- \\ & 103 \end{aligned}$ | $\begin{aligned} & 104- \\ & 113 \end{aligned}$ | $\begin{gathered} 114- \\ 123 \end{gathered}$ | $\begin{gathered} 124- \\ 133 \end{gathered}$ | 134+ |
| 2012 | 3055 | 0.10 | 0.05 | 0.08 | 0.15 | 0.15 | 0.17 | 0.06 | 0.01 | 0.00 | 0.00 | 0.00 | 0.03 | 0.08 | 0.09 | 0.03 | 0.00 |
| 2013 | 4762 | 0.19 | 0.16 | 0.09 | 0.10 | 0.16 | 0.16 | 0.09 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.01 | 0.02 | 0.01 | 0.00 |
| 2014 | 3506 | 0.02 | . 05 | 0.13 | 0.22 | 0.22 | 0.12 | 0.08 | 0.03 | 0.00 | 0.00 | 0.00 | 0.02 | 0.03 | 0.03 | 0.02 | 0.0 |
| 2015 | 167 | 0.01 | . 04 | 0.09 | 0.23 | 0.37 | 0.14 | 0.05 | 0.01 | 0.00 | 0.00 | 0.00 | 0.01 | 0.02 | 0.02 | 0.01 | 0.0 |
| 2016 | 2114 | 0.01 | 0.01 | 0.03 | 0.12 | 0.29 | 0.36 | 0.08 | 0.02 | 0.00 | 0.00 | 0.00 | 0.01 | 0.03 | 0.03 | 0.02 | 0.00 |
| 2017 | 2748 | 0.02 | 0.03 | 0.03 | 0.06 | 0.19 | 0.33 | 0.18 | 0.02 | 0.00 | 0.00 | 0.00 | 0.00 | 0.02 | 0.07 | 0.03 | 0.0 |
| 2018 | 1628 | 0.03 | 0.06 | 0.12 | 0.11 | 0.09 | 0.17 | 0.18 | 0.04 | 0.00 | 0.00 | 0.01 | 0.01 | 0.15 | 0.07 | 0.08 | 0.02 |
| 2019 | 236 | 0.13 | 0.06 | 0.06 | 0.13 | 0.08 | 0.05 | 0.01 | 0.01 | 0 | 0 | 0.00 | 0.04 | 0.11 | 0.14 | 0.1 | 0.05 |

Table 10. The observed proportion of tagged crab by each size class released and recovered after 1-3 year of liberty 1980-2019 periods.

Year at liberty 1

|  | $64-73$ | $74-83$ | $84-93$ | $94-103$ | $104-113$ | $114-123$ | $124-33$ | $>134$ | n |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $64-73$ | 0 | 0.2 | 0.8 | 0 | 0 | 0 | 0 | 0 | 5 |
| $74-83$ |  | 0 | 0.44 | 0.47 | 0.09 | 0 | 0 | 0 | 47 |
| $84-93$ |  |  | 0 | 0.32 | 0.62 | 0.05 | 0.01 | 0 | 146 |
| $94-103$ |  |  |  | 0.03 | 0.62 | 0.34 | 0.01 | 0.00 | 317 |
| $104-113$ |  |  |  |  | 0.31 | 0.59 | 0.09 | 0 | 241 |
| $114-123$ |  |  |  |  |  | 0.42 | 0.47 | 0.11 | 210 |
| $124-133$ |  |  |  |  |  |  | 0.69 | 0.31 | 81 |
| $>134$ |  |  |  |  |  |  |  | 1 | 26 |

Year at liberty 2

|  | $64-73$ | $74-83$ | $84-93$ | $94-103$ | $104-113$ | $114-123$ | $124-33$ | $>134$ | n |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $64-73$ | 0 | 0 | 0.09 | 0.55 | 0.36 | 0 | 0 | 0 | 11 |
| $74-83$ |  | 0 | 0 | 0.11 | 0.85 | 0.04 | 0 | 0 | 113 |
| $84-93$ |  |  | 0 | 0.04 | 0.32 | 0.61 | 0.03 | 0 | 114 |
| $94-103$ |  |  |  | 0.02 | 0.36 | 0.41 | 0.20 | 0 | 94 |
| $104-113$ |  |  |  |  | 0.06 | 0.71 | 0.22 | 0 | 108 |
| $114-123$ |  |  |  |  |  | 0.17 | 0.72 | 0.11 | 65 |
| $124-133$ |  |  |  |  |  |  | 0.36 | 0.64 | 25 |
| $>134$ |  |  |  |  |  |  |  | 1 | 8 |

Year at liberty 3

|  | $64-73$ | $74-83$ | $84-93$ | $94-103$ | $104-113$ | $114-123$ | $124-33$ | $>134$ | n |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $64-73$ | 0 | 0 | 0 | 0 | 0.5 | 0.5 | 0 | 0 | 22 |
| $74-83$ | 0 | 0 | 0 | 0 | 0.26 | 0.66 | 0.082 | 0 | 73 |
| $84-93$ | 0 | 0 | 0 | 0.04 | 0.26 | 0.53 | 0.17 | 0 | 53 |
| $94-103$ | 0 | 0 | 0 | 0 | 0.06 | 0.67 | 0.27 | 0 | 52 |
| $104-113$ | 0 | 0 | 0 | 0 | 0 | 0.26 | 0.62 | 0.12 | 34 |
| $114-123$ | 0 | 0 | 0 | 0 | 0 | 0 | 0.79 | 0.21 | 14 |
| $124-133$ | 0 | 0 | 0 | 0 | 0 | 0 | 0.1 | 0.9 | 10 |
| $>134$ | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 |

Table 11. Summary of initial input parameter values and bounds for a length-based population model of Norton Sound red king crab. Parameters with "log_" indicate log scaled parameters

