# Update on Chinook salmon mortality due to bycatch in the EBS pollock fishery 

## Summary

This paper updates past analysis from the North Pacific Fishery Management Council (NMFS/NPFMC 2015, Ianelli and Stram 2015) using recent Chinook salmon prohibited species catch (PSC) and length composition data collected by observers from the pollock fishery along with genetics as summarized in annual Technical Memoranda (e.g., lii et al. 2018). The length data were converted to age composition estimates using the same strata configurations and available age data from the previous analyses. These estimates were then applied to the same model to arrive at adult equivalency (AEQ) estmates (and estimated uncertainty). These were applied to the latest available genetics data (by A and B seasons) to estimate AEQ to regional origins. Finally, these estimates are compared with the recent run-size estimates provided by ADFG. Results indicate that the ratio of AEQ relative to regional run strengths for coastal west Alaska and Yukon river stocks remains low since Amendments 91 and 110 went into effect.

## Methods

A method was developed to estimate how salmon bycatch numbers would propagate to adult equivalent spawning salmon. That is, how many and in what year would the salmon have returned had they not been taken as bycatch. A stochastic "adult equivalence" (AEQ) model was developed, which accounts for sources of uncertainty and allows for estimating the impact on run strengths from selected regions (Ianelli and Stram 2015). The steps in this process are briefly outlined as:

1. Compile statistics on Chinook salmon bycatch by region and season in the pollock fishery including
a. Total bycatch by season and main sector (Table 1)
b. Length and sex composition of the bycatch (sample sizes are shown in Table 2)
c. Date and location
2. Compile available age composition data organized by strata (here historical age-length keys were used for A and B seasons between two main fishing areas of the Bering Sea; Fig. 1).
3. Convert the seasonal and regional length compositions (shown by absolute measurements in Fig. 2 and by proportion in Fig. 3) into age estimates for each year, area, season using the age-length keys from step 2 (Fig. 4 and Fig. 5; Tables 3 and 4).
4. Provide demographic characteristics of Chinook salmon for use in the AEQ model (these include the oceanic survival-at-age and maturity-at-age and were the same values as used in Ianelli and Stram 2015).
5. Update the season-specific genetics information (the "Bayes" estimates were used from Iii et al. (2013, 2015, 2018), Guthrie et al. (2013, 2014, 2016) for the period 2011-2016 (Table 5).
6. Run the AEQ model with these inputs (extending the estimates back to 1994-2017) and compile/summarize results.
7. Compare a subset (where data are available) of the AEQ results against corresponding runstrength estimates.

The model on the reduction in Chinook salmon returns in year $t, A E Q_{t}$, can thus be expressed (without stock specificity) as:

$$
\begin{equation*}
A E Q_{t}=\sum_{a=3}^{7} c_{t, a} \gamma_{a}+\sum_{j=3}^{6} \sum_{a=j+1}^{7}\left[\gamma_{a} c_{t-(a-j), j} \prod_{i=j}^{a-1}\left(1-\gamma_{i}\right) s_{i}\right] \tag{1}
\end{equation*}
$$

where $c_{t, a}$ is the bycatch of age $a$ salmon in year $t, S_{a}$ is the proportion of salmon surviving from age $a$ to $a+1$, and $\gamma_{a}$ is the proportion of salmon at sea that would have returned to spawn at age $a$. In words, the first term to the right of the equals sign is simply the number of mature Chinook salmon in the bycatch in the current year whereas the second term accounts for the Chinook salmon caught in previous years that would have been mature in the current year. All age 7 Chinook salmon in the bycatch were assumed to be returning to spawn in the year they were caught (i.e. $\gamma_{7}=1$ ) and they represent the oldest fish in the model. We assume that 7 year-old Chinook salmon taken in the fall were returning to spawn that year. In fact, these fish would have been more likely to return the following year. This assumption simplified the model and data preparation. Also, relatively few fish this age were caught late in the season.

Given estimates of AEQ, the model partitions these into regional stock groups (RSG) groups. This was done by assigning the stratum-specific AEQ estimates to each of the nine identified RSGs (e.g., Table 5; Guthrie et al., 2013 for RSG and genetic stock identification, or GSI, determinations). We assumed that given the number of samples used for GSI within each year $(t)$ and stratum $(i)$ that the numbers assigned to RSG $k$ can be assumed to follow a multinomial distribution with parameters

$$
\begin{equation*}
p_{t, i, 1}, \quad, p_{t, i, 9} \quad \sum_{k} p_{t, i, k}=1 \tag{2}
\end{equation*}
$$

