Aleutian Islands Golden King Crab (Lithodes aequispinus) Model-Based Stock Assessment in spring 2016

Draft report for the May 2016 Crab Plan Team Meeting

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## Executive Summary

1. Stock

Golden king crab, Lithodes aequispinus, Aleutian Islands, east of $174^{\circ} \mathrm{W}$ longitude (EAG) and west of $174^{\circ} \mathrm{W}$ longitude (WAG).

## 2. Catches

The Aleutian Islands golden king crab commercial fishery developed in the early 1980s; the harvest peaked in 1986/87 at 5.900 and 8.800 million pounds, respectively, for EAG and WAG. Catches have been steady since 1996/97 following implementation of total allowable catches (TACs) of 3.000 (EAG) and 2.700 (WAG) million pounds. The TACs were increased to 3.150 and 2.835 million pounds for the two respective regions for the 2008/09 fishing year following an Alaska Board of Fisheries (BOF) decision. These levels are below the limit TACs determined under Tier 5 criteria (considering 1991-1995 mean catch as the limit catch) under the most recent crab management plan. The TACs were further increased by another BOF decision to 3.310 million pounds for EAG and 2.980 million pounds for WAG beginning with the 2012/13 fishing year. The fishery has harvested close to TAC levels since 1996/97. Catch rates increased in both EAG and WAG fisheries in the mid-2000s; however, in recent years WAG catch rates have declined.

## 3. Stock biomass

Estimated mature male biomass (MMB) for EAG under scenario 1 decreased from peak levels during the mid-1990s of the directed fishery and then systematically increased and stabilized in recent years. Estimated MMB under scenario 1 for WAG decreased during the late 1980s and 1990s and systematically increased during 2000s and decreased during last few years since 2009. The lowest levels of MMB for EAG were observed in 1996-1997 and that for WAG were in 1991-1992. Stock trends reflected the fishery standardized CPUE trends in both regions.

## 4. Recruitment

The numbers of recruits to the model size groups under scenario 1 have fluctuated in both EAG and WAG. For EAG, the model recruitment was high in 1988, 1991 and

2010-2011, and lowest in 1986, while model recruitment for WAG was highest in 1986 and 1993 and lowest in 2008.

## 5. Management performance

The model has not yet been used for making any management decisions.

## 6. Basis for the OFL

We provide the OFL estimates under the Tier 4 approach for EAG, WAG, and the two regions sum together (i.e. for the entire Aleutian Islands, AI), respectively. The length-based model developed for Tier 4 analysis estimates MMB on February 15 each year for the period 1986 through 2015 and projects to February 15, 2016 for OFL and ABC determination. The Tier 4 approach proposes a maximum $\mathrm{F}_{\mathrm{OFL}}$ of $\boldsymbol{\boldsymbol { \gamma }} \boldsymbol{M}$. The following OFLs and ABCs were determined under Tier 4 based on using the 1986-2015 mean MMB as the reference biomass ( $M M B_{r e f}$ ). The total OFL and ABC estimates are provided for ten scenarios for EAG, WAG, and AI, respectively in the following five tables. We treat scenario 1 as the base scenario. If the model is accepted, we recommend the OFL and ABC estimates for scenario 1 (base), scenario 10 (dome shaped selectivity), or scenario 16 (restricted time series of total catch and total catch size composition).

EAG (Tier 4):
Biomass, total OFL, and ABC for the next fishing season in million pounds.


## WAG (Tier 4):

Biomass, total OFL, and ABC for the next fishing season in million pounds.

| Scenario | Tier | $M M B_{\text {ref }}$ | $\begin{array}{ll} \hline \text { Current } & \mathrm{N} \\ \text { MMB } & M \end{array}$ | $\begin{array}{ll} \hline \mathrm{MMB} / & \\ M M B_{r e f} & \mathrm{~F} \end{array}$ | $\mathrm{F}_{\text {OFL }}$ | Years to define $M M B_{r e f}$ | $\gamma$ | M | OFL | ABC $\left(\mathrm{P}^{*}=0.49\right)$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 4a | 12.141 | 12.226 | 1.01 | 0.23 | 1986-2015 | 1 | 0.23 | 2.041 | 2.031 |
| 2 | 4b | 11.892 | 11.731 | 0.99 | 0.226 | 1986-2015 | 1 | 0.23 | 1.918 | 1.895 |
| 3 | 4 a | 11.957 | 11.969 | 1.00 | 0.23 | 1986-2015 | 1 | 0.23 | 2.010 | 2.000 |
| 4 | 4 a | 11.299 | 11.588 | 1.03 | 0.23 | 1986-2015 | 1 | 0.23 | 1.798 | 1.789 |
| 5 | 4a | 10.485 | 10.950 | 1.04 | 0.23 | 1986-2015 | 1 | 0.23 | 1.502 | 1.494 |
| 7 | 4a | 11.156 | 11.630 | 1.04 | 0.23 | 1986-2015 | 1 | 0.23 | 1.983 | 1.974 |
| 10 | 4 a | 19.265 | 19.809 | 1.03 | 0.23 | 1986-2015 | 1 | 0.23 | 2.554 | 2.488 |
| 12 | 4 a | 12.242 | 13.412 | 1.10 | 0.23 | 1986-2015 | 1 | 0.23 | 2.253 | 2.242 |
| 14 | 4 a | 19.359 | 21.256 | 1.10 | 0.23 | 1986-2015 | 1 | 0.23 | 2.809 | 2.795 |
| 16 | 4 a | 12.287 | 12.384 | 1.01 | 0.23 | 1986-2015 | 1 | 0.23 | 2.013 | 2.002 |
| Biomass in 1,000 t; total OFL and ABC for the next fishing season in t . |  |  |  |  |  |  |  |  |  |  |
| Scenario | Tier | $M M B_{\text {ref }}$ | Current <br> MMB | MMB/ <br> $M M B_{r e f}$ | $\mathrm{F}_{\mathrm{OFL}}$ | Years to define $M M B_{\text {ref }}$ | $\boldsymbol{V}$ | M | OFL | ABC$(\mathrm{P} *=0.49)$ |
|  |  |  |  |  |  |  |  |  |  |  |
| 1 | 4a | 5.507 | 5.546 | 61.01 | 10.23 | 1986-2015 | 1 | 0.23 | 925.882 | 921.389 |
| 2 | 4 b | 5.394 | 5.321 | 10.99 | 0.226 | 1986-2015 | 1 | 0.23 | 869.988 | 859.640 |
| 3 | 4a | 5.424 | 5.429 | $9 \quad 1.00$ | 0.23 | 1986-2015 | 1 | 0.23 | 911.578 | 907.304 |
| 4 | 4a | 5.125 | 5.256 | 61.03 | 0.23 | 1986-2015 | 1 | 0.23 | 815.478 | 811.430 |
| 5 | 4a | 4.756 | 4.967 | 71.04 | 40.23 | 1986-2015 | 1 | 0.23 | 681.168 | 677.799 |
| 7 | 4 a | 5.060 | 5.275 | 51.04 | 40.23 | 1986-2015 | 1 | 0.23 | 899.695 | 895.246 |
| 10 | 4 a | 8.738 | 8.986 | 61.03 | 0.23 | 1986-2015 | 1 | 0.23 | 1,158.281 | 1,128.612 |
| 12 | 4 a | 5.553 | 6.084 | 41.10 | 0.23 | 1986-2015 | 1 | 0.23 | 1,021.916 | 1,016.889 |
| 14 | 4a | 8.781 | 9.642 | 21.10 | 0.23 | 1986-2015 | 1 | 0.23 | 1,274.247 | 1,268.009 |
| 16 | 4 a | 5.573 | 5.617 | $7 \quad 1.01$ | 0.23 | 1986-2015 | 1 | 0.23 | 912.973 | 908.270 |

## Aleutian Islands (sum of OFL and ABC for EAG and WAG):

Total OFL and ABC for the next fishing season.

| Scenario | OFL <br> (million <br> pounds) | ABC <br> $\left(\mathrm{P}^{*}=0.49\right)$ <br> (million | OFL <br> $(1,000 \mathrm{t})$ | ABC <br> $\left(\mathrm{P}^{*}=0.49\right)$ <br> $(1,000 \mathrm{t})$ |
| :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |
| pounds) |  |  |  |  |

7. Probability density functions of OFL

Assuming a lognormal distribution of total OFL, we determined the cumulative distributions of OFL and selected the median as the OFL.
8. The basis for the ABC recommendation

See the ABC section
9. A summary of results of any rebuilding analysis:

Not applicable.

## A. Summary of Major Changes

1. Changes (if any) to management of the fishery

None.
2. Changes to input data
(a) Data update: The 2014/15 commercial fishery retained and total catch, observer nominal total CPUE and fishing effort (pot lifts) to calculate total catches for 1990/91-2014/15, and groundfish discarded catch by size for 1989/89-2014/15 were added. The commercial retained size frequency and observer sample size frequency data were recalculated weighting by sampled vessel's catch.
(b) New data: EAG male tag-recapture data by size and time-at-large for 1991, 1997, 2000, 2003, and 2006 releases were considered for the WAG model analysis.
(c) Observer pot sample legal size crab CPUE data were standardized by the generalized linear model (GLM) with the negative binomial link function, separately for 1995/96-2004/05 and 2005/06-2014/15 periods.
(d) Fish ticket retained catch CPUE were standardized by the GLM using a lognormal link function considering a suite of explanatory variables. The 19851998 data were used in the fit and the indices were used in model scenario 3.
3. Changes to assessment methodology

None. The same model has been improved.
4. Changes to assessment results

Not applicable because the model has not been used previously.

## B. Response to September 2015 CPT comments

Comment 1: the impact of the way the initial conditions are specified:

- estimated as log-deviations about an initial abundance (as in Scenario 1 for the EAG)
- based on projecting the model from unfished equilibrium at some earlier time (e.g.1960), with recruitments estimated for each year after that time and catches from 1981.

Response:

1. The initial abundance in 1985 was determined by two methods: a) projecting the model from unfished equilibrium in 1960 with recruitment estimated for each year after 1960 up to 1985 (Equation A.6) and considering retained catches during 1981/82-1984/85; and b) using the exponential formula (Equation A.4) to generate the initial abundance. Initial abundance and the difference in log initial abundance are compared for the two methods for EAG and WAG, respectively:
(a)

(b)



Figure 1. Plots of the initial (1985/86) abundances [top plots in (a) and (b)] and differences of the $\log$ initial abundances [bottom plots in (a) and (b)] derived by equilibrium and exponential formulations vs. the length-class mid length for EAG [(a)] and WAG [(b)]. An $M$ value of $0.23 \mathrm{yr}^{-1}$ was used.

The initial abundances for length-classes above 128 mm CL are similar for both regions (Figure 1). There are few zero entries for certain size bins when exponential formulation was used, but not when equilibrium condition was used. So, equilibrium initial condition appears to be a better formulation.

Comment 2. The impact of the choice of the value for natural mortality. The scenarios can be based on factors included in current Scenario 1 (two total selectivity curves, two catchabilities, and one retention curve). Each of the following scenarios should be conducted for all options to specify the initial conditions and for natural mortality.

## 2(a) Drop the groundfish bycatch size-composition data, pre-specify selectivity for the groundfish fishery (e.g., uniform across sizes), and assign a low weight to the groundfish bycatch mass data.

Response
Natural mortality, $M$ :
The natural mortality was estimated within the integrated model, separately for EAG and WAG stocks without using any penalty function for $M$. The groundfish (or trawl) bycatch length composition data were not considered for $M$ estimation because the fits were bad (see Figures 8 and 28 for EAG and WAG, respectively). The $M$ was optimized within the range 0.1 to $0.4 \mathrm{yr}^{-1}$. We tried several different initial $M$ values and found that the $M$ estimates were stable (Table C). The best $M$ estimate for EAG was $0.2462 \mathrm{yr}^{-1}(S E=$ $\left.0.0378 \mathrm{yr}^{-1}\right)$ and that for WAG was $0.2160 \mathrm{yr}^{-1}\left(S E=0.0343 \mathrm{yr}^{-1}\right)$. The best $M$ estimates were obtained for a starting value of $0.26 \mathrm{yr}^{-1}$ in the optimization (Table C). Then we explored the total and component negative log likelihood values for different levels of $M$. The plots in Figure 2 show well defined minima in the total negative log likelihood values at the best $M$ estimates for EAG and WAG, respectively, which are greater than the currently used $M$ of $0.18 \mathrm{yr}^{-1}$. For the base and most other model scenarios, an average of the two $M$ estimates for EAG and WAG, which is $0.23 \mathrm{yr}^{-1}$ (rounded to two decimal places), was used. There are no evidences for the stocks in the two regions to have different $M$ values. Hence, an average $M$ value would be an appropriate choice for a common $M$ to use in the stock assessment for the two fisheries.

Table C. Estimates of $M$ by the integrated model for different initial $M$ values in EAG and WAG. $S E=$ asymptotic standard error of $M$.

|  | EAG |  | WAG |  |
| :---: | :---: | :---: | :---: | :---: |
| Initial $M$ yr $^{-1}$ | Estimated $M \mathrm{yr}^{-1}$ | $S E \mathrm{yr}^{-1}$ | Estimated $M \mathrm{yr}^{-1}$ | $S E \mathrm{yr}^{-1}$ |
| 0.1 | 0.24624 | 0.03777 | 0.21595 | 0.03433 |
| 0.15 | 0.24624 | 0.03777 | 0.21593 | 0.03432 |
| 0.20 | 0.24624 | 0.03777 | 0.21595 | 0.03433 |
| 0.26 | 0.24624 | 0.03777 | 0.21595 | 0.03433 |
| 0.30 | 0.24624 | 0.03777 | 0.21590 | 0.03431 |
| 0.35 | 0.24624 | 0.03777 | 0.21595 | 0.03433 |
| 0.40 | 0.24624 | 0.03777 | 0.21595 | 0.03433 |



Figure 2. Total and components negative log-likelihoods vs. $M$ for scenario 1 model fit to 1985/86-2014/15 golden king crab data in the EAG (top panel) and WAG (bottom panel). The negative log likelihood values were zero adjusted. The dotted vertical line is for $M=0.18 \mathrm{yr}^{-1}$.

2(a): All EAG and WAG scenarios used a low groundfish bycatch likelihood weight, 0.2 (CV= 3.344). Scenarios 5 and 6 provide the OFL and ABC estimates when groundfish size composition was dropped. Dropping groundfish size composition and pre-specifying its selectivity to 1 for all length-classes resulted in lower OFL and ABC estimates than that for other scenarios for WAG, but not for EAG. Contribution of groundfish bycatch removal to the total removal during 1989/90-2014/15 was very small, 0.03\% for EAG and $0.05 \%$ for WAG.

2 (b) Estimate a new selectivity pattern in 1994/95 rather than 1998/99 as in Scenario 2 because the total catch size-composition data for the EAG for the years before 1995 were only collected from Catcher Processor (CP) vessels and appear to contain far more small crab than for the years after 1995.

## Response

Scenarios 7 (equilibrium initial cond.) provides the OFL estimates for EAG and WAG, respectively, when the total selectivity and catchability patterns were split into 1985/86$1994 / 95$, 1995/96-2004/05, and 2005/06-2014/15. This recommendation produced similar fit to that of scenario 1. However, this scenario reduced the pot fishery fishing mortality, total and retained catch OFLs from that of scenario 1 (Table 29).

## 2 (c) Consider dome-shaped rather than asymptotic selectivity.

Response

1. Scenarios $10,11,14$, and 15 provide the OFL estimates for EAG and WAG, respectively; when the dome shaped selectivity (Equation A.11) was used (Table 29).
2. Area shrinkage: Scenarios 12 and 13 provide the OFL estimates considering the asymptotic total selectivity and the shrinkage of area over the years in the CPUE calculation (Equation A.17) whereas scenarios 14 and 15 provide those estimates made under the same condition considering the dome shaped total selectivity (Table 29).

The OFL estimates are higher and the F estimates are lower (Table 29, Figures 22 and 40 for EAG and WAG, respectively) for the dome shaped total selectivity compared to asymptotic total selectivity. However, the trends in F values are similar. The shrinkage of area with dome shaped selectivity dramatically increased the OFL estimates for WAG, but not for EAG.

The SSC concurred with the CPT comments.

## C. Introduction

1. Scientific name: Golden king crab, Lithodes aequispinus.
2. Distribution: In Alaska, golden king crab is distributed in the Aleutian Islands, on the continental slope of the eastern Bering Sea, and around the Gulf of Alaska to southeastern Alaska.
3. Evidence of stock structure: There is no direct evidence of separate stock structure in the Aleutian Islands. But different CPUE trends suggest different factors may influence stock productivity in EAG and WAG.
4. Life history characteristics relevant to management: There is a paucity of information on golden king crab life history characteristics due in part to the deep depth distribution ( $\sim 200-1000 \mathrm{~m}$ ) and the asynchronous nature of life history events (Otto and Cummiskey 1985; Somerton and Otto 1986). The reproductive cycle is thought to last approximately 24 months and at any one time, ovigerous females can be found carrying egg clutches in highly disparate developmental states (Otto and Cummiskey 1985). Females carry large, yolk-rich, eggs, which hatch into lecithotrophic (non-feeding) larvae that are negatively phototactic (Adams and Paul 1999). Molting and mating are also asynchronous and protracted (Shirley and Zhou 1997; Otto and Cummiskey 1985) with some indications of seasonality (Hiramoto 1985). Molt increment for large males (adults) in Southeast Alaska is 16.3 mm CL per molt (Koeneman and Buchanan 1985), and for legal males in the EAG was estimated at 14.4 mm CL (Watson et al. 2002). Annual molting probability of males decreases with increasing size, which results in a protracted inter-molt period and creates difficulty in determining annual molt probability (Watson et al. 2002). Male size-at-maturity varies among stocks (Webb 2014), but declines with increasing latitude from about 130 mm CL in the Aleutian Islands to 90 mm CL in Saint Matthew Island section (Somerton and Otto 1986). Along with a lack of annual survey data, limited stock-specific life history stock information prevents development of the standard length-based assessment model.
5. and
6. Brief summary of management history and annual ADFG harvest strategy: Since 1996, the Alaska Department of Fish and Game (ADF\&G) has divided management of the Aleutian Islands golden king crab fishery at $174^{\circ} \mathrm{W}$ longitude (ADF\&G 2002). Hereafter, the east of $174^{\circ} \mathrm{W}$ longitude stock segment is referred to as EAG and the west of $174^{\circ} \mathrm{W}$ longitude stock segment is referred to as WAG. The stocks in the two areas were managed with a constant annual guideline harvest level or total allowable (retained) catch ( 3.000 million pounds for EAG and 2.700 million pounds for WAG). In 2008, however, the total allowable catch was increased by the BOF to 3.150 and 2.830 million pounds for EAG and WAG, respectively (an approximately $5 \%$ increase in TAC). Additional management measures include a male-only fishery and a minimum legal size limit ( 152.4 mm CW, or approximately 136 mm CL ), which is at least one annual molt increment larger than the $50 \%$ maturity length of 120.8 mm CL for males (Otto and Cummiskey, 1985). In the model scenarios, a knife-edge $50 \%$ maturity length of 121 mm CL was used for mature male biomass (MMB) estimation. Daily catch and catch-per-unit effort (CPUE) are determined in-season to monitor fishery performance and progress towards the respective TACs. Figures 3 to 5 provide the historical time series of catches, CPUE, and the geographic distribution of catch during recent fishing seasons. Increases in CPUE were observed during the late 1990s through the early 2000s, and with the implementation of crab rationalization in 2005. This is likely due to changes in gear in the late 1990s (crab fishermen, personal communication, July 1, 2008) and, after rationalization, to increased soak time (Siddeek et al. 2015), and decreased competition owing to the reduced number of vessels fishing. Decreased competition could allow crab vessels to target only the most productive fishing areas. In 2012, the BOF increased the TAC levels to 3.31 million pounds for EAG and 2.98 million pounds for WAG beginning with the 2012/13 fishing year.
7. Summary of the history of the basis and estimates $M M B_{M S Y}$ or proxy $M M B_{M S Y}$ : The Tier 4 assessment model has not yet been accepted.

## D. Data

1. Summary of new information:

Data are updated by adding the 2014/15 commercial fishery retained catch by size, estimated total catch by size, groundfish male discard catch by size, and observer CPUE index with standard error to the time series. The details are given in the following table:
2. Available catch and tagging data.

| Years | Retained Catch | Total Catch | Groundfish Discarded Catch | Observer CPUE Index | Fishery Retained Catch CPUE | Tag Releases | Tag Recaptures |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Data <br> Types | By length | By length (Observer nominal total CPUE and effort were used to estimate total catch) | By length | Annual CPUE indices with standard errors were estimated by negative binomial GLM | Annual CPUE indices with standard errors were estimated by lognormal GLM for scenario 3 |  | Releaserecapture length and time-atliberty. There are 1717 records. |
| 1985/86 |  |  |  |  |  |  |  |
| 1986/87 |  |  |  |  |  |  |  |
| 1987/88 |  |  |  |  |  |  |  |
| 1988/89 |  |  |  |  |  |  |  |
| 1989/90 |  |  |  |  |  |  |  |
| 1990/91 |  |  |  |  |  |  |  |
| 1991/92 |  |  |  |  |  |  |  |
| 1992/93 |  |  |  |  |  |  |  |
| 1993/94 |  |  |  |  |  |  |  |
| 1994/95 |  |  |  |  |  |  |  |
| 1995/96 |  |  |  |  |  |  |  |
| 1996/97 |  |  |  |  |  |  |  |
| 1997/98 |  |  |  |  |  |  |  |
| 1998/99 |  |  |  |  |  |  |  |
| 1999/00 |  |  |  |  |  |  |  |
| 2000/01 |  |  |  |  |  |  |  |
| 2001/02 |  |  |  |  |  |  |  |
| 2002/03 |  |  |  |  |  |  |  |
| 2003/04 |  |  |  |  |  |  |  |
| 2004/05 |  |  |  |  |  |  |  |
| 2005/06 |  |  |  |  |  |  |  |
| 2007/08 |  |  |  |  |  |  |  |
| 2008/09 |  |  |  |  |  |  |  |
| 2009/10 |  |  |  |  |  |  |  |
| 2010/11 |  |  |  |  |  |  |  |
| $\begin{aligned} & 2011 / 12 \\ & 2012 / 13 \end{aligned}$ |  |  |  |  |  |  |  |
| 2013/14 |  |  |  |  |  |  |  |
| 2014/15 |  |  |  |  |  |  |  |

a. A time series of retained and total catch, groundfish fishery discard mortality, and pot fishery effort (Table 1 for EAG and Table 15 for WAG). The estimation methods are described in Appendix B.
b. Time series of pot fishery and observer nominal retained and total CPUE, observer sample size, estimated observer CPUE index (Table 2 for EAG and Table 16 for WAG), and estimated commercial fishery CPUE index (Table 3 for EAG and Table 17 for WAG). The estimation methods, CPUE fits and diagnostic plots are given in Appendix B.
c. Information on length compositions (Figures 6 to 8 for length compositions and 9 for mean lengths for EAG; and 25 to 27 for length compositions and 28 for mean lengths for WAG).
d. Survey biomass estimates are not available for the area because no systematic surveys, covering the entire fishing area, have occurred.
f. Other time series data: None.
3. Length-weight relationship: $\mathrm{W}=\mathrm{al}^{\mathrm{b}}$ where $\mathrm{a}=2.988 * 10^{-4}, \mathrm{~b}=3.135$.
4. Information on any data sources available, but excluded from the assessment: None.

## E. Analytic Approach

## 1. History of modeling approaches for this stock

The model is under development, and yet to be accepted for OFL and ABC setting. The main stumbling block for model acceptance is the scaling of biomass which appears to be low, especially for WAG. In the September 2015 meeting, CPT proposed a number of ways to improving the model fit and scaling biomass: (a) estimate the initial abundance in 1985/86 by equilibrium condition; (b) determine $M$ in the model; and (c) consider dome shaped total selectivity. We considered all these suggestions in this modeling.

## 2. Model Description

a. The underlying population dynamics model is male-only and length-based (Appendix A). This model combines commercial retained catch, total catch, groundfish (trawl) fishery discarded catch, standardized observer legal size catch-per-unit-effort (CPUE) indices, fishery retained catch size composition, total catch size composition, groundfish discard catch size composition, and tag recaptures by release-recapture length to estimate stock assessment parameters. To include a long time series of CPUE indices for stock abundance contrast, we also considered the 1985/86-1998/99 legal size standardized CPUE indices as a separate likelihood component in scenario 3.

We fitted the observer and commercial fishery CPUE indices with GLM estimated standard errors and an additional constant variance, the latter was estimated by the model fit.

There were significant changes in fishing practice due to changes in management regulations (e.g., since 1996/97 constant TAC and since 2005/06 crab rationalization), pot configuration (escape web on the pot door increased to 9 -inch since 1999), and improved observer recording in Aleutian Islands golden king crab fisheries since 1998. These changes prompted us to consider two sets of catchability and total selectivity parameters with only one set of retention parameters for the periods 1985/86-2004/05 and 2005/06-2014/15. However, in order to respond to one of the September 2015 CPT comments, we considered three catchabilities, three sets of total selectivity, and one set of retention curve in one scenario (scenario 7).

The data series used in the current assessment for EAG ranges from 1985/86 to $2014 / 15$ for retained catch biomass and size composition; 1995/96 to 2014/15 for standardized legal size crab observer CPUE index; 1989/90 to 2014/15 for groundfish fishery male bycatch biomass and size composition; 1985/86 to 1998/99 for standardized crab fish ticket CPUE index; 1990/91 to

2014/15 for total catch biomass and total catch length composition; and 1991, 1997, 2000, 2003, and 2006 releases and up to 2012 recapture time period for tagging information.