| Parameter | Parameter description | Lower | Upper |
| :---: | :---: | :---: | :---: |
| $\log _{\text {_ }} \mathrm{q}_{1}$ | Commercial fishery catchability (1977-93) | -20.5 | 20 |
| $\log _{\text {_ }}$ q 2 | Commercial fishery catchability (1994-2007) | -20.5 | 20 |
| log_q3 | Commercial fishery catchability (2008-2019) | -20.5 | 20 |
| $\log _{\sim} \mathrm{N}_{76}$ | Initial abundance | 2.0 | 15.0 |
| $\mathrm{R}_{0}$ | Mean Recruit | 2.0 | 12.0 |
| $\log \sigma_{R}{ }^{2}$ | Recruit standard deviation | -40.0 | 40.0 |
| $\mathrm{a}_{1-7}$ | Intimal length proportion | 0 | 10.0 |
| $\mathrm{r}_{1}$ | Proportion of length class 1 for recruit | 0 | 10.0 |
| $\log _{2} \alpha$ | Inverse logistic molting parameter | -5.0 | -1.0 |
| $\log \beta$ | Inverse logistic molting parameter | 1.0 | 5.5 |
| $\log _{\text {_ }} \phi_{\text {st1 }}$ | Logistic trawl selectivity parameter | -5.0 | 1.0 |
| $\log _{-} \phi_{w a}$ | Inverse logistic winter pot selectivity parameter | -5.0 | 1.0 |
| $\log _{-} \phi_{w b}$ | Inverse logistic winter pot selectivity parameter | 0.0 | 6.0 |
| $\mathrm{Sw}_{1,2}$ | Winter pot selectivity of length class 1,2 | 0.1 | 1.0 |
|  |  |  |  |
| $\log \phi_{l}$ | Logistic commercial catch selectivity parameter | -5.0 | 1.0 |
| $\log _{-}$acr | Logistic summer commercial retention selectivity $(1976-2007)$ | -5.0 | 1.0 |
| $\log _{-} \mathrm{bcr}$ | Logistic summer commercial retention selectivity (1976-2007) | 0.0 | 6.0 |
| $\log _{-}$acr | Logistic summer commercial retention selectivity (2008-2019) | -5.0 | 1.0 |
| $\log _{-} \mathrm{bc}$ r | Logistic summer commercial retention selectivity (2008-2019) | 0.0 | 6.0 |
| log_awr | Logistic winter commercial retention selectivity parameter | -5.0 | 1.0 |
| log_bwr | Logistic winter commercial retention selectivity parameter | 0.0 | 6.0 |
| $w^{2}{ }_{t}$ | Additional variance for standard CPUE | 0.0 | 6.0 |
| ms | Natural mortality multipliers | 0.5 | 5.0 |
| q. 1 | Survey q for NMFS trawl 1976-91 | 0.1 | 1.0 |
| q. 2 | Survey q for NMFS NBS trawl 2010,17,19 | 0.1 | 1.0 |
| $\sigma$ | Growth transition sigma | 0.0 | 30.0 |
| $\beta_{1}$ | Growth transition mean | 0.0 | 20.0 |
| $\beta_{2}$ | Growth transition increment | 0.0 | 20.0 |

*: Parameter was unestimable because model estimated trawl survey selectivity was 1.0 across all size classes.

|  | Mosel 21,0 |  | Model 18.0e |  | Model 21.2 |  | Model 21.2 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| name | Estimate | std.dev | Estimate | std.dev | Estimate | std.dev | Estimate | std.dev |
| $\log _{-} \mathrm{q}_{1}$ | -7.218 | 0.198 | -6.866 | 0.117 | -7.202 | 0.196 | -6.866 | 0.117 |
| $\log _{\_} \mathrm{q}_{2}$ | -6.713 | 0.152 |  |  | -6.700 | 0.151 |  |  |
| $\log _{\_} \mathrm{q}_{3}$ | -6.770 | 0.147 |  |  | -6.786 | 0.141 |  |  |
| $\log _{-} \mathrm{N}_{76}$ | 9.137 | 0.139 | 9.086 | 0.122 | 9.134 | 0.138 | 9.084 | 0.122 |
| $\mathrm{R}_{0}$ | 6.413 | 0.081 | 6.414 | 0.080 | 6.413 | 0.081 | 6.412 | 0.080 |
| $\mathrm{a}_{1}$ | 0.976 | 4.459 | 1.054 | 4.469 | 0.973 | 4.460 | 1.060 | 4.469 |
| $\mathrm{a}_{2}$ | 1.672 | 4.193 | 1.779 | 4.202 | 1.685 | 4.191 | 1.768 | 4.204 |
| $\mathrm{a}_{3}$ | 3.454 | 3.931 | 3.524 | 3.943 | 3.461 | 3.930 | 3.519 | 3.945 |
| $\mathrm{a}_{4}$ | 3.977 | 3.909 | 4.008 | 3.921 | 3.977 | 3.908 | 4.010 | 3.923 |
| $\mathrm{a}_{5}$ | 4.250 | 3.900 | 4.268 | 3.913 | 4.248 | 3.899 | 4.271 | 3.914 |
| $\mathrm{a}_{6}$ | 3.508 | 3.929 | 3.520 | 3.941 | 3.506 | 3.928 | 3.522 | 3.943 |
| $\mathrm{a}_{7}$ | 2.066 | 4.195 | 2.065 | 4.208 | 2.066 | 4.194 | 2.065 | 4.210 |
| r1 | 10.000 | 0.227 | 10.000 | 0.241 | 10.000 | 0.229 | 10.000 | 0.238 |
| r2 | 9.616 | 0.281 | 9.613 | 0.293 | 9.615 | 0.283 | 9.615 | 0.291 |
| log_a | -2.728 | 0.089 | -2.726 | 0.089 | -2.726 | 0.089 | -2.727 | 0.089 |
| $\log _{-} \mathrm{b}$ | 4.834 | 0.015 | 4.833 | 0.015 | 4.834 | 0.015 | 4.833 | 0.015 |
| $\log _{-} \phi_{\text {st1 }}$ | -5.000 | 0.030 | -5.000 | 0.032 | -5.000 | 0.030 | -5.000 | 0.032 |
| $\log _{\sim} \phi_{\text {w }}{ }^{\text {a }}$ | -2.393 | 0.428 | -2.394 | 0.430 | -2.396 | 0.430 | -2.390 | 0.429 |
| $\log _{-} \phi_{w b}$ | 4.776 | 0.066 | 4.776 | 0.067 | 4.776 | 0.067 | 4.777 | 0.066 |
| Sw1 | 0.060 | 0.033 | 0.060 | 0.033 | 0.060 | 0.033 | 0.060 | 0.033 |
| Sw2 | 0.424 | 0.144 | 0.425 | 0.145 | 0.424 | 0.145 | 0.425 | 0.144 |
| Sw3 | 0.731 | 0.232 | 0.735 | 0.233 | 0.730 | 0.233 | 0.737 | 0.232 |
| $\log _{-} \phi_{l}$ | -2.062 | 0.045 | -2.070 | 0.039 | -2.070 | 0.039 | -2.064 | 0.043 |
| log_фral | -0.856 | 0.142 | -0.792 | 0.123 | -0.785 | 0.123 | -0.870 | 0.143 |
| log_фrb1 | -0.490 | 0.283 | 4.647 | 0.007 | 4.648 | 0.007 | -0.479 | 0.285 |
| log_¢ra2 | 4.643 | 0.008 |  |  |  |  | 4.643 | 0.008 |
| $\log _{-}$¢rb2 | 4.656 | 0.013 |  |  |  |  | 4.655 | 0.013 |
| log_ ${ }^{\text {w }}$ rra | -0.927 | 0.600 | -0.925 | 0.604 | -0.926 | 0.603 | -0.923 | 0.607 |
| $\log _{\sim} \phi$ wrb | 4.651 | 0.040 | 4.651 | 0.040 | 4.651 | 0.040 | 4.651 | 0.040 |
| $w^{2}{ }_{t}$ | 0.090 | 0.028 | 0.000 | 0.000 | 0.089 | 0.028 | 0.000 | 0.000 |
| q. 1 | 0.769 | 0.138 | 0.805 | 0.131 | 0.772 | 0.138 | 0.807 | 0.131 |
| q. 2 | 0.940 | 0.192 | 0.913 | 0.185 | 0.941 | 0.192 | 0.910 | 0.185 |
| $\sigma$ | 3.818 | 0.209 | 3.819 | 0.208 | 3.814 | 0.209 | 3.822 | 0.208 |
| $\beta_{1}$ | 11.776 | 0.697 | 11.795 | 0.695 | 11.773 | 0.695 | 11.801 | 0.697 |
| $\beta_{2}$ | 7.819 | 0.171 | 7.812 | 0.171 | 7.821 | 0.171 | 7.810 | 0.171 |
| ms78 | 3.453 | 0.272 | 3.386 | 0.266 | 3.461 | 0.272 | 3.377 | 0.265 |