For the years where GSI information is missing (data from 1994-2006 and 2017 which are absent from Table 6), the estimated proportions by RSGs were based on mean stratum-specific values from the years when GSI data were available. These additional parameters were constrained based on the estimated within-stratum inter-annual variability. That is, if the proportions assigned to RSGs varied as estimated from the genetics data, then that variability was propagated to the years when genetic data were unavailable. This was a compromise which acknowledges sampling uncertainty for those years and correctly weights the information (due to sample size) between years when GSI information was available. For example, the new observer data collection system for genetic samples has resulted in more precise estimates of GSI in recent years hence those years have greater influence on stratum-specific GSI results. Adjusting the AEQ for RSG requires estimation over a range of years when GSI results are available. This was accomplished here by applying the appropriate GSI results (i.e. estimates of proportions within RSGs) for the years as lagged by AEQ. This step is needed to apportion the AEQ results to stock of origin based on genetic samples which consist of mature and immature fish. By splitting the AEQ estimates to relative contributions of bycatch from previous years, and applying GSI data from those years, they can then be realigned and renormalized to get proportions from systems by year. For years in which GSI information was unavailable, mean GSI data (with an error term which accounted for year-effect variability) were used.
Since Chinook salmon bycatch occurs in both the A and B season of the pollock fishery, data from these seasons were modeled separately. For each separate run, Monte-Carlo Markov Chain (MCMC) samples from the posterior distribution were obtained based on chain lengths of 1 million (after burn in) and selecting every $200^{\text {th }}$ parameter draw. Output resulted in 5,000 samples from each season (summed over strata) and then summed to get annual AEQ totals by RSG. The model was implemented using ADMB (Fournier et al., 2012) software.

Separate estimates of run-strengths (from 1994-2017) were used assuming uncertainties in run size:
$\dot{S}_{t, k}=S_{t, k} e^{\varepsilon_{t}} \quad \varepsilon_{t} \sim N\left(0, \sigma_{S}^{2}\right)$
where $\sigma_{s}^{2}$ was a pre-specified level of run-size variance (assumed to correspond to a conservative coefficient of variation of $10 \%$ for this study). The measure that relates the historical bycatch levels to the subsequent returning salmon run $k$ in year $t$, the "impact", is thus:

$$
\begin{equation*}
u_{t, k}=\frac{A E Q_{t, k}}{A E Q_{t, k}+\dot{S}_{t, k}} \tag{4}
\end{equation*}
$$

where $A E Q_{t, k}$ and $S_{t, k}^{\&}$ are the adult-equivalent bycatch and stock size (run return) estimates. The calculation of $A E Q_{t, k}$ includes the bycatch of salmon returning to spawn in year $t$ and the bycatch from previous years for the same brood year (i.e., at younger, immature ages). Note that the allocation of the AEQ to RSGs is necessarily independent of the age composition of the bycatch.

## Results and conclusions

Results of the AEQ overall were similar to past analyses which evaluated data through 2012 (Fig. 6). Adding in the stock identification data shows the AEQs broken out by regional stock groups are quite similar over time (Fig. 7 and 8).

This analysis was a relatively straightforward update from the work used for Amendment 110 (NMFS/NPFMC 2015). This work was done without testing alternative stratifications of data nor incorporating any new age data that may be available from the bycatch. The key new pieces were the stock identification results (by season) and the total bycatch numbers, and corresponding length frequencies. The updated bycatch numbers remain low relative to the 2005-2007 period (Fig 6). However, there appears to be a slight increasing trend since 2013 (Fig. 6), similar to the trend seen in updated run strengths (Fig. 7). The run strengths were quite similar to previous estimates and combined with the AEQ estimates, results suggest that the impact rate has remained low but there appears to be a slight upturn for the 2017 bycatch levels.

The extent that sampling levels could be improved might be worth re-examining (Faunce 2015). For example, increasing the length composition sample size might provide added insights on biological patterns in the bycatch (present practice of only measuring fish from which genetics samples are taken has greatly reduced these data collections). However, given that observer efforts are fully prescribed, some tradeoff analyses might be warranted. This analysis could also be improved by obtaining more contemporary length-at-age data for the age-length keys as there is some evidence of changes in growth over time (Ohlberger et al. 2018).

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## Tables

Table 1. Chinook salmon bycatch in the pollock fishery by season (A and B), area (NW=west of $170^{\circ} \mathrm{W}$; SE=east of $170^{\circ} \mathrm{W}$ ), and sector ( $\mathrm{CV}=$ shorebased catcher vessels, At sea means mothership operations, catcher-processors, and CDQ). Source: NMFS Alaska Region, Juneau.