The data series used for the WAG ranges are the same as those for EAG.
b. Software: AD Model Builder (Fournier et al. 2012).
c. -f . Details are given in Appendix A.
g. Critical assumptions and consequences of assumption failures: Because of the lack of an annual stock survey we relied heavily on standardized CPUE indices (Appendix B) and catch and size composition information to determine the stock abundance trends in both regions. We assumed that the observer and fish ticket CPUE indices are proportional to crab abundances. We kept $M$ constant at $0.23 \mathrm{yr}^{-1}$ (the mean value from EAG and WAG $M$ estimates), assumed directed pot fishery discard mortality proportion at 0.20 $\mathrm{yr}^{-1}$, assumed overall groundfish fishery mortality proportion at $0.65 \mathrm{yr}^{-1}$ [mean of groundfish pot fishery mortality $\left(0.5 \mathrm{yr}^{-1}\right)$ and groundfish trawl fishery mortality $\left(0.8 \mathrm{yr}^{-1}\right)$ ], groundfish fishery selectivity at full selection for all length classes (selectivity $=1.0$ ), and discard of legal size males in the directed pot fishery was not considered. These fixed values invariably reduced the number of model parameters to be estimated and helped in convergence. We assumed different $q$ 's (scaling parameter for standardized CPUE in the model) and logistic selectivity patterns (Equation A.10) for different periods for the pot fishery. We also assumed a dome shaped selectivity (Equation A.11) (see the scenarios Table $D$ in the subsequent section).
h. Changes to any of the above since the previous assessment: Does not apply for this assessment since the model has not yet been approved.
i. Model code has been checked and validated. The code is available from the authors.

## 3. Model Selection and Evaluation

a. Description of alternative model configurations:

We considered 16 scenarios for EAG and WAG. We presented OFL and ABC results for preferred ten scenarios in the executive summary tables. We considered scenario 1 as the base scenario. It considers:
i) Estimating initial abundance by the equilibrium condition;
ii) Two catchability and two sets of logistic total selectivity for the periods 1985/86-2004/05 and 2005/06-2014/15, and a single set of logistic retention curve parameters;
iii) Full selectivity (selectivity $=1.0$ ) for groundfish (trawl) bycatch;
iv) Stock dynamics $M=0.23 \mathrm{yr}^{-1}$, pot fishery handling mortality $=0.2 \mathrm{yr}^{-1}$; and ground fish bycatch handling mortality for trawl $=0.8 \mathrm{yr}^{-1}$ and for fish pot $=0.5 \mathrm{yr}^{-1}$;
v) Calculating size transition matrix using tagging data by the normal probability function with the logistic molt probability sub-model. The tagrecaptures were treated as Bernoulli trials (i.e., stage-1 weighting); and
vi) Rescaling initial length composition sample sizes using Equation A. 21 with a set of maximum effective sample sizes (retained catch $=200$, total catch $=150$, and groundfish (trawl) discarded catch $=25$ ).
vii) The salient features and variations from the base scenario of all other scenarios are listed in Table D. The stage- 1 weighting refers to initial weighting of effective sample sizes and stage- 2 weighting refers to iterative reweighting of effective sample sizes. The detail weights with coefficient of variations (CVs) assigned to each type of data are listed in Table A2.

Table D. Features of model scenarios.

| Scenario | Initial Condition estimation | Size- composition weighting | $\begin{gathered} \hline \text { Catchability } \\ \text { and total } \\ \text { selectivity sets } \\ \hline \end{gathered}$ | Total selectivity type | CPUE data type | Treatment of trawl/total size composition and catch data | Natural mortality $\left(M y r^{-1}\right)$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | Equilibrium | Stage-1 | 2 | logistic | Observer | Trawl bycatch size-composition data included | 0.23 |
| 2 | Exponential | Stage-1 | 2 | logistic | Observer | Trawl bycatch size-composition data included | 0.23 |
| 3 | Equilibrium | Stage-1 | 2 | logistic | Observer and fish ticket | Trawl bycatch size-composition data included | 0.23 |
| 4 | Equilibrium | Stage-1 | 2 | logistic | Observer | Trawl bycatch size-composition data down weighted (0.5) | 0.23 |
| 5 | Equilibrium | Stage-1 | 2 | logistic | Observer | Trawl bycatch size-composition data not included | 0.23 |
| 6 | Exponential | Stage-1 | 2 | logistic | Observer | Trawl bycatch size-composition data not included | 0.23 |
| 7 | Equilibrium | Stage-1 | 3 | logistic | Observer | Trawl bycatch size-composition data included | 0.23 |
| 8 | Equilibrium | Stage-1 | 2 | logistic | Observer | Trawl bycatch size-composition data included | 0.18 |
| 9 | Equilibrium | Stage-2 | 2 | logistic | Observer | Trawl bycatch size-composition data included | 0.23 |
| 10 | Equilibrium | Stage-1 | 2 | dome shaped | Observer | Trawl bycatch size-composition data included | 0.23 |
| 11 | Equilibrium | Stage-1 | 2 | dome shaped | Observer | Trawl bycatch size-composition data included | 0.18 |
| 12 | Equilibrium | Stage-1 | 2 | logistic | Observer and fishing area | Trawl bycatch size-composition data included | 0.23 |
| 13 | Equilibrium | Stage-1 | 2 | logistic | Observer and fishing area | Trawl bycatch size-composition data included | 0.18 |
| 14 | Equilibrium | Stage-1 | 2 | dome shaped | Observer and fishing area | Trawl bycatch size-composition data included | 0.23 |
| 15 | Equilibrium | Stage-1 | 2 | dome shaped | Observer and fishing area | Trawl bycatch size-composition data included | 0.18 |
| 16 | Equilibrium | Stage-1 | 2 | logistic | Observer | Total size composition and catch data for 1990/911995/96 (EAG) or -1994/95 (WAG) not included | 0.23 |

Note: proportion of fishing area by year used in scenarios 12 to 15 are provided in Tables 5 and 19 for EAG and WAG, respectively.
viii) The entire time period, 1985/86-2014/15, was used to determine the mean MMB as $M M B_{\text {ref }}$ (a proxy for $M M B_{M S Y}$ ) for $\mathrm{MMB}_{\text {current }} / M M B_{\text {ref }}$ estimation under Tier 4 for all scenarios.
b. Progression of results: Model was not previously used, so, not applicable.
c. Model has not yet been approved. So labeling the previous year approved model as model 0 is not applicable.
d. Evidence of search for balance between realistic and simpler models: Unlike annually surveyed stocks, Aleutian Islands golden king crab stock biomass is difficult to track and few biological parameters are assumed based on knowledge from red king crab (e.g., handling mortality rate of $0.2 \mathrm{yr}^{-1}$ ) due to a lack of species/stock specific information. We fixed a number of model parameters after initially running the model with free parameters to reduce the number of parameters to be estimated (e.g., groundfish bycatch selectivity parameters were fixed). The 16 scenarios also considered different configuration of parameters to select parsimonious models. The detailed results of the preferred ten scenarios are provided in tables and figures. The total and retained catch OFLs, maximum total pot fishing mortality during 1985/862014/15, and the reduction in terminal (2015) and 1985 MMB from the initial condition (i.e., virgin MMB in 1960 for equilibrium initial abundance or 1985 initial abundance for exponential initial condition) for the entire 16 scenarios for EAG and WAG are provided in Table 29. The reduction in terminal MMB from the initial condition is higher for $M=0.18 \mathrm{yr}^{-1}$ than $0.23 \mathrm{yr}^{-1}$. The exponential formulation of initial condition (i.e., 1985 abundance estimate) produced lower reduction in terminal MMB for WAG, but exceeded the initial MMB for EAG. This is not surprising because 1985 initial abundance is not the virgin stock abundance. The reduction in 1986 MMB from equilibrium initial MMB is higher for EAG (reduced to 0.43 to 0.61 of initial MMB) than that for WAG (reduced to 0.56 to 0.78 of initial MMB).
e. Convergence status and criteria: ADMB default convergence criteria were used.
f. Table of the sample sizes assumed for the size compositional data:

We estimated the initial input effective sample sizes (i.e., stage-1) from the original number of length measurements using Equation A. 21 for all scenarios except scenario 9. For scenario 9 (iterative reweighting), we estimated the stage-1 sample sizes using

Equation A. 21 without setting limits to maximum sample sizes. We estimated the stage-2 effective sample sizes from estimated input effective sample sizes (i.e., stage1 sample sizes) using Francis' (2011) mean length based method (i.e., Francis TA1.8 method, Punt in press) as follows:

Observed mean length for year $t$,

$$
\begin{equation*}
\overline{l_{t}}=\sum_{i=1}^{n} l_{t, i} \times P_{t, i} \tag{1}
\end{equation*}
$$

Predicted mean length for year $t$,

$$
\begin{equation*}
\hat{\bar{l}_{t}}=\sum_{i=1}^{n} l_{t, i} \times \hat{P}_{t, i} \tag{2}
\end{equation*}
$$

Variance of the predicted mean length in year $t$,

$$
\begin{equation*}
\operatorname{var}\left(\hat{\bar{l}}_{t}\right)=\frac{\sum_{i=1}^{n} \hat{P}_{t, i}\left(l_{t, i}-\hat{\bar{l}}_{t}\right)^{2}}{s_{t}} \tag{3}
\end{equation*}
$$

Francis' reweighting parameter W,

$$
\begin{equation*}
W=\frac{1}{\operatorname{var}\left\{\frac{\hat{\bar{l}}_{t}-\hat{l}_{t}}{\left.\sqrt{\operatorname{var}\left(\hat{l}_{t}\right)}\right\}}\right\}} \tag{4}
\end{equation*}
$$

where $\widehat{P}_{t, i}$ and $P_{t, i}$ are the estimated and observed proportions of the catch during year t in size-class $\mathrm{i}, \hat{\bar{l}}_{t}$ and $\overline{l_{t}}$ are predicted and observed mean length of the catch during year t , and W is the reweighting multiplier of stage- 1 sample sizes.

We provide the initial input sample sizes and stage-2 effective sample sizes in Tables 4 and 18 for EAG and WAG, respectively. We multiplied the input sample sizes by the estimated W for a number of iterative fittings until the estimated W approached 1.0 .
g. Provide the basis for data weighting, including whether the input effective sample sizes are tuned and the survey CV adjusted: Described previously (f).
h. Do parameter estimates make sense? The estimated parameter values are within the bounds and various plots suggest that the parameter values are reasonable for a fixed $M$ of $0.23 \mathrm{yr}^{-1}$ for these stocks.
i. Model selection criteria: We used a number of diagnostic criteria to select the base model over alternative models: CPUE fits, observed vs. predicted tag recapture numbers by time at large and release size, observed and predicted mean lengths by time at large and release length class (for EAG), and catch and bycatch fits. Figures are provided for the preferred scenarios in the Results section.
j. Residual analysis: We illustrated residual fits by bubble plots for size composition predictions in various figures in the Results section.
k. Model evaluation: Only one model with a number of scenarios is presented and the evaluations are presented in the Results section below.

## 4. Results

1. List of effective sample sizes and weighting factors:

The input effective sample sizes are listed in Tables 4 and 18 and weights for different data sets are provided in Table A2 for various scenarios respectively for EAG and WAG. These weights (with the corresponding coefficient of variations) adequately fitted the length compositions and no further changes were examined.

We used weighting factors for catch biomass, recruitment deviation, pot fishery F , and groundfish fishery F. We set the retained catch biomass weight to a large value (500) because retained catches are more reliable than any other data sets. We scaled the total catch biomass weight in accordance with the observer annual sample sizes with a maximum of 300 . The total catches were derived from observer total CPUE and effort. In some years, observer sample sizes were low (Tables 2 and 16). We chose a small ground fish bycatch weight (0.2) based on the September CPT suggestion to lower its weight. We used the best fit criteria to choose the lower weight for the groundfish bycatch. Groundfish bycatch in the golden king crab fisheries is very minor. We set the CPUE weights to 1 for all scenarios. We included a constant (model estimated) variance in addition to input CPUE variance for the CPUE fit. We used the Burnham et al. (1987) suggested formula for $\ln (\mathrm{CPUE})$ [and $\ln (\mathrm{MMB})$ ] variance estimation (Equation A.16). However, the estimated additional variance values were small for observer CPUE indices, but relatively large for the fish ticket CPUE indices for EAG. But both values are small for WAG (scenario 3). Nevertheless, the CPUE index variances estimated from the
negative binomial and lognormal GLMs were adequate to fit the model confirmed by the fit diagnostics (Fox and Weisberg 2011). Parameter estimates are provided in Tables 6 to 8 for EAG and 20 to 22 for WAG for eleven scenarios. The numbers of estimable parameters are listed in Table A1 of Appendix A. The weights with the corresponding coefficient of variations specifications are detailed in Tables A2 of Appendix A for EAG and WAG.
2. Include tables showing differences in likelihood: Tables 14 and 28 list the total and component negative log likelihood values and their differences between scenarios of similar sample sizes for EAG and WAG, respectively.
3. Tables of estimates:
a. The parameter estimates with one standard deviation for eleven scenarios which included ten preferred scenarios are summarized respectively in Tables 6 to 8 for EAG and 20 to 22 for WAG. We have also provided the boundaries for parameter searches in those tables, and the estimates were within the bounds.
b. All scenarios considered molt probability parameters in addition to the linear growth increment and normal growth variability parameters to determine the size transition matrix.
c. The mature male and legal male abundance time series for arbitrarily selected five scenarios ( $1,5,10,14$, and 16) among the ten scenarios are summarized in Tables 9 to 13 for EAG and Tables 23 to 27 for WAG.
d. The recruitment estimates for those five scenarios are summarized in Tables 9 to 13 for EAG and Tables 23 to 27 for WAG.
e. The likelihood component values and the total likelihood values for preferred ten scenarios with scenario 9 (iteratively reweighting the effective sample sizes) are summarized in Table 14 for EAG and Table 28 for WAG. Scenarios 10 and 9 have the minima among the total negative log likelihoods for models with base data for EAG and WAG, respectively.
4. Graphs of estimates:
a. Total selectivity and retention curves of the pre- and post-rationalization periods for eight of the preferred ten scenarios are illustrated in Figure 10 for EAG and Figure 29 for WAG. Total selectivity for the pre-rationalization period was used
in the tagging model. The groundfish bycatch selectivity appeared flat in the preliminary analysis, indicating that all size groups were vulnerable to the gear. This is also shown in the size compositions of groundfish bycatch (Figures 8 and 27 for EAG and WAG, respectively). Thus, we set the groundfish bycatch selectivity to 1.0 for all length-classes in the subsequent analysis.
b. The mature male biomass time series for eleven scenarios (preferred 10 plus scenario 9) are depicted in Figures 20 and 37 for EAG and WAG, respectively. Mature male biomass tracked the CPUE trends well for all scenarios for EAG and WAG. The biomass variance was estimated using Burnham et al. (1987) suggested formula (Equation A. 16 in Appendix A). We determined the mature male biomass values on 15 February and considered the entire time series (1985/86-2014/15) for $M M B_{r e f}$ calculation for Tier 4 approach.
c. The full selection pot fishery F over time for eleven scenarios (preferred 10 plus scenario 9) are shown in Figures 21 and 38 for EAG and WAG, respectively. The F peaked in late 1980s and 1990s and systematically declined in the EAG and generally declined in the WAG in subsequent years, but with a slightly increasing trend in the WAG in the recent years.
d. F vs. MMB: We did not provide this figure because the model has not yet been approved.
e. Stock-Recruitment relationship: None.
f. The temporal changes in total number of recruits to the modeled population for eleven scenarios (preferred 10 plus scenario 9) are illustrated in Figure 18 for EAG and in Figure 35 for WAG. The recruitment distribution to the model size group (101-185 mm CL) is shown in Figures 19 and 36 for EAG and WAG, respectively for the eleven scenarios.
5. Evaluation of the fit to the data:
g. Fits to catches: The fishery retained, total, and groundfish bycatch (observed vs. estimated) plots for eleven scenarios are illustrated in Figures 22 and 39 for EAG and WAG, respectively. All predicted fits were very closer to observed values, especially for retained catch and groundfish bycatch mortality. However, pre 1995 total catch data did not fit well.
h. Survey data plot: We did not consider the pot survey data for the analysis.
i. CPUE index data: The predicted vs. input CPUE indices for eleven scenarios are shown in Figure 17 for EAG and Figure 34 for WAG. Scenario 3 tracks indices back to 1985/86. All scenarios appear to fit the CPUE indices satisfactorily. The CPUE variance was estimated using Burnham et al. (1987) suggested formula (Equation A. 16 in Appendix A).
j. Tagging data: The predicted vs. observed tag recaptures by length-class for years 1 to 6 recaptures are depicted in Figure 14 for EAG and Figure 33 for WAG. The predictions appear reasonable. Observed and predicted mean lengths of recaptures vs. release length for different periods of recaptures for EAG tagging data are tracking reasonably well (Figure 15). Note that we used the EAG tagging information for size transition matrix estimation for both stocks (EAG and WAG). The size transition matrices estimated using EAG tagging data in the EAG and WAG models were similar. For illustrative purpose, the estimated size transition matrix elements for the base scenario (scenario 1) are compared between EAG and WAG. The matrix elements appear very similar (Figure 16).
k. Molt probability: The predicted molt probabilities vs. CL for the eleven scenarios are depicted in Figures 24 and 41 for EAG and WAG, respectively. The fits appear to be satisfactory.

1. Fit to catch size compositions: Retained, total, and groundfish discard length compositions are shown in Figures 6 to 8 for EAG and 25 to 27 for WAG. The corresponding observed vs. predicted mean lengths with $95 \%$ confidence intervals are depicted in Figures 9 and 28 for EAG and WAG, respectively. The retained and total catch size composition fits appear satisfactory. The predicted mean lengths track the observed mean lengths well; however, the $95 \%$ confidence intervals for retained catch and total catch mean lengths appear narrower than that for the groundfish (trawl) mean length; furthermore, pre-1995 confidence intervals for all categories are wider than that of post-1995.
We illustrate the standardized residual plots as bubble plots of size composition over time for retained catch (Figures 11 and 30 for EAG and WAG, respectively), for total catch (Figures 12 and 31 for EAG and WAG, respectively), and for
groundfish discard catch (Figures 13 and 32 for EAG and WAG, respectively) for four selected scenarios $(1,5,10,14)$. The retained catch bubble plots appear random for the selected four scenarios.
m . Marginal distributions for the fits to the composition data: We did not provide this plot in this report.
n. Plots of implied versus input effective sample sizes and time series of implied effective sample sizes: We did not provide the plots, but provided the estimated values in Tables 4 and 18 for EAG and WAG, respectively.
o. Tables of RMSEs for the indices: We did not provide this table in this report.
p. Quantile-quantile (q-q) plots: We did not provide this plot in this report.
2. Retrospective and historical analysis: The retrospective fits for the ten scenarios are shown in Figure 23 for EAG and in Figure 40 for WAG. The retrospective patterns did not show severe departure when terminal year's data were removed systematically and hence the current formulation of the model appears stable.
3. Uncertainty and sensitivity analysis:
a. The main task was to determine a plausible size transition matrix to project the population over time. In a previous study, we investigated the sensitivity of the model to determine the size transition matrix by using or not using a molt probability function (Siddeek et al. 2016). The model fit is better when the molt probability model is included. Therefore, we included a molt probability submodel for the size transition matrix calculation in all scenarios.
b. We also determined likelihood values at different $M$ values and plotted component negative likelihood against $M$. It appears that the trends in negative log likelihood of retained length frequency, recruitment deviation, and CPUE were similar to that of the total for changes in $M$ for EAG (Figure 2). For WAG, the trends in negative likelihood for retained length composition and recruitment deviation were similar to that of the total for changes in $M$, but the minimum negative likelihood of CPUE was attained at a lower $M$ value (Figure 2).
4. Conduct 'jitter analysis': We did not conduct the (random) jitter analysis on model parameters. However, we performed a system search on input values within the
bound only for the $M$ parameter in the "Response to September 2015 CPT comments" section on page 8 . The $M$ estimates were robust to different starting $M$ values.

## F. Calculation of the OFL

Specification of the Tier level:
The Aleutian Islands golden king crab stocks are currently managed under Tier 5 (average catch OFL) control rule. Our analysis attempts to upgrade this stock to the Tier 4 level. The Tier 4 calculation procedure is described below:

1. List of parameters and stock size required by the control rule are:

An average mature male biomass $(M M B)$ for a specified time period, $M M B_{\text {ref }}$ (a proxy for $M M B_{M S Y}$ ) current $M M B$; an $M$ value; and a $\gamma$ value.
2. Specification of the total catch OFL:
(a) if $M M B_{\text {current }} \geq M M B_{\text {ref }}, F_{O F L}=\gamma M$;
(b) if $M M B_{\text {current }}<M M B_{\text {ref }}$ and $M M B_{\text {current }}>0.25 M M B_{\text {ref }}$,
$F_{O F L}=\gamma M \frac{\left(\frac{M M B_{\text {current }}}{M M B_{r e f}}=\alpha\right)}{(1-\alpha)}$
(c) if $M M B_{\text {current }} \leq 0.25 M M B_{\text {ref }}, F_{\text {OIFL }}=0$,
where $\mathrm{MMB}_{\text {current }}$ is the mature male biomass in the current year, $M M B_{\text {ref }}$ is average mature male biomass, and $\gamma$ is a multiplying factor of $M$.

The OFL is estimated by an iterative procedure accounting for intervening total removals (see Appendix A for the formulas).

For the selection of $M M B_{\text {ref }}$, we chose the period from 1986 to 2015 . This resulted in a MMBref range of 6.231 to 7.157 thousand metric tons for EAG and 4.756 to 8.781 thousand metric tons for WAG for the preferred ten scenarios. The current MMB (in 2015) range was 8.090 to 9.712 thousand metric tons for EAG and 4.967 to 9.642
thousand metric tons for WAG for the ten scenarios, resulting in an $F_{\text {OFL }}$ of 0.23 for EAG and slightly less for scenario 2 for WAG. The total OFL for EAG ranged from 1.360 to 1.721 thousand metric tons and for WAG from 0.681 to 1.274 thousand metric tons for the ten scenarios. The $\gamma$ value was set to 1.0 and an $M$ value of 0.23 was used for OFL calculation (see tables in the Executive Summary).
3. Specification of the retained catch portion of the total catch OFL:

We applied the $F_{O F L}$ with the retention curve to calculate the retained catch portion of the total catch OFL. The retained catch OFLs for EAG ranged from 1.327 to 1.679 thousand metric tons and that for WAG ranged from 0.638 to 1.202 thousand metric tons for the ten scenarios.

Recommendation for $F_{O F L}$, OFL total catch, and the retained catch portion of the OFL for coming year:
We recommend them for scenarios 1,10 , and 16 , respectively.
Scenario 1:
EAG: $F_{\text {OFL }}=0.23$; OFL total catch $=1.622$ thousand metric tons; retained catch portion of the $\mathrm{OFL}=1.583$ thousand metric tons.

WAG: $F_{O F L}=0.23$; OFL total catch $=0.926$ thousand metric tons; retained catch portion of the OFL $=0.871$ thousand metric tons.
Scenario 10:
EAG: $F_{\text {OFL }}=0.23$; OFL total catch $=1.721$ thousand metric tons; retained catch portion of the OFL $=1.677$ thousand metric tons.
WAG: $F_{O F L}=0.23$; OFL total catch $=1.158$ thousand metric tons; retained catch portion of the $\mathrm{OFL}=1.089$ thousand metric tons.
Scenario 16:
EAG: $F_{\text {OFL }}=0.23$; OFL total catch $=1.603$ thousand metric tons; retained catch portion of the $\mathrm{OFL}=1.565$ thousand metric tons.

WAG: $F_{O F L}=0.23$; OFL total catch $=0.913$ thousand metric tons; retained catch portion of the $\mathrm{OFL}=0.861$ thousand metric tons.

## G. Calculation of the ABC

Specification of the probability distribution of the total catch OFL:

We estimated the cumulative probability distribution of OFL assuming a $\log$ normal distribution of OFL. We calculated the OFL at the 0.5 probability and the ABC at the 0.49 probability. The ABC estimates varied for different scenarios.

Under Tier 4 approach, the ABC estimates calculated at 0.49 probability ranged 1.352 to 1.713 thousand metric tons for EAG and 0.678 to 1.268 thousand metric tons for WAG for the ten scenarios.

## H. Rebuilding Analysis

Not applicable.

## I. Data Gaps and Research Priorities

1. The recruit abundances were estimated from commercial catch sampling data. The implicit assumption in the analysis was that the estimated recruits come solely from the same exploited stock through growth and mortality. The current analysis did not consider the possibility that additional recruitment may occur through immigration from neighboring areas and possibly separate sub-stocks. Extensive tagging experiments or resource surveys are needed to investigate stock distributions.
2. We estimated $M$ in the model. However, an independent estimate of $M$ is needed for comparison. Tagging is one possibility.
3. An extensive tagging study will also provide independent estimates of molting probability and growth. We used the historical tagging data to determine the size transition matrix.
4. An arbitrary $20 \%$ handling mortality rate on discarded males was used, which was obtained from the red king crab literature (Kruse et al. 2000; Siddeek 2002). An experimentally-based independent estimate of handling mortality is needed for golden king crab.
5. The Aleutian king crab research foundation has recently initiated crab survey programs in the Aleutian Islands. This program needs to be strengthened and
continued for golden king crab research to address some of the data gap and expand the data sources.