|  | Model 21.0 |  | Model 21.1 |  | Model 21.4 |  | Model 21.5 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| name | Estimate | std.dev | Estimate | std.dev | Estimate | std.dev | Estimate | std.dev |
| $\log _{-} \mathrm{q}_{1}$ | -7.218 | 0.198 | -6.410 | 0.148 | -7.078 | 0.192 | -7.239 | 0.197 |
| $\log _{-} \mathrm{q}_{2}$ | -6.713 | 0.152 | -6.177 | 0.158 | -6.540 | 0.156 | -6.664 | 0.153 |
| $\log _{-} \mathrm{q}_{3}$ | -6.770 | 0.147 | -6.238 | 0.153 | -6.668 | 0.152 | -6.753 | 0.146 |
| $\log _{-} \mathrm{N}_{76}$ | 9.137 | 0.139 | 8.452 | 0.040 | 9.495 | 0.165 | 9.328 | 0.162 |
| $\mathrm{R}_{0}$ | 6.413 | 0.081 | 5.813 | 0.044 | 7.083 | 0.159 | 6.652 | 0.129 |
| $\mathrm{a}_{1}$ | 0.976 | 4.459 | 1.553 | 4.369 | 2.554 | 4.590 | 1.246 | 4.494 |
| $\mathrm{a}_{2}$ | 1.672 | 4.193 | 2.276 | 4.095 | 2.871 | 4.386 | 1.937 | 4.231 |
| $\mathrm{a}_{3}$ | 3.454 | 3.931 | 4.013 | 3.827 | 4.415 | 4.146 | 3.686 | 3.974 |
| $\mathrm{a}_{4}$ | 3.977 | 3.909 | 4.477 | 3.802 | 4.735 | 4.127 | 4.153 | 3.953 |
| $\mathrm{a}_{5}$ | 4.250 | 3.900 | 4.668 | 3.793 | 4.874 | 4.118 | 4.389 | 3.944 |
| $\mathrm{a}_{6}$ | 3.508 | 3.929 | 3.771 | 3.828 | 3.949 | 4.149 | 3.607 | 3.973 |
| $\mathrm{a}_{7}$ | 2.066 | 4.195 | 1.905 | 4.159 | 2.070 | 4.427 | 2.070 | 4.240 |
| r1 | 10.000 | 0.227 | 10.000 | 0.171 | 10.000 | 0.311 | 10.000 | 0.372 |
| r2 | 9.616 | 0.281 | 9.576 | 0.236 | 9.523 | 0.356 | 9.654 | 0.406 |
| $\log _{\text {_ }}$ a | -2.728 | 0.089 | -2.779 | 0.096 | -2.741 | 0.094 | -2.722 | 0.089 |
| $\log _{-} \mathrm{b}$ | 4.834 | 0.015 | 4.832 | 0.017 | 4.816 | 0.016 | 4.827 | 0.015 |
| $\log _{-} \phi_{\text {st1 }}$ | -5.000 | 0.030 | -5.000 | 0.011 | -2.422 | 0.099 | -5.000 | 0.078 |
| $\log _{-} \phi_{w a}$ | -2.393 | 0.428 | -2.059 | 0.332 | -1.825 | 0.437 | -2.275 | 0.453 |
| $\log _{-} \phi_{w b}$ | 4.776 | 0.066 | 4.807 | 0.034 | 4.866 | 0.027 | 4.817 | 0.049 |
| Sw1 | 0.060 | 0.033 | 0.068 | 0.034 | 0.044 | 0.021 | 0.058 | 0.030 |
| Sw2 | 0.424 | 0.144 | 0.504 | 0.129 | 0.357 | 0.084 | 0.406 | 0.115 |
| Sw3 | 0.731 | 0.232 | 0.848 | 0.185 | 0.727 | 0.140 | 0.772 | 0.194 |
| $\log _{\_} \phi_{l}$ | -2.062 | 0.045 | -2.038 | 0.046 | -1.949 | 0.042 | -2.024 | 0.045 |
| log_фra1 | -0.856 | 0.142 | -0.871 | 0.142 | -0.896 | 0.141 | -0.866 | 0.142 |
| log_фrb1 | -0.490 | 0.283 | -0.540 | 0.270 | -0.459 | 0.286 | -0.473 | 0.286 |
| log_фra2 | 4.643 | 0.008 | 4.644 | 0.009 | 4.652 | 0.010 | 4.646 | 0.009 |
| log_фrb2 | 4.656 | 0.013 | 4.663 | 0.011 | 4.655 | 0.014 | 4.655 | 0.013 |
| log_фwra | -0.927 | 0.600 | -1.113 | 0.472 | -0.882 | 0.663 | -0.897 | 0.660 |
| log_фwrb | 4.651 | 0.040 | 4.671 | 0.033 | 4.648 | 0.042 | 4.648 | 0.042 |
| $w^{2}{ }_{t}$ | 0.090 | 0.028 | 0.120 | 0.035 | 0.085 | 0.027 | 0.085 | 0.027 |
| q, 1 | 0.769 | 0.138 | 1.713 | 0.211 | 0.740 | 0.130 | 0.696 | 0.127 |
| q, 2 | 0.940 | 0.192 | 1.250 | 0.256 | 0.927 | 0.190 | 0.914 | 0.187 |
| $\sigma$ | 3.818 | 0.209 | 3.967 | 0.194 | 3.796 | 0.205 | 3.790 | 0.209 |
| $\beta_{1}$ | 11.776 | 0.697 | 12.775 | 0.702 | 12.670 | 0.723 | 12.021 | 0.706 |
| $\beta_{2}$ | 7.819 | 0.171 | 7.454 | 0.170 | 7.595 | 0.17748 | 7.7681 | 0.1738 |
| ms 78 | 3.453 | 0.272 |  |  |  |  | 2.3206 | 0.3898 |
| M |  |  |  |  | 0.42 | 0.03 | 0.26 | 0.03 |

