# Saint Matthew Island Blue King Crab Stock Assesssment 2020 

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## Executive Summary

1. Stock: Blue king crab, Paralithodes platypus, Saint Matthew Island (SMBKC), Alaska.
2. Catches: Peak historical harvest was $4,288 \mathrm{t}$ ( 9.454 million pounds) in $1983 / 84^{1}$. The fishery was closed for 10 years after the stock was declared overfished in 1999. Fishing resumed in 2009/10 with a fishery-reported retained catch of 209 t ( 0.461 million pounds), less than half the 529.3 t ( 1.167 million pound) TAC. Following three more years of modest harvests supported by a fishery catch per unit effort (CPUE) of around 10 crab per pot lift, the fishery was again closed in 2013/14 due to declining trawl-survey estimates of abundance and concerns about the health of the stock. The directed fishery resumed again in $2014 / 15$ with a TAC of 300 t ( 0.655 million pounds), but the fishery performance was relatively poor with a retained catch of 140 t ( 0.309 million pounds). The retained catch in 2015/16 was even lower at 48 t ( 0.105 million pounds) and the fishery has remained closed since 2016/17.
3. Stock biomass: The 1978-2019 NMFS trawl survey mean biomass is $5,605 \mathrm{t}$ with the 2019 value being the 15 th lowest ( $3,170 \mathrm{t}$; the tenth lowest since 2000). This 2019 biomass of $\geq 90 \mathrm{~mm}$ carapace length (CL) male crab is $57 \%$ of the long term mean at 6.99 million pounds (with a CV of $34 \%$ ), and an $83 \%$ increase from the 2018 biomass. The most recent 3 -year average of the NMFS survey is $40 \%$ of the mean value, indicating a decline in biomass compared to historical survey estimates, notably in 2010 and 2011 that were over four times the current average. However, the 2019 value is substantially larger than the two previous years ( $3,170 \mathrm{t}$ compared to $1,731 \mathrm{t}$ in 2018 and $1,794 \mathrm{t}$ in 2017). Due to cancellation of the 2020 bottom trawl surveys there is no additional abundance data in the model for 2020. The ADFG pot survey last occured in 2018, when the relative biomass index was the lowest in the time series ( $12 \%$ of the mean from the 11 surveys conducted since 1995). The assessment model estimates temper this increase and suggest that the stock (in survey biomass units) is presently at about $26 \%$ of the long term model-predicted survey biomass average, similar to the last three years. The trend from these values suggests a steady state in the last few years, which does not fit the 2019 observed survey data point well.
4. Recruitment: Recruitment is based on estimated number of male crab within the $90-104 \mathrm{~mm}$ CL size class in each year. The 2019 trawl-survey area-swept estimate of 0.403 million male SMBKC in this size class is the twelfth lowest in the 42 years since 1978 and follows two of the lowest previously observed values in 2017 and 2018. The recent six-year (2014-2019) average recruitment is only $47 \%$ of the long-term mean. In the pot-survey, the abundance of this size group in 2017 was also the second-lowest in the time series ( $22 \%$ of the mean for the available pot-survey data) whereas in 2018 the value was the lowest observed at only $10 \%$ of the mean value.
5. Management performance: In this assessment, estimated total male catch is the sum of fisheryreported retained catch, estimated male discard mortality in the directed fishery, and estimated male

[^0]bycatch mortality in the groundfish fisheries. Based on the reference model for SMBKC, the estimate for mature male biomass was below the minimum stock-size threshold (MSST) in 2018/19 and is in an "overfished" condition, despite a directed fishery closure since the 2016/17 season (and hence overfishing has not occurred) (Tables 1, 3, and 4). Computations which indicate the relative impact of fishing (i.e., the "dynamic $B_{0}$ ") suggests, that the current spawning stock biomass has been reduced to $55 \%$ of what it would have been in the absence of fishing, assuming the same level of recruitment as estimated.