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Table 1. Time series of annual retained catch (number and weight of crabs), estimated total male catch (number and weight of crabs on the deck), pot fishery effort (number of pot lifts), and estimated groundfish fishery discard mortality (number and weight of crabs) (handling mortality rates of $50 \%$ for pot and $80 \%$ for trawl gear were applied, only to the male portion) for the EAG golden king crab stock. The crab numbers are for the size range $101-185+\mathrm{mm}$ CL. 1985 refers to the 1985/86 fishing year. NA: no observer sampling to compute total catch. The directed fishery data included cost-recovery beginning in 2013/14.

| Year | Retained <br> Catch <br> (no.) | Retained <br> Catch <br> Biomass <br> (t) | Total <br> Catch <br> (no.) | Total <br> Catch <br> Biomass <br> (t) | Pot Fishery <br> Effort (no. <br> pot lifts) | Groundfish <br> Discard <br> Mortality(no.) | Groundfish <br> Discard <br> Mortality (t) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | $1,251,267$ | 2,695 |  |  | 117,718 |  |  |
| 1986 | $1,374,943$ | 2,818 |  |  | 155,240 |  |  |
| 1987 | 968,614 | 1,893 |  |  | 146,501 |  |  |
| 1988 | $1,156,046$ | 2,397 |  |  | 155,518 |  |  |
| 1989 | $1,419,777$ | 2,753 | $1,48,518$ | 2,738 | 106,262 | 388 | 0.61 |
| 1990 | 892,699 | 1,632 | $1,148,518$ | 1,190 | 1.98 |  |  |
| 1991 | $1,083,243$ | 2,018 | $4,385,096$ | 5,910 | 133,428 | 0 | 0.00 |
| 1992 | $1,127,291$ | 2,115 | $4,331,508$ | 5,589 | 133,778 | 779 | 1.01 |
| 1993 | 767,918 | 1,415 | NA | NA | 106,890 | 719 | 0.95 |
| 1994 | $1,086,560$ | 2,029 | $1,712,658$ | 3,257 | 191,455 | 311 | 0.29 |
| 1995 | $1,150,168$ | 2,211 | $2,742,782$ | 3,742 | 177,773 | 569 | 0.78 |
| 1996 | 848,045 | 1,615 | $1,452,362$ | 2,064 | 113,460 | 46 | 0.04 |
| 1997 | 780,481 | 1,474 | $1,788,351$ | 2,555 | 106,403 | 76 | 0.10 |
| 1998 | 740,011 | 1,407 | $2,011,777$ | 2,804 | 83,378 | 587 | 0.76 |
| 1999 | 709,332 | 1,229 | $1,556,398$ | 2,287 | 79,129 | 284 | 0.35 |
| 2000 | 704,363 | 1,352 | $1,706,999$ | 2,564 | 71,551 | 387 | 0.47 |
| 2001 | 730,030 | 1,394 | $1,352,904$ | 2,105 | 62,639 | 934 | 1.47 |
| 2002 | 643,668 | 1,236 | $1,119,586$ | 1,808 | 52,042 | 707 | 0.68 |
| 2003 | 643,074 | 1,287 | $1,111,206$ | 1,825 | 58,883 | 392 | 0.43 |
| 2004 | 637,536 | 1,261 | 965,443 | 1,627 | 34,848 | 59 | 0.12 |
| 2005 | 623,971 | 1,262 | 927,444 | 1,724 | 24,569 | 252 | 0.28 |
| 2006 | 650,587 | 1,375 | 860,688 | 1,632 | 26,195 | 679 | 0.70 |
| 2007 | 633,253 | 1,316 | 911,185 | 1,802 | 22,653 | 697 | 0.69 |
| 2008 | 666,947 | 1,406 | 929,694 | 1,799 | 24,466 | 808 | 0.85 |
| 2009 | 679,886 | 1,433 | 936,938 | 1,761 | 26,298 | 718 | 1.14 |
| 2010 | 670,698 | 1,398 | 935,574 | 1,729 | 25,851 | 2,415 | 2.41 |
| 2011 | 668,828 | 1,428 | 920,866 | 1,747 | 17,915 | 1,208 | 1.15 |
| 2012 | 687,666 | 1,482 | 990,519 | 1,939 | 20,827 | 2,058 | 3.61 |
| 2013 | 720,220 | 1,529 | 978,645 | 1,829 | 21,388 | 894 | 2.04 |
| 2014 | 719,064 | 1,536 | $1,012,683$ | 1,951 | 17,002 | 1,327 | 2.31 |
|  |  |  |  |  |  |  |  |

Table 2. Time series of nominal annual pot fishery retained, observer retained, and observer total catch-per-unit-effort (CPUE, number of crabs per pot lift), observer sample size (number of sampled pots), and GLM estimated observer CPUE Index for the EAG golden king crab stock. Observer retained CPUE includes retained and non-retained legal size crabs and 1990 refers to the 1990/91 fishing year.
$\left.\begin{array}{cccccc}\hline & \begin{array}{c}\text { Pot } \\ \text { Fishery } \\ \text { Yominal } \\ \text { Retained } \\ \text { CPUE }\end{array} & \begin{array}{c}\text { Obs. } \\ \text { Nominal } \\ \text { Retained } \\ \text { CPUE }\end{array} & \begin{array}{c}\text { Obs. } \\ \text { Nominal } \\ \text { Total } \\ \text { CPUE }\end{array} & \begin{array}{c}\text { Obs. } \\ \text { Sample } \\ \text { Size }\end{array} & \begin{array}{c}\text { Obs. } \\ \text { (no.pot } \\ \text { lifts) }\end{array} \\ \hline 1990 & 8.90 & 2.17 & 13.00 & 138 & \\ \text { CPUE } \\ \text { Index }\end{array}\right]$

Table 3. Time series of GLM estimated CPUE Indices and standard errors for the fish ticket based retained catch-per-pot lift for the EAG golden king crab stock. The GLM was fitted to the 1985/86 to 1998/99 time series of data and used in scenario 3. 1985 refers to the 1985/86 fishing year.

| Year | CPUE <br> Index | Standard <br> Error |
| :---: | :---: | :---: |
| 1985 | 1.67 | 0.08 |
| 1986 | 1.22 | 0.06 |
| 1987 | 0.96 | 0.06 |
| 1988 | 1.03 | 0.05 |
| 1989 | 1.04 | 0.04 |
| 1990 | 0.83 | 0.05 |
| 1991 | 0.84 | 0.05 |
| 1992 | 0.93 | 0.05 |
| 1993 | 0.90 | 0.06 |
| 1994 | 0.80 | 0.05 |
| 1995 | 0.77 | 0.05 |
| 1996 | 0.83 | 0.05 |
| 1997 | 1.20 | 0.06 |
| 1998 | 1.36 | 0.06 |

Table 4. The number of length measurements and stage-2 effective sample size iteratively estimated by Francis method for retained, total, and groundfish discard catches of golden king crab during 1985 to 2014 in EAG. NA: not available.

| Year | Retained <br> Length | Retained <br> Stage-2 <br> Sample <br> Sizfective (no) <br> Sample <br> Size (no) | Total <br> Length <br> Sample <br> Size <br> (no) | Total <br> Stage-2 <br> Effective <br> Sample <br> Size (no) | Groundfish <br> Length <br> Sample Size <br> (no) | Groundfish <br> Stage-2 <br> Effective <br> Sample Size <br> (no) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 2,108 | 100 |  |  |  |  |
| 1986 | 1,226 | 58 |  |  |  |  |
| 1987 | 1,304 | 62 |  |  |  |  |
| 1988 | 13,456 | 636 |  |  | 107 | 0.56 |
| 1989 | 59,054 | 2,789 |  |  | 486 | 2.54 |
| 1990 | 22,720 | 1,073 | 2,600 | 5 | NA | NA |
| 1991 | 23,343 | 1,103 | 6,654 | 12 | NA |  |
| 1992 | 20,629 | 974 | 5,469 | 9 | 9 | 0.05 |
| 1993 | 6,254 | 295 | NA | NA | 6 | 0.03 |
| 1994 | 4,666 | 220 | 1,235 | 2 | 77 | 0.40 |
| 1995 | 11,514 | 544 | 116,305 | 201 | 47 | 0.25 |
| 1996 | 5,889 | 278 | 85,021 | 147 | 19 | 0.10 |
| 1997 | 12,775 | 603 | 77,565 | 134 | 62 | 0.32 |
| 1998 | 7,960 | 376 | 83,374 | 144 | 263 | 1.38 |
| 1999 | 5,820 | 275 | 73,728 | 128 | 291 | 1.52 |
| 2000 | 4,907 | 232 | 28,334 | 49 | 1,682 | 8.80 |
| 2001 | 4,539 | 214 | 35,606 | 62 | 606 | 3.17 |
| 2002 | 4,119 | 195 | 24,536 | 42 | 945 | 4.94 |
| 2003 | 3,629 | 171 | 22,859 | 40 | 699 | 3.66 |
| 2004 | 3,123 | 148 | 19,481 | 34 | 132 | 0.69 |
| 2005 | 2,290 | 108 | 12,451 | 22 | 1,129 | 5.91 |
| 2006 | 2,629 | 124 | 9,464 | 16 | 795 | 4.16 |
| 2007 | 2,776 | 131 | 12,530 | 22 | 1,002 | 5.24 |
| 2008 | 2,542 | 120 | 15,715 | 27 | 3,515 | 18.39 |
| 2009 | 2,355 | 111 | 13,972 | 24 | 385 | 2.02 |
| 2010 | 2,353 | 111 | 15,291 | 26 | 1,253 | 6.55 |
| 2011 | 2,507 | 118 | 18,994 | 33 | 1,686 | 8.82 |
| 2012 | 2,926 | 138 | 20,648 | 36 | 149 | 0.78 |
| 2013 | 2,605 | 123 | 22,819 | 39 | 52 | 0.26 |
| 2014 | 2,075 | 98 | 22,365 | 39 | 54 | 0.28 |

Table 5. Total fished area (square miles) and relative proportion of area fished (relative to the maximum area fished) during 1985/86-2014/15 in EAG.

| Year | Fished Area <br> (sq.miles) | Relative <br> Proportion | Year | Fished Area <br> (sq.miles) | Relative <br> Proportion |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 257,180 | 0.57 | 2005 | 134,060 | 0.30 |
| 1986 | 281,826 | 0.63 | 2006 | 140,498 | 0.31 |
| 1987 | 244,721 | 0.54 | 2007 | 164,523 | 0.37 |
| 1988 | 278,705 | 0.62 | 2008 | 157,924 | 0.35 |
| 1989 | 450,286 | 1.00 | 2009 | 139,148 | 0.31 |
| 1990 | 188,461 | 0.42 | 2010 | 155,592 | 0.35 |
| 1991 | 257,978 | 0.57 | 2011 | 125,924 | 0.28 |
| 1992 | 234,338 | 0.52 | 2012 | 141,013 | 0.31 |
| 1993 | 158,490 | 0.35 | 2013 | 147,034 | 0.33 |
| 1994 | 257,087 | 0.57 | 2014 | 136,227 | 0.30 |
| 1995 | 280,213 | 0.62 |  |  |  |
| 1996 | 269,754 | 0.60 |  |  |  |
| 1997 | 263,378 | 0.58 |  |  |  |
| 1998 | 166,178 | 0.37 |  |  |  |
| 1999 | 176,031 | 0.39 |  |  |  |
| 2000 | 171,383 | 0.38 |  |  |  |
| 2001 | 144,729 | 0.32 |  |  |  |
| 2002 | 138,283 | 0.31 |  |  |  |
| 2003 | 120,167 | 0.27 |  |  |  |
| 2004 | 87,066 | 0.19 |  |  |  |

Table 6. Parameter estimates and standard deviations with the 2015 (February 15) MMB for scenarios 1, 2, 3, and 4 for the golden king crab data from the EAG, 1985/86-2014/15. Recruitment and fishing mortality deviations and initial size frequency determination parameters were omitted from this list.

|  | Scenario 1 |  | Scenario 2 |  | Scenario 3 |  | Scenario 4 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Parameter | Estimate | Std Dev | Estimate | Std Dev | Estimate | Std Dev | Estimate | Std Dev | Limits |
| $\log _{-} \omega_{1}$ ( growth incr. intercept) | 2.54 | 0.02 | 2.54 | 0.02 | 2.54 | 0.02 | 2.54 | 0.02 | 1.0, 4.5 |
| $\omega_{2}$ (growth incr. slope) | -9.30 | 1.77 | -9.20 | 1.78 | -9.37 | 1.77 | -9.59 | 1.77 | -12.0,-5.0 |
| $\log _{2}$ ( molt prob. slope) | -2.50 | 0.07 | -2.51 | 0.07 | -2.50 | 0.07 | -2.48 | 0.07 | -4.61,-1.39 |
| $\log _{-} \mathrm{b}$ (molt prob. L50) | 4.95 | 0.01 | 4.95 | 0.01 | 4.95 | 0.01 | 4.95 | 0.01 | 3.87,5.05 |
| $\sigma$ (growth variability std) | 3.68 | 0.10 | 3.68 | 0.10 | 3.67 | 0.10 | 3.67 | 0.10 | 0.1,12.0 |
| log_total sel delta $\theta$, 1985-04 | 3.38 | 0.13 | 3.37 | 0.13 | 3.37 | 0.13 | 3.38 | 0.13 | 0.,4.4 |
| $\log _{-}$total sel delta $\theta$, 2005-14 | 3.10 | 0.19 | 3.08 | 0.19 | 3.10 | 0.19 | 3.12 | 0.19 | 0.,4.4 |
| $\log _{-}$ret. sel delta $\theta$, 1985-14 | 1.86 | 0.08 | 1.85 | 0.08 | 1.86 | 0.08 | 1.86 | 0.08 | 0.,4.4 |
| log_tot sel $\theta_{50}, 1985-04$ | 4.83 | 0.02 | 4.83 | 0.02 | 4.83 | 0.02 | 4.83 | 0.02 | 4.0,5.0 |
| log_tot sel $\theta_{50}, 2005-14$ | 4.92 | 0.02 | 4.92 | 0.02 | 4.92 | 0.02 | 4.92 | 0.02 | 4.0,5.0 |
| $\log _{-}$ret. sel $\theta_{50}, 1985-14$ | 4.91 | 0.002 | 4.91 | 0.002 | 4.91 | 0.002 | 4.91 | 0.002 | 4.0,5.0 |
| $\log _{\_} \beta_{\mathrm{r}}$ (rec.distribution par.) | -0.73 | 0.25 | -0.72 | 0.25 | -0.72 | 0.25 | -0.78 | 0.28 | -12.0, 12.0 |
| $\operatorname{logq} 2$ (catchability 1985-04) | -0.63 | 0.11 | -0.64 | 0.11 | -0.65 | 0.09 | -0.67 | 0.12 | -9.0, 2.25 |
| logq3 (catchability 2005-14) | -0.96 | 0.21 | -1.01 | 0.21 | -0.97 | 0.19 | -0.96 | 0.21 | -9.0, 2.25 |
| log_newsh1 (N1985) |  |  | 2.21 | 0.07 |  |  |  |  | 0.01, 10.0 |
| log_mean_rec (mean rec.) | 0.90 | 0.05 | 0.97 | 0.06 | 0.90 | 0.05 | 0.91 | 0.05 | 0.01, 5.0 |
| log_mean_Fpot (Pot fishery F) | -1.07 | 0.10 | -0.94 | 0.11 | -1.09 | 0.09 | -1.09 | 0.10 | -15.0, -0.01 |
| log_mean_Fground (GF byc. F) | -9.51 | 0.39 | -9.53 | 0.39 | -9.51 | 0.39 | -9.51 | 0.39 | -15.0, -1.6 |
| $\sigma_{e}^{2}$ (observer CPUE additional var) | 0.02 | 0.01 | 0.02 | 0.01 | 0.02 | 0.01 | 0.02 | 0.01 | 0.0, 0.15 |
| $\sigma_{e}^{2}$ (fishery CPUE additional var) |  |  |  |  | 0.05 | 0.02 |  |  | 0.0,1.0 |
| 2015 MMB | 10,124 | 1,714 | 10,635 | 1,909 | 10,125 | 1,652 | 9,802 | 1,658 |  |

Table 7. Parameter estimates and standard deviations with the 2015 (February 15) MMB for scenarios 5, 10, 12, and 14 for the golden king crab data from the EAG, 1985/86-2014/15. Recruitment and fishing mortality deviations and initial size frequency determination parameters were omitted from this list.

|  | Scenario 5 |  | Scenario 10 |  | Scenario 12 |  | Scenario 14 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Parameter | Estimate | Std Dev | Estimate | Std Dev | Estimate | Std Dev | Estimate | Std Dev | Limits |
| $\log _{-} \omega_{1}$ ( growth incr. intercept) | 2.53 | 0.02 | 2.57 | 0.02 | 2.54 | 0.02 | 2.57 | 0.02 | 1.0, 4.5 |
| $\omega_{2}$ ( growth incr. slope) | -9.96 | 1.78 | -7.45 | 1.88 | -9.36 | 1.77 | -7.21 | 1.98 | -12.0, -5.0 |
| $\log _{-} \mathrm{a}$ (molt prob. slope) | -2.44 | 0.07 | -2.59 | 0.08 | -2.50 | 0.07 | -2.59 | 0.09 | -4.61, -1.39 |
| $\log _{\text {_ }} \mathrm{b}$ (molt prob. L50) | 4.95 | 0.01 | 4.97 | 0.01 | 4.95 | 0.01 | 4.97 | 0.01 | 3.869, 5.05 |
| $\sigma$ (growth variability std) | 3.66 | 0.10 | 3.72 | 0.10 | 3.67 | 0.10 | 3.73 | 0.11 | 0.1, 12.0 |
| d1 (incr. dome sel slope 1985-04.) |  |  | 0.07 | 0.01 |  |  | 0.07 | 0.01 | 0.01,1.0 |
| d2 (decr. dome sel slope 1985-04.) |  |  | -0.12 | 0.01 |  |  | -0.12 | 0.01 | -1.0,-0.1 |
| d3 (incr. dome sel slope 2005-14.) |  |  | 0.14 | 0.03 |  |  | 0.14 | 0.03 | 0.01,1.0 |
| d4 (decr. dome sel slope 2005-14.) |  |  | -0.11 | 0.15 |  |  | -0.15 | 0.16 | -1.0,0.01 |
| log_total sel delta $\theta$, 1985-04 | 3.36 | 0.14 |  |  | 3.37 | 0.13 |  |  | 0., 4.4 |
| $\log _{-}$total sel delta $\theta, 2005-14$ | 3.11 | 0.20 |  |  | 3.15 | 0.19 |  |  | 0., 4.4 |
| $\log _{\text {_ }}$ ret. sel delta $\theta, 1985-14$ | 1.85 | 0.08 | 1.91 | 0.08 | 1.85 | 0.08 | 1.90 | 0.08 | 0., 4.4 |
| log_tot sel $\theta_{50}, 1985-04$ | 4.82 | 0.02 | 5.92 | 72.61 | 4.83 | 0.02 | 5.29 | 0.31 | 4.0, 6.0 |
| log_tot sel $\theta_{50}, 2005-14$ | 4.92 | 0.03 | 4.93 | 0.05 | 4.93 | 0.03 | 4.93 | 0.03 | 4.0, 5.3 |
| log_tot sel $\theta_{95}, 1985-04$ |  |  | 4.97 | 0.03 |  |  | 4.97 | 0.03 | 4.9, 5.3 |
| log_tot sel $\theta_{95}, 2005-14$ |  |  | 5.17 | 0.06 |  |  | 5.17 | 0.04 | -6.0,5.3 |
| log_ret. sel $\theta_{50}, 1985-14$ | 4.91 | 0.002 | 4.92 | 0.002 | 4.91 | 0.002 | 4.92 | 0.002 | 4.0, 5.0 |
| $\log _{\_} \beta_{\mathrm{r}}$ (rec.distribution par.) | -1.00 | 0.34 | -0.63 | 0.26 | -0.73 | 0.25 | -0.62 | 0.27 | -12.0, 12.0 |
| logq2 (catchability 1985-04) | -0.73 | 0.13 | -0.70 | 0.10 | -0.84 | 0.16 | -1.00 | 0.16 | -9.0, 2.25 |
| $\operatorname{logq} 3$ (catchability 2005-14) | -0.98 | 0.22 | -1.06 | 0.18 | -1.04 | 0.22 | -1.24 | 0.19 | -9.0, 2.25 |
| $\log _{-} \eta$ (stock mixing parameter) |  |  |  |  | -0.24 | 0.14 | -0.35 | 0.15 | -9.0, 0.01 |
| log_mean_rec (mean rec.) | 0.94 | 0.05 | 0.90 | 0.06 | 0.87 | 0.05 | 0.87 | 0.06 | 0.01, 5.0 |
| log_mean_Fpot (Pot fishery F) | -1.12 | 0.11 | -1.10 | 0.09 | -1.01 | 0.10 | -1.02 | 0.09 | -15.0, -0.01 |
| log_mean_Fground (GF byc. F) | -9.53 | 0.39 | -9.56 | 0.39 | -9.45 | 0.39 | -9.49 | 0.39 | -15.0, -1.6 |
| $\sigma_{e}^{2}$ (CPUE additional var) | 0.02 | 0.01 | 0.02 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.0, 0.15 |
| 2015 MMB | 9,837 | 1,734 | 10,410 | 1,851 | 8,883 | 1,587 | 8,664 | 1,609 |  |

Table 8 Parameter estimates and standard deviations with the 2015 (February 15) MMB for scenarios 7, 9, and 16 for the golden king crab data from the EAG, 1985/86-2014/15. Recruitment and fishing mortality deviations and initial size frequency determination parameters were omitted from this list.

|  | Scenario 7 |  | Scenario 9 |  | Scenario 16 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Parameter | Estimate | Std Dev | Estimate | Std Dev | Estimate | Std Dev | Limits |
| $\log _{-} \omega_{1}$ ( growth incr. intercept) | 2.54 | 0.02 | 2.54 | 0.02 | 2.54 | 0.02 | 1.0, 4.5 |
| $\omega_{2}$ ( growth incr. slope) | -9.21 | 1.82 | -8.78 | 1.75 | -8.94 | 1.79 | -12.0, -5.0 |
| $\log _{-} \mathrm{a}$ (molt prob. slope) | -2.50 | 0.09 | -2.52 | 0.06 | -2.54 | 0.08 | -4.61, -1.39 |
| $\log _{-} \mathrm{b}$ (molt prob. L50) | 4.95 | 0.01 | 4.95 | 0.005 | 4.95 | 0.01 | 3.869, 5.05 |
| $\sigma$ (growth variability std) | 3.68 | 0.10 | 3.67 | 0.10 | 3.68 | 0.10 | 0.1, 12.0 |
| log_total sel delta $\theta$, 1985-94 | 3.18 | 0.21 |  |  |  |  | 0., 4.4 |
| log_total sel delta $\theta$, 1995-04 | 3.59 | 0.19 |  |  |  |  | 0., 4.4 |
| log_total sel delta $\theta$, 1985-04 |  |  | 3.37 | 0.10 | 3.26 | 0.12 | 0., 4.4 |
| log_total sel delta日, 2005-14 | 3.11 | 0.19 | 3.13 | 0.13 | 3.10 | 0.19 | 0., 4.4 |
| $\log _{-}$ret. sel delta $\theta, 1985-14$ | 1.87 | 0.08 | 1.87 | 0.04 | 1.89 | 0.08 | 0., 4.4 |
| log_tot sel $\theta_{50}, 1985-94$ | 4.82 | 0.03 |  |  |  |  | 4.0, 5.0 |
| log_tot sel $\theta_{50}, 1995-04$ | 4.87 | 0.04 |  |  |  |  | 4.0, 5.0 |
| log_tot sel $\theta_{50}, 1985-04$ |  |  | 4.83 | 0.01 | 4.85 | 0.02 | 4.0, 5.0 |
| log_tot sel $\theta_{50}, 2005-14$ | 4.93 | 0.02 | 4.93 | 0.01 | 4.93 | 0.02 | 4.0, 5.0 |
| log_ret. sel $\theta_{50}, 1985-14$ | 4.91 | 0.002 | 4.91 | 0.001 | 4.91 | 0.002 | 4.0, 5.0 |
| $\log \_\beta_{\mathrm{r}}$ (rec.distribution par.) | -0.66 | 0.26 | -1.00 | 0.21 | -0.72 | 0.26 | $-12.0,12.0$ |
| Logq1 (catchability 1985-94) | -0.63 | 10164.00 |  |  |  |  | $-9.0,2.25$ |
| $\operatorname{logq2}$ (catchability 1995-04) | -0.40 | 0.21 |  |  |  |  | -9.0, 2.25 |
| $\operatorname{logq2}$ (catchability 1985-04) |  |  | -0.66 | 0.09 | -0.57 | 0.11 | -9.0, 2.25 |
| logq3 (catchability 2005-14) | -0.88 | 0.21 | -0.98 | 0.16 | -0.94 | 0.21 | -9.0, 2.25 |
| log_mean_rec (mean rec.) | 0.87 | 0.05 | 0.92 | 0.05 | 0.90 | 0.05 | 0.01, 5.0 |
| log_mean_Fpot (Pot fishery F) | -0.98 | 0.12 | -1.08 | 0.08 | -1.06 | 0.09 | -15.0, -0.01 |
| $\log _{-}$mean_Fground (GF byc. F) | -9.45 | 0.39 | -9.54 | 0.39 | -9.50 | 0.39 | -15.0, -1.6 |
| $\sigma_{e}^{2}$ (CPUE additional var) | 0.02 | 0.01 | 0.02 | 0.01 | $0.02$ | $0.01$ | 0.0, 0.15 |
| $2015 \text { MMB }$ | 9,504 | 1,651 | 10,230 | 1,621 | 10,165 | 1,762 |  |

Table 9. Annual abundance estimates of model recruits (millions of crabs), legal male biomass with standard deviation $(\mathrm{t})$, and mature male biomass with standard deviation ( t ) for scenario 1 for golden king crab in the EAG. Legal male biomass was estimated on July 1 (start of fishing year) and mature male biomass for fishing year y was estimated on February 15 , fishing year $\mathrm{y}+1$ after the fishing year y fishery total catch removal. 1985 refers to the 1985/86 fishing year. Recruits estimates for 1961 to 2015 are restricted to 1985-2015.