| Sector | A Season |  |  | B Season |  |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | CV | At sea | A sub-total | CV | At sea | B sub-total |  |
| 1991 | 10,192 | 26,646 | 36,838 | 1,667 | 548 | 2,215 | 39,053 |
| 1992 | 6,725 | 16,688 | 23,413 | 1,604 | 8,654 | 10,258 | 33,671 |
| 1993 | 3,017 | 12,398 | 15,415 | 2,614 | 18,590 | 21,204 | 36,619 |
| 1994 | 8,346 | 18,939 | 27,285 | 1,206 | 3,399 | 4,605 | 31,890 |
| 1995 | 2,040 | 6,942 | 8,982 | 781 | 3,640 | 4,421 | 13,403 |
| 1996 | 15,228 | 20,757 | 35,985 | 9,944 | 9,544 | 19,488 | 55,473 |
| 1997 | 4,954 | 5,393 | 10,347 | 22,551 | 11,423 | 33,974 | 44,321 |
| 1998 | 4,334 | 10,784 | 15,118 | 27,218 | 8,909 | 36,127 | 51,245 |
| 1999 | 3,103 | 3,248 | 6,351 | 2,662 | 2,964 | 5,626 | 11,977 |
| 2000 | 878 | 2,544 | 3,422 | 718 | 821 | 1,539 | 4,961 |
| 2001 | 8,555 | 9,928 | 18,483 | 3,779 | 11,182 | 14,961 | 33,444 |
| 2002 | 10,336 | 11,457 | 21,793 | 9,560 | 3,141 | 12,701 | 34,494 |
| 2003 | 15,367 | 17,242 | 32,609 | 6,998 | 5,979 | 12,977 | 45,586 |
| 2004 | 11,576 | 11,529 | 23,105 | 22,231 | 6,364 | 28,595 | 51,700 |
| 2005 | 13,797 | 13,491 | 27,288 | 34,826 | 5,204 | 40,030 | 67,318 |
| 2006 | 35,638 | 22,653 | 58,291 | 22,648 | 1,731 | 24,379 | 82,670 |
| 2007 | 36,463 | 33,770 | 70,233 | 41,338 | 10,680 | 52,018 | 122,251 |
| 2008 | 10,692 | 5,823 | 16,515 | 4,245 | 588 | 4,833 | 21,348 |
| 2009 | 6,241 | 3,643 | 9,884 | 2,207 | 485 | 2,692 | 12,576 |
| 2010 | 3,735 | 3,894 | 7,629 | 1,932 | 135 | 2,067 | 9,696 |
| 2011 | 4,442 | 2,695 | 7,137 | 13,950 | 4,412 | 18,362 | 25,499 |
| 2012 | 7,988 | 3,148 | 11,136 | 9,955 | 146 | 10,101 | 21,237 |
| 2013 | 6,592 | 4,595 | 11,187 | 4,105 | 542 | 4,647 | 15,834 |
| 2014 | 6,420 | 5,116 | 11,536 | 2,712 | 783 | 3,495 | 15,031 |
| 2015 | 7,789 | 4,522 | 12,311 | 2,492 | 3,180 | 5,672 | 17,983 |
| 2016 | 8,040 | 8,776 | 16,816 | 1,984 | 3,117 | 5,101 | 21,917 |
| 2017 | 9,057 | 12,538 | 21,595 | 5,991 | 2,339 | 8,330 | 29,925 |
| 2018 | 2,682 | 3,825 | 6,507 | 0 | 0 | 0 | 6,507 |

Table 2. The number of Chinook salmon measured for lengths in the pollock fishery by season (A and B), and area ((NW=west of $170^{\circ} \mathrm{W}$; $\mathrm{SE}=$ east of $\left.170^{\circ} \mathrm{W}\right)$. Source: NMFS Alaska Fisheries Science Center observer data.

| Season | A | B |  |  |
| ---: | ---: | ---: | ---: | ---: |
| Area | All | NW | SE | Total |
| 1991 | 5,098 | 112 | 278 | 5,488 |
| 1992 | 3,927 | 20 | 2,008 | 5,955 |
| 1993 | 2,648 | 184 | 1,230 | 4,062 |
| 1994 | 6,150 | 372 | 1,259 | 7,781 |
| 1995 | 2,324 | 39 | 1,009 | 3,372 |
| 1996 | 13,221 | 178 | 7,872 | 21,271 |
| 1997 | 4,831 | 1154 | 12,625 | 18,610 |
| 1998 | 4,904 | 229 | 12,000 | 17,133 |
| 1999 | 3,127 | 628 | 2,067 | 5,822 |
| 2000 | 2,013 | 223 | 687 | 2,923 |
| 2001 | 8,211 | 1841 | 3,555 | 13,607 |
| 2002 | 9,448 | 137 | 6,367 | 15,952 |
| 2003 | 15,707 | 1838 | 5,235 | 22,780 |
| 2004 | 11,355 | 3062 | 9,627 | 24,044 |
| 2005 | 13,929 | 5935 | 10,460 | 30,324 |
| 2006 | 25,165 | 1231 | 12,842 | 39,238 |
| 2007 | 26,822 | 3457 | 22,528 | 52,807 |
| 2008 | 7,294 | 344 | 2,193 | 9,831 |
| 2009 | 4,969 | 249 | 1,159 | 6,377 |
| 2010 | 3,485 | 137 | 828 | 4,450 |
| 2011 | 751 | 163 | 1,720 | 2,634 |
| 2012 | 820 | 17 | 359 | 1,196 |
| 2013 | 850 | 57 | 458 | 1,365 |
| 2014 | 1,193 | 68 | 292 | 1,553 |
| 2015 | 1,333 | 288 | 341 | 1,962 |
| 2016 | 1,877 | 83 | 448 | 2,408 |
| 2017 | 2,337 | 57 | 701 | 3,095 |