Table 12. Estimated molting probability incorporated transition matrix.

| Pre-molt | Post-molt Length Class |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Length | 64-73 | 74-83 | 84-93 | 94-103 | 104-113 | 114-123 | 124-133 | 134+ |
| Class |  |  |  |  |  |  |  |  |
| 64-73 | 0.02 | 0.11 | 0.79 | 0.08 | 0.00 | 0.00 | 0.00 | 0.00 |
| 74-83 |  | 0.04 | 0.25 | 0.68 | 0.02 | 0.00 | 0.00 | 0.00 |
| 84-93 |  |  | 0.08 | 0.43 | 0.48 | 0.00 | 0.00 | 0.00 |
| 94-103 |  |  |  | 0.16 | 0.58 | 0.26 | 0.00 | 0.00 |
| 104-113 |  |  |  |  | 0.29 | 0.60 | 0.11 | 0.00 |
| 114-123 |  |  |  |  |  | 0.49 | 0.48 | 0.03 |
| 124-133 |  |  |  |  |  |  | 0.70 | 0.30 |
| 134+ |  |  |  |  |  |  |  | 1.00 |


| Pre-molt | Post-molt Length Class |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 64-73 | 74-83 | 84-93 | 94-103 | 104-113 | 114-123 | 124-133 | 134+ |
| 64-73 | 0.02 | 0.11 | 0.79 | 0.08 | 0.00 | 0.00 | 0.00 | 0.00 |
| 74-83 |  | 0.04 | 0.25 | 0.68 | 0.02 | 0.00 | 0.00 | 0.00 |
| 84-93 |  |  | 0.08 | 0.43 | 0.48 | 0.00 | 0.00 | 0.00 |
| 94-103 |  |  |  | 0.16 | 0.58 | 0.26 | 0.00 | 0.00 |
| 104-113 |  |  |  |  | 0.29 | 0.60 | 0.10 | 0.00 |
| 114-123 |  |  |  |  |  | 0.49 | 0.48 | 0.03 |
| 124-133 |  |  |  |  |  |  | 0.71 | 0.29 |
| 134+ |  |  |  |  |  |  |  | 1.00 |

Model 21.1

| Pre-molt | Post-molt Length Class |  |  |  |  |  |  |  |  |
| :---: | :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | :---: |
|  | $64-73$ | $74-83$ | $84-93$ | $94-103$ | $104-113$ | $114-123$ | $124-133$ | $134+$ |  |
| $64-73$ | 0.03 | 0.09 | 0.77 | 0.11 | 0.00 | 0.00 | 0.00 | 0.00 |  |
| $74-83$ |  | 0.05 | 0.24 | 0.68 | 0.03 | 0.00 | 0.00 | 0.00 |  |
| $84-93$ |  |  | 0.10 | 0.44 | 0.46 | 0.01 | 0.00 | 0.00 |  |
| $94-103$ |  |  |  | 0.18 | 0.59 | 0.23 | 0.00 | 0.00 |  |
| $104-113$ |  |  |  |  | 0.33 | 0.59 | 0.08 | 0.00 |  |
| $114-123$ |  |  |  |  |  |  |  |  |  |
| $124-133$ |  |  |  | 0.55 | 0.43 | 0.02 |  |  |  |
| $134+$ |  |  |  |  |  |  |  |  |  |


| Pre-molt | Post-molt Length Class |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 64-73 | 74-83 | 84-93 | 94-103 | 104-113 | 114-123 | 124-133 | 134+ |
| 64-73 | 0.02 | 0.11 | 0.79 | 0.08 | 0.00 | 0.00 | 0.00 | 0.00 |
| 74-83 |  | 0.04 | 0.25 | 0.68 | 0.02 | 0.00 | 0.00 | 0.00 |
| 84-93 |  |  | 0.08 | 0.43 | 0.48 | 0.00 | 0.00 | 0.00 |
| 94-103 |  |  |  | 0.16 | 0.58 | 0.26 | 0.00 | 0.00 |
| 104-113 |  |  |  |  | 0.29 | 0.60 | 0.11 | 0.00 |
| 114-123 |  |  |  |  |  | 0.49 | 0.48 | 0.03 |
| 124-133 |  |  |  |  |  |  | 0.70 | 0.30 |
| 134+ |  |  |  |  |  |  |  | 1.00 |


| Pre-molt | Post-molt Length Class |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 64-73 | 74-83 | 84-93 | 94-103 | 104-113 | 114-123 | 124-133 | 134+ |
| 64-73 | 0.02 | 0.11 | 0.79 | 0.08 | 0.00 | 0.00 | 0.00 | 0.00 |
| 74-83 |  | 0.04 | 0.25 | 0.68 | 0.02 | 0.00 | 0.00 | 0.00 |
| 84-93 |  |  | 0.08 | 0.43 | 0.48 | 0.00 | 0.00 | 0.00 |
| 94-103 |  |  |  | 0.16 | 0.58 | 0.26 | 0.00 | 0.00 |
| 104-113 |  |  |  |  | 0.29 | 0.60 | 0.10 | 0.00 |
| 114-123 |  |  |  |  |  | 0.49 | 0.48 | 0.03 |
| 124-133 |  |  |  |  |  |  | 0.71 | 0.29 |
| 134+ |  |  |  |  |  |  |  | 1.00 |

Model 21.4

| Pre-molt | Post-molt Length Class |  |  |  |  |  |  |  |  |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | :---: |
|  | $64-73$ | $74-83$ | $84-93$ | $94-103$ | $104-113$ | $114-123$ | $124-133$ | $134+$ |  |
| $64-73$ | 0.03 | 0.08 | 0.79 | 0.10 | 0.00 | 0.00 | 0.00 | 0.00 |  |
| $74-83$ |  | 0.05 | 0.21 | 0.71 | 0.03 | 0.00 | 0.00 | 0.00 |  |
| $84-93$ |  |  | 0.10 | 0.41 | 0.49 | 0.01 | 0.00 | 0.00 |  |
| $94-103$ |  |  |  | 0.18 | 0.57 | 0.25 | 0.00 | 0.00 |  |
| $104-113$ |  |  |  |  | 0.33 | 0.58 | 0.09 | 0.00 |  |
| $114-123$ |  |  |  |  | 0.54 | 0.44 | 0.02 |  |  |
| $124-133$ |  |  |  |  |  | 0.76 | 0.24 |  |  |
| $\mathbf{1 3 4 +}$ |  |  |  |  |  |  |  |  |  |


| Pre-molt | Post-molt Length Class |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 64-73 | 74-83 | 84-93 | 94-103 | 104-113 | 114-123 | 124-133 | 134+ |
| 64-73 | 0.02 | 0.10 | 0.79 | 0.08 | 0.00 | 0.00 | 0.00 | 0.00 |
| 74-83 |  | 0.05 | 0.24 | 0.69 | 0.02 | 0.00 | 0.00 | 0.00 |
| 84-93 |  |  | 0.09 | 0.42 | 0.48 | 0.00 | 0.00 | 0.00 |
| 94-103 |  |  |  | 0.16 | 0.57 | 0.26 | 0.00 | 0.00 |
| 104-113 |  |  |  |  | 0.30 | 0.60 | 0.10 | 0.00 |
| 114-123 |  |  |  |  |  | 0.50 | 0.47 | 0.03 |
| 124-133 |  |  |  |  |  |  | 0.72 | 0.28 |
| 134+ |  |  |  |  |  |  |  | 1.00 |