| Year | Recruits to the Model ( $\geq \mathbf{1 0 1} \mathbf{~ m m}$ CL) | Mature Male Biomass ( $\geq 121 \mathrm{~mm} \mathrm{CL}$ ) | Standard <br> Deviation | Legal Male Biomass ( $\geq 136$ mm CL) | Standard <br> Deviation |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 1.98 |  |  | 10,126 | 723 |
| 1986 | 1.47 | 8,817 | 526 | 8,493 | 496 |
| 1987 | 2.99 | 6,855 | 374 | 6,610 | 355 |
| 1988 | 4.61 | 5,828 | 328 | 5,686 | 303 |
| 1989 | 1.52 | 5,095 | 328 | 4,949 | 289 |
| 1990 | 2.52 | 4,864 | 307 | 4,568 | 283 |
| 1991 | 4.78 | 4,784 | 324 | 4,637 | 307 |
| 1992 | 2.13 | 4,415 | 373 | 4,323 | 343 |
| 1993 | 2.05 | 5,005 | 383 | 4,732 | 348 |
| 1994 | 3.04 | 5,370 | 311 | 5,180 | 293 |
| 1995 | 1.51 | 4,836 | 282 | 4,717 | 268 |
| 1996 | 2.32 | 4,413 | 279 | 4,218 | 271 |
| 1997 | 3.20 | 4,165 | 296 | 4,049 | 286 |
| 1998 | 2.48 | 4,177 | 321 | 4,064 | 311 |
| 1999 | 3.15 | 4,713 | 377 | 4,524 | 363 |
| 2000 | 2.73 | 5,235 | 436 | 5,078 | 420 |
| 2001 | 2.08 | 5,884 | 495 | 5,689 | 477 |
| 2002 | 3.50 | 6,375 | 558 | 6,171 | 537 |
| 2003 | 2.09 | 6,669 | 610 | 6,532 | 591 |
| 2004 | 1.57 | 7,386 | 713 | 7,143 | 682 |
| 2005 | 3.31 | 7,533 | 787 | 7,319 | 756 |
| 2006 | 2.77 | 7,145 | 829 | 7,019 | 802 |
| 2007 | 2.67 | 7,490 | 929 | 7,260 | 889 |
| 2008 | 3.07 | 7,862 | 1,022 | 7,622 | 981 |
| 2009 | 2.03 | 8,094 | 1,093 | 7,878 | 1,052 |
| 2010 | 4.10 | 8,383 | 1,147 | 8,124 | 1,104 |
| 2011 | 4.21 | 8,349 | 1,178 | 8,184 | 1,143 |
| 2012 | 3.34 | 9,221 | 1,350 | 8,955 | 1,293 |
| 2013 | 1.55 | 10,337 | 1,528 | 9,999 | 1,462 |
| 2014 | 1.82 | 10,826 | 1,654 | 10,489 | 1,590 |
| 2015 | 2.99 | 10,124 | 1,714 | 9,909 | 1,666 |

Table 10. Annual abundance estimates of model recruits (millions of crabs), legal male biomass with standard deviation ( t ), and mature male biomass with standard deviation ( t ) for scenario 5 for golden king crab in the EAG. Legal male biomass was estimated on July 1 (start of fishing year) and mature male biomass for fishing year $y$ was estimated on February 15, fishing year $y+1$ after the fishing year y fishery total catch removal. 1985 refers to the 1985/86 fishing year. Recruits estimates for 1961 to 2015 are restricted to 1985-2015.

| Year | Recruits to the Model ( $\geq \mathbf{1 0 1} \mathbf{~ m m}$ CL) | $\begin{gathered} \text { Mature Male } \\ \text { Biomass } \\ (\geq \mathbf{1 2 1 ~ m m ~ C L}) \end{gathered}$ | Standard <br> Deviation | Legal Male <br> Biomass ( $\geq \mathbf{1 3 6}$ mm CL) | Standard <br> Deviation |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 1.95 |  |  | 10,333 | 738 |
| 1986 | 1.44 | 8,934 | 538 | 8,626 | 507 |
| 1987 | 3.05 | 6,914 | 383 | 6,682 | 364 |
| 1988 | 4.24 | 5,810 | 337 | 5,693 | 312 |
| 1989 | 1.77 | 5,048 | 339 | 4,925 | 299 |
| 1990 | 3.04 | 4,702 | 329 | 4,460 | 303 |
| 1991 | 4.30 | 4,689 | 350 | 4,584 | 329 |
| 1992 | 2.51 | 4,565 | 417 | 4,472 | 383 |
| 1993 | 2.10 | 5,071 | 449 | 4,853 | 408 |
| 1994 | 2.80 | 5,555 | 352 | 5,387 | 336 |
| 1995 | 1.82 | 5,049 | 328 | 4,945 | 318 |
| 1996 | 2.20 | 4,535 | 330 | 4,380 | 328 |
| 1997 | 3.36 | 4,369 | 356 | 4,262 | 351 |
| 1998 | 2.67 | 4,337 | 386 | 4,258 | 383 |
| 1999 | 3.49 | 4,968 | 466 | 4,804 | 459 |
| 2000 | 2.95 | 5,611 | 556 | 5,481 | 545 |
| 2001 | 2.25 | 6,480 | 648 | 6,299 | 637 |
| 2002 | 3.06 | 7,133 | 744 | 6,939 | 728 |
| 2003 | 2.10 | 7,437 | 802 | 7,301 | 786 |
| 2004 | 2.21 | 7,886 | 889 | 7,679 | 865 |
| 2005 | 2.79 | 7,921 | 970 | 7,741 | 942 |
| 2006 | 2.54 | 7,741 | 1,033 | 7,580 | 1,001 |
| 2007 | 2.25 | 7,842 | 1,091 | 7,637 | 1,054 |
| 2008 | 3.05 | 7,921 | 1,133 | 7,711 | 1,097 |
| 2009 | 3.61 | 7,802 | 1,165 | 7,637 | 1,127 |
| 2010 | 2.65 | 8,087 | 1,213 | 7,901 | 1,167 |
| 2011 | 3.10 | 8,819 | 1,270 | 8,559 | 1,222 |
| 2012 | 3.38 | 9,100 | 1,321 | 8,894 | 1,278 |
| 2013 | 2.80 | 9,380 | 1,394 | 9,170 | 1,346 |
| 2014 | 2.60 | 9,768 | 1,529 | 9,513 | 1,475 |
| 2015 | 2.55 | 9,837 | 1,734 | 9,597 | 1,675 |

Table 11. Annual abundance estimates of model recruits (millions of crabs), legal male biomass with standard deviation ( $t$ ), and mature male biomass with standard deviation ( t ) for scenario 10 for golden king crab in the EAG. Legal male biomass was estimated on July 1 (start of fishing year) and mature male biomass for fishing year y was estimated on February 15, fishing year y+1 after the fishing year y fishery total catch removal. 1985 refers to the 1985/86 fishing year. Recruits estimates for 1961 to 2015 are restricted to 1985-2015.

| Year | Recruits to the Model ( $\mathbf{~} \mathbf{1 0 1} \mathbf{~ m m}$ CL) | $\begin{gathered} \text { Mature Male } \\ \text { Biomass } \\ (\geq 121 \mathrm{~mm} \mathrm{CL}) \end{gathered}$ | Standard <br> Deviation | Legal Male Biomass ( $\geq \mathbf{1 3 6}$ mm CL) | Standard <br> Deviation |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 2.20 |  |  | 10,746 | 976 |
| 1986 | 1.43 | 9,584 | 798 | 9,285 | 778 |
| 1987 | 2.78 | 7,702 | 659 | 7,462 | 647 |
| 1988 | 5.03 | 6,571 | 578 | 6,441 | 563 |
| 1989 | 1.48 | 5,636 | 530 | 5,535 | 511 |
| 1990 | 2.55 | 5,415 | 517 | 5,159 | 507 |
| 1991 | 4.78 | 5,317 | 515 | 5,220 | 514 |
| 1992 | 2.12 | 4,922 | 535 | 4,883 | 525 |
| 1993 | 2.03 | 5,446 | 520 | 5,252 | 512 |
| 1994 | 3.12 | 5,780 | 449 | 5,673 | 457 |
| 1995 | 1.57 | 5,225 | 417 | 5,176 | 424 |
| 1996 | 2.33 | 4,795 | 419 | 4,663 | 426 |
| 1997 | 3.29 | 4,571 | 434 | 4,507 | 439 |
| 1998 | 2.57 | 4,584 | 459 | 4,530 | 464 |
| 1999 | 3.21 | 5,140 | 523 | 5,021 | 525 |
| 2000 | 2.84 | 5,712 | 592 | 5,629 | 594 |
| 2001 | 2.09 | 6,390 | 661 | 6,284 | 664 |
| 2002 | 3.49 | 6,921 | 733 | 6,806 | 732 |
| 2003 | 2.05 | 7,222 | 786 | 7,172 | 786 |
| 2004 | 1.58 | 7,890 | 884 | 7,744 | 873 |
| 2005 | 3.38 | 7,984 | 949 | 7,861 | 937 |
| 2006 | 2.83 | 7,580 | 987 | 7,532 | 977 |
| 2007 | 2.66 | 7,926 | 1,097 | 7,784 | 1,074 |
| 2008 | 3.07 | 8,315 | 1,201 | 8,171 | 1,178 |
| 2009 | 2.09 | 8,535 | 1,275 | 8,422 | 1,254 |
| 2010 | 4.23 | 8,800 | 1,336 | 8,650 | 1,312 |
| 2011 | 4.05 | 8,808 | 1,377 | 8,743 | 1,360 |
| 2012 | 3.13 | 9,719 | 1,556 | 9,564 | 1,521 |
| 2013 | 1.64 | 10,737 | 1,713 | 10,543 | 1,675 |
| 2014 | 1.94 | 11,083 | 1,804 | 10,903 | 1,772 |
| 2015 | 2.46 | 10,410 | 1,851 | 10,308 | 1,827 |

Table 12. Annual abundance estimates of model recruits (millions of crabs), legal male biomass with standard deviation ( t ), and mature male biomass with standard deviation ( t ) for scenario 14 for golden king crab in the EAG. Legal male biomass was estimated on July 1 (start of fishing year) and mature male biomass for fishing year y was estimated on February 15, fishing year y+1 after the fishing year y fishery total catch removal. 1985 refers to the 1985/86 fishing year. Recruits estimates for 1961 to 2015 are restricted to 1985-2015.

| Year | Recruits to the Model ( $\mathbf{~} \mathbf{1 0 1} \mathbf{~ m m}$ CL) | $\begin{gathered} \text { Mature Male } \\ \text { Biomass } \\ (\geq 121 \mathrm{~mm} \mathrm{CL}) \end{gathered}$ | Standard <br> Deviation | Legal Male Biomass ( $\geq \mathbf{1 3 6}$ mm CL) | Standard <br> Deviation |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 2.26 |  |  | 11,208 | 1,487 |
| 1986 | 1.42 | 10,059 | 1,325 | 9,777 | 1,291 |
| 1987 | 2.78 | 8,171 | 1,151 | 7,938 | 1,117 |
| 1988 | 5.17 | 6,983 | 1,005 | 6,859 | 972 |
| 1989 | 1.49 | 5,992 | 895 | 5,904 | 857 |
| 1990 | 2.57 | 5,772 | 827 | 5,530 | 796 |
| 1991 | 4.84 | 5,647 | 780 | 5,569 | 764 |
| 1992 | 2.19 | 5,231 | 761 | 5,213 | 742 |
| 1993 | 2.11 | 5,758 | 716 | 5,591 | 702 |
| 1994 | 3.15 | 6,107 | 635 | 6,029 | 638 |
| 1995 | 1.58 | 5,578 | 592 | 5,550 | 596 |
| 1996 | 2.25 | 5,149 | 578 | 5,038 | 581 |
| 1997 | 3.13 | 4,892 | 571 | 4,846 | 573 |
| 1998 | 2.42 | 4,813 | 565 | 4,778 | 567 |
| 1999 | 2.97 | 5,223 | 594 | 5,133 | 594 |
| 2000 | 2.56 | 5,640 | 635 | 5,586 | 636 |
| 2001 | 1.84 | 6,124 | 687 | 6,051 | 687 |
| 2002 | 3.07 | 6,444 | 743 | 6,364 | 742 |
| 2003 | 1.87 | 6,547 | 790 | 6,525 | 788 |
| 2004 | 1.49 | 6,974 | 870 | 6,873 | 859 |
| 2005 | 3.16 | 6,950 | 917 | 6,866 | 906 |
| 2006 | 2.44 | 6,527 | 940 | 6,512 | 929 |
| 2007 | 2.30 | 6,822 | 1,015 | 6,725 | 994 |
| 2008 | 2.78 | 7,063 | 1,101 | 6,980 | 1,079 |
| 2009 | 1.92 | 7,134 | 1,168 | 7,085 | 1,146 |
| 2010 | 3.68 | 7,312 | 1,208 | 7,226 | 1,185 |
| 2011 | 3.59 | 7,291 | 1,231 | 7,277 | 1,215 |
| 2012 | 3.02 | 8,016 | 1,379 | 7,939 | 1,347 |
| 2013 | 1.48 | 8,865 | 1,508 | 8,764 | 1,471 |
| 2014 | 1.78 | 9,244 | 1,563 | 9,138 | 1,535 |
| 2015 | 2.38 | 8,664 | 1,609 | 8,628 | 1,589 |

Table 13. Annual abundance estimates of model recruits (millions of crabs), legal male biomass with standard deviation ( $t$ ), and mature male biomass with standard deviation (t) for scenario 16 for golden king crab in the EAG. Legal male biomass was estimated on July 1 (start of fishing year) and mature male biomass for fishing year y was estimated on February 15, fishing year y+1 after the fishing year y fishery total catch removal. 1985 refers to the $1985 / 86$ fishing year. Recruits estimates for 1961 to 2015 are restricted to 1985-2015.

|  | Recruits to the <br> Model ( $\mathbf{\geq 1 0 1 ~ m m}$ <br> CL) | Mature Male <br> Biomass <br> ( $\geq \mathbf{1 2 1 ~ m m ~ C L ) ~}$ | Standard <br> Deviation | Legal Male <br> Biomass ( $\mathbf{1 3 6}$ <br> mm CL) | Standard <br> Deviation |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 1.98 |  |  | 10,129 | 724 |
| 1986 | 1.50 | 8,877 | 531 | 8,511 | 501 |
| 1987 | 2.94 | 6,926 | 386 | 6,645 | 365 |
| 1988 | 5.13 | 5,917 | 343 | 5,739 | 316 |
| 1989 | 1.71 | 5,199 | 355 | 5,021 | 312 |
| 1990 | 2.44 | 5,235 | 363 | 4,860 | 329 |
| 1991 | 3.83 | 5,320 | 395 | 5,099 | 371 |
| 1992 | 2.31 | 4,956 | 452 | 4,795 | 414 |
| 1993 | 2.21 | 5,106 | 435 | 4,831 | 395 |
| 1994 | 2.57 | 5,412 | 350 | 5,185 | 324 |
| 1995 | 1.24 | 4,909 | 291 | 4,725 | 269 |
| 1996 | 2.60 | 4,274 | 263 | 4,055 | 254 |
| 1997 | 3.07 | 3,833 | 275 | 3,712 | 268 |
| 1998 | 2.44 | 3,932 | 303 | 3,773 | 293 |
| 1999 | 3.22 | 4,447 | 356 | 4,219 | 342 |
| 2000 | 2.62 | 4,957 | 413 | 4,762 | 396 |
| 2001 | 2.13 | 5,641 | 472 | 5,396 | 453 |
| 2002 | 3.57 | 6,106 | 531 | 5,866 | 511 |
| 2003 | 2.17 | 6,436 | 588 | 6,260 | 566 |
| 2004 | 1.61 | 7,217 | 699 | 6,928 | 663 |
| 2005 | 3.32 | 7,449 | 781 | 7,192 | 746 |
| 2006 | 2.79 | 7,126 | 831 | 6,963 | 799 |
| 2007 | 2.72 | 7,486 | 935 | 7,216 | 891 |
| 2008 | 3.08 | 7,870 | 1,034 | 7,586 | 987 |
| 2009 | 2.01 | 8,126 | 1,112 | 7,861 | 1,065 |
| 2010 | 4.14 | 8,415 | 1,170 | 8,108 | 1,121 |
| 2011 | 4.28 | 8,374 | 1,203 | 8,165 | 1,163 |
| 2012 | 3.35 | 9,245 | 1,381 | 8,923 | 1,315 |
| 2013 | 1.54 | 10,378 | 1,569 | 9,971 | 1,493 |
| 2014 | 1.82 | 10,867 | 10,704 | 10,470 | 1,630 |
| 2015 | 2.45 |  |  | 9,909 | 1,707 |
|  |  |  |  |  |  |

Table 14. Negative log-likelihood values of the fits for scenarios ( Sc ) 1 (equilibrium initial cond.), 2 (exponential formula initial cond.), 3 (added fish ticket CPUE likelihood), 4 (down weight groundfish size composition), 5 (drop groundfish size composition), 7 (three q and total sel.), 9 (stage-2 effective sample size), 10 (dome shaped total selectivity), 12 (area shrinkage), 14 (dome shaped total sel. and area shrinkage), and 16 (restricted time series of total catch and size comp.) for golden king crab in the EAG. Differences in likelihood values are given for scenarios with the same number of data points (base). Likelihood components with zero entry in the entire rows are omitted. Grey highlighted values are minima for scenarios with comparable base number of data points. RetdcatchB= retained catch biomass.

| Likelihood Component | Sc 1 | Sc 2 | Sc 3 | Sc 4 | Sc 5 | Sc 7 | Sc 9 | Sc 10 | Sc 12 | Sc 14 | Sc16 | Sc 2 <br> Sc 1 | Sc 4 Sc 1 | $\begin{aligned} & \text { Sc } 7 \\ & - \\ & \text { Sc } 1 \\ & \hline \end{aligned}$ | $\begin{array}{ll} \hline \text { Sc } 9 \\ - & \\ \text { Sc } 1 \\ \hline \end{array}$ | Sc 10 Sc 1 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Number of free parameters | 134 | 122 | 135 | 134 | 134 | 137 | 134 | 138 | 135 | 139 | 134 |  |  |  |  |  |
| Data | base | base | base+ fishery CPUE | base | basegroundfish length comp | base | base | base | Base+ area | Base + area | $\begin{gathered} \hline \text { Base- } \\ \text { pre 1996 } \\ \text { Total } \\ \text { catch \& } \\ \text { length } \\ \text { comp } \\ \hline \end{gathered}$ |  |  |  |  |  |
| Retlencomp | -889.22 | -890.22 | -889.13 | -889.47 | -889.26 | -889.80 | -892.08 | -887.75 | -888.79 | -886.46 | -890.25 | -1.00 | -0.25 | -0.59 | -2.86 | 1.47 |
| Totallencomp | -866.78 | -866.83 | -867.17 | -867.69 | -868.29 | -867.35 | -865.16 | -868.53 | -866.25 | -867.93 | -731.75 | -0.05 | -0.91 | -0.57 | 1.62 | -1.74 |
| GroundFish <br> discdlencomp | -678.72 | -678.71 | -678.95 | -336.01 | 0.00 | -679.51 | -665.38 | -684.40 | -679.16 | -686.15 | -679.27 | 0.01 | 342.71 | -0.78 | 13.34 | -5.68 |
| Observer cpue | -12.59 | -12.43 | -12.81 | -12.94 | -13.25 | -12.54 | -9.19 | -12.28 | -14.75 | -15.65 | -11.47 | 0.16 | -0.35 | 0.05 | 3.40 | 0.31 |
| RetdcatchB | 8.10 | 8.11 | 8.76 | 7.82 | 7.59 | 8.16 | 8.95 | 8.45 | 8.28 | 8.90 | 6.02 | 0.01 | -0.28 | 0.05 | 0.85 | 0.34 |
| TotalcatchB | 31.77 | 31.79 | 32.67 | 31.18 | 30.71 | 32.29 | 34.55 | 32.70 | 31.99 | 33.46 | 12.15 | 0.02 | -0.60 | 0.51 | 2.77 | 0.93 |
| GdiscdcatchB | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Rec_dev | 7.08 | 6.11 | 6.76 | 5.24 | 4.07 | 6.75 | 10.40 | 7.39 | 6.85 | 7.10 | 6.91 | -0.97 | -1.84 | -0.33 | 3.31 | 0.31 |
| Pot F_dev | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | -0.01 | 0.00 | 0.00 | 0.00 | 0.00 |
| Gbyc_F_dev | 0.08 | 0.08 | 0.08 | 0.08 | 0.08 | 0.08 | 0.08 | 0.08 | 0.08 | 0.08 | 0.08 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Tag | 2,690.79 | 2,690.99 | 2,690.89 | 2,690.15 | 2,689.64 | 2690.86 | 2,693.07 | 2,684.31 | 2,691.06 | 2,684.19 | 2689.57 | 0.20 | -0.64 | 0.07 | 2.28 | -6.48 |
| Fishery cpue | , | , | -1.19 | , | ,68.64 |  | , | , | , | , |  |  |  |  |  |  |
| Total | 290.54 | 288.90 | 289.94 | 628.37 | 961.31 | 288.97 | 315.24 | 279.99 | 289.34 | 277.58 | 402.03 | -1.63 | 337.83 | -1.57 | 24.71 | -10.54 |

Table 15. Time series of annual retained catch (number and weight of crabs), estimated total male catch (number and weight of crabs on the deck), pot fishery effort (number of pot lifts), and estimated groundfish fishery discard mortality (number and weight of crabs) (handling mortality rates of $50 \%$ for pot and $80 \%$ for trawl gear were applied, only to the male portion) for the WAG golden king crab stock. The crab numbers are for the size range 101-185+ mm CL. 1985 refers to the 1985/86 fishing year.

|  | Retained <br> Catch <br> (no.) | Retained <br> Catch <br> Biomass <br> $(\mathbf{t})$ | Total <br> Catch <br> $($ (no. $)$ | Total <br> Catch <br> Biomass <br> $(\mathbf{t})$ | Pot <br> Fishery <br> Effort (no. <br> pot lifts) | Groundfish <br> Discard <br> Mortality(no.) | Groundfish <br> Discard <br> Mortality <br> $(\mathbf{t )}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 981,949 | 2,010 |  |  | 118,563 |  |  |
| 1986 | $2,052,652$ | 4,230 |  |  | 277,780 |  |  |
| 1987 | $1,248,732$ | 2,514 |  |  | 160,229 |  |  |
| 1988 | $1,285,914$ | 2,454 |  |  | 166,409 |  |  |
| 1989 | $1,610,281$ | 3,047 |  | 202,541 | 51 | 0.08 |  |
| 1990 | 889,017 | 1,630 | $2,753,326$ | 3,691 | 108,533 | 374 | 0.57 |
| 1991 | 747,852 | 1,355 | $1,827,434$ | 2,572 | 101,429 | 16 | 0.03 |
| 1992 | 543,541 | 1,025 | $1,113,229$ | 1,520 | 69,443 | 318 | 0.43 |
| 1993 | 352,339 | 665 | $2,001,547$ | 2,822 | 127,764 | 0 | 0.00 |
| 1994 | 845,058 | 1,617 | $3,634,246$ | 4,953 | 195,138 | 82 | 0.12 |
| 1995 | 619,636 | 1,185 | $1,567,028$ | 2,132 | 115,248 | 628 | 0.71 |
| 1996 | 652,801 | 1,231 | $1,269,315$ | 1,767 | 99,267 | 559 | 1.04 |
| 1997 | 558,446 | 1,062 | $1,236,592$ | 1,799 | 86,811 | 211 | 0.37 |
| 1998 | 505,407 | 931 | 782,551 | 1,087 | 35,975 | 1,182 | 1.85 |
| 1999 | 658,377 | 1,235 | $1,467,177$ | 2,093 | 107,040 | 1,091 | 1.42 |
| 2000 | 723,794 | 1,378 | $1,612,997$ | 2,233 | 101,239 | 692 | 0.80 |
| 2001 | 686,738 | 1,282 | $1,503,857$ | 2,138 | 105,512 | 303 | 0.43 |
| 2002 | 664,823 | 1,214 | $1,335,068$ | 1,893 | 78,979 | 700 | 0.92 |
| 2003 | 676,633 | 1,245 | $1,192,551$ | 1,862 | 66,236 | 200 | 0.31 |
| 2004 | 685,465 | 1,262 | $1,249,016$ | 1,880 | 56,846 | 699 | 0.95 |
| 2005 | 639,368 | 1,230 | $1,079,095$ | 1,780 | 30,116 | 1,798 | 3.46 |
| 2006 | 523,701 | 1,048 | 894,219 | 1,547 | 26,870 | 1,311 | 2.28 |
| 2007 | 600,595 | 1,230 | 965,889 | 1,609 | 29,950 | 943 | 1.50 |
| 2008 | 587,661 | 1,208 | 997,465 | 1,730 | 26,200 | 3,979 | 6.45 |
| 2009 | 628,332 | 1,333 | 900,797 | 1,676 | 26,489 | 2,173 | 4.31 |
| 2010 | 626,246 | 1,338 | 868,127 | 1,588 | 29,994 | 1,056 | 2.48 |
| 2011 | 616,118 | 1,332 | 817,532 | 1,514 | 26,326 | 1,576 | 2.25 |
| 2012 | 672,916 | 1,404 | $1,000,311$ | 1,822 | 32,716 | 2,216 | 3.74 |
| 2013 | 686,883 | 1,440 | $1,037,749$ | 1,901 | 41,835 | 2,569 | 3.85 |
| 2014 | 635,312 | 1,257 | 935,794 | 1,591 | 41,548 | 1,635 | 2.46 |
|  |  |  |  |  |  |  |  |