Table 3. Age specific Chinook salmon mean bycatch estimates by season and calendar age based on the mean of 1000 bootstrap samples of available length and age data, 1991-2006. Note that totals may differ from official totals due to random variability of the bootstrap sampling procedure.

| Year/season | Age 3 | Age 4 | Age 5 | Age 6 | Age 7 | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1991 | 5,624 | 15,901 | 13,486 | 3,445 | 347 | 38,802 |
| A | 5,406 | 14,764 | 12,841 | 3,270 | 313 | 36,593 |
| B | 218 | 1,137 | 646 | 174 | 34 | 2,209 |
| 1992 | 5,136 | 9,528 | 14,538 | 3,972 | 421 | 33,596 |
| A | 1,017 | 4,633 | 13,498 | 3,798 | 408 | 23,355 |
| B | 4,119 | 4,895 | 1,040 | 174 | 13 | 10,241 |
| 1993 | 2,815 | 16,565 | 12,992 | 3,673 | 401 | 36,446 |
| A | 1,248 | 3,654 | 7,397 | 2,778 | 290 | 15,368 |
| B | 1,567 | 12,910 | 5,595 | 895 | 111 | 21,078 |
| 1994 | 849 | 5,300 | 20,533 | 4,744 | 392 | 31,817 |
| A | 436 | 3,519 | 18,726 | 4,211 | 326 | 27,218 |
| B | 413 | 1,781 | 1,807 | 533 | 66 | 4,599 |
| 1995 | 498 | 3,895 | 4,827 | 3,796 | 367 | 13,382 |
| A | 262 | 1,009 | 3,838 | 3,534 | 327 | 8,969 |
| B | 236 | 2,885 | 989 | 263 | 40 | 4,413 |
| 1996 | 5,091 | 18,590 | 26,202 | 5,062 | 421 | 55,366 |
| A | 863 | 7,187 | 23,118 | 4,431 | 349 | 35,947 |
| B | 4,228 | 11,403 | 3,085 | 632 | 71 | 19,418 |
| 1997 | 5,855 | 23,972 | 7,233 | 5,710 | 397 | 43,167 |
| A | 456 | 2,013 | 3,595 | 3,899 | 271 | 10,234 |
| B | 5,399 | 21,958 | 3,638 | 1,811 | 126 | 32,933 |
| 1998 | 19,168 | 16,169 | 11,751 | 2,514 | 615 | 50,216 |
| A | 1,466 | 2,254 | 8,639 | 2,079 | 512 | 14,950 |
| B | 17,703 | 13,915 | 3,112 | 435 | 103 | 35,266 |
| 1999 | 870 | 5,343 | 4,424 | 1,098 | 21 | 11,757 |
| A | 511 | 1,639 | 3,151 | 898 | 18 | 6,217 |
| B | 360 | 3,704 | 1,272 | 200 | 3 | 5,540 |
| 2000 | 662 | 1,923 | 1,800 | 518 | 34 | 4,939 |
| A | 365 | 1,167 | 1,406 | 453 | 26 | 3,416 |
| B | 298 | 757 | 395 | 66 | 8 | 1,522 |
| 2001 | 6,512 | 12,365 | 11,948 | 1,994 | 190 | 33,009 |
| A | 2,840 | 3,458 | 9,831 | 1,798 | 171 | 18,098 |
| B | 3,672 | 8,907 | 2,117 | 196 | 19 | 14,910 |
| 2002 | 3,843 | 13,893 | 10,655 | 5,469 | 489 | 34,349 |
| A | 1,580 | 5,063 | 9,234 | 5,328 | 478 | 21,683 |
| B | 2,263 | 8,830 | 1,421 | 141 | 11 | 12,666 |
| 2003 | 5,703 | 16,723 | 20,124 | 3,791 | 298 | 46,639 |
| A | 2,941 | 9,408 | 17,411 | 3,437 | 267 | 33,464 |
| B | 2,763 | 7,315 | 2,713 | 354 | 31 | 13,175 |
| 2004 | 6,935 | 23,740 | 18,371 | 4,406 | 405 | 53,858 |
| A | 1,111 | 5,520 | 13,090 | 3,763 | 354 | 23,838 |
| B | 5,824 | 18,220 | 5,282 | 643 | 51 | 30,020 |
| 2005 | 10,466 | 30,717 | 21,886 | 4,339 | 304 | 67,711 |
| A | 1,407 | 6,993 | 15,563 | 3,361 | 226 | 27,550 |
| B | 9,059 | 23,724 | 6,323 | 978 | 78 | 40,161 |
| 2006 | 11,835 | 31,455 | 32,452 | 6,636 | 490 | 82,869 |
| A | 3,604 | 17,574 | 30,447 | 6,404 | 465 | 58,494 |
| B | 8,231 | 13,881 | 2,005 | 232 | 25 | 24,374 |