Table 13. Estimated selectivity, mortality, molting probabilities, and proportions of legal crab by length class (mm CL) for Norton Sound male red king crab. Model 21.0

| Molting Probability |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 21.0 | 19.0 e | 21.1 | 21.2 | 21.3 | 21.4 | 21.5 |  |
| $64-73$ | 0.98 | 0.98 | 0.97 | 0.98 | 0.98 | 0.97 | 0.98 |  |
| $74-83$ | 0.96 | 0.96 | 0.95 | 0.96 | 0.96 | 0.95 | 0.95 |  |
| $84-93$ | 0.92 | 0.92 | 0.91 | 0.92 | 0.92 | 0.90 | 0.92 |  |
| $94-103$ | 0.86 | 0.86 | 0.84 | 0.86 | 0.85 | 0.83 | 0.85 |  |
| $104-113$ | 0.75 | 0.75 | 0.74 | 0.75 | 0.75 | 0.72 | 0.74 |  |
| $114-123$ | 0.62 | 0.61 | 0.61 | 0.62 | 0.61 | 0.58 | 0.60 |  |
| $124-133$ | 0.45 | 0.45 | 0.45 | 0.45 | 0.45 | 0.42 | 0.44 |  |
| $134+$ | 0.30 | 0.30 | 0.31 | 0.30 | 0.30 | 0.27 | 0.29 |  |

Trawl Selectivity

|  | 21.0 | 19.0 e | 21.1 | 21.2 | 21.3 | 21.4 | 21.5 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $64-73$ | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 0.67 | 1.00 |
| $74-83$ | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 0.83 | 1.00 |
| $84-93$ | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 0.92 | 1.00 |
| $94-103$ | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 0.97 | 1.00 |
| $104-113$ | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 0.99 | 1.00 |
| $114-123$ | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 0.99 | 1.00 |
| $124-133$ | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| $134+$ | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |

Winter Pot Selectivity

|  | 21.0 | 19.0 e | 21.1 | 21.2 | 21.3 | 21.4 | 21.5 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $64-73$ | 0.06 | 0.06 | 0.07 | 0.06 | 0.06 | 0.04 | 0.06 |
| $74-83$ | 0.42 | 0.43 | 0.50 | 0.42 | 0.43 | 0.36 | 0.41 |
| $84-93$ | 0.73 | 0.73 | 0.85 | 0.73 | 0.74 | 0.73 | 0.77 |
| $94-103$ | 0.86 | 0.86 | 0.95 | 0.86 | 0.86 | 0.99 | 0.93 |
| $104-113$ | 0.72 | 0.72 | 0.85 | 0.72 | 0.72 | 0.97 | 0.82 |
| $114-123$ | 0.50 | 0.50 | 0.62 | 0.50 | 0.51 | 0.86 | 0.63 |
| $124-133$ | 0.29 | 0.29 | 0.31 | 0.29 | 0.29 | 0.55 | 0.38 |
| $134+$ | 0.14 | 0.14 | 0.11 | 0.14 | 0.14 | 0.20 | 0.18 |

Winter Pot retention probability

|  | 21.0 | 19.0 e | 21.1 | 21.2 | 21.3 | 21.4 | 21.5 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $64-73$ | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| $74-83$ | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| $84-93$ | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| $94-103$ | 0.08 | 0.08 | 0.06 | 0.08 | 0.08 | 0.08 | 0.08 |
| $104-113$ | 0.82 | 0.82 | 0.64 | 0.82 | 0.82 | 0.85 | 0.84 |
| $114-123$ | 1.00 | 1.00 | 0.98 | 1.00 | 1.00 | 1.00 | 1.00 |
| $124-133$ | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| $134+$ | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |

Summer Commercial selectivity

|  | 21.0 | 19.0 e | 21.1 | 21.2 | 21.3 | 21.4 | 21.5 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $64-73$ | 0.12 | 0.13 | 0.10 | 0.13 | 0.12 | 0.04 | 0.09 |
| $74-83$ | 0.33 | 0.34 | 0.29 | 0.34 | 0.33 | 0.16 | 0.26 |
| $84-93$ | 0.63 | 0.64 | 0.60 | 0.65 | 0.64 | 0.45 | 0.57 |
| $94-103$ | 0.86 | 0.86 | 0.84 | 0.87 | 0.86 | 0.77 | 0.84 |
| $104-113$ | 0.96 | 0.96 | 0.95 | 0.96 | 0.96 | 0.93 | 0.95 |
| $114-123$ | 0.99 | 0.99 | 0.99 | 0.99 | 0.99 | 0.98 | 0.99 |
| $124-133$ | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| $134+$ | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |

Summer Commercial retention probability 1976-2007

|  | 21.0 | 19.0 e | 21.1 | 21.2 | 21.3 | 21.4 | 21.5 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $64-73$ | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| $74-83$ | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| $84-93$ | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| $94-103$ | 0.09 | 0.07 | 0.09 | 0.07 | 0.10 | 0.07 | 0.08 |
| $104-113$ | 0.88 | 0.87 | 0.87 | 0.87 | 0.87 | 0.82 | 0.86 |
| $114-123$ | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| $124-133$ | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| $134+$ | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |

Summer Commercial retention probability 2008-2019

|  | 21.0 | 19.0 e | 21.1 | 21.2 | 21.3 | 21.4 | 21.5 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $64-73$ | 0.00 |  | 0.00 |  | 0.00 | 0.00 | 0.00 |
| $74-83$ | 0.00 |  | 0.00 |  | 0.00 | 0.00 | 0.00 |
| $84-93$ | 0.00 |  | 0.00 |  | 0.00 | 0.00 | 0.00 |
| $94-103$ | 0.02 |  | 0.01 |  | 0.02 | 0.01 | 0.02 |
| $104-113$ | 0.88 |  | 0.82 |  | 0.89 | 0.89 | 0.89 |
| $114-123$ | 1.00 |  | 1.00 |  | 1.00 | 1.00 | 1.00 |
| $124-133$ | 1.00 |  | 1.00 |  | 1.00 | 1.00 | 1.00 |
| $134+$ | 1.00 |  | 1.00 |  | 1.00 | 1.00 | 1.00 |

Table 14. Annual abundance estimates (million crab) and mature male biomass (Feb 01) (MMB, million lb) for Norton Sound red king crab estimated by a length-based analysis.