Table 16. Time series of nominal annual pot fishery retained, observer retained, and observer total catch-per-unit-effort (CPUE, number of crabs per pot lift), observer sample size (number of sampled pots), and GLM estimated observer CPUE Index for the WAG golden king crab stock. 1990 refers to the 1990/91 fishing year. Observer retained CPUE includes retained and nonretained legal size crabs.

|  | Pot <br> Fishery <br> Nominal <br> Retained <br> CPUE | Obs. <br> Nominal <br> Retained <br> CPUE | Obs. <br> Nominal <br> Total <br> CPUE | Obs. <br> Sample <br> Size <br> (no.pot <br> Year | Obs. <br> CPUE <br> Index |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1990 | 6.98 | 11.83 | 26.67 | 340 |  |
| 1991 | 7.43 | 7.78 | 19.18 | 857 |  |
| 1992 | 5.90 | 6.39 | 16.83 | 690 |  |
| 1993 | 4.43 | 6.54 | 17.23 | 174 |  |
| 1994 | 4.08 | 6.71 | 19.23 | 1,270 |  |
| 1995 | 4.65 | 4.96 | 14.28 | 5,598 | 1.17 |
| 1996 | 6.07 | 5.42 | 13.54 | 7,194 | 0.95 |
| 1997 | 6.56 | 6.52 | 15.03 | 3,985 | 0.96 |
| 1998 | 11.40 | 9.42 | 23.09 | 1,876 | 1.07 |
| 1999 | 6.32 | 5.93 | 14.49 | 4,523 | 0.91 |
| 2000 | 6.97 | 6.40 | 16.64 | 4,740 | 0.85 |
| 2001 | 6.51 | 5.99 | 14.66 | 4,454 | 0.83 |
| 2002 | 8.42 | 7.47 | 17.37 | 2,509 | 0.92 |
| 2003 | 10.22 | 9.29 | 18.17 | 3,334 | 1.16 |
| 2004 | 12.06 | 11.14 | 22.45 | 2,619 | 1.27 |
| 2005 | 21.23 | 23.74 | 35.94 | 1,365 | 1.12 |
| 2006 | 19.64 | 23.96 | 33.41 | 1,183 | 1.03 |
| 2007 | 20.05 | 21.04 | 32.46 | 1,082 | 0.97 |
| 2008 | 22.43 | 24.59 | 38.17 | 979 | 1.11 |
| 2009 | 23.72 | 26.53 | 34.05 | 892 | 1.16 |
| 2010 | 20.88 | 22.34 | 29.03 | 867 | 1.02 |
| 2011 | 23.40 | 23.81 | 31.12 | 837 | 1.07 |
| 2012 | 20.57 | 22.82 | 30.76 | 1,109 | 1.08 |
| 2013 | 16.42 | 16.95 | 24.96 | 1,223 | 0.77 |
| 2014 | 15.29 | 15.28 | 22.67 | 1,137 | 0.77 |
|  |  |  |  |  |  |

Table 17. Time series of GLM estimated CPUE Indices and standard errors for the fish ticket based retained catch-per-pot lift for the WAG golden king crab stock. The GLM was fitted to the 1985/86 - 1998/99 time series of data and used in scenario 3. 1985 refers to the 1985/86 fishing year.

|  | CPUE <br> Index | Standard Error |
| :---: | :---: | :---: |
| 1985 | 2.02 | 0.07 |
| 1986 | 1.72 | 0.05 |
| 1987 | 1.21 | 0.05 |
| 1988 | 1.35 | 0.04 |
| 1989 | 1.14 | 0.03 |
| 1990 | 0.88 | 0.04 |
| 1991 | 0.72 | 0.04 |
| 1992 | 0.72 | 0.05 |
| 1993 | 0.68 | 0.06 |
| 1994 | 0.82 | 0.04 |
| 1995 | 0.88 | 0.04 |
| 1996 | 0.84 | 0.04 |
| 1997 | 0.77 | 0.03 |
| 1998 | 1.05 | 0.04 |

Table 18. The number of length measurements and stage- 2 effective sample size iteratively estimated by Francis method for retained, total, and groundfish discard catches of golden king crab during 1985 to 2014 in WAG. NA: not available.

| Year | Retained <br> Length | Retained <br> Stage-2 <br> Sffective <br> Sample <br> Sizele (no) | Total <br> Length <br> Sample <br> Size <br> (no) | Total <br> Stage-2 <br> Effective <br> Sample <br> Size (no) | Groundfish <br> Length <br> Sample Size <br> (no) | Groundfish <br> Stage-2 <br> Effective <br> Sample Size <br> (no) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 1,730 | 5 |  |  |  |  |
| 1986 | 1,952 | 6 |  |  |  |  |
| 1987 | 377 | 1 |  |  |  |  |
| 1988 | 61,604 | 179 |  |  |  |  |
| 1989 | 99,433 | 289 |  |  | 33 | 2.83 |
| 1990 | 48,582 | 141 | 6,907 | 29 | 154 | 13.2 |
| 1991 | 50,886 | 148 | 12,165 | 51 | 2 | 0.17 |
| 1992 | 43,491 | 126 | 8,401 | 35 | 34 | 2.91 |
| 1993 | 13,343 | 39 | 4,630 | 19 | NA | NA |
| 1994 | 27,068 | 79 | 23,668 | 99 | 2 | 0.17 |
| 1995 | 16,956 | 49 | 91,652 | 384 | 15 | 1.29 |
| 1996 | 23,385 | 68 | 76,743 | 321 | 46 | 3.94 |
| 1997 | 29,437 | 86 | 53,100 | 222 | 12 | 1.03 |
| 1998 | 25,304 | 74 | 34,535 | 145 | 269 | 23.05 |
| 1999 | 21,387 | 62 | 61,770 | 259 | 247 | 21.17 |
| 2000 | 23,956 | 70 | 71,869 | 301 | 149 | 12.77 |
| 2001 | 18,583 | 54 | 63,707 | 267 | 100 | 8.57 |
| 2002 | 18,937 | 55 | 41,863 | 175 | 143 | 12.25 |
| 2003 | 13,810 | 40 | 38,578 | 162 | 27 | 2.31 |
| 2004 | 13,295 | 39 | 34,832 | 146 | 44 | 3.77 |
| 2005 | 11,675 | 34 | 24,111 | 101 | 20 | 1.71 |
| 2006 | 11,631 | 34 | 26,988 | 113 | 188 | 16.11 |
| 2007 | 8,272 | 24 | 26,643 | 112 | 291 | 24.94 |
| 2008 | 10,530 | 31 | 25,190 | 106 | 174 | 14.91 |
| 2009 | 9,690 | 28 | 29,909 | 125 | 141 | 12.08 |
| 2010 | 9,818 | 29 | 24,817 | 104 | 35 | 3 |
| 2011 | 10,639 | 31 | 26,054 | 109 | 53 | 4.54 |
| 2012 | 6,542 | 19 | 32,921 | 138 | 78 | 6.68 |
| 2013 | 2,609 | 8 | 29,736 | 125 | 64 | 5.48 |
| 2014 | 2,929 | 9 | 25,491 | 107 | 45 | 3.86 |
|  |  |  |  |  |  |  |

Table 19. Total fished area (square miles) and relative proportion of area fished (relative to the maximum area fished) during 1985/86-2014/15 in WAG.

| Year | Fished Area <br> (sq.miles) | Relative <br> Proportion | Year | Fished Area <br> (sq.miles) | Relative <br> Proportion |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 35,082 | 0.60 | 2008 | 26,226 | 0.45 |
| 1986 | 54,680 | 0.93 | 2009 | 29,622 | 0.51 |
| 1987 | 45,732 | 0.78 | 2010 | 37,396 | 0.64 |
| 1988 | 55,326 | 0.94 | 2011 | 26,242 | 0.45 |
| 1989 | 57,678 | 0.98 | 2012 | 30,759 | 0.52 |
| 1990 | 57,164 | 0.97 | 2013 | 43,957 | 0.75 |
| 1991 | 51,506 | 0.88 | 2014 | 41,569 | 0.71 |
| 1992 | 53,059 | 0.90 |  |  |  |
| 1993 | 32,066 | 0.55 |  |  |  |
| 1994 | 50,503 | 0.86 |  |  |  |
| 1995 | 51,717 | 0.88 |  |  |  |
| 1996 | 46,477 | 0.79 |  |  |  |
| 1997 | 53,534 | 0.91 |  |  |  |
| 1998 | 42,970 | 0.73 |  |  |  |
| 1999 | 51,592 | 0.88 |  |  |  |
| 2000 | 58,644 | 1.00 |  |  |  |
| 2001 | 50,478 | 0.86 |  |  |  |
| 2002 | 51,968 | 0.89 |  |  |  |
| 2003 | 53,717 | 0.92 |  |  |  |
| 2004 | 52,812 | 0.90 |  |  |  |
| 2005 | 29,039 | 0.50 |  |  |  |
| 2006 | 34,008 | 0.58 |  |  |  |
| 2007 | 35,811 | 0.61 |  |  |  |

Table 20. Parameter estimates and standard deviations with the 2015 (February 15) MMB for scenarios 1, 2, 3, and 4 for the golden king crab data from the WAG, 1985/86-2014/15. Recruitment and fishing mortality deviations and initial size frequency determination parameters were omitted from this list.

|  | Scenario 1 |  | Scenario 2 |  | Scenario 3 |  | Scenario 4 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Parameter | Estimate | Std Dev | Estimate | Std Dev | Estimate | Std Dev | Estimate | Std Dev | Limits |
| $\log _{-} \omega_{1}$ ( growth incr. intercept) | 2.53 | 0.02 | 2.53 | 0.02 | 2.53 | 0.02 | 2.53 | 0.02 | 1.0, 3.85 |
| $\omega_{2}$ (growth incr. slope) | -10.54 | 1.74 | -10.51 | 1.74 | -10.64 | 1.72 | -10.02 | 1.75 | -60.0,-2.0 |
| log_a (molt prob. slope) | -2.46 | 0.06 | -2.47 | 0.07 | -2.45 | 0.06 | -2.48 | 0.07 | -4.61,-1.39 |
| log_b (molt prob. L50) | 4.95 | 0.004 | 4.95 | 0.01 | 4.95 | 0.004 | 4.95 | 0.01 | 3.869,6.0 |
| $\sigma$ (growth variability std) | 3.66 | 0.10 | 3.66 | 0.10 | 3.66 | 0.10 | 3.67 | 0.10 | 0.1,9.0 |
| log_total sel delta $\theta$, 1985-04 | 3.27 | 0.11 | 3.27 | 0.11 | 3.21 | 0.11 | 3.26 | 0.11 | 0.,4.4 |
| $\log _{-}$total sel delta $\theta$, 2005-14 | 2.87 | 0.18 | 2.89 | 0.18 | 2.87 | 0.17 | 2.97 | 0.17 | 0.,4.4 |
| $\log _{-}$ret. sel delta $0,1985-14$ | 1.73 | 0.06 | 1.73 | 0.06 | 1.72 | 0.06 | 1.73 | 0.06 | 0.,4.4 |
| log_tot sel $\theta_{50}, 1985-04$ | 4.83 | 0.01 | 4.83 | 0.01 | 4.82 | 0.01 | 4.83 | 0.01 | 3.98,5.1 |
| log_tot sel $\theta_{50}, 2005-14$ | 4.86 | 0.01 | 4.87 | 0.01 | 4.86 | 0.01 | 4.88 | 0.02 | 3.98,5.5 |
| log_ret. sel $\theta_{50}, 1985-14$ | 4.91 | 0.002 | 4.91 | 0.002 | 4.91 | 0.002 | 4.91 | 0.002 | 4.85,4.98 |
| $\log \beta_{\mathrm{r}}$ (rec.distribution par.) | -0.71 | 0.30 | -0.70 | 0.30 | -0.82 | 0.28 | -0.97 | 0.29 | -12.0, 12.0 |
| $\operatorname{logq2}$ (catchability 1985-04) | -0.37 | 0.10 | -0.36 | 0.10 | -0.38 | 0.06 | -0.30 | 0.09 | -9.0, 2.25 |
| logq3 (catchability 2005-14) | -0.92 | 0.14 | -0.89 | 0.14 | -0.91 | 0.13 | -0.76 | 0.15 | -9.0, 2.25 |
| log_newsh1 (N1985) |  |  | 2.49 | 0.12 |  |  |  |  | 0.01, 10.0 |
| log_mean_rec (mean rec.) | 0.83 | 0.05 | 0.76 | 0.05 | 0.81 | 0.05 | 0.82 | 0.05 | 0.01, 5.0 |
| log_mean_Fpot (Pot fishery F) | -1.06 | 0.08 | -0.84 | 0.08 | -1.06 | 0.07 | -0.98 | 0.08 | -9.0, -0.01 |
| log_mean_Fground (GF byc. F) | -8.87 | 0.38 | -8.86 | 0.38 | -8.86 | 0.38 | -8.81 | 0.38 | -15.0, -2.0 |
| $\sigma_{e}^{2}$ (observer CPUE additional var) | 0.02 | 0.01 | 0.02 | 0.01 | 0.02 | 0.01 | 0.02 | 0.01 | 0.0, 0.15 |
| $\sigma_{e}^{2} \quad$ (fishery CPUE additional var) |  |  |  |  | 0.001 | 0.003 |  |  | 0.0,1.0 |
| 2015 MMB | 5,343 | 999 | 5,117 | 962 | 5,237 | 941 | 4,864 | 877 |  |

Table 21. Parameter estimates and standard deviations with the 2015 (February 15) MMB for scenarios 5, 10, 12 , and 14 for the golden king crab data from the WAG, 1985/86-2014/15. Recruitment and fishing mortality deviations and initial size frequency determination parameters were omitted from this list.

|  | Scenario 5 |  | Scenario 10 |  | Scenario 12 |  | Scenario 14 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Parameter | Estimate | Std Dev | Estimate | Std Dev | Estimate | Std Dev | Estimate | Std Dev | Limits |
| $\log _{-} \omega_{1}$ ( growth incr. intercept) | 2.54 | 0.02 | 2.58 | 0.02 | 2.53 | 0.02 | 2.58 | 0.02 | 1.0, 3.85 |
| $\omega_{2}$ ( growth incr. slope) | -9.16 | 1.78 | -8.23 | 1.85 | -10.57 | 1.74 | -8.26 | 1.84 | -60.0,-2.0 |
| $\log _{\text {_ }}$ (molt prob. slope) | -2.51 | 0.07 | -2.55 | 0.07 | -2.46 | 0.06 | -2.55 | 0.07 | -4.61,-1.39 |
| $\log _{\text {_ }} \mathrm{b}$ (molt prob. L50) | 4.95 | 0.01 | 4.98 | 0.01 | 4.95 | 0.004 | 4.98 | 0.01 | 3.869,6.0 |
| $\sigma$ (growth variability std) | 3.68 | 0.10 | 3.72 | 0.10 | 3.66 | 0.10 | 3.72 | 0.10 | 0.1,9.0 |
| d1 (incr. dome sel slope 1985-04.) |  |  | 0.07 | 0.01 |  |  | 0.07 | 0.01 | $0.01,1.0$ |
| d2 (decr. dome sel slope 1985-04.) |  |  | -0.14 | 0.01 |  |  | -0.14 | 0.01 | -1.0,-0.1 |
| d3 (incr. dome sel slope 2005-14.) |  |  | 0.18 | 0.02 |  |  | 0.18 | 0.02 | 0.01,1.0 |
| d4 (decr. dome sel slope 2005-14.) |  |  | -0.05 | 0.01 |  |  | -0.05 | 0.01 | -1.0,0.01 |
| log_total sel delta $\theta$, 1985-04 | 3.24 | 0.10 |  |  | 3.27 | 0.11 |  |  | 0., 4.4 |
| $\log _{-}$total sel delta $\theta$, 2005-14 | 3.06 | 0.15 |  |  | 2.88 | 0.18 |  |  | 0., 4.4 |
| $\log _{-}$ret. sel delta $\theta$, 1985-14 | 1.73 | 0.06 | 1.81 | 0.06 | 1.73 | 0.06 | 1.81 | 0.06 | 0., 4.4 |
| log_tot sel $\theta_{50}, 1985-04$ | 4.84 | 0.01 | 5.22 | 0.11 | 4.83 | 0.01 | 5.21 | 0.11 | 3.98, 5.3 |
| log_tot sel $\theta_{50}, 2005-14$ | 4.90 | 0.02 | 4.93 | 0.02 | 4.87 | 0.01 | 4.93 | 0.02 | 3.98, 5.5 |
| log_tot sel $\theta_{95}, 1985-04$ |  |  | 4.96 | 0.02 |  |  | 4.96 | 0.02 | 4.9, 5.2 |
| log_tot sel $\theta_{95}, 2005-14$ |  |  | -5.71 | 870.52 |  |  | -5.71 | 849.52 | -6.0,5.2 |
| log_ret. sel $\theta_{50}, 1985-14$ | 4.91 | 0.002 | 4.92 | 0.002 | 4.91 | 0.002 | 4.92 | 0.002 | 4.85, 4.98 |
| $\log _{\_} \beta_{\mathrm{r}}$ (rec.distribution par.) | -1.28 | 0.27 | -0.56 | 0.31 | -0.68 | 0.31 | -0.54 | 0.31 | -12.0, 12.0 |
| $\operatorname{logq} 2$ (catchability 1985-04) | -0.22 | 0.09 | -0.70 | 0.15 | -0.42 | 0.10 | -0.74 | 0.15 | -9.0, 2.25 |
| $\operatorname{logq} 3$ (catchability 2005-14) | -0.55 | 0.16 | -1.07 | 0.16 | -1.17 | 0.23 | -1.34 | 0.22 | -9.0, 2.25 |
| $\log _{-} \eta$ (stock mixing parameter) |  |  |  |  | -0.41 | 0.27 | -0.43 | 0.24 | -9.0, 0.01 |
| log_mean_rec (mean rec.) | 0.80 | 0.05 | 0.97 | 0.09 | 0.84 | 0.05 | 0.98 | 0.09 | 0.01, 5.0 |
| log_mean_Fpot (Pot fishery F) | -0.88 | 0.08 | -1.19 | 0.11 | -1.07 | 0.08 | -1.19 | 0.11 | -9.0, -0.01 |
| log_mean_Fground (GF byc. F) | -8.76 | 0.38 | -9.26 | 0.41 | -8.88 | 0.38 | -9.27 | 0.41 | -15.0, -2.0 |
| $\sigma_{e}^{2}$ (CPUE additional var) | 0.01 | 0.01 | 0.02 | 0.01 | 0.02 | 0.01 | 0.01 | 0.01 | 0.0, 0.15 |
| 2015 MMB | 4,287 | 742 | 8,920 | 2,263 | 5,932 | 1,125 | 9,638 | 2,353 |  |

Table 22. Parameter estimates and standard deviations with the 2015 (February 15) MMB for scenarios 7, 9 , and 16 for the golden king crab data from the WAG, 1985/86-2014/15. Recruitment and fishing mortality deviations and initial size frequency determination parameters were omitted from this list.

|  | Scenario 7 |  | Scenario 9 |  | Scenario 16 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Parameter | Estimate | Std Dev | Estimate | Std Dev | Estimate | Std Dev | Limits |
| $\log _{-} \omega_{1}$ ( growth incr. intercept) | 2.54 | 0.02 | 2.52 | 0.02 | 2.53 | 0.02 | 1.0, 3.85 |
| $\omega_{2}$ ( growth incr. slope) | -11.39 | 1.69 | -11.46 | 1.72 | -10.01 | 1.76 | -60.0,-2.0 |
| $\log _{-} \mathrm{a}$ (molt prob. slope) | -2.35 | 0.06 | -2.45 | 0.06 | -2.51 | 0.07 | -4.61, -1.39 |
| log_b (molt prob. L50) | 4.96 | 0.004 | 4.94 | 0.005 | 4.95 | 0.005 | 3.869, 6.0 |
| $\sigma$ (growth variability std) | 3.64 | 0.10 | 3.66 | 0.10 | 3.66 | 0.10 | 0.1, 9.0 |
| log_total sel delta $\theta$, 1985-94 | 3.47 | 0.55 |  |  |  |  | 0., 4.4 |
| log_total sel delta $\theta$, 1995-04 | 3.36 | 0.13 |  |  |  |  | 0., 4.4 |
| log_total sel delta $\theta$, 1985-04 |  |  | 3.52 | 0.09 | 3.22 | 0.10 | 0., 4.4 |
| log_ total sel delta $\theta$, 2005-14 | 2.87 | 0.18 | 2.85 | 0.10 | 2.95 | 0.17 | 0., 4.4 |
| $\log _{-}$ret. sel delta $\theta, 1985-14$ | 1.70 | 0.06 | 1.72 | 0.08 | 1.73 | 0.07 | 0., 4.4 |
| log_tot sel $\theta_{50}, 1985-94$ | 4.57 | 0.14 |  |  |  |  | 3.98, 5.1 |
| log_tot sel $\theta_{50}, 1995-04$ | 4.87 | 0.02 |  |  |  |  | 3.98,5.1 |
| log_tot sel $\theta_{50}, 1985-04$ |  |  | 4.85 | 0.02 | 4.86 | 0.01 | 3.98,5.1 |
| log_tot sel $\theta_{50}, 2005-14$ | 4.86 | 0.01 | 4.88 | 0.01 | 4.87 | 0.02 | 3.98,5.5 |
| log_ret. sel $\theta_{50}, 1985-14$ | 4.91 | 0.002 | 4.91 | 0.002 | 4.91 | 0.002 | 4.85,4.98 |
| $\log _{\_} \beta_{\mathrm{r}}$ (rec.distribution par.) | -0.70 | 0.28 | -0.60 | 0.18 | -0.72 | 0.32 | $-12.0,12.0$ |
| Logq1 (catchability 1985-94) | -1.09 | 11141.00 |  |  |  |  | $-9.0,2.25$ |
| $\operatorname{logq} 2$ (catchability 1995-04) | -0.09 | 0.14 |  |  |  |  | -9.0, 2.25 |
| logq2 (catchability 1985-04) |  |  | -0.21 | 0.10 | -0.26 | 0.10 | -9.0, 2.25 |
| logq3 (catchability 2005-14) | -0.87 | 0.14 | -0.67 | 0.13 | -0.91 | 0.15 | -9.0, 2.25 |
| log_mean_rec (mean rec.) | 0.81 | 0.05 | 0.81 | 0.05 | 0.83 | 0.05 | 0.01, 5.0 |
| log_mean_Fpot (Pot fishery F) | -1.01 | 0.08 | -0.89 | 0.08 | -1.04 | 0.08 | -9.0, -0.01 |
| $\log _{-}$mean_Fground (GF byc. F) | -8.77 | 0.38 | -8.76 | 0.38 | -8.88 | 0.39 | -15.0, -2.0 |
| $\sigma_{e}^{2}$ (CPUE additional var) | 0.02 | 0.01 | 0.02 | 0.01 | 0.02 | 0.01 | 0.0, 0.15 |
| $2015 \text { MMB }$ | 5,020 | 978 | 4,248 | 793 | 5,409 | 1,036 |  |

Table 23. Annual abundance estimates of model recruits (millions of crabs), legal male biomass with standard deviation ( t ), and mature male biomass with standard deviation ( t ) for scenario 1 for golden king crab in the WAG. Legal male biomass was estimated on July 1 (start of fishing year) and mature male biomass for fishing year y was estimated on February 15 , fishing year $\mathrm{y}+1$ after the fishing year y fishery total catch removal. 1985 refers to the 1985/86 fishing year. Recruits estimates for 1961 to 2015 are restricted to 1985-2015.

| Year | Recruits to the Model ( $\mathbf{~} \mathbf{1 0 1} \mathbf{~ m m}$ CL) | $\begin{gathered} \text { Mature Male } \\ \text { Biomass } \\ (\geq 121 \mathrm{~mm} \mathrm{CL}) \end{gathered}$ | Standard <br> Deviation | Legal Male Biomass $(\geq 136 \mathrm{~mm} \mathrm{CL})$ | Standard <br> Deviation |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 3.12 |  |  | 10,812 | 899 |
| 1986 | 3.80 | 10,166 | 667 | 9,835 | 612 |
| 1987 | 2.67 | 7,041 | 443 | 6,797 | 389 |
| 1988 | 2.42 | 6,433 | 367 | 6,160 | 319 |
| 1989 | 2.47 | 5,538 | 272 | 5,323 | 242 |
| 1990 | 1.69 | 3,712 | 235 | 3,555 | 206 |
| 1991 | 1.66 | 3,454 | 212 | 3,294 | 194 |
| 1992 | 1.34 | 3,178 | 222 | 3,056 | 206 |
| 1993 | 3.80 | 3,174 | 224 | 3,060 | 208 |
| 1994 | 1.54 | 3,530 | 221 | 3,494 | 199 |
| 1995 | 1.86 | 3,658 | 206 | 3,452 | 189 |
| 1996 | 2.23 | 3,715 | 216 | 3,580 | 203 |
| 1997 | 1.44 | 3,708 | 250 | 3,602 | 229 |
| 1998 | 1.90 | 3,929 | 239 | 3,780 | 224 |
| 1999 | 2.54 | 4,056 | 243 | 3,949 | 229 |
| 2000 | 2.65 | 3,885 | 259 | 3,783 | 246 |
| 2001 | 2.62 | 4,077 | 316 | 3,937 | 298 |
| 2002 | 3.28 | 4,566 | 399 | 4,405 | 376 |
| 2003 | 2.39 | 5,213 | 493 | 5,062 | 468 |
| 2004 | 3.11 | 6,096 | 629 | 5,882 | 592 |
| 2005 | 2.75 | 6,584 | 720 | 6,413 | 684 |
| 2006 | 2.27 | 7,211 | 796 | 6,997 | 757 |
| 2007 | 3.67 | 7,825 | 831 | 7,592 | 793 |
| 2008 | 0.97 | 7,991 | 848 | 7,813 | 810 |
| 2009 | 1.68 | 8,584 | 852 | 8,260 | 813 |
| 2010 | 1.64 | 7,828 | 819 | 7,637 | 788 |
| 2011 | 2.08 | 7,036 | 774 | 6,854 | 743 |
| 2012 | 1.95 | 6,356 | 738 | 6,181 | 708 |
| 2013 | 2.07 | 5,876 | 754 | 5,683 | 722 |
| 2014 | 1.82 | 5,453 | 833 | 5,274 | 800 |
| 2015 | 2.29 | 5,343 | 999 | 5,163 | 957 |