Table 4. Age specific Chinook salmon mean bycatch estimates by season and calendar age based on the mean of 1000 bootstrap samples of available length and age data, 2007-2017. Note that totals may differ from official totals due to random variability of the bootstrap sampling procedure.

| Year/season | Age 3 | Age 4 | Age 5 | Age 6 | Age 7 | Total |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{2 0 0 7}$ | $\mathbf{1 6 , 1 7 4}$ | $\mathbf{6 6 , 0 2 4}$ | $\mathbf{3 3 , 2 8 6}$ | $\mathbf{5 , 5 7 9}$ | $\mathbf{3 5 7}$ | $\mathbf{1 2 1 , 4 1 9}$ |
| A | 5,791 | 29,269 | 28,648 | 5,059 | 317 | 69,084 |
| B | 10,384 | 36,755 | 4,638 | 520 | 40 | 52,336 |
| $\mathbf{2 0 0 8}$ | $\mathbf{6 7 0}$ | $\mathbf{4 , 6 9 1}$ | $\mathbf{1 1 , 0 9 8}$ | $\mathbf{1 , 9 4 3}$ | $\mathbf{1 5 6}$ | $\mathbf{1 8 , 5 5 8}$ |
| A | 159 | 2,936 | 10,647 | 1,913 | 154 | 15,808 |
| B | 511 | 1,755 | 451 | 30 | 2 | 2,749 |
| $\mathbf{2 0 0 9}$ | $\mathbf{3 2 8}$ | $\mathbf{3 , 7 0 6}$ | $\mathbf{5 , 6 5 4}$ | $\mathbf{1 , 3 9 2}$ | $\mathbf{1 1 1}$ | $\mathbf{1 1 , 1 9 0}$ |
| A | 94 | 2,768 | 5,453 | 1,377 | 110 | 9,802 |
| B | 234 | 938 | 200 | 14 | 2 | 1,389 |
| $\mathbf{2 0 1 0}$ | $\mathbf{3 8 8}$ | $\mathbf{2 , 1 1 5}$ | $\mathbf{4 , 8 5 2}$ | $\mathbf{1 , 2 1 7}$ | $\mathbf{9 4}$ | $\mathbf{8 , 6 6 6}$ |
| A | 160 | 1,417 | 4,733 | 1,209 | 93 | 7,612 |
| B | 227 | 698 | 119 | 8 | 1 | 1,054 |
| $\mathbf{2 0 1 1}$ | $\mathbf{3 , 2 7 1}$ | $\mathbf{7 , 6 9 7}$ | $\mathbf{4 , 4 0 5}$ | $\mathbf{9 4 6}$ | $\mathbf{8 2}$ | $\mathbf{1 6 , 4 0 1}$ |
| A | 141 | 2,242 | 3,888 | 927 | 80 | 7,277 |
| B | 3,130 | 5,455 | 518 | 19 | 1 | 9,124 |
| $\mathbf{2 0 1 2}$ | $\mathbf{8 5 7}$ | $\mathbf{3 , 0 7 5}$ | $\mathbf{4 , 6 0 4}$ | $\mathbf{9 4 9}$ | $\mathbf{8 6}$ | $\mathbf{9 , 5 7 0}$ |
| A | 87 | 2,166 | 4,502 | 946 | 85 | 7,785 |
| B | 770 | 908 | 102 | 3 | 1 | 1,785 |
| $\mathbf{2 0 1 3}$ | $\mathbf{1 , 2 2 6}$ | $\mathbf{5 , 1 2 4}$ | $\mathbf{3 , 4 9 9}$ | $\mathbf{6 6 8}$ | $\mathbf{5 8}$ | $\mathbf{1 0 , 5 7 5}$ |
| A | 465 | 3,605 | 3,327 | 662 | 58 | 8,116 |
| B | 761 | 1,519 | 172 | 6 | 0 | 2,459 |
| $\mathbf{2 0 1 4}$ | $\mathbf{1 , 2 8 9}$ | $\mathbf{5 , 3 5 9}$ | $\mathbf{5 , 8 8 2}$ | $\mathbf{6 9 8}$ | $\mathbf{5 8}$ | $\mathbf{1 3 , 2 8 5}$ |
| A | 941 | 4,157 | 5,699 | 689 | 57 | 11,543 |
| B | 347 | 1,202 | 183 | 9 | 0 | 1,742 |
| $\mathbf{2 0 1 5}$ | $\mathbf{1 , 8 5 0}$ | $\mathbf{7 , 8 5 9}$ | $\mathbf{4 , 6 9 7}$ | $\mathbf{9 2 5}$ | $\mathbf{6 1}$ | $\mathbf{1 5 , 3 9 1}$ |
| A | 1,241 | 5,811 | 4,423 | 917 | 60 | 12,453 |
| B | 608 | 2,048 | 273 | 8 | 0 | 2,938 |
| $\mathbf{2 0 1 6}$ | $\mathbf{1 , 5 3 8}$ | $\mathbf{9 , 0 6 3}$ | $\mathbf{7 , 8 2 1}$ | $\mathbf{1 , 0 1 0}$ | $\mathbf{7 7}$ | $\mathbf{1 9 , 5 0 8}$ |
| A | 884 | 7,446 | 7,679 | 1,005 | 76 | 17,091 |
| B | 654 | 1,617 | 142 | 5 | 0 | 2,418 |
| $\mathbf{2 0 1 7}$ | $\mathbf{2 , 4 0 6}$ | $\mathbf{1 0 , 1 3 5}$ | $\mathbf{1 1 , 7 8 7}$ | $\mathbf{1 , 9 1 7}$ | $\mathbf{1 4 8}$ | $\mathbf{2 6 , 3 9 3}$ |
| A | 872 | 8,144 | 11,638 | 1,909 | 147 | 22,710 |
| B | 1,534 | 1,991 | 149 | 8 | 1 | 3,683 |
|  |  |  |  |  |  |  |