MMB

| Year | Model | Model | Model | Model | Model | Model | Model |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 21.0 | 19.0 e | 21.1 | 21.2 | 21.3 | 21.4 | 21.5 |
| 1976 | 17.70 | 16.60 | 8.36 | 17.62 | 16.59 | 22.07 | 20.79 |
| 1977 | 19.39 | 18.37 | 9.97 | 19.32 | 18.34 | 21.57 | 21.80 |
| 1978 | 16.80 | 16.00 | 9.43 | 16.74 | 15.98 | 17.04 | 18.17 |
| 1979 | 11.83 | 11.29 | 6.93 | 11.78 | 11.28 | 11.41 | 12.51 |
| 1980 | 6.29 | 5.97 | 3.35 | 6.25 | 5.97 | 5.92 | 6.63 |
| 1981 | 4.13 | 4.01 | 2.25 | 4.09 | 4.01 | 4.03 | 4.35 |
| 1982 | 3.04 | 3.03 | 1.35 | 3.01 | 3.04 | 3.24 | 3.30 |
| 1983 | 3.85 | 3.81 | 1.83 | 3.83 | 3.80 | 4.21 | 4.20 |
| 1984 | 4.42 | 4.35 | 2.02 | 4.41 | 4.34 | 4.71 | 4.76 |
| 1985 | 4.92 | 4.81 | 2.13 | 4.91 | 4.79 | 5.12 | 5.28 |
| 1986 | 5.23 | 5.02 | 2.12 | 5.21 | 5.00 | 5.22 | 5.55 |
| 1987 | 5.19 | 4.89 | 1.95 | 5.17 | 4.88 | 5.00 | 5.44 |
| 1988 | 5.23 | 4.86 | 1.94 | 5.22 | 4.84 | 4.98 | 5.47 |
| 1989 | 5.09 | 4.71 | 1.96 | 5.08 | 4.69 | 4.79 | 5.28 |
| 1990 | 4.87 | 4.52 | 1.95 | 4.85 | 4.50 | 4.59 | 5.01 |
| 1991 | 4.60 | 4.33 | 1.96 | 4.58 | 4.31 | 4.37 | 4.72 |
| 1992 | 4.37 | 4.19 | 2.13 | 4.35 | 4.18 | 4.19 | 4.45 |
| 1993 | 4.02 | 3.95 | 2.16 | 4.01 | 3.93 | 3.87 | 4.06 |
| 1994 | 3.37 | 3.38 | 1.88 | 3.35 | 3.36 | 3.23 | 3.37 |
| 1995 | 2.78 | 2.85 | 1.60 | 2.77 | 2.84 | 2.69 | 2.77 |
| 1996 | 2.40 | 2.50 | 1.39 | 2.39 | 2.49 | 2.38 | 2.40 |
| 1997 | 2.43 | 2.55 | 1.44 | 2.42 | 2.54 | 2.44 | 2.43 |
| 1998 | 3.07 | 3.22 | 1.86 | 3.06 | 3.20 | 3.23 | 3.11 |
| 1999 | 4.01 | 4.20 | 2.40 | 4.01 | 4.17 | 4.18 | 4.08 |
| 2000 | 4.35 | 4.52 | 2.69 | 4.34 | 4.49 | 4.26 | 4.33 |
| 2001 | 4.03 | 4.16 | 2.51 | 4.03 | 4.14 | 3.82 | 3.96 |
| 2002 | 4.03 | 4.08 | 2.49 | 4.03 | 4.06 | 3.92 | 4.01 |
| 2003 | 4.36 | 4.32 | 2.65 | 4.36 | 4.30 | 4.36 | 4.39 |
| 2004 | 4.62 | 4.50 | 2.73 | 4.62 | 4.48 | 4.66 | 4.67 |
| 2005 | 4.47 | 4.31 | 2.60 | 4.48 | 4.29 | 4.47 | 4.52 |
| 2006 | 4.23 | 4.06 | 2.43 | 4.24 | 4.04 | 4.27 | 4.31 |
| 2007 | 4.32 | 4.17 | 2.45 | 4.34 | 4.16 | 4.51 | 4.46 |
| 2008 | 4.71 | 4.60 | 2.70 | 4.73 | 4.60 | 5.01 | 4.90 |
| 2009 | 5.07 | 4.99 | 2.86 | 5.09 | 4.99 | 5.39 | 5.28 |
| 2010 | 5.19 | 5.15 | 2.93 | 5.21 | 5.16 | 5.43 | 5.37 |
| 2011 | 4.79 | 4.78 | 2.72 | 4.81 | 4.79 | 4.83 | 4.91 |
| 2012 | 4.28 | 4.30 | 2.55 | 4.29 | 4.31 | 4.26 | 4.35 |
| 2013 | 4.05 | 4.09 | 2.46 | 4.05 | 4.11 | 4.16 | 4.14 |
| 2014 | 4.43 | 4.52 | 2.72 | 4.43 | 4.54 | 4.85 | 4.63 |
| 2015 | 4.95 | 5.10 | 3.11 | 4.96 | 5.11 | 5.39 | 5.17 |
| 2016 | 4.49 | 4.67 | 2.85 | 4.50 | 4.67 | 4.64 | 4.60 |
| 2017 | 3.63 | 3.80 | 2.42 | 3.64 | 3.81 | 3.53 | 3.62 |
| 2018 | 2.75 | 2.92 | 1.99 | 2.76 | 2.92 | 2.60 | 2.69 |
| 2019 | 2.32 | 2.47 | 1.90 | 2.33 | 2.47 | 2.22 | 2.24 |
| 2020 | 2.93 | 3.08 | 2.68 | 2.93 | 3.09 | 3.05 | 2.87 |
| 2021 | 4.38 | 4.54 | 4.07 | 4.37 | 4.56 | 4.37 | 4.21 |
| 2022 | 5.18 | 5.33 | 4.90 | 5.17 | 5.35 | 4.70 | 4.79 |