Table 1: Status and catch specifications (1000 t) for the reference model.

| Year | MSST | Biomass <br> $\left(M M B_{\text {mating }}\right)$ | TAC | Retained <br> catch | Total <br> male catch | OFL | ABC |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $2016 / 17$ | 1.97 | 2.23 | 0.00 | 0.00 | 0.001 | 0.14 | 0.11 |
| $2017 / 18$ | 1.85 | 2.05 | 0.00 | 0.00 | 0.003 | 0.12 | 0.10 |
| $2018 / 19$ | 1.74 | 1.15 | 0.00 | 0.00 | 0.001 | 0.04 | 0.03 |
| $2019 / 20$ | 1.67 | 1.06 | 0.00 | 0.00 | 0.001 | 0.04 | 0.03 |
| $2020 / 21$ |  | 1.12 |  |  |  | 0.05 | 0.04 |

Table 2: Status and catch specifications (million pounds) for the reference model.

| Year | MSST | Biomass <br> $\left(M M B_{\text {mating }}\right)$ | TAC | Retained <br> catch | Total <br> male catch | OFL | ABC |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $2016 / 17$ | 4.3 | 4.91 | 0.000 | 0.000 | 0.002 | 0.31 | 0.25 |
| $2017 / 18$ | 4.1 | 2.85 | 0.000 | 0.000 | 0.007 | 0.27 | 0.22 |
| $2018 / 19$ | 3.84 | 2.54 | 0.000 | 0.000 | 0.002 | 0.08 | 0.07 |
| $2019 / 20$ | 3.68 | 2.34 | 0.000 | 0.000 | 0.002 | 0.096 | 0.08 |
| $2020 / 21$ |  | 2.48 |  |  |  | 0.112 | 0.08 |

6. Basis for the OFL: Estimated mature-male biomass (MMB) on 15 February is used as the measure of biomass for this Tier 4 stock, with males measuring $\geq 105 \mathrm{~mm}$ CL considered mature. The $B_{M S Y}$ proxy is obtained by averaging estimated MMB over a specific reference period, and current CPT/SSC guidance recommends using the full assessment time frame (1978-2019) as the default reference period.

Table 3: Basis for the OFL (1000 t) from the reference model.

| Year | Tier | $B_{M S Y}$ | Biomass <br> $\left(M M B_{\text {mating }}\right)$ | $B / B_{M S Y}$ | $F_{O F L}$ |  | Basis for $B_{M S Y}$ | Natural <br> mortality |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $2016 / 17$ | 4 b | 3.67 | 2.23 | 0.61 | 0.09 | 1 | $1978-2016$ | 0.18 |
| $2017 / 18$ | 4 b | 3.86 | 2.05 | 0.53 | 0.08 | 1 | $1978-2017$ | 0.18 |
| $2018 / 19$ | 4 b | 3.7 | 1.15 | 0.35 | 0.043 | 1 | $1978-2017$ | 0.18 |
| $2019 / 20$ | 4 b | 3.48 | 1.06 | 0.31 | 0.042 | 1 | $1978-2018$ | 0.18 |
| $2020 / 21$ | 4 b | 3.34 | 1.12 | 0.34 | 0.047 | 1 | $1978-2019$ | 0.18 |

## A. Summary of Major Changes

## Changes in Management of the Fishery

There are no new changes in management of the fishery.

## Changes to the Input Data

Data used in this assessment have been updated to include the most recently available fishery data. This assessment includes no new survey data points due to the cancellation of the 2020 NMFS trawl-survey. The triennial ADF\&G pot surveys were last conducted in 2018, and are back on a triennial cycle, with the next survey planned for 2021. Due to the lack of bycatch in other crab fisheries and new survey data there is no new size compositon data. The assessment was updated with 2010-2019 groundfish trawl and fixed gear bycatch estimates based on NMFS Alaska Regional Office (AKRO) data. The directed fishery has been closed since 2016/17, so no recent fishery data are available.

## Changes in Assessment Methodology

This assessment uses the General Model for Alaska Crab Stocks (GMACS) framework. The model is configured to track three stages of length categories and was first presented in May 2011 by W.Gaeuman, ADF\&G, and accepted by the CPT in May 2012. A difference from the original approach and that used here is that natural and fishing mortalities are continuous within 5 discrete time blocks within a year (using the appropriate catch equation rather than assuming an applied pulse removal). The time blocks within a year in GMACS are controlled by changing the proportion of natural mortality that is applied each block. Diagnostic output includes estimates of the "dynamic $B_{0}$ " which simply computes the ratio of the estimated spawning biomass relative to the spawning biomass that would have occurred had there been no historical fishing mortality. Details of this implementation and other model details are provided in Appendix A.