Table 5. The stock composition estimates (using the "Bayes" estimates) as presented in Auke Bay Lab publications on Chinook salmon bycatch by season (Iii et al. 2013, 2015, 2018, Guthrie et al. 2013, 2014, 2016), 2011-2016.

|  |  | $\begin{aligned} & v \\ & u \\ & 3 \\ & \overrightarrow{0} \\ & 0 \end{aligned}$ | Z 0 0 0 0 0 | $\begin{aligned} & \text { ह } \\ & \text { y } \\ & \text { n } \\ & \text { n } \end{aligned}$ | $\begin{aligned} & \text { E } \\ & \text { d } \\ & \text { un } \\ & \text { z } \end{aligned}$ | $\begin{aligned} & \text { B } \\ & 0 \\ & 3 \\ & Z \end{aligned}$ | $\begin{aligned} & \dot{0} \\ & \stackrel{0}{2} \\ & 0 \end{aligned}$ | $\begin{aligned} & \mathbb{Z} \\ & 0 \\ & 0 \\ & \text { y } \end{aligned}$ | $\begin{aligned} & v \\ & u \\ & w \\ & \tilde{N} \\ & \tilde{\sim} \\ & 0 \\ & 0 \end{aligned}$ | O | $\begin{aligned} & \pi \\ & u \\ & \underset{\sim}{2} \\ & 0 \\ & < \\ & 3 \end{aligned}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| A season |  |  |  |  |  |  |  |  |  |  | \% in A season |  |
| 2011 | 0.2\% | 53.9\% | 1.8\% | 7.4\% | 21.8\% | 0.6\% | 0.0\% | 0.0\% | 3.1\% | 7.2\% | 4.0\% | 28\% |
| 2012 | 0.5\% | 67.8\% | 1.2\% | 3.1\% | 16.2\% | 0.2\% | 0.0\% | 0.1\% | 1.7\% | 7.3\% | 1.9\% | 68\% |
| 2013 | 0.9\% | 50.2\% | 1.1\% | 7.2\% | 19.1\% | 0.5\% | 0.1\% | 0.0\% | 1.9\% | 17.0\% | 2.0\% | 63\% |
| 2014 | 0.6\% | 54.7\% | 3.3\% | 4.1\% | 22.7\% | 0.1\% | 0.0\% | 0.0\% | 0.6\% | 10.2\% | 3.7\% | 77\% |
| 2015 | 0.6\% | 45.9\% | 1.0\% | 3.6\% | 14.5\% | 2.8\% | 0.2\% | 0.0\% | 3.9\% | 19.1\% | 8.4\% | 67\% |
| 2016 | 0.6\% | 39.0\% | 1.7\% | 2.2\% | 16.9\% | 0.6\% | 0.0\% | 0.0\% | 3.9\% | 26.1\% | 8.9\% | 77\% |
| B season |  |  |  |  |  |  |  |  |  |  | \% in B season |  |
| 2011 | 1.0\% | 73.7\% | 1.3\% | 0.7\% | 3.4\% | 3.6\% | 0.6\% | 0.1\% | 1.4\% | 7.8\% | 6.4\% | 72\% |
| 2012 | 2.4\% | 51.9\% | 0.2\% | 1.0\% | 0.1\% | 3.8\% | 0.1\% | 0.1\% | 8.2\% | 15.3\% | 16.9\% | 32\% |
| 2013 | 0.9\% | 51.9\% | 1.9\% | 1.4\% | 5.9\% | 6.9\% | 0.1\% | 0.0\% | 1.9\% | 14.3\% | 14.8\% | 37\% |
| 2014 | 0.4\% | 31.7\% | 1.7\% | 1.6\% | 0.1\% | 18.4\% | 0.1\% | 0.1\% | 3.6\% | 24.5\% | 17.9\% | 23\% |
| 2015 | 0.5\% | 39.6\% | 1.7\% | 2.7\% | 10.6\% | 4.0\% | 0.1\% | 0.0\% | 4.5\% | 21.8\% | 14.5\% | 33\% |
| 2016 | 0.2\% | 16.5\% | 0.4\% | 0.7\% | 1.1\% | 5.8\% | 1.8\% | 0.0\% | 6.5\% | 37.0\% | 29.9\% | 23\% |
| Total |  |  |  |  |  |  |  |  |  |  |  |  |
| 2011 | 0.8\% | 68.2\% | 1.4\% | 2.6\% | 8.5\% | 2.8\% | 0.4\% | 0.1\% | 1.9\% | 7.6\% | 5.7\% |  |
| 2012 | 1.1\% | 62.8\% | 0.9\% | 2.4\% | 11.1\% | 1.3\% | 0.0\% | 0.1\% | 3.7\% | 9.8\% | 6.7\% |  |
| 2013 | 0.9\% | 50.8\% | 1.4\% | 5.1\% | 14.2\% | 2.9\% | 0.1\% | 0.0\% | 1.9\% | 16.0\% | 6.7\% |  |
| 2014 | 0.6\% | 49.3\% | 2.9\% | 3.5\% | 17.5\% | 4.4\% | 0.0\% | 0.0\% | 1.3\% | 13.5\% | 7.0\% |  |
| 2015 | 0.6\% | 43.8\% | 1.2\% | 3.3\% | 13.2\% | 3.2\% | 0.2\% | 0.0\% | 4.1\% | 20.0\% | 10.4\% |  |
| 2016 | 0.5\% | 33.8\% | 1.4\% | 1.8\% | 13.2\% | 1.8\% | 0.4\% | 0.0\% | 4.5\% | 28.7\% | 13.8\% |  |

Table 6. Mean values of stochastic simulation results of AEQ Chinook mortality attributed to the pollock fishery by region, 1994-2017. These simulations include stochasticity in natural mortality ( $\mathrm{CV}=0.1$ ), bycatch age composition (via bootstrap samples), maturation rate ( $\mathrm{CV}=0.1$ ), and stock composition (as detailed above). NOTE: these results are based on the assumption that the genetics findings from the 2005-2016 data represent the historical pattern of bycatch stock composition (by strata). Italicized column is the sum of the western Alaska stocks AEQ estimate.