Legal abundance (> 103mm CL)

| Year | Model | Model | Model | Model | Model | Model | Model |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | 21.0 | 19.0 e | 21.1 | 21.2 | 21.3 | 21.4 | 21.5 |
| 1976 | 7.50 | 7.05 | 3.65 | 7.46 | 7.04 | 9.70 | 8.89 |
| 1977 | 7.18 | 6.82 | 3.76 | 7.15 | 6.81 | 8.21 | 8.12 |
| 1978 | 5.59 | 5.34 | 3.16 | 5.57 | 5.33 | 5.77 | 6.06 |
| 1979 | 3.73 | 3.57 | 2.18 | 3.71 | 3.57 | 3.65 | 3.96 |
| 1980 | 1.97 | 1.88 | 1.04 | 1.95 | 1.88 | 1.87 | 2.07 |
| 1981 | 1.36 | 1.35 | 0.74 | 1.35 | 1.35 | 1.37 | 1.44 |
| 1982 | 1.19 | 1.20 | 0.55 | 1.18 | 1.20 | 1.34 | 1.32 |
| 1983 | 1.57 | 1.54 | 0.76 | 1.56 | 1.54 | 1.79 | 1.73 |
| 1984 | 1.77 | 1.73 | 0.82 | 1.76 | 1.73 | 1.96 | 1.93 |
| 1985 | 1.93 | 1.88 | 0.84 | 1.92 | 1.87 | 2.10 | 2.11 |
| 1986 | 2.00 | 1.91 | 0.82 | 1.99 | 1.90 | 2.08 | 2.15 |
| 1987 | 1.94 | 1.82 | 0.73 | 1.94 | 1.81 | 1.96 | 2.07 |
| 1988 | 1.93 | 1.78 | 0.71 | 1.93 | 1.78 | 1.93 | 2.05 |
| 1989 | 1.84 | 1.70 | 0.70 | 1.84 | 1.69 | 1.81 | 1.94 |
| 1990 | 1.75 | 1.63 | 0.69 | 1.74 | 1.62 | 1.73 | 1.83 |
| 1991 | 1.64 | 1.55 | 0.69 | 1.63 | 1.55 | 1.63 | 1.71 |
| 1992 | 1.53 | 1.48 | 0.73 | 1.53 | 1.48 | 1.53 | 1.59 |
| 1993 | 1.40 | 1.39 | 0.73 | 1.40 | 1.38 | 1.40 | 1.43 |
| 1994 | 1.17 | 1.18 | 0.63 | 1.16 | 1.18 | 1.16 | 1.18 |
| 1995 | 0.97 | 1.00 | 0.54 | 0.97 | 1.00 | 0.98 | 0.98 |
| 1996 | 0.87 | 0.91 | 0.49 | 0.87 | 0.91 | 0.90 | 0.89 |
| 1997 | 0.93 | 0.97 | 0.54 | 0.93 | 0.97 | 0.97 | 0.95 |
| 1998 | 1.24 | 1.30 | 0.72 | 1.24 | 1.29 | 1.38 | 1.29 |
| 1999 | 1.60 | 1.67 | 0.91 | 1.60 | 1.66 | 1.73 | 1.65 |
| 2000 | 1.61 | 1.67 | 0.96 | 1.61 | 1.66 | 1.63 | 1.63 |
| 2001 | 1.45 | 1.49 | 0.86 | 1.45 | 1.48 | 1.44 | 1.45 |
| 2002 | 1.50 | 1.50 | 0.87 | 1.50 | 1.50 | 1.54 | 1.53 |
| 2003 | 1.68 | 1.65 | 0.96 | 1.68 | 1.64 | 1.76 | 1.73 |
| 2004 | 1.76 | 1.70 | 0.98 | 1.76 | 1.69 | 1.86 | 1.81 |
| 2005 | 1.64 | 1.58 | 0.90 | 1.65 | 1.57 | 1.72 | 1.70 |
| 2006 | 1.56 | 1.50 | 0.86 | 1.57 | 1.49 | 1.66 | 1.63 |
| 2007 | 1.66 | 1.60 | 0.92 | 1.67 | 1.60 | 1.83 | 1.75 |
| 2008 | 1.84 | 1.80 | 1.03 | 1.84 | 1.80 | 2.05 | 1.94 |
| 2009 | 1.97 | 1.94 | 1.10 | 1.97 | 1.95 | 2.19 | 2.09 |
| 2010 | 1.96 | 1.95 | 1.09 | 1.97 | 1.95 | 2.14 | 2.06 |
| 2011 | 1.73 | 1.73 | 0.97 | 1.74 | 1.74 | 1.81 | 1.80 |
| 2012 | 1.54 | 1.54 | 0.90 | 1.54 | 1.55 | 1.59 | 1.58 |
| 2013 | 1.52 | 1.54 | 0.91 | 1.52 | 1.55 | 1.65 | 1.59 |
| 2014 | 1.76 | 1.80 | 1.08 | 1.76 | 1.81 | 2.05 | 1.89 |
| 2015 | 1.94 | 2.00 | 1.21 | 1.94 | 2.00 | 2.20 | 2.06 |
| 2016 | 1.63 | 1.70 | 1.03 | 1.64 | 1.70 | 1.74 | 1.69 |
| 2017 | 1.24 | 1.30 | 0.82 | 1.25 | 1.30 | 1.23 | 1.25 |
| 2018 | 0.93 | 0.99 | 0.67 | 0.94 | 0.99 | 0.90 | 0.92 |
| 2019 | 0.83 | 0.87 | 0.67 | 0.83 | 0.88 | 0.81 | 0.81 |
| 2020 | 1.20 | 1.25 | 1.06 | 1.20 | 1.26 | 1.31 | 1.20 |
| 2021 | 1.82 | 1.87 | 1.62 | 1.81 | 1.88 | 1.85 | 1.76 |
| 2022 | 1.99 | 2.05 | 1.81 | 1.99 | 2.05 | 1.84 | 1.85 |
|  |  |  |  |  |  |  |  |

Table 15. Summary of catch (million lb) for Norton Sound red king crab. Assumed average crab weight is 2.0 lb for winter subsistence catch and 1.0 lb for Winter subsistence discards.