## Changes in Assessment Results

Both surveys indicate a decline over the past few years. The "reference" model is that which was selected for use in 2019. The base model presented here is the reference model with updated groundfish bycatch data for the 2019/20 crab season (model 16.0 base). One additional model is presented for consideration, which is a small variant of the base model, model 16.0a (fixR), which fixes recruitment in the most recent year to the average of the last seven years to avoid unrealistically high recruitment estimates. Additionally, retrospective analyses without the terminal year of survey data and runs with "fake" survey data were performed to assess the uncertainty in the 2020 biomass estimates and reference point calculations due to the lack of a 2020 survey; the methods and results are detailed in Appendix C.
In addition to the two models for considerations, one additional model is presented here to assess sensitivity of data inputs to the model, attempting to deal with the disparity between the two survey time series (no pot). The no pot configuration runs the base model 16.0 without the ADF\&G pot survey data, therefore only having the NMFS trawl survey as the abundance index.

## B. Responses to SSC and CPT

## CPT and SSC Comments on Assessments in General

Comment: Regarding general code development, the SSC and CPT outstanding requests continue to be as follows:

1. add the ability to conduct retrospective analyses

Retrospective runs/simulations are presented here in Appendix C as part of the analyses done to assess uncertainty in the model output (Figure 28). The ability to automate these in GMACS is still under developement but the author was able to do them by manually editing the data files.
2. Continued exploration of data weighting (Francis and other approaches) and evaluation of models with and without the 1998 natural mortality spike. The authors are encouraged to bring other models forward for CPT and SSC consideration
We continued with the iterative re-weighting for composition data (Table 16). We did not address models without the natural mortality spike. These have been considered previously.

Comment: Regarding potential model scenarios for Sept. 2020, the SSC and CPT requests are:

1. Explore model without $A D f f G$ pot survey data

Model 20.1 explores this sensitivity to the data inputs and is shown here in the model scenarios.
2. Random walk or exploration of catchability

The intial model of time blocks for Q did not show much potential for this in May 2020, therefore it was not a focus for the Sept. 2020 runs. More coding work is needed to make a true random walk for catchability GMACS and this will be added to GMACS model development, hopefully during the Jan 2021 modeling workshop.

Comment: Explore potential explanations for the discrepancy in the time trends of the two types of survey data, including movement hypotheses using spatial models (not necessarily VAST)
Limited progress due to time availability and current world events. This will be a large focus on upcoming work on this model as the scenario without the ADF\&G pot survey data (20.1) shows the differences in the current status of the stock between the two abundance surveys (Figure 13).

Comment: Explore May 2020 model with VAST estimates
Progress is underway to refine the SMBKC VAST estimates using preliminary code that incorporates the island effect. Jon Richar (NMFS) is working on these estimates. At the time of this final SAFE there are no additional improvements to this data set and therefore the VAST model is not presented as a model option. Future work on VAST models for this stock includes VAST data output for the NMFS trawl survey incorporating the island effect and VAST output using both survey data sets together.
Comment: Please use the correct model number (e.g., if 19.0 is the same model as was first adopted in 16.0 then it is still 16.0.)

Completed. Base model is 16.0.

## C. Introduction

## Scientific Name

The blue king crab is a lithodid crab, Paralithodes platypus (Brant 1850).

## Distribution

Blue king crab are sporadically distributed throughout the North Pacific Ocean from Hokkaido, Japan, to southeastern Alaska (Figure 1). In the eastern Bering Sea small populations are distributed around St. Matthew Island, the Pribilof Islands, St. Lawrence Island, and Nunivak Island. Isolated populations
also exist in some other cold water areas of the Gulf of Alaska (NPFMC 1998). The St. Matthew Island Section for blue king crab is within Area Q2 (Figure 2), which is the Northern District of the Bering Sea king crab registration area and includes the waters north of Cape Newenham ( $58^{\circ} 39^{\prime} \mathrm{N}$. lat.) and south of Cape Romanzof ( $61^{\circ} 49^{\prime}$ N. lat.).

## Stock Structure

The Alaska Department of Fish and Game (ADF\&G) Gene Conservation Laboratory, has detected regional population differences between blue king crab collected from St. Matthew Island and the Pribilof Islands ${ }^{2}$. The NMFS tag-return data from studies on blue king crab in the Pribilof Islands and St. Matthew Island support the idea that legal-sized males do not migrate between the two areas (Otto and Cummiskey 1990). St. Matthew Island blue king crab tend to be smaller than their Pribilof conspecifics, and the two stocks are managed separately.