Table 24. Annual abundance estimates of model recruits (millions of crabs), legal male biomass with standard deviation ( t ), and mature male biomass with standard deviation ( t ) for scenario 5 for golden king crab in the WAG. Legal male biomass was estimated on July 1 (start of fishing year) and mature male biomass for fishing year y was estimated on February 15 , fishing year $\mathrm{y}+1$ after the fishing year y fishery total catch removal. 1985 refers to the 1985/86 fishing year. Recruits estimates for 1961 to 2015 are restricted to 1985-2015.

| Year | Recruits to the $\begin{gathered} \text { Model }(\geq 101 \\ \text { mm CL) } \end{gathered}$ | $\begin{gathered} \text { Mature Male } \\ \text { Biomass } \\ (\geq 121 \mathrm{~mm} \mathrm{CL}) \end{gathered}$ | Standard <br> Deviation | $\begin{gathered} \text { Legal Male } \\ \text { Biomass }(\geq 136 \\ \text { mm CL) } \end{gathered}$ | Standard <br> Deviation |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 3.18 |  |  | 10,562 | 903 |
| 1986 | 3.77 | 10,009 | 687 | 9,663 | 624 |
| 1987 | 2.69 | 6,928 | 444 | 6,656 | 381 |
| 1988 | 1.87 | 6,355 | 368 | 6,044 | 315 |
| 1989 | 2.98 | 5,460 | 259 | 5,197 | 230 |
| 1990 | 1.69 | 3,383 | 224 | 3,244 | 195 |
| 1991 | 1.38 | 3,330 | 211 | 3,125 | 186 |
| 1992 | 1.85 | 3,115 | 215 | 2,956 | 197 |
| 1993 | 3.52 | 3,001 | 224 | 2,900 | 204 |
| 1994 | 1.59 | 3,494 | 221 | 3,419 | 195 |
| 1995 | 1.70 | 3,591 | 199 | 3,352 | 179 |
| 1996 | 2.20 | 3,616 | 197 | 3,456 | 184 |
| 1997 | 1.26 | 3,497 | 204 | 3,382 | 190 |
| 1998 | 2.00 | 3,688 | 207 | 3,517 | 195 |
| 1999 | 2.21 | 3,703 | 207 | 3,599 | 194 |
| 2000 | 2.35 | 3,532 | 209 | 3,406 | 196 |
| 2001 | 2.47 | 3,537 | 236 | 3,389 | 220 |
| 2002 | 2.50 | 3,818 | 278 | 3,657 | 259 |
| 2003 | 2.40 | 4,298 | 322 | 4,124 | 302 |
| 2004 | 2.40 | 4,781 | 391 | 4,597 | 365 |
| 2005 | 2.36 | 5,166 | 468 | 4,982 | 439 |
| 2006 | 2.47 | 5,476 | 550 | 5,285 | 517 |
| 2007 | 2.29 | 5,913 | 614 | 5,723 | 578 |
| 2008 | 1.67 | 6,188 | 632 | 5,980 | 599 |
| 2009 | 1.91 | 6,324 | 618 | 6,103 | 589 |
| 2010 | 1.90 | 5,955 | 595 | 5,784 | 568 |
| 2011 | 1.55 | 5,644 | 565 | 5,470 | 538 |
| 2012 | 1.94 | 5,362 | 537 | 5,175 | 512 |
| 2013 | 1.82 | 4,809 | 536 | 4,661 | 510 |
| 2014 | 2.12 | 4,415 | 581 | 4,253 | 554 |
| 2015 | 2.23 | 4,287 | 742 | 4,137 | 702 |

Table 25. Annual abundance estimates of model recruits (millions of crabs), legal male biomass with standard deviation ( t ), and mature male biomass with standard deviation ( t ) for scenario 10 for golden king crab in the WAG. Legal male biomass was estimated on July 1 (start of fishing year) and mature male biomass for fishing year y was estimated on February 15, fishing year y+1 after the fishing year y fishery total catch removal. 1985 refers to the 1985/86 fishing year. Recruits estimates for 1961 to 2015 are restricted to 1985-2015.

| Year | Recruits to the <br> Model ( $\geq \mathbf{1 0 1}$ <br> mm CL) | Mature Male <br> Biomass <br> $\mathbf{( \geq 1 2 1 ~ m m ~ C L ) ~}$ | Standard <br> Deviation | Legal Male <br> Biomass ( $\geq \mathbf{1 3 6}$ <br> mm CL) | Standard <br> Deviation |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 4.98 |  |  |  |  |
| 1986 | 3.17 | 12,847 | 1,586 | 13,550 | 1,786 |
| 1987 | 3.37 | 10,246 | 1,285 | 10,080 | 1,559 |
| 1988 | 2.67 | 9,167 | 1,150 | 9,089 | 1,272 |
| 1989 | 2.99 | 8,122 | 1,044 | 8,054 | 1,138 |
| 1990 | 2.11 | 6,150 | 986 | 6,100 | 1,038 |
| 1991 | 1.85 | 5,877 | 994 | 5,836 | 981 |
| 1992 | 1.51 | 5,623 | 1,013 | 5,613 | 995 |
| 1993 | 4.44 | 5,536 | 996 | 5,540 | 1,018 |
| 1994 | 2.13 | 5,836 | 982 | 5,954 | 1,005 |
| 1995 | 2.48 | 6,147 | 1,075 | 6,124 | 999 |
| 1996 | 2.63 | 6,449 | 1,166 | 6,502 | 1,093 |
| 1997 | 1.96 | 6,694 | 1,252 | 6,764 | 1,193 |
| 1998 | 2.28 | 7,039 | 1,330 | 7,080 | 1,280 |
| 1999 | 3.21 | 7,285 | 1,375 | 7,347 | 1,357 |
| 2000 | 3.26 | 7,190 | 1,430 | 7,254 | 1,400 |
| 2001 | 3.17 | 7,553 | 1,556 | 7,595 | 1,455 |
| 2002 | 3.71 | 8,210 | 1,718 | 8,263 | 1,581 |
| 2003 | 2.68 | 8,982 | 1,871 | 9,073 | 1,746 |
| 2004 | 3.62 | 9,883 | 2,037 | 9,940 | 1,906 |
| 2005 | 3.31 | 10,309 | 2,124 | 10,411 | 2,066 |
| 2006 | 2.68 | 10,952 | 2,247 | 11,025 | 2,156 |
| 2007 | 4.34 | 11,648 | 2,337 | 11,702 | 2,270 |
| 2008 | 1.25 | 11,862 | 2,378 | 11,972 | 2,356 |
| 2009 | 2.11 | 12,549 | 2,437 | 12,514 | 2,398 |
| 2010 | 2.02 | 11,728 | 2,357 | 11,755 | 2,444 |
| 2011 | 2.42 | 10,855 | 2,253 | 10,829 | 2,361 |
| 2012 | 2.36 | 10,071 | 2,158 | 10,024 | 2,244 |
| 2013 | 2.70 | 9,461 | 2,121 | 9,401 | 2,144 |
| 2014 | 2.39 | 8,963 | 2,137 | 8,929 | 2,105 |
| 2015 | 2.64 | 8,920 | 2,263 | 8,892 | 2,129 |
|  |  |  |  |  | 2,257 |
|  |  |  |  |  |  |

Table 26. Annual abundance estimates of model recruits (millions of crabs), legal male biomass with standard deviation ( t ), and mature male biomass with standard deviation ( t ) for scenario 14 for golden king crab in the WAG. Legal male biomass was estimated on July 1 (start of fishing year) and mature male biomass for fishing year y was estimated on February 15, fishing year y+1 after the fishing year y fishery total catch removal. 1985 refers to the 1985/86 fishing year. Recruits estimates for 1961 to 2015 are restricted to 1985-2015.

| Year | Recruits to the <br> Model ( $\geq \mathbf{1 0 1}$ <br> mm CL) | Mature Male <br> Biomass <br> $\mathbf{( \geq 1 2 1 ~ m m ~ C L ) ~}$ | Standard <br> Deviation | Legal Male <br> Biomass ( $\mathbf{1 3 6}$ <br> mm CL) | Standard <br> Deviation |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |
| 1985 | 4.98 | 12,847 | 1,586 | 13,673 | 1,894 |
| 1986 | 3.18 | 10,246 | 1,285 | 10,181 | 1,572 |
| 1987 | 3.38 | 9,167 | 1,150 | 9,181 | 1,286 |
| 1988 | 2.67 | 8,122 | 1,044 | 8,139 | 1,150 |
| 1989 | 3.00 | 6,150 | 986 | 6,180 | 1,051 |
| 1990 | 2.12 | 5,877 | 994 | 5,915 | 994 |
| 1991 | 1.85 | 5,623 | 1,013 | 5,695 | 1,009 |
| 1992 | 1.52 | 5,536 | 996 | 5,621 | 1,033 |
| 1993 | 4.50 | 5,836 | 982 | 6,043 | 1,020 |
| 1994 | 2.07 | 6,147 | 1,075 | 6,230 | 1,014 |
| 1995 | 2.50 | 6,449 | 1,166 | 6,583 | 1,110 |
| 1996 | 2.57 | 6,694 | 1,252 | 6,826 | 1,205 |
| 1997 | 1.96 | 7,039 | 1,330 | 7,103 | 1,287 |
| 1998 | 2.27 | 7,285 | 1,375 | 7,350 | 1,358 |
| 1999 | 3.11 | 7,190 | 1,430 | 7,235 | 1,396 |
| 2000 | 3.18 | 7,553 | 1,556 | 7,514 | 1,444 |
| 2001 | 3.14 | 8,210 | 1,718 | 8,122 | 1,557 |
| 2002 | 3.62 | 8,982 | 1,871 | 8,896 | 1,709 |
| 2003 | 2.65 | 9,883 | 2,037 | 9,724 | 1,852 |
| 2004 | 3.64 | 10,309 | 2,124 | 10,185 | 1,993 |
| 2005 | 3.27 | 10,952 | 2,247 | 10,823 | 2,070 |
| 2006 | 2.58 | 11,648 | 2,337 | 11,503 | 2,187 |
| 2007 | 4.50 | 11,862 | 2,378 | 11,769 | 2,289 |
| 2008 | 1.28 | 12,549 | 2,437 | 12,407 | 2,346 |
| 2009 | 2.16 | 11,728 | 2,357 | 11,731 | 2,408 |
| 2010 | 2.19 | 10,855 | 2,253 | 10,888 | 2,336 |
| 2011 | 2.73 | 10,071 | 2,158 | 10,233 | 2,229 |
| 2012 | 2.55 | 9,461 | 2,121 | 9,828 | 2,144 |
| 2013 | 2.92 | 8,963 | 9,137 | 9,533 | 2,138 |
| 2014 | 2.46 | 2,263 | 9,629 | 2,354 |  |
| 2015 | 2.67 |  |  |  |  |
|  |  |  |  |  |  |

Table 27. Annual abundance estimates of model recruits (millions of crabs), legal male biomass with standard deviation ( t ), and mature male biomass with standard deviation ( t ) for scenario 16 for golden king crab in the WAG. Legal male biomass was estimated on July 1 (start of fishing year) and mature male biomass for fishing year y was estimated on February 15, fishing year $\mathrm{y}+1$ after the fishing year y fishery total catch removal. 1985 refers to the 1985/86 fishing year. Recruits estimates for 1961 to 2015 are restricted to 1985-2015.

| Year | $\begin{aligned} & \text { Recruits to the } \\ & \text { Model }(\geq 101 \\ & \text { mm CL) } \end{aligned}$ | $\begin{gathered} \text { Mature Male } \\ \text { Biomass } \\ (\geq 121 \mathrm{~mm} \mathrm{CL}) \end{gathered}$ | Standard Deviation | $\begin{gathered} \text { Legal Male } \\ \text { Biomass }(\geq 136 \\ \text { mm CL) } \end{gathered}$ | Standard Deviation |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 3.18 |  |  | 10,799 | 896 |
| 1986 | 3.82 | 10,270 | 671 | 9,853 | 614 |
| 1987 | 2.67 | 7,200 | 459 | 6,863 | 396 |
| 1988 | 2.64 | 6,631 | 381 | 6,259 | 328 |
| 1989 | 2.33 | 5,776 | 298 | 5,477 | 261 |
| 1990 | 1.68 | 4,080 | 274 | 3,831 | 238 |
| 1991 | 2.16 | 3,842 | 258 | 3,613 | 233 |
| 1992 | 1.32 | 3,610 | 269 | 3,440 | 242 |
| 1993 | 1.92 | 3,826 | 241 | 3,627 | 221 |
| 1994 | 2.03 | 4,106 | 219 | 3,958 | 200 |
| 1995 | 1.93 | 3,471 | 196 | 3,310 | 178 |
| 1996 | 2.20 | 3,422 | 193 | 3,243 | 180 |
| 1997 | 1.38 | 3,481 | 237 | 3,308 | 207 |
| 1998 | 1.91 | 3,707 | 217 | 3,501 | 199 |
| 1999 | 2.50 | 3,822 | 218 | 3,670 | 203 |
| 2000 | 2.64 | 3,650 | 233 | 3,497 | 218 |
| 2001 | 2.67 | 3,823 | 291 | 3,626 | 271 |
| 2002 | 3.42 | 4,314 | 382 | 4,088 | 355 |
| 2003 | 2.69 | 5,004 | 492 | 4,779 | 462 |
| 2004 | 3.14 | 5,997 | 660 | 5,698 | 612 |
| 2005 | 2.71 | 6,682 | 771 | 6,409 | 723 |
| 2006 | 2.33 | 7,381 | 845 | 7,073 | 798 |
| 2007 | 3.68 | 7,974 | 877 | 7,663 | 832 |
| 2008 | 0.97 | 8,138 | 891 | 7,878 | 847 |
| 2009 | 1.66 | 8,696 | 891 | 8,290 | 846 |
| 2010 | 1.64 | 7,927 | 855 | 7,675 | 819 |
| 2011 | 2.14 | 7,104 | 805 | 6,875 | 771 |
| 2012 | 1.98 | 6,406 | 767 | 6,183 | 733 |
| 2013 | 2.11 | 5,934 | 786 | 5,683 | 747 |
| 2014 | 1.83 | 5,516 | 868 | 5,277 | 827 |
| 2015 | 2.30 | 5,409 | 1,036 | 5,168 | 982 |

Table 28. Negative log-likelihood values of the fits for scenarios ( Sc ) 1 (equilibrium initial cond.), 2 (exponential formula initial cond.), 3 (added fish ticket CPUE likelihood), 4 (down weight groundfish size composition), 5 (drop groundfish size composition), 7 (three q and total sel.), 9 (stage-2 effective sample size), 10 (dome shaped total selectivity), 12 (area shrinkage), 14 (dome shaped total sel. and area shrinkage), and 16 (restricted time series of total catch and size comp.) for golden king crab in the WAG. Differences in likelihood values are given for scenarios with the same number of data points (base). Likelihood components with zero entry in the entire rows are omitted. Grey highlighted values are minima for scenarios with comparable base number of data points. RetdcatchB= retained catch biomass.



Table 29. Predicted total and retained catch OFL ( t ) under Tier 4 assumption for various scenarios for EAG and WAG, respectively. $\mathrm{Sc}=$ scenario; $\mathrm{F}_{\max }=$ estimated maximum instantaneous total pot fishery mortality during 1985/86-2014/15; $\mathrm{MMB}_{2015}$ or $\mathrm{MMB}_{1986} / \mathrm{MMB}_{\text {initial }}=$ ratio of terminal MMB or 1986 MMB relative to initial MMB ( $=$ MMB $_{1961}$ for equilibrium condition or $\mathrm{MMB}_{1986}$ for exponential initial condition); and $\mathrm{LF}=$ length composition.

|  | EAG |  |  |  |  | WAG |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Sc | Total Catch OFL <br> (t) | Retained Catch OFL (t) | $\begin{aligned} & \mathbf{F}_{\text {max }} \\ & \mathbf{y r}^{-1} \end{aligned}$ | $\begin{gathered} \hline \text { MMB }_{2015} \\ / \\ \text { MMB }_{\text {initial }} \end{gathered}$ | $\begin{gathered} \hline \text { MMB }_{1986} \\ / \\ \text { MMB }_{\text {initial }} \end{gathered}$ | Total Catch OFL <br> (t) | Retained Catch OFL (t) | $\begin{aligned} & \mathbf{F}_{\text {max }} \\ & \mathbf{y r}^{-1} \end{aligned}$ | $\begin{gathered} \mathbf{M M B}_{2015} \\ / \\ \text { MMB }_{\text {initial }} \end{gathered}$ | $\begin{gathered} \mathbf{M M B}_{1986} \\ / \\ \text { MMB }_{\text {initial }} \end{gathered}$ | $\underset{\mathbf{y r}^{-1}}{\mathbf{M}}$ | Remarks |
| 1 | 1,622 | 1,583 | 1.02 | 0.53 | 0.51 | 926 | 871 | 1.25 | 0.39 | 0.71 | 0.23 | Equilibrium initial condition, asymptotic selectivity |
| 2 | 1,720 | 1,679 | 1.02 | 1.07 | 1.00 | 870 | 819 | 1.25 | 0.61 | 1.00 | 0.23 | Exponential initial condition, asymptotic selectivity |
| 3 | 1,627 | 1,588 | 0.99 | 0.53 | 0.50 | 911 | 858 | 1.16 | 0.38 | 0.63 | 0.23 | Equilibrium initial condition, asymptotic selectivity, fish ticket CPUE |
| 4 | 1,562 | 1,519 | 1.02 | 0.53 | 0.51 | 815 | 767 | 1.28 | 0.38 | 0.73 | 0.23 | Equilibrium initial condition, asymptotic selectivity, down weighted groundfish bycatch LF. |
| 5 | 1,567 | 1,515 | 1.01 | 0.55 | 0.51 | 681 | 638 | 1.31 | 0.38 | 0.76 | 0.23 | Equilibrium initial condition, asymptotic selectivity, removed groundfish bycatch LF. |
| 6 | 1,692 | 1,634 | 1.01 | 1.11 | 1.00 | 693 | 650 | 1.35 | 0.44 | 1.00 | 0.23 | Exponential initial condition, asymptotic selectivity, removed groundfish bycatch LF. |
| 7 | 1,510 | 1,474 | 0.97 | 0.52 | 0.52 | 900 | 848 | 1.02 | 0.38 | 0.72 | 0.23 | Equilibrium initial condition, three catchability and asymptotic total selectivity. |
| 8 | 1,159 | 1,132 | 1.07 | 0.44 | 0.43 | 660 | 622 | 1.32 | 0.32 | 0.59 | 0.18 | Equilibrium initial condition, asymptotic selectivity |
| 9 | 1,596 | 1,567 | 1.01 | 0.57 | 0.49 | 702 | 661 | 1.35 | 0.32 | 0.78 | 0.23 | Equilibrium initial condition, asymptotic selectivity, iterative effective sample sizes. |
| 10 | 1,721 | 1,677 | 1.00 | 0.52 | 0.53 | 1,158 | 1,089 | 1.15 | 0.48 | 0.69 | 0.23 | Equilibrium initial condition, dome shaped selectivity. |
| 11 | 1,202 | 1,172 | 1.06 | 0.43 | 0.44 | 684 | 644 | 1.28 | 0.33 | 0.56 | 0.18 | Equilibrium initial condition, dome shaped selectivity. |
| 12 | 1,360 | 1,327 | 1.00 | 0.50 | 0.54 | 1,02 | 965 | 1.25 | 0.42 | 0.69 | 0.23 | Equilibrium initial condition, asymptotic selectivity, fishing area shrinkage proportion used for CPUE prediction. |
| 13 | 1,360 | 1,327 913 | 1.00 1.05 | 0.50 0.41 | 0.54 0.45 | 1,022 740 | 699 | 1.25 1.31 | 0.42 0.34 | 0.69 0.58 | 0.23 0.18 | Equilibrium initial condition, asymptotic selectivity, fishing area shrinkage proportion used for CPUE prediction. |
| 14 | 1,406 | 1,369 | 0.99 | 0.49 | 0.61 | 1,274 | 1,202 | 1.14 | 0.51 | 0.69 | 0.23 | Equilibrium initial condition, dome shaped selectivity, fishing area shrinkage proportion used for CPUE prediction. |
| 15 | 964 | 939 | 1.05 | 0.39 | 0.48 | 766 | 1,202 723 | 1.28 | 0.51 0.51 | 0.69 | 0.23 0.18 | Equilibrium initial condition, dome shaped selectivity, fishing area shrinkage proportion used for CPUE prediction. |
| 16 | 1,603 | 1,565 | 0.99 | 0.53 | 0.51 | 913 | 861 | 1.21 | 0.39 | 0.71 | 0.23 | Equilibrium initial condition, asymptotic selectivity, total LF and total catch time series restricted to 1996/97-2014/15 for EAG or 1995/962014/15 for WAG. |



Figure 3. Historical commercial harvest (from fish tickets; metric tons) and catch-per-unit effort (CPUE, number of crabs per pot lift) of golden king crab in the EAG, 1985/86-2014/15 fisheries (note: 1985 refers to the 1985/86 fishing year).


Figure 4. Historical commercial harvest (from fish tickets; metric tons) and catch-per-unit effort (CPUE, number of crabs per pot lift) of golden king crab in the WAG, 1985/86-2014/15 fisheries (note: 1985 refers to the 1985/86 fishing year ).


Figure 5. Aleutian Islands golden king crab harvest by ADF\&G statistical areas for 2014/15.


Figure 6. Predicted (line) vs. observed (bar) retained catch relative length frequency distributions for scenarios 1 (black line), 2 (red line), 3 (blue line), 4 (yellow line), 5 (orange line), 7 (dark red dashed line), 9 (dark blue line), 10 (green line), 12 (dark green line), 14 (violet line), and 16 (gray line) data of golden king crab in the EAG, 1985/86 to 2014/15.


Figure 7. Predicted (line) vs. observed (bar) total catch relative length frequency distributions for scenarios 1 (black line), 2 (red line), 3 (blue line), 4 (yellow line), 5 (orange line), 7 (dark red dashed line), 9 (dark blue line), 10 (green line), 12 (dark green line), 14 (violet line), and 16 (gray line) data of golden king crab in the EAG, 1990/91 to 2014/15.


Figure 8. Predicted (line) vs. observed (bar) groundfish (trawl) discarded catch relative length frequency distributions for scenarios 1 (black line), 2 (red line), 3 (blue line), 4 (yellow line), 5 (orange line), 7 (dark red dashed line), 9 (dark blue line), 10 (green line), 12 (dark green line), 14 (violet line), and 16 (gray line) data of golden king crab in the EAG, 1989/90 to 2014/15.


Figure 9. Predicted (green line) vs. observed (black line) mean lengths (with $+/-1.96$ SE) of retained (top right), total (top left), and groundfish (trawl) discarded (bottom right) catch compositions for scenario 1 data of golden king crab in the EAG, 1985/86 to 2014/15.

Note: $\overline{l_{t}}=\sum_{i=1}^{n} l_{t, i} \times P_{t, i}$ and $S E^{2}=\sum_{i=1}^{n} l_{t, i}^{2} \times \operatorname{Var}\left(P_{t, i}\right)$, where $\bar{l}_{t}$ is the mean mid length for year $t ; P_{t, i}$ is the proportion of ith size-class in year $t ; S E$ is the standard error; $n$ is the number of length classes; and $\operatorname{Var}\left(P_{t, i}\right)$ is computed using Equation A. 20 .


Figure 10. Estimated total (black solid line) and retained selectivity (red dotted line) for pre- and post- rationalization periods under scenarios (Sc) $1,2,3,4,5,10,12$, and 14 fits of EAG golden king crab data.


Scenario 10
Figure 11. Bubble plots of standardized residuals of retained catch length composition for scenarios $1,5,10$, and 14 fits for EAG golden king crab, 1985/86-2014/15. Filled circles are the positive and unfilled circles are the negative standardized residuals. The area of the circle is the relative magnitude of the residual.


Figure 12. Bubble plots of standardized residuals of total catch length composition for scenarios 1, 5, 10, and 14 fits for EAG golden king crab, 1990/91-2014/15. Filled circles are the positive and unfilled circles are the negative standardized residuals. The area of the circle is the relative magnitude of the residual.


Figure 13. Bubble plots of standardized residuals of groundfish (trawl) bycatch length composition for scenarios 1 , 5, 10 , and 14 fits for EAG golden king crab, 1989/90-2014/15. Filled circles are the positive and unfilled circles are the negative standardized residuals. The area of the circle is the relative magnitude of the residual.


Figure 14. Observed tag recaptures (open circle) vs. predicted tag recaptures (solid line) by size bin for years 1 to 6 recaptures for scenario 1 fit of EAG golden king crab.


Figure 15. Observed and predicted mean length (with two SE) of recaptures vs. release length for years 1 to 6 recaptures for scenario 1 fit of EAG golden king crab.


Figure 16. Comparison of estimated growth matrix components (proportions) between EAG (black circles) and WAG (green line) for scenario 1 fits to golden king crab data. The number at the top of each plot is the mid length ( mm CL ) of the contributing length-class. The proportions in each plot are the proportions falling into different length-classes as a result of growth of the contributing lengthclass.


Figure 17. Comparison of input CPUE indices (open circles with $+/-2$ SE) with predicted CPUE indices (colored solid lines) for scenarios (Sc) 1, $2,3,4,5,7,9,10,12,14$, and 16 fits for EAG golden king crab data, 1985/86-2014/15. Model estimated additional standard error was added to each input standard error.