|  | BC- | Coast <br> W AK | Middle <br> Yukon | Upper <br> Yukon | Combined <br> West. AK | N AK <br> Penin | NW <br> GOA | Russia | SEAK | Other | Total |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1994 | 8,211 | 16,681 | 510 | 1,067 | 18,258 | 4,507 | 835 | 219 | 972 | 14 | 33,016 |
| 1995 | 4,983 | 10,367 | 320 | 687 | 11,374 | 2,916 | 475 | 133 | 596 | 8 | 20,485 |
| 1996 | 7,151 | 15,406 | 483 | 1,075 | 16,964 | 4,585 | 612 | 191 | 869 | 9 | 30,381 |
| 1997 | 8,499 | 15,998 | 472 | 893 | 17,363 | 3,707 | 1,032 | 227 | 973 | 20 | 31,821 |
| 1998 | 8,946 | 15,634 | 443 | 740 | 16,817 | 2,993 | 1,246 | 238 | 994 | 25 | 31,259 |
| 1999 | 6,884 | 11,643 | 324 | 505 | 12,472 | 2,011 | 1,010 | 183 | 754 | 21 | 23,335 |
| 2000 | 4,176 | 6,969 | 192 | 291 | 7,452 | 1,149 | 625 | 111 | 455 | 13 | 13,981 |
| 2001 | 4,382 | 8,636 | 260 | 525 | 9,421 | 2,204 | 481 | 117 | 512 | 8 | 17,125 |
| 2002 | 6,173 | 12,764 | 393 | 839 | 13,996 | 3,555 | 599 | 165 | 737 | 10 | 25,235 |
| 2003 | 8,027 | 16,466 | 505 | 1,069 | 18,040 | 4,524 | 796 | 214 | 955 | 13 | 32,569 |
| 2004 | 9,479 | 18,839 | 570 | 1,162 | 20,571 | 4,885 | 1,020 | 253 | 1,112 | 19 | 37,339 |
| 2005 | 11,794 | 22,156 | 652 | 1,232 | 24,040 | 5,109 | 1,439 | 314 | 1,350 | 28 | 44,074 |
| 2006 | 15,090 | 30,213 | 917 | 1,886 | 33,016 | 7,941 | 1,595 | 403 | 1,775 | 29 | 59,849 |
| 2007 | 18,092 | 36,543 | 1,114 | 2,315 | 39,972 | 9,766 | 1,870 | 482 | 2,137 | 33 | 72,352 |
| 2008 | 15,078 | 29,362 | 879 | 1,748 | 31,989 | 7,311 | 1,704 | 402 | 1,752 | 32 | 58,268 |
| 2009 | 7,706 | 14,873 | 443 | 871 | 16,187 | 3,638 | 888 | 205 | 892 | 16 | 29,532 |
| 2010 | 2,909 | 6,061 | 187 | 403 | 6,651 | 1,710 | 276 | 78 | 348 | 5 | 11,977 |
| 2011 | 2,276 | 5,798 | 160 | 337 | 6,295 | 1,387 | 275 | 72 | 306 | 5 | 10,616 |
| 2012 | 2,642 | 7,460 | 189 | 354 | 8,003 | 1,417 | 377 | 92 | 370 | 8 | 12,909 |
| 2013 | 2,570 | 6,396 | 171 | 337 | 6,904 | 1,324 | 340 | 83 | 336 | 7 | 11,564 |
| 2014 | 2,743 | 5,941 | 189 | 397 | 6,527 | 1,718 | 297 | 78 | 335 | 5 | 11,703 |
| 2015 | 3,135 | 6,187 | 201 | 429 | 6,817 | 1,811 | 299 | 81 | 370 | 4 | 12,517 |
| 2016 | 4,509 | 7,147 | 251 | 529 | 7,927 | 2,336 | 323 | 99 | 479 | 5 | 15,678 |
| 2017 | 5,551 | 10,292 | 348 | 779 | 11,419 | 3,398 | 372 | 131 | 627 | 5 | 21,503 |