| Year | Summer Com | Winter Com | Winter Sub | Discards Winter Sub | Total |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1977 | 0.52 | 0.000 | 0.000 | 0 | 0.520 |
| 1978 | 2.09 | 0.024 | 0.025 | 0.008 | 2.147 |
| 1979 | 2.93 | 0.001 | 0.000 | 0 | 2.931 |
| 1980 | 1.19 | 0.000 | 0.000 | 0 | 1.190 |
| 1981 | 1.38 | 0.000 | 0.001 | 0 | 1.381 |
| 1982 | 0.23 | 0.000 | 0.003 | 0.001 | 0.234 |
| 1983 | 0.37 | 0.001 | 0.021 | 0.006 | 0.398 |
| 1984 | 0.39 | 0.002 | 0.022 | 0.005 | 0.419 |
| 1985 | 0.43 | 0.003 | 0.017 | 0.002 | 0.452 |
| 1986 | 0.48 | 0.005 | 0.014 | 0.004 | 0.503 |
| 1987 | 0.33 | 0.003 | 0.012 | 0.002 | 0.347 |
| 1988 | 0.24 | 0.001 | 0.005 | 0.001 | 0.247 |
| 1989 | 0.25 | 0.000 | 0.012 | 0.002 | 0.264 |
| 1990 | 0.19 | 0.010 | 0.024 | 0.004 | 0.228 |
| 1991 | 0 | 0.010 | 0.015 | 0.002 | 0.027 |
| 1992 | 0.07 | 0.021 | 0.023 | 0.003 | 0.117 |
| 1993 | 0.33 | 0.005 | 0.002 | 0 | 0.337 |
| 1994 | 0.32 | 0.017 | 0.008 | 0.001 | 0.346 |
| 1995 | 0.32 | 0.022 | 0.011 | 0.002 | 0.355 |
| 1996 | 0.22 | 0.005 | 0.003 | 0.001 | 0.229 |
| 1997 | 0.09 | 0.000 | 0.001 | 0.001 | 0.092 |
| 1998 | 0.03 | 0.002 | 0.017 | 0.012 | 0.061 |
| 1999 | 0.02 | 0.007 | 0.015 | 0.003 | 0.045 |
| 2000 | 0.3 | 0.008 | 0.011 | 0.004 | 0.323 |
| 2001 | 0.28 | 0.003 | 0.001 | 0 | 0.284 |
| 2002 | 0.25 | 0.007 | 0.004 | 0.003 | 0.264 |
| 2003 | 0.26 | 0.017 | 0.008 | 0.005 | 0.290 |
| 2004 | 0.34 | 0.001 | 0.002 | 0.001 | 0.344 |
| 2005 | 0.4 | 0.006 | 0.008 | 0.003 | 0.417 |
| 2006 | 0.45 | 0.000 | 0.002 | 0.001 | 0.453 |
| 2007 | 0.31 | 0.008 | 0.021 | 0.011 | 0.350 |
| 2008 | 0.39 | 0.015 | 0.019 | 0.009 | 0.433 |
| 2009 | 0.4 | 0.012 | 0.010 | 0.002 | 0.424 |
| 2010 | 0.42 | 0.012 | 0.014 | 0.002 | 0.448 |
| 2011 | 0.4 | 0.009 | 0.013 | 0.003 | 0.425 |
| 2012 | 0.47 | 0.025 | 0.015 | 0.004 | 0.514 |
| 2013 | 0.35 | 0.061 | 0.015 | 0.014 | 0.440 |
| 2014 | 0.39 | 0.035 | 0.007 | 0.002 | 0.434 |
| 2015 | 0.40 | 0.099 | 0.019 | 0.005 | 0.523 |
| 2016 | 0.42 | 0.080 | 0.011 | 0.001 | 0.512 |
| 2017 | 0.41 | 0.078 | 0.012 | 0.001 | 0.501 |
| 2018 | 0.30 | 0.029 | 0.008 | 0.001 | 0.338 |
| 2019 | 0.08 | 0.032 | 0.003 | 0.001 | 0.116 |
| 2020 | 0 | Conf. | 0.001 | 0.000 | Conf |
| 2021 | 0 | 0.001 | 0.009 | 0.000 | 0.010 |

Figure 1. King crab fishing districts and sections of Statistical Area Q.


Figure 2. Closed water regulations in effect for the Norton Sound commercial crab fishery. Line around the coastline delineates the $3-\mathrm{mil}$ state waters zone.


Figure 3. Model estimated annual molting probability, and selectivity for trawl survey, winter pot survey, summer commercial fishery, and summer and winter commercial retention. X-axis is carapace length (mm).



Figure 4. Model estimated abundances of total, legal (CL>104mm) and recruit (CL 64-94nn) males during1976-2021.



Figure 5. Estimated MMB during 1976-2022 (Model 21.0 solid black, Model 21.1 dash red). Horizontal line Bmsy (Average MMB of 1980-2022).



Figure 6. Observed (open circle) (White: NMFS, Red ADF\&G) and model trawl survey male abundances with $95 \%$ lognormal Confidence Intervals (crab $\geq 64 \mathrm{mmCL}$ ).



Figure 7. Observed (open circle) with $95 \%$ lognormal Confidence Intervals with additional variance (red), and model estimated (Model 21.0 line black, Model 21.1 dash red) standardized CPUE


Figure 8. Predicted vs. observed (bar New Shell: blue, Old Shell: blue) length class proportions for the summer commercial harvest 1977-2019.

Model 21.0, 19.0e, 21.2, 21.3


Models 21.0, 21.1, 21.4, 21.5


Figure 9. Predicted (line) vs. observed (bar New Shell: blue, Old Shell: blue) length class proportions for trawl survey 1976-2021.


CL mm


CL mm

Figure 10 Predicted (line) vs. observed (bar New Shell: blue, Old Shell: blue) length class proportions for winter pot survey 1982-2012.


CL mm


Figure 11. Predicted (line) vs. observed (bar New Shell: left blue, Old Shell: right green) length class proportions for summer commercial total and discards (1987-1994, 2012-2019) and winter commercial retained fishery 2015-2018

Summer discards, total, winter retain: observed vs predicted


Summer total


Winter retain


Summer discards, total, winter retain: observed vs predicted


Summer total



Figure 12. Predicted (line) vs. observed (bar) length class proportions for tag recovery data.
Tag recovery data observed vs predicted
Recovery after 1 year


Recovery after 2 years


Recovery after 3 years

$\begin{array}{ll}\text { - } & 21.0 \\ - & 19.0 \mathrm{e}\end{array}$

- 19.0 e
$\ldots=21.2$
$-\quad 21.2$
$--\quad 21.3$


Recovery after 2 years



Figure 13. Input vs. model implied effective sample size. Figures in the first column show implied effective sample size ( x -axis) vs. frequency ( y -axis). Vertical solid line is the harmonic mean of implied sample size. Figures in the second column show input sample sizes (x-axis) vs. implied effective sample sizes ( y -axis). Dashed line indicates the linear regression slope, and solid line is $1: 1$ line. Figures in the third column show years (x-axis) vs. implied effective sample sizes (y-axis). Horizontal solid line is the harmonic mean of implied sample size.





Figure 14. Bubble plots of predicted and observed length proportions.
Black circle indicates model underestimates than observed, and white circle indicates model overestimates than observed. Size of circle indicates degree of deviance (larger circle $=$ larger deviance). In ideal model fit case, distribution of sizes and colors of circles should be random (i.e., no systematic model misfits).







Figure 15. Standardized Pearson residual plot of trawl, summer commercial retain, winter survey, observer, length size classes 1-8.







Figure 16. QQ Plot of Trawl survey and Commercial CPUE.




Commercial CPUE






Figure 17. Retrospective Analyses of Norton Sound Red King Crab MMB from 2011 to 2022. Solid black line: 2022 assessment model results.

## Retrospective Analysis Model 21.0



Retrospective Analysis Model 21.4


Retrospective Analysis Model 21.5



[^0]:    ${ }^{\text {a }}$ Deadloss included in total. ${ }^{\text {b }}$ Millions of pounds. ${ }^{\mathrm{c}}$ Information not available.