Figure 18. Estimated number of male recruits (millions of crab $\geq 101 \mathrm{~mm} \mathrm{CL}$ ) to the golden king crab assessment model for scenarios (Sc) 1, 2, 3, $4,5,7,9,10,12,14$, and 16 fits in EAG, 1961-2015. The number of recruits are centralized using ( R -mean R )/mean R for comparing different scenarios' results.


Figure 19. Recruit distribution to the golden king crab assessment model size group for scenarios (Sc) $1,2,3,4,5,7,9,10,12,14$, and 16 fits in EAG.


Figure 20. Trends in golden king crab mature male biomass for scenarios (Sc) $1,2,3,4,5,7,9,10,12,14$, and 16 model fits in the EAG, 1960/61-2014/15. Mature male crab is $\geq 121 \mathrm{~mm}$ CL. Scenario 1 estimates have two standard errors confidence limits.


Figure 21. Trends in pot fishery full selection total fishing mortality of golden king crab for scenarios (Sc) 1,2, 3, 4, 5, 7, 9, 10, 12, 14, and 16 model fits in the EAG, 1981-2014 (note: 1981 refers to the 1981/82 fishing year).


Figure 22. Observed (open circle) vs. predicted (solid line) retained catch (top left), total catch (top right), and trawl (or groundfish) bycatch (bottom left) of golden king crab for scenarios (Sc) 1, 2, 3, 4, 5, 7, 9, 10, 12, 14, and 16 fits in the EAG, 1985-2014. (note: 1985 refers to the 1985/86 fishing year).


Figure 23. Retrospective fits of the model for removal of terminal year's data for scenarios (Sc) 1, $2,3,4,5,7,10,12,14$, and 16 fits for golden king crab in the EAG, 1985-2014 (note: 1985 refers to the1985/86 fishing year).


Figure 24. Estimated molt probability vs. carapace length of golden king crab for scenarios ( Sc ) $1,2,3,4,5,7,9,10,12,14$, and 16 model fits in the EAG, 1985/86-2014/15.


Figure 25. Predicted (line) vs. observed (bar) retained catch relative length frequency distributions for scenarios 1 (black line), 2 (red line), 3 (blue line), 4 (yellow line), 5 (orange line), 7 (dark red dashed line), 9 (dark blue line), 10 (green line), 12 (dark green line), 14 (violet line), and 16 (gray line) data of golden king crab in the WAG, 1985/86 to 2014/15.


Figure 26. Predicted (line) vs. observed (bar) total catch relative length frequency distributions for scenarios 1 (black line), 2 (red line), 3 (blue line), 4 (yellow line), 5 (orange line), 7 (dark red dashed line), 9 (dark blue line), 10 (green line), 12 (dark green line), 14 (violet line), and 16 (gray line) data of golden king crab in the WAG, 1990/91 to 2014/15.


Figure 27. Predicted (line) vs. observed (bar) groundfish (trawl) discarded catch relative length frequency distributions for scenarios 1 (black line), 2 (red line), 3 (blue line), 4 (yellow line), 5 (orange line), 7 (dark red dashed line), 9 (dark blue line), 10 (green line), 12 (dark green line), 14 (violet line), and 16 (gray line) data of golden king crab in the WAG, 1989/90 to 2014/15.


Figure 28. Predicted (green line) vs. observed (black line) mean lengths (with +/- 1.96 SE) of retained (top right), total (top left), and groundfish (trawl) discarded (bottom right) catch compositions for scenario 1 data of golden king crab in the WAG, 1985/86 to 2014/15.


Figure 29. Estimated total (black solid line) and retained selectivity (red dotted line) for pre- and post- rationalization periods under scenarios (Sc) $1,2,3,4,5,10,12$, and 14 fits of WAG golden king crab data.


Figure 30. Bubble plots of standardized residuals of retained catch length composition for scenarios 1, 5, 10 and 14 fits for WAG golden king crab, 1985/86-2014/15. Filled circles are the positive and unfilled circles are the negative standardized residuals. The area of the circle is the relative magnitude of the residual.


Figure 31. Bubble plots of standardized residuals of total catch length composition for scenarios 1, 5, 10, and 14 fits for WAG golden king crab, 1990/91-2014/15. Filled circles are the positive and unfilled circles are the negative standardized residuals. The area of the circle is the relative magnitude of the residual.


Figure 32. Bubble plots of standardized residuals of groundfish (trawl) bycatch length composition for scenarios 1, 5, 10, and 14 fits for WAG golden king crab, 1989/90-2014/15. Filled circles are the positive and unfilled circles are the negative standardized residuals. The area of the circle is the relative magnitude of the residual.


Figure 33. Observed tag recaptures (open circle) vs. predicted tag recaptures (solid line) by size bin for years 1 to 6 recaptures for scenario 1 fit of WAG golden king crab data. The tagging experiments were conducted in EAG.


Figure 34. Comparison of input CPUE indices (open circles with $+/-2$ SE) with predicted CPUE indices (colored solid lines) for scenarios (Sc) 1 , $2,3,4,5,7,9,10,12,14$, and 16 fits for WAG golden king crab data, 1985/86-2014/15. Model estimated additional standard error was added to each input standard error.


Figure 35. Estimated number of male recruits (millions of crab $\geq 101 \mathrm{~mm} \mathrm{CL}$ ) to the golden king crab assessment model for scenarios (Sc) 1, 2, 3, $4,5,7,9,10,12,14$, and 16 fits in WAG, 1961-2015. The number of recruits are centralized using ( R -mean R )/mean R for comparing different scenarios' results.


Figure 36. Recruit distribution to the golden king crab assessment model size group for scenarios (Sc) 1, 2, 3, 4, 5, 7, 9, 10, 12, 14, and 16 fits in WAG.


Figure 37. Trends in golden king crab mature male biomass for scenarios (Sc) 1, 2, 3, 4, 5, 7, 9, 10, 12, 14, and 16 model fits in the WAG, 1960/61-2014/15. Mature male crab is $\geq 121 \mathrm{~mm}$ CL. Scenario 1 estimates have two standard errors confidence limits. The MMB trend lines for scenarios 10 (green line, dome shaped selectivity) and 14 (violet line, area shrinkage factor on CPUE prediction and dome shaped selectivity) are higher than the rest.


Figure 38. Trends in pot fishery full selection total fishing mortality of golden king crab for scenarios (Sc) 1, 2, 3, 4, 5, 7, 9, 10, 12, 14, and 16 model fits in the WAG, 1981-2014 (note: 1981 refers to the 1981/82 fishing year).


Figure 39. Observed (open circle) vs. predicted (solid line) retained catch (top left), total catch (top right), and trawl (or groundfish) bycatch (bottom left) of golden king crab for scenarios (Sc) 1, 2, 3, 4, 5, 7, 9, 10, 12, 14, and 16 fits in the WAG, 1985-2014. (note: 1985 refers to the 1985/86 fishing year).


Figure 40. Retrospective fits of mature male biomass by the model when terminal year's data were systematically removed until $2012 / 13$ for scenarios (Sc) $1,2,3,4,5,7,10,12,14$, and 16 fits for golden king crab in the WAG, 1985-2014 (note: 1985 refers to the1985/86 fishing year).


Figure 41 . Estimated molt probability vs. carapace length of golden king crab under scenarios 1 , $2,3,4,5,7,9,10,12,14$, and 16 for WAG.

## Appendix A: Integrated model

Aleutian Islands Golden King Crab (Lithodes aequispinus) Stock Assessment Model
Development- East of $174^{\circ} \mathrm{W}$ (EAG) and west of $174^{\circ} \mathrm{W}$ (WAG) Aleutian Island stocks

## Basic population dynamics

The annual [male] abundances by size are modeled using the equation:

$$
\begin{equation*}
N_{t+1, j}=\sum_{i=1}^{j}\left[N_{t, i} e^{-M}-\left(\hat{C}_{t, i}+\widehat{D}_{t, i}+\widehat{T r}_{t, i}\right) e^{\left(y_{t}-1\right) M}\right] X_{i, j}+R_{t+1, j} \tag{A.1}
\end{equation*}
$$

where $N_{t, i}$ is the number of [male] crab in length class $i$ on 1 July (start of fishing year) of year $t ; \hat{C}_{t, i}, \hat{D}_{t, i}$, and $\widehat{T} r_{t, i}$ are respectively the predicted fishery retained, pot fishery discard dead, and groundfish fishery discard dead catches in length class $i$ during year $t$; $\widehat{D}_{t, i}$ is estimated from the intermediate total ( $\hat{T}_{t, i}$ temp $)$ catch and the retained ( $\hat{C}_{t, i}$ ) catch by the Equation 2c. $X_{i, j}$ is the probability in length-class $i$ growing into length-class $j$ during the year; $y_{t}$ is elapsed time period from 1 July to the mid -point of fishing period in year $t ; M$ is instantaneous rate of natural mortality, and $R_{t+1, j}$ recruitment to length class j in year $\mathrm{t}+1$.

The catches are predicted using the equations

$$
\begin{align*}
& \widehat{T}_{t, j, t e m p}=\frac{F_{t} S_{t, j}^{T}}{z_{t, j}} N_{t, j} e^{-y_{t} M}\left(1-e^{-z_{t, j}}\right)  \tag{A.2a}\\
& \hat{C}_{t, j}=\frac{F_{t} S_{t, j}^{T} s_{t, j}^{r}}{z_{t, j}} N_{t, j} e^{-y_{t} M}\left(1-e^{-z_{t, j}}\right)  \tag{A.2b}\\
& \widehat{D}_{t, j}=0.2\left(\hat{T}_{t, j, t e m p}-\hat{C}_{t, j}\right)  \tag{A.2c}\\
& \widehat{T r}_{t, j}=0.65 \frac{F_{t}^{T r} \frac{s_{j}^{T r}}{z_{t, j}} N_{t, j} e^{-y_{t} M}\left(1-e^{-z_{t, j}}\right)}{\widehat{T}_{t, j}=\hat{C}_{t, j}+\widehat{D}_{t, j}} \tag{A.2d}
\end{align*}
$$

where $Z_{t, j}$ is total fishery-related mortality on animals in length-class $j$ during year $t$ :

$$
\begin{equation*}
Z_{t, j}=F_{t} s_{t, j}^{T} s_{t, j}^{r}+0.2 F_{t} s_{t, j}^{T}\left(1-s_{t, j}^{r}\right)+0.65 F_{t}^{T r} s_{j}^{T r} \tag{A.3}
\end{equation*}
$$

$F_{t}$ is the full selection fishing mortality in the pot fishery, $F_{t}^{T r}$ is the full selection fishing mortality in the trawl fishery, $s_{t, j}^{T}$ is the total selectivity for animals in length-class $j$ by the
pot fishery during year $t, s_{j}^{T r}$ is the selectivity for animals in length-class j by the trawl fishery, $s_{t, j}^{r}$ is the probability of retention for animals in length-class $j$ by the pot fishery during year t . Pot bycatch mortality of 0.2 and groundfish bycatch mortality of 0.65 (average of trawl (0.8) and fish pot (0.5) mortality) were assumed.

The initial conditions $(\mathrm{t}=1985)$ are computed using the equation

$$
\begin{equation*}
N_{1985, i}=\tilde{N}_{1985} e^{\varepsilon_{i}} / \sum_{j} e^{\varepsilon_{j}} \tag{A.4}
\end{equation*}
$$

where $\tilde{N}_{1985}$ is the total abundance in 1985 , and $\varepsilon_{i}$ are parameters which determine the initial (1985) length-structure (one of $\varepsilon_{i}=0$ to ensure identifiability).

We also used the equilibrium initial condition using the following relations:
The equilibrium stock abundance is
$\underline{N}=\mathbf{X} . S . \underline{N}+\underline{R}$
The equilibrium abundance in $1960, \underline{\mathrm{~N}}_{1960}$, is
$\underline{N}_{1960}=(\mathbf{I}-\mathbf{X S})^{-1} \underline{R}$
where $\mathbf{X}$ is the growth matrix, $\mathbf{S}$ is a matrix with diagonal elements given by $e^{-M}, \mathbf{I}$ is the identity matrix, and $\underline{R}$ is the product of average recruitment and relative proportion of total recruitment to each size-class.

We used the mean number of recruits from 1996 to 2014 in equation (A.6) to obtain the equilibrium solution under only natural mortality ( 0.23 ) in year 1960, and then projected the equilibrium abundance under natural mortality with recruitment estimated for each year after 1960 up to 1985 with removal of retained catches during 1981/82 to 1984/85.

## Growth Matrix

The growth matrix $\mathbf{X}$ is modeled as follows:

$$
X_{i, j}= \begin{cases}0 & \text { if } j<i  \tag{A.7}\\ P_{i, j}+\left(1-m_{i}\right) & \text { if } j=i \\ P_{i, j} & \text { if } j>i\end{cases}
$$

where:
$P_{i, j}=m_{i}\left\{\begin{array}{lr}\int_{-\infty}^{j_{2}-L_{i}} N\left(x \mid \mu_{i}, \sigma^{2}\right) d x & \text { if } j=i \\ \int_{j_{1}-L_{i}}^{j_{i}-L_{i}} N\left(x \mid \mu_{i}, \sigma^{2}\right) d x & \text { if } i<j<n, \\ \int_{j_{1}-L_{i}}^{\infty} N\left(x \mid \mu_{i}, \sigma^{2}\right) d x & \text { if } i=n\end{array}\right.$

$$
N\left(x \mid \mu_{i}, \sigma^{2}\right)=\frac{1}{\sqrt{2 \pi \sigma^{2}}} e^{-\left(\frac{x-\mu_{i}}{\sqrt{2} \sigma}\right)^{2}}, \text { and }
$$

$\mu_{\mathrm{i}}$ is the mean growth increment for crabs in size-class $i$ :
$\mu_{\mathrm{i}}=\omega_{1}+\omega_{2} * \bar{L}_{\mathrm{i}}$.
$\omega_{1} \quad, \omega_{2}, \quad$ and $\sigma$ are estimable parameters, and $j_{1}$ and $j_{2}$ are the lower and upper limits of the receiving length-class $j$ (in mm CL ), and $\bar{L}_{\mathrm{i}}$ is the mid-point of the contributing length interval $i$. The quantity $m_{i}$ is the molt probability for size-class $i$ :
$\mathrm{m}_{\mathrm{i}}=\frac{1}{1+\mathrm{e}^{\mathrm{c}\left(\mathrm{T}_{\mathrm{i}}-\mathrm{d}\right)}}$
where $c$ and $d$ are parameters.

## Selectivity and retention

a) Selectivity and retention are both assumed to be logistic functions of length. Selectivity depends on the fishing period for the pot fishery:

$$
\begin{equation*}
S_{i}=\frac{1}{1+e^{\left[-\ln \left(19 \frac{\tau_{i}-\theta_{50}}{\theta_{95}-\theta_{50}}\right]\right.}} \tag{A.10}
\end{equation*}
$$

where $\theta_{95}$ and $\theta_{50}$ are the parameters of the selectivity/ retention pattern (Mark Maunder, unpublished generic crab model). In the program, we re-parameterized the denominator $\left(\theta_{95}-\theta_{50}\right)$ to $\log ($ delta $\theta)$ so that the difference is always positive.
b) A dome shaped total selectivity is considered for certain scenarios.

$$
\begin{equation*}
\mathrm{S}_{\mathrm{i}}=\left[\frac{1}{1+\mathrm{e}^{\left[-\mathrm{dj}\left(\mathrm{\tau}_{\mathrm{i}}-\theta_{50}\right)\right]}} \times\left\{1-\frac{1}{\left.1+\mathrm{e}^{\left[\mathrm{dk}\left(\tau_{\mathrm{i}}-\theta_{95}\right)\right.}\right]}\right] \frac{1}{\mathrm{x}}\right. \tag{A.11}
\end{equation*}
$$

where dj and dk are two sets of slopes for the first (increasing) and second (decreasing) logistic curves for the pre- and post-rationalization periods; $\theta_{50}$ and $\theta_{95}$ are inflection
points for the first (increasing) and second (decreasing) curves; and X is the maximum of the first two terms on the right hand side (Quinn and Deriso, 1999).

## Recruitment

Recruitment to length -class $i$ during year $t$ is modeled as $R_{t, i}=\bar{R} e^{\epsilon_{i}} \Omega_{i}$ where $\Omega_{i}$ is a normalized gamma function
$\operatorname{gamma}\left(x \mid \alpha_{r}, \beta_{r}\right)=\frac{x^{\alpha_{r}-1} e^{\frac{x}{\beta_{r}}}}{\beta_{r}{ }^{\alpha_{r}} \Gamma_{\left(\alpha_{r}\right)}}$
with $\alpha_{r}$ and $\beta_{r}$ (restricted to the first six length- classes).

## Parameter estimation

Table A1 lists the parameters of the model indicating which are estimated and which are pre-specified. The objective function includes contributions related to the fit of the model to the available data and penalties (priors on the various parameters).

Tables A2 lists the values for the weight parameters, which weight (with the corresponding coefficient of variations in parentheses) the components of the objective function for EAG and WAG, respectively.

## Likelihood components

## Catches

The contribution of the catch data (retained, total, and groundfish discarded) to the objective function is given by:

$$
\begin{align*}
& L L_{r}^{\text {catch }}=\lambda_{r} \sum_{t}\left\{\ell \operatorname{n}\left(\sum_{j} \hat{C}_{t, j} w_{j}+c\right)-\ln \left(\sum_{j} C_{t, j} w_{j}+c\right)\right\}^{2}  \tag{A.13a}\\
& L L_{T}^{\text {catch }}=\lambda_{T} \sum_{t}\left\{\ln \left(\sum_{j} \hat{T}_{t, j} w_{j}+c\right)-\ln \left(\sum_{j} T_{t, j} w_{j}+c\right)\right\}^{2}  \tag{A.13b}\\
& L L_{G D}^{\text {catch }}=\lambda_{G D} \sum_{t}\left\{\ln \left(\sum_{j} \widehat{T r}_{t, j} w_{j}+c\right)-\ln \left(\sum_{j} T r_{t, j} w_{j}+c\right)\right\}^{2} \tag{A.13c}
\end{align*}
$$

where $\lambda_{r}, \lambda_{T}$ and $\lambda_{G D}$ are weights assigned to likelihood components for the retained, pot total and groundfish discard catches; $w_{j}$ is the average mass of a crab is length-class $j$; $C_{t, j}, T_{t, j}$, and $\operatorname{Tr}_{t, j}$ are, respectively, the observed numbers of crab in size class $j$ for retained, pot total, and groundfish fishery discarded crab during year $t$, and $c$ is a small constant value.

## Catch-rate indices

The catch-rate indices are assumed to be lognormally distributed about the model prediction. Account is taken of variation in additional to that related to sampling variation:

$$
\begin{equation*}
L L_{r}^{C P U E}=\lambda_{r, C P U E}\left\{0.5 \sum_{t} \ln \left[2 \pi\left(\sigma_{r, t}^{2}+\sigma_{e}^{2}\right)\right]+\sum_{t} \frac{\left(\ln \left(C P U E_{t}^{r}+c\right)-\ln \left(C \widehat{P U E}_{t}^{r}+c\right)\right)^{2}}{2\left(\sigma_{r, t}^{2}+\sigma_{e}^{2}\right)}\right\} \tag{A.14}
\end{equation*}
$$

where $C P U E_{t}^{r}$ is the standardized retain catch-rate index for year $t, \sigma_{r, t}$ is standard error of the logarithm of $C P U E_{t}^{r}$, and $C \widehat{P U E}_{t}^{r}$ is the model-estimate corresponding to $C P U E_{t}^{r}$ :

$$
\begin{equation*}
\widehat{C P U E}{ }_{t}^{r}=q_{k} \sum_{j} S_{j}^{T} S_{j}^{r}\left(N_{t, j}-0.5\left[\widehat{C_{t, j}}+\widehat{D_{t, j}}+\widehat{T r_{t, j}}\right]\right) e^{-y_{t} M} \tag{A.15}
\end{equation*}
$$

where $q_{k}$ is the catchability coefficient during the k-th time period (e.g., pre- and postrationalization time periods), $\sigma_{e}$ is the extent of over-dispersion, $c$ is a small constant to prevent zero values (0.001), and $\lambda_{r, C P U E}$ is the weight assigned to the catch-rate data. We used the same likelihood formula (A.14) for fish ticket retained catch rate indices for scenario 3 model.
Following Burnham et al. (1987), we computed the $\ln (C P U E)$ variance by:

$$
\begin{equation*}
\sigma_{r, t}^{2}=\ln \left(1+C V_{r, t}^{2}\right) \tag{A.16}
\end{equation*}
$$

We also considered area shrinkage component in the catch rate estimation for some scenarios following (Zhou et al. (2011):

$$
\begin{equation*}
\widehat{\operatorname{CPUE}}_{\mathrm{t}}^{\mathrm{r}}=\mathrm{q}_{\mathrm{k}}\left(\frac{\mathrm{at}}{\mathrm{~A}}\right)^{\eta} \sum_{\mathrm{j}} \mathrm{~S}_{\mathrm{j}}^{\mathrm{T}} S_{\mathrm{j}}^{\mathrm{r}}\left(\mathrm{~N}_{\mathrm{t}, \mathrm{j}}-0.5\left[\widehat{\mathrm{C}_{\mathrm{t}, \mathrm{j}}}+\widehat{\mathrm{D}_{\mathrm{t}, \mathrm{j}}}+\widehat{\mathrm{Tr}_{\mathrm{t}, \mathrm{j}}}\right]\right) \mathrm{e}^{-\mathrm{y}_{\mathrm{t}} \mathrm{M}} \tag{A.17}
\end{equation*}
$$

where $a_{t}$ is the area fished during year $t$ and $A$ is the maximum area where stock has been caught during 1985-2014. However we used only the 1995-2014 proportions in the model scenarios (because observer standardized CPUE indices are available only for 1995-2014). We treated the WAG and EAG stocks separate for this analysis and assumed that the shrinkage of fishing area occurred separately at each stock inhabiting area. The parameter $\eta$ determines the mixing process in each area: no mixing when $\eta=1$, full mixing when $\eta=0$, and partial mixing when $0<\eta<1$.

## Length-composition data

The length-composition data are included in the likelihood function using the robust normal for proportions likelihood, i.e. generically:

$$
\begin{equation*}
L L_{r}^{L F}=0.5 \sum_{t} \sum_{j} \ln \left(2 \pi \sigma_{t, j}^{2}\right)-\sum_{t} \sum_{j} \ln \left[\exp \left(-\frac{\left(P_{t, j}-\hat{P}_{t, j}\right)^{2}}{2 \sigma_{t, j}^{2}}\right)+0.01\right] \tag{A.18}
\end{equation*}
$$

where $P_{t, j}$ is the observed proportion of crabs in length-class $j$ in the catch during year $t$, $\hat{P}_{t, j}$ is the model-estimate corresponding to $P_{t, j}$, i.e.:

$$
\begin{gather*}
\hat{L}_{t, j}^{r}=\frac{\hat{C}_{t, j}}{\sum_{j}^{n} \hat{C}_{t, j}} \\
\hat{L}_{t, j}^{T}=\frac{\hat{T}_{t, j}}{\sum_{j}^{n} \hat{T}_{t, j}} \\
\hat{L}_{t, j}^{G F}=\frac{\widehat{\operatorname{Tr}}_{t, j}}{\sum_{j}^{n} \widehat{\operatorname{Tr}}_{t, j}} \tag{A.19}
\end{gather*}
$$

$\sigma_{t, j}^{2}$ is the variance of $P_{t, j}$ :

$$
\begin{equation*}
\sigma_{t, j}^{2}=\left[\left(1-P_{t, j}\right) P_{t, j}+\frac{0.1}{n}\right] / S_{t} \tag{A.20}
\end{equation*}
$$

$S_{t}$ is the effective sample size for year $t$.

The input effective sample sizes were rescaled from actual numbers of length measurements for all scenarios except scenario 9 (iterative reweighting) as follows:

$$
\begin{align*}
& S_{t}^{r}=\min (0.01 * \text { number of length measurements in year } t, 200) \\
& S_{t}^{T}=\min (0.001 * \text { number of length measurements in year } t, 150)  \tag{A.21}\\
& S_{t}^{G F}=\min (0.1 * \text { number of length measurements in year } t, 25)
\end{align*}
$$

Iterative reweighting of effective sample sizes at stage- 2 for scenario 9 was done using Francis' (2011) method. Stage-1 weighting for this procedure was done using equation A.21, but without implementing the maximum limits of 200,150 , and 25 for retained catch, total catch, and groundfish bycatch, respectively.

Note: The likelihood calculation for retained length composition starts from length-class 6 (mid length 128 mm CL ) because the length-classes 1 to 5 mostly contain zero data.

## Tagging data

Let $V_{j, t, y}$ be the number of males that were released in year $t$ that were in length-class $j$ when they were released and were recaptured after $y$ years, and $\underline{\underline{V}}_{j, t, y}$ be the vector of recaptures by length-class from the males that were released in year $t$ that were in lengthclass $j$ when they were released and were recaptured after $y$ years. The multinomial likelihood of the tagging data is then:

$$
\begin{equation*}
\ell \mathrm{n} L=\sum_{t} \sum_{j} \sum_{y} \sum_{i} \tilde{V}_{j, t, y, i} \ln \hat{\rho}_{j, t, y, i} \tag{A.22}
\end{equation*}
$$

where $\hat{\rho}_{j, t, y, i}$ is the proportion in length-class $i$ of the recaptures of males which were released during year $t$ that were in length-class $j$ when they were released and were recaptured after $y$ years:

$$
\begin{equation*}
\underline{\hat{\rho}}_{j, t, y} \propto \underline{s}^{T}[\mathbf{X}]^{y} \underline{\Omega}^{(j)} \tag{A.23}
\end{equation*}
$$

where $\underline{\Omega}^{(j)}$ is a vector with $V_{j, t, y}$ at element $j$ and 0 otherwise, and $s^{T}$ is the total selectivity vector (Punt et al. 1997).