## Life History

Like the red king crab, Paralithodes camtshaticus, the blue king crab is considered a shallow water species by comparison with other lithodids such as golden king crab, Lithodes aequispinus, and the scarlet king crab, Lithodes couesi (Donaldson and Byersdorfer 2005). Adult male blue king crab are found at an average depth of 70 m (NPFMC 1998). The reproductive cycle appears to be annual for the first two reproductive cycles and biennial thereafter (Jensen and Armstrong 1989), and mature crab seasonally migrate inshore where they molt and mate. Unlike red king crab, juvenile blue king crab do not form pods, but instead rely on cryptic coloration for protection from predators and require suitable habitat such as cobble and shell hash. Somerton and MacIntosh (1983) estimated SMBKC male size at sexual maturity to be 77 mm carapace length (CL). Paul et al. (1991) found that spermatophores were present in the vas deferens of $50 \%$ of the St. Matthew Island blue king crab males examined with sizes of $40-49 \mathrm{~mm}$ CL and in $100 \%$ of the males at least 100 mm CL. Spermataphore diameter also increased with increasing CL with an asymptote at ~ 100 mm CL. It was noted, however, that although spermataphore presence indicates physiological sexual maturity, it may not be an indicator of functional sexual maturity. For purposes of management of the St. Matthew Island blue king crab fishery, the State of Alaska uses 105 mm CL to define the lower size bound of functionally mature males (Pengilly and Schmidt 1995). Otto and Cummiskey (1990) report an average growth increment of 14.1 mm CL for adult SMBKC males.

## Management History

The SMBKC fishery developed subsequent to baseline ecological studies associated with oil exploration (Otto 1990). Ten U.S. vessels harvested 545 t ( 1.202 million pounds) in 1977, and harvests peaked in 1983 when 164 vessels landed $4,288 \mathrm{t}$ ( 9.454 million pounds) (Fitch et al. 2012; Table 7).

The fishing seasons were generally short, often lasting only a few days. The fishery was declared overfished and closed in 1999 when the stock biomass estimate was below the minimum stock-size threshold (MSST) of $4,990 \mathrm{t}$ ( 11.0 million pounds) as defined by the Fishery Management Plan (FMP) for the Bering Sea/Aleutian Islands King and Tanner crabs (NPFMC 1999). Zheng and Kruse (2002) hypothesized a high level of SMBKC natural mortality from 1998 to 1999 as an explanation for the low catch per unit effort (CPUE) in the 1998/99 commercial fishery and the low numbers across all male crab size groups caught in the annual NMFS eastern Bering Sea trawl survey from 1999 to 2005 (see survey data in next section). In November 2000, Amendment 15 to the FMP for Bering Sea/Aleutian Islands king and Tanner crabs was approved to implement a rebuilding plan for the SMBKC stock (NPFMC 2000). The rebuilding plan included a State of Alaska regulatory harvest strategy ( $5 A A C 34.917$ ), area closures, and gear modifications. In addition, commercial crab fisheries near St. Matthew Island were scheduled in fall and early winter to reduce the potential for bycatch mortality of vulnerable molting and mating crab.

[^1]NMFS declared the stock rebuilt on 21 September 2009, and the fishery was reopened after a 10-year closure on 15 October 2009 with a TAC of 529 t ( 1.167 million pounds), closing again by regulation on 1 February 2010. Seven participating vessels landed a catch of 209 t ( 0.461 million pounds) with a reported effort of 10,697 pot lifts and an estimated CPUE of 9.9 retained individual crab per pot lift. The fishery remained open the next three years with modest harvests and similar CPUE, but large declines in the NMFS trawlsurvey estimate of stock abundance raised concerns about the health of the stock. This prompted ADF\&G to close the fishery again for the $2013 / 14$ season. The fishery was reopened for the $2014 / 15$ season with a low TAC of 297 t ( 0.655 million pounds) and in $2015 / 16$ the TAC was further reduced to 186 t ( 0.411 million pounds) then completely closed the 2016/17 season.