## Figures



Fig. 1. Summary distribution of age samples by length collected by the NMFS groundfish observer program during 1997-1999 and analyzed by University of Washington scientists (Myers et al. 2003) for the A-season (top panel) and B season (bottom panel).

Fig. 2. Length frequency measurements collected by NMFS observers by season and year of Chinook salmon occurring as bycatch in the pollock fishery. This figure indicates the change in sampling intensity for length measurements of Chinook salmon bycatch.


Fig. 3 Length frequency proportions by season and year of Chinook salmon occurring as bycatch in the pollock fishery.


Fig. 4. Chinook salmon bycatch age composition by year and relative age with older (top) and younger (bottom) by estimated age. Vertical spread of blobs represent uncertainty as estimated from the two-stage bootstrap re-sampling procedure.


Fig. 5. Time series of Chinook adult equivalent bycatch from the pollock fishery, 2007-2016 compared to the annual totals under different assumptions about ocean mortality rates.


Fig. 5. (continued) Time series of Chinook adult equivalent bycatch from the pollock fishery, 2007-2016 compared to the annual totals under different assumptions about ocean mortality rates.


Fig. 6. Time series of Chinook salmon adult equivalent bycatch estimates from the pollock fishery, 1994-2017 with constant vertical scale (top set of figures) and where vertical scales vary between stock groupings (bottom set of figures).


Fig. 7. Time series of Chinook salmon run strength estimates for western Alaska (includes coastal west Alaska stocks plus lower and middle Yukon River) and for the Canadian portion of the upper Yukon River, 1994-2017. Source K. Howard ADFG.


Figure 8. Estimated impact of the BS pollock fishery on the Upper Yukon stock (top) and combined west Alaska (which includes the "middle Yukon"; bottom), 1994-2017. Vertical axis is the ratio of AEQ over the point estimates of total run sizes.


Fig. 9. Combined western Alaska Chinook salmon adult equivalent mortality estimates (vertical scale) compared to combined in-river returns (horizontal scale, thousands), 1994-2017. This represesnts the genetic stock ID estimates applied to AEQ for the lower and middle Yukon River plus the "coastal west Alaska" stocks (Kuskokwim, Nushagak, and Norton Sound). Recent years are indicated in red.