This likelihood function is predicted on the assumption that all recaptures are in the pot fishery and the reporting rate is independent of the size of crab. The expected number of recaptures in length-class $l$ is given by:

$$
\begin{equation*}
r_{l}=\sum_{t} \sum_{j} \frac{s_{l}\left[\mathbf{X}^{t}\right]_{j, l}}{\sum_{l^{\prime}} s_{l^{\prime}}\left[\mathbf{X}^{t}\right]_{j, l^{\prime}}} \sum_{k} V_{j, k, t} \tag{A.24}
\end{equation*}
$$

The last term, $\sum_{k} V_{j, k, t}$, is the number of recaptured male crab that were released in
length-class $j$ after t time-steps. The term $\sum_{j} \frac{s_{l}\left[\mathbf{X}^{t}\right]_{j, l}}{\sum_{l^{\prime}} s_{l}\left[\mathbf{X}^{t}\right]_{j, l}} \sum_{k} V_{j, k, t}$ is the predicted number of animals recaptured in length-class $l$ that were at liberty for t time-steps.

## Penalties

Penalties are imposed on the deviations of annual pot fishing mortality about mean pot fishing mortality, annual trawl fishing mortality about mean trawl fishing mortality, recruitment about mean recruitment, and the posfunction (fpen) :

$$
\begin{align*}
& P_{1}=\lambda_{F} \sum_{t}\left(\ell \mathrm{n} F_{t}-\ell \mathrm{n} \bar{F}\right)^{2}  \tag{A.25}\\
& P_{2}=\lambda_{F^{T r}} \sum_{t}\left(\ell \mathrm{n} F_{t}^{T_{r}}-\ell \mathrm{n} \bar{F}^{T_{r}}\right)^{2}  \tag{A.26}\\
& P_{3}=\lambda_{R} \sum_{t}\left(\ell \mathrm{n} \varepsilon_{t}\right)^{2}  \tag{A.27}\\
& P_{5}=\lambda_{\text {posfn }} * \text { fpen } \tag{A.28}
\end{align*}
$$

## Standardized Residual of Length Composition

$$
\begin{equation*}
\text { Std. } \text { Res }_{t, j}=\frac{P_{t, j}-\widehat{t_{t, j}}}{\sqrt{2 \sigma_{t, j}^{2}}} \tag{A.29}
\end{equation*}
$$

## Output Quantities

## Harvest rate

Total pot fishery harvest rate:

$$
\begin{equation*}
E_{t}=\frac{\sum_{j=1}^{n}\left(\hat{C}_{j, t}+\widehat{D}_{j, t}\right)}{\sum_{j=1}^{n} N_{j, t}} \tag{A.30}
\end{equation*}
$$

Exploited legal male biomass at the start of year $t$ :
$L M B_{t}=\sum_{j=\text { legal size }}^{n} s_{j}^{T} s_{j}^{r} N_{j, t} w_{j}$
where $w_{j}$ is the weight of an animal in length-class j .
Mature male biomass on 15 February spawning time (NPFMC 2007) in the following year:

$$
\begin{equation*}
M M B_{t}=\sum_{j=\text { mature size } e}^{n}\left\{N_{j, t} e^{y^{\prime} M}-\left(\hat{C}_{j, t}+\widehat{D}_{j, t}+\widehat{\operatorname{Tr}}_{j, t}\right) e^{\left(y_{t}-y \prime\right) M}\right\} w_{j} \tag{A.32}
\end{equation*}
$$

where $y^{\prime}$ is the elapsed time from 1 July to 15 February in the following year.

For estimating the next year limit harvest levels from current year stock abundances, a limit $F^{\prime}$ value is needed. Current crab management plan specifies five different Tier formulas for different stocks depending on the strength of information available for a stock, for computing $F^{\prime}$ (NPFMC 2007). For the golden king crab, the following Tier 4 formula is applied to compute $F^{\prime}$ :
(a) If $M M B_{t} \geq M \bar{M} B, \quad F^{\prime}=\gamma M$,
(b) If $M M B_{t}<M \bar{M} B$ and $M M B_{t}>0.25 M \bar{M} B$,

$$
\begin{equation*}
F^{\prime}=\gamma M \frac{\left(\frac{M M B_{t}}{M \overline{M B}}-\alpha\right)}{(1-\alpha)} \tag{A.33}
\end{equation*}
$$

(c) If $M M B_{t} \leq 0.25 M \bar{M} B, F^{\prime}=0$
where $\gamma$ is a constant multiplier of $M, \alpha$ is a parameter, and $M \bar{M} B$ is the mean mature male biomass estimated for a selected time period and used as a $B_{M S Y}$ proxy for the Tier 4 stock.

Because projected $M M B_{t}$ is depended on the intervening retained and discard catch (i.e., $M M B_{t}$ is estimated after the fishery), an iterative procedure is applied using Equations A. 32 and A. 33 with retained and discard catch predicted from Equations A.2b-d. The next year limit harvest catch is estimated using Equations A.2b-d with the estimated $F^{\prime}$ value.

Table A1. Estimated parameters of the population dynamics model

| Parameter | Number of parameters |
| :---: | :---: |
| Initial conditions: |  |
| Initial total numbers, $\tilde{N}_{1985}$ | 1 |
| Length-specific proportions, $\varepsilon_{i}$ | $n-1$ |
| Length specific equilibrium abundance $N_{e q}$ | $n$ |
| Fishing mortalities: |  |
| Pot fishery, $F_{t}$ | 1985-2014 |
| Mean pot fishery fishing mortality, $\bar{F}$ | 1 |
| Trawl fishery, $F_{t}{ }^{T r}$ | 1989-2014 (the mean F for 1989 to 1994 was used to estimate trawl discards back to 1985 . |
| Mean trawl fishery fishing mortality, $\bar{F}^{\text {Tr }}$ | 1 |
| Selectivity and retention: |  |
| Pot fishery total selectivity $\theta_{50}^{T}$ | 2 (1985-2004; 2005+) |
| Pot fishery total selectivity difference, delta $\theta^{T}$ | 2 (1985-2004; 2005+) |
| Pot fishery retention $\theta_{50}^{r}$ | 2 (1985-2004; 2005+) |
| Pot fishery retention difference delta $\theta^{r}$ | 2 (1985-2004; 2005+) |
| Growth: |  |
| Expected growth increment, $\omega_{1}, \omega_{2}$ | 2 |
| Variability in growth increment, $\sigma$ | 1 |
| Molt probability (size transition matrix with tag data) $a$ | 1 |
| Molt probability (size transition matrix with tag data) $b$ | 1 |
| Natural mortality, $M$ | Pre-specified, $0.23 \mathrm{yr}^{-1}$ |
| Recruitment: |  |
| Distribution to length-class, $\alpha_{r}, \beta_{r}$ | 2 |
| Median recruitment, $\bar{R}$ | 1 |
| Recruitment deviations, $\varepsilon_{t}$ | 55 (1961-2015) or 30 (1986-2015) |
| $\mathrm{F}_{\text {OFL }}$ | 1 |
| Fishery catchability, $q$ | $\begin{aligned} & 2 \text { (1985-2004; 2005+) or } 3 \text { (1985-1994; } \\ & 1995-2004 ; 2005+) \end{aligned}$ |
| Likelihood weights (coefficient of variation) | Pre-specified, varies for different scenarios |

Table A2. Specifications for the weights with corresponding coefficient of variations* in parentheses for each scenario for EAG and WAG. select. phase $=$ selectivity phase.


Table A2 continued.

| Trawl fishing mortality $\operatorname{dev}, \lambda_{F^{\text {tr }}}$ | Initially 1000 , relaxed to 0.001 at phases $\geq$ select. phase | Initially 1000 , relaxed to 0.001 at phases $\geq$ select. phase | Initially 1000, relaxed to 0.001 at phases $\geq$ select. phase | Initially 1000, relaxed to 0.001 at phases $\geq$ select. phase | Initially 1000, relaxed to 0.001 at phases $\geq$ select. phase | Initially 1000, relaxed to 0.001 at phases $\geq$ select. phase | Initially 1000, relaxed to 0.001 at phases $\geq$ select. phase |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Recruitment, $\lambda_{R}$ | 2 (0.533) | 2 | 2 | 2 | 2 | 2 | 2 |
| Posfunction (to keep abundance estimates always positive), | 1000 (0.022) | 1000 | 1000 | 1000 | 1000 | 1000 | 1000 |
| $\lambda_{\text {posfn }}$ |  |  |  |  |  |  |  |
| Tagging likelihood | EAG individual tag returns | EAG individual tag returns | EAG individual tag returns | EAG individual tag returns | EAG individual tag returns | EAG individual tag returns | EAG individual tag returns |

* Coefficient of Variation, $\quad C V=\sqrt{e^{\frac{1}{2 \times \text { Weight }}-1}}$

Table A2 continued.

| Weight | Value |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | Scenario 8 | Scenario 9 | Scenario 10 | Scenario 11 | Scenario 12 |
| Catch: |  |  |  |  |  |
| Retained catch. $\lambda_{r}$ | 500 (0.032) | 500 | 500 | 500 | 500 |
| Total catch, $\lambda_{T}$ | Number of sampled pots scaled to a max $300$ | Number of sampled pots scaled to a max | Number of sampled pots scaled to a max | Number of sampled pots scaled to a max | Number of sampled pots scaled to a max |
|  |  | 300 | 300 | 300 | 300 |
| Groundfish bycatch, $\lambda_{G D}$ | 0.2 (3.344) | 0.2 | 0.2 | 0.2 | 0.2 |
| Catch-rate: <br> Observer legal size crab catchrate, $\lambda_{r, \text { CPUE }}$ | 1(0.805) | 1 | 1 | 1 | 1 |
| Penalty weights: |  |  |  |  |  |
| Pot fishing mortality dev, $\lambda_{F}$ | Initially 1000, relaxed to 0.001 at phases $\geq$ select.phase | Initially 1000 , relaxed to 0.001 at phases $\geq$ select.phase | Initially 1000 , relaxed to 0.001 at phases $\geq$ select. phase | Initially 1000 , relaxed to 0.001 at phases $\geq$ select. phase | Initially 1000 , relaxed to 0.001 at phases $\geq$ select.phase |
| Trawl fishing mortality dev, $\lambda_{F^{T r}}$ | Initially 1000 , relaxed to 0.001 at phases $\geq$ select. phase | Initially 1000 , relaxed to 0.001 at phases $\geq$ select. phase | Initially 1000 , relaxed to 0.001 at phases $\geq$ select. phase | Initially 1000 , <br> relaxed to 0.001 <br> at phases $\geq$ <br> select. phase | Initially 1000 , relaxed to 0.001 at phases $\geq$ select.phase |
| Recruitment, $\lambda_{R}$ | 2(0.533) | 2 | 2 | 2 | 2 |
| Posfunction (to keep abundance estimates always positive), $\lambda_{\text {posfn }}$ | 1000 (0.022) | 1000 | 1000 | 1000 | 1000 |
| Tagging likelihood | EAG individual tag returns | EAG individual tag returns | EAG individual tag returns | EAG individual tag returns | EAG individual tag returns |

Table A2 continued.

| Weight | Value |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
|  | Scenario 13 | Scenario 14 | Scenario 15 | Scenario 16 |
| Catch: |  |  |  |  |
| Total catch, $\lambda_{T}$ | 500 (0.032) | 500 | 500 | 500 |
|  | Number of sampled pots scaled to a max 300 | Number of sampled pots scaled to a max 300 | Number of sampled pots scaled to a max | Number of sampled pots scaled to a max 300 |
|  |  |  | 300 |  |
| Groundfish bycatch, $\lambda_{G D}$ Catch-rate: <br> Observer legal size crab catch-rate, $\lambda_{r, \text { CPUE }}$ | 0.2 (3.344) | 0.2 | 0.2 | 0.2 |
|  |  |  |  |  |
|  | 1(0.805) | 1 | 1 | 1 |
| Penalty weights: |  |  |  |  |
| Pot fishing mortality dev, | Initially 1000, relaxed to | Initially 1000, relaxed to | Initially 1000, relaxed | Initially 1000, relaxed to |
| $\lambda_{F}$ | 0.001 at phases $\geq$ select. phase | 0.001 at phases $\geq$ select. phase | to 0.001 at phases $\geq$ select. phase | 0.001 at phases $\geq$ select. phase |
| Trawl fishing mortality $\operatorname{dev}, \lambda_{F^{T r}}$ | Initially 1000 , relaxed to 0.001 at phases $\geq$ select. phase | Initially 1000 , relaxed to 0.001 at phases $\geq$ select. phase | Initially 1000 , relaxed to 0.001 at phases $\geq$ select. phase | Initially 1000 , relaxed to 0.001 at phases $\geq$ select. phase |
| Recruitment, $\lambda_{R}$ | 2(0.533) | 2 | 2 | 2 |
| Posfunction (to keep abundance estimates always positive), $\lambda_{\text {posfn }}$ | 1000 (0.022) | 1000 | 1000 | 1000 |
| Tagging likelihood | EAG individual tag returns | EAG individual tag returns | EAG individual tag returns | EAG individual tag returns |

## Appendix B: Catch and CPUE data

The commercial catch and length frequency distribution were estimated from ADF\&G landing records and dockside sampling (Bowers et al., 2008, 2011). The annual retained catch, total catch, and groundfish (or trawl) discarded mortality are provided in Table 1 for EAG and Table 14 for WAG. The weighted length frequency data were used to distribute the catch into different $(5-\mathrm{mm})$ size intervals. The length frequency data for a year were weighted by each sampled vessel's catch as follows. The $i$-th length-class frequency was estimated as:

$$
\begin{equation*}
\sum_{j=1}^{k} C_{j} \frac{L F_{j, i}}{\sum_{i=1}^{n} L F_{j, i}} \tag{B.1}
\end{equation*}
$$

where $k=$ number of sampled vessels in an year, $L F_{j, i}=$ number of crabs in the $i$-th length-class in the sample from $j$-th vessel, $\mathrm{n}=$ number of size classes, $C_{j}=$ number of crabs caught by $j$-th vessel. Then the relative frequency for the year was calculated and applied to the annual retained catch (in number of crabs) to obtain retained catch by length-class.

The annual total catch (in number of crabs) was estimated by the observer nominal (unstandardized) total CPUE considering all vessels multiplied by the total fishing effort (number of pot lifts). The weighted length frequency of the observer samples across the fleet was estimated using Equation B.1. Observer measurement of crab ranged from 20 to 220 mm CL. To restrict the total number of crabs to the model assumed size range (101-185+ mm CL), the proportion of observer total relative length frequency corresponding to this size range was multiplied by the total catch (number of crabs). This total number of crabs was distributed into length-classes using the weighted relative length frequency. Thus crab sizes < 101 mm CL were excluded from the model. Note that the total crab catch by size that went into the model did not consider retained and discard components separately. However, once the model estimated the annual total catch, then retained catch was deducted from this total and multiplied by handling mortality [we used a $20 \%$ handling mortality (Siddeek et al. 2005) to obtain the directed fishery discarded (dead) catch].

Observer data have been collected since 1988 (Moore et al. 2000; Barnard et al. 2001; Barnard and Burt 2004; Gaeuman 2011), but data were not comprehensive in the initial years, so a shorter time series of data for the period 1990/91-2014/15 was selected for this analysis. During 1990/91-1994/95, observers were only deployed on catcherprocessor vessels. During 1995/96-2004/05, observers were deployed on all fishing vessels during their fishing activity. Observers have been deployed on all fishing vessels since 2005/06, but catcher-only vessels are required to carry observers for a minimum of $50 \%$ of their fishing activity during a season; catcher-processor vessels are still required to carry observers during all fishing activity. Onboard observers count and measure all crabs caught and categorize catch as females, sublegal males, retained legal males, and non-retained legal males in a sampled pot. Prior to the 2009/10 season, depending on season, area, and type of fishing vessel, observers were also instructed to sample additional pots in which all crab were only counted and categorized as females, sublegal males, retained legal males, and non-retained legal males, but were not measured. Annual mean nominal CPUEs of retained and total crabs were estimated considering all sampled pots within each season (Tables 2 and 15). For model-fitting the CPUE time series was further restricted to 1995/962014/15 because the reliability of categorization of crabs by observers improved after 1995. Length-specific CPUE data collected by observers provides information on a wider size range of the stock than did the commercial catch length frequency data obtained from mostly legal-sized landed males.

There were significant changes in fishing practice due to changes in management regulations (e.g., since 1996/97 constant TAC and since 2005/06 crab rationalization), pot configuration (escape web on the pot door increased to $9 "$ since 1999), and improved observer recording in Aleutian Islands golden king crab fisheries since 1998. These changes prompted us to consider two separate observer CPUE time series, 1995/96-2004/05 and 2005/06-2014/15, to estimate CPUE indices for model input.

To include a long time series of CPUE indices for stock abundance contrast, we also considered the 1985/86-1998/99 legal size standardized CPUE as a separate likelihood component in scenario 3. Because of the lack of soak time data previous to 1990, we estimated the CPUE index considering a limited set of explanatory variables (e.g., vessel, captain, area, month) and fitting the lognormal GLM to fish ticket data (Tables 3 and 16).

## Observer CPUE index:

The CPUE standardization followed the GLM fitting procedure (Maunder and Punt 2004; Starr 2012). We considered the negative binomial GLM on positive and zero catches to select the explanatory variables. The response variable CPUE is the observer sample catch record for a pot haul. The negative binomial model uses the $\log$ link function for the GLM fit. Therefore, we assumed the null model to be

$$
\begin{equation*}
\ln \left(\mathrm{CPUE}_{\mathrm{i}}\right)=\text { Year }_{\mathrm{y}_{\mathrm{i}}} \tag{B.2}
\end{equation*}
$$

The maximum set of model terms offered to the stepwise selection procedure was:
$\ln \left(\right.$ CPUE $\left._{\mathrm{I}}\right)=$ Year $_{\mathrm{y}_{\mathrm{i}}}+\mathrm{ns}\left(\right.$ Soak $_{\mathrm{si}}$, df $)+$ Month $_{\mathrm{m}_{\mathrm{i}}}+$ Area $_{\mathrm{ai}}+$ Vessel $_{\mathrm{vi}}+$
Captain $_{\text {ci }}+$ Gear $_{\text {gi }}+n s\left(\right.$ Depth $_{\text {di }}$, df $)+n s\left(\right.$ VesSoak $_{\text {vsi }}$, df $)$,
where ns=cubic spline , $\mathrm{df}=$ degree of freedom, and all variables are selfexplanatory.

We used a log link function and a dispersion parameter ( $\theta$ ) in the GLM fitting process (Siddeek et al., in press).

The final models for EAG were:
$\ln ($ CPUE $)=$ Year + Gear + Captain $+\mathrm{ns}($ Soak, 3$)$
for the 1995/96-2004/05 period $\left[\theta=1.33, \mathrm{R}^{2}=0.23\right.$ with $\mathrm{ns}($ Soak, 3 ) forced in]
$\ln ($ CPUE $)=$ Year + Captain + ns $($ Soak, 16$)+$ Gear
for the $2005 / 06-2014 / 15 \operatorname{period}\left(\theta=2.25, R^{2}=0.11\right)$.

The final models for WAG were:
$\ln ($ CPUE $)=$ Year + Captain + Gear $+\mathrm{ns}($ Soak, 8$)$
for the 1995/96-2004/05 period ( $\theta=0.98, R^{2}=0.18$ ), and

$$
\begin{equation*}
\ln (\text { CPUE })=\text { Year }+ \text { Captain }+ \text { Gear }+ \text { ns }(\text { Soak, } 16) \tag{B.7}
\end{equation*}
$$

for the 2005/06-2014/15 period $\left[\theta=1.16, \mathrm{R}^{2}=0.05\right.$ with $\mathrm{ns}($ Soak, 16) forced in]

Standardized nominal CPUE data are presented in Tables 2 and 15 respectively, for EAG and WAG.

Figures B. 1 and B. 10 depict the trends in nominal and standardized CPUE indices for the two CPUE time series for EAG and WAG, respectively. Figures B.2-B. 3 and B.11-B. 12 show the diagnostic plots for the fits for EAG and WAG, respectively. The deviance and QQ plots support good fits to EAG and WAG data by GLM using the negative binomial error distribution. Figures B.4-B. 9 and B.13-B. 18 depict CDI and influence plots of the predictor variables for EAG and WAG, respectively.

Fish Ticket CPUE index:
We also fitted the lognormal GLM for fish ticket retained CPUE time series 1985/86 - 1998/99 offering year, month, vessel, captain, and area as explanatory variables. The final model for EAG was:
$\ln ($ CPUE $)=$ Year + Captain + Vessel + Month, R $^{2}=0.45$
and for WAG was:
$\ln ($ CPUE $)=$ Year + Captain + Vessel, R $^{2}=0.46$
The $\mathrm{R}^{2}$ for the fish ticket data fits are much higher compared to that for observer data fits.

Figures B. 19 and B. 21 depict the trends in nominal and standardized CPUE indices for the fish ticket CPUE time series for EAG and WAG, respectively. Figures B. 20 and B. 22 show the QQ plots for the fits for EAG and WAG, respectively. The QQ plots support reasonable fits to EAG and WAG data by GLM using the lognormal error distribution.


Figure B.1. Trends in non-standardized [arithmetic (nominal)] and standardized (negative binomial GLM) CPUE indices with +/- 2 SE for Aleutian Islands golden king crab from EAG (east of $174{ }^{\circ} \mathrm{W}$ longitude). Top panel: 1995/96-2004/05 observer data and bottom panel: 2005/06-2014/15 observer data. Standardized indices: black line and non-standardized indices: red line.


Figure B.2. Deviance residuals vs. explanatory and response variables of the best negative binomial fit model for legal male crab CPUE. Deviance residuals for factor variables are shown as box plots and only the linear part of the cubic splines are specified on the x -axis for soak time variable. Observer data from EAG for 1995/96-2004/05 (top) and 2005/06-2014/15 (bottom) periods were used. The solid green lines are the loess smoother through the plotted values.


Figure B.3. Studentized residual plots for negative binomial GLM fit for EAG golden king crab observer legal size male crab CPUE data. Top panel is for 1995/96-2004/05 data and the bottom panel is for 2005/06-2014/15 data.


Figure B.4. CDI plot for Captain for the negative binomial fit of 1995/96-2004/05 data for EAG.


Figure B.5. CDI plot for Gear for the negative binomial fit of 1995/96-2004/05 data for EAG.


Figure B.6. Influence plot of predictor variables in the final model for the negative binomial fit of 1995/96-2004/05 data for EAG.


Figure B.7. CDI plot for Captain for the negative binomial fit of 2005/06-2014/15 data for EAG.


Figure B.8. CDI plot for Gear for the negative binomial fit of 2005/06-2014/15 data for EAG.


Figure B.9. Influence plot of predictor variables in the final model for the negative binomial fit of 2005/06-2014/15 data for EAG.


Figure B.10. Trends in non-standardized [arithmetic (nominal)] and standardized (negative binomial GLM) CPUE indices with +/- 2 SE for Aleutian Islands golden king crab from WAG (east of $174^{\circ} \mathrm{W}$ longitude). Top panel: 1995/96-2004/05 observer data and bottom panel: 2005/06-2014/15 observer data. Standardized indices: black line and non-standardized indices: red line.


Figure B.11. Deviance residuals vs. explanatory and response variables of the best negative binomial fit model for legal male crab CPUE. Deviance residuals for factor variables are shown as box plots and only the linear part of the cubic splines are specified on the x -axis for soak time variable. Observer data from WAG for 1995/96-2004/05 (top) and 2005/06-2014/15 (bottom) periods were used. The solid lines are the loess smoother through the plotted values.


Figure B.12. Studentized residual plots for negative binomial GLM fit for WAG golden king crab observer legal size male crab CPUE data. Top panel is for 1995/96-2004/05 and bottom panel is for 2005/06-2014/15 data sets, respectively.


Figure B.13. CDI plot for Captain for the negative binomial fit of 1995/96-2004/05 data for WAG.


Figure B.14. CDI plot for Gear for the negative binomial fit of 1995/96-2004/05 data for WAG.


Figure B.15. Influence plot of predictor variables in the final model for the negative binomial fit of 1995/96-2004/05 data for WAG.


Figure B.16. CDI plot for Captain for the negative binomial fit of 2005/06-2014/15 data for WAG.


Figure B.17. CDI plot for Gear for the negative binomial fit of 2005/06-2014/15 data for WAG.


Figure B.18. Influence plot of predictor variables in the final model for the negative binomial fit of 2005/06-2014/15 data for WAG.


Figure B.19. Trends in non-standardized [arithmetic (nominal)] and standardized (lognormal GLM) CPUE indices with +/- 2 SE for Aleutian Islands golden king crab from EAG (east of $174^{\circ}$ W longitude). The 1985/86-1998/99 fish ticket data set was used. Standardized indices: black line and non-standardized indices: red line.

## Log Normal Fit, EAG 1985/86-1998/99



Figure B.20. Studentized residual plots for lognormal GLM fit for EAG golden king crab fish ticket CPUE data. The 1985/86-1998/99 fish ticket data set was used.


Figure B.21. Trends in non-standardized [arithmetic (nominal)] and standardized (lognormal GLM) CPUE indices with +/- 2 SE for Aleutian Islands golden king crab from WAG (east of $174^{\circ}$ W longitude). The 1985/86-1998/99 fish ticket data set was used. Standardized indices: black line and non-standardized indices: red line.

Log Normal Fit, WAG 1985/86-1998/99


Figure B.22. Studentized residual plots for lognormal GLM fit for WAG golden king crab fish ticket CPUE data. The 1985/86-1998/99 fish ticket data set was used.