Although historical observer data are limited due to low sampling effort, bycatch of female and sublegal male crab from the directed blue king crab fishery off St. Matthew Island was relatively high historically, with estimated total bycatch in terms of number of crab captured sometimes more than twice as high as the catch of legal crab (Moore et al. 2000; ADF\&G Crab Observer Database). Pot-lift sampling by ADF\&G crab observers (Gaeuman 2013; ADF\&G Crab Observer Database) indicates similar bycatch rates of discarded male crab since the reopening of the fishery (Table 5), with total male discard mortality in the 2012/13 directed fishery estimated at about $12 \%$ ( 88 t or 0.193 million pounds) of the reported retained catch weight, assuming $20 \%$ handling mortality.

These data suggest a reduction in the bycatch of females, which may be attributable to the later timing of the contemporary fishery and the more offshore distribution of fishery effort since reopening in 2009/10 ${ }^{3}$. Some bycatch of discarded blue king crab has also been observed historically in the eastern Bering Sea snow crab fishery, but in recent years it has generally been negligible. The St. Matthew Island golden king crab fishery, the third commercial crab fishery to have taken place in the area, typically occurred in areas with depths exceeding blue king crab distribution. The NMFS observer data suggest that variable, but mostly limited, SMBKC bycatch has also occurred in the eastern Bering Sea groundfish fisheries (Table 6).

## D. Data

## Summary of New Information

Data used in this assessment were updated to include the most recently available fishery and survey estimates. The only new data in the 2020 assessment model is updated bycatch estimates, no new survey or size composition data were added. The assessment uses updated 1993-2019 groundfish and fixed gear bycatch estimates based on NMFS AKRO data. The directed fishery has been closed since the 2016/17 season, and therefore no directed fishery catch data are available. The data used in each of the new models is shown in Figure 3.

## Major Data Sources

Major data sources used in this assessment include annual directed-fishery retained-catch statistics from fish tickets (1978/79-1998/99, 2009/10-2012/13, and 2014/15-2015/16; Table 7); results from the annual NMFS eastern Bering Sea trawl survey (1978-2019; Table 8); results from the ADF\&G SMBKC pot survey (every third year during 1995-2013, then 2015-2018; Table 9); mean somatic mass given length category by year (Table 10); size-frequency information from ADF\&G crab-observer pot-lift sampling (1990/91-1998/99, 2009/10-2012/13, and 2014/15-2016/17; Table 5); and the NMFS groundfish-observer bycatch biomass estimates (1992/93-2019/20; Table 6).

Figure 4 maps stations from which SMBKC trawl-survey and pot-survey data were obtained. Further information concerning the NMFS trawl survey as it relates to commercial crab species is available in Daly et al. (2014); see Gish et al. (2012) for a description of ADF\&G SMBKC pot-survey methods. It should be

[^2]noted that the two surveys cover different geographic regions and that each has in some years encountered proportionally large numbers of male blue king crab in areas not covered by the other survey (Figure 5). Crabobserver sampling protocols are detailed in the crab-observer training manual (ADF\&G 2013). Groundfish SMBKC bycatch data come from the NMFS Regional office and have been compiled to coincide with the SMBKC management area.

## Other Data Sources

The growth transition matrix used is based on Otto and Cummiskey (1990), as in the past. Other relevant data sources, including assumed population and fishery parameters, are presented in Appendix A, which also provides a detailed description of the model configuration used for this assessment.

## E. Analytic Approach

## History of Modeling Approaches for this Stock

A four-stage catch-survey-analysis (CSA) assessment model was used before 2011 to estimate abundance and biomass and prescribe fishery quotas for the SMBKC stock. The four-stage CSA is similar to a full length-based analysis, the major difference being coarser length groups, which are more suited to a small stock with consistently low survey catches. In this approach, the abundance of male crab with a CL $\geq 90$ mm is modeled in terms of four crab stages: stage 1: $90-104 \mathrm{~mm}$ CL; stage 2: $105-119 \mathrm{~mm}$ CL; stage 3 : newshell 120-133 mm CL; and stage 4: oldshell $\geq 120 \mathrm{~mm}$ CL and newshell $\geq 134 \mathrm{~mm}$ CL. Motivation for these stage definitions comes from the fact that for management of the SMBKC stock, male crab measuring $\geq 105 \mathrm{~mm}$ CL are considered mature, whereas 120 mm CL is considered a proxy for the legal size of 5.5 in carapace width, including spines. Additional motivation for these stage definitions comes from an estimated average growth increment of about 14 mm per molt for SMBKC (Otto and Cummiskey 1990).
Concerns about the pre-2011 assessment model led to the CPT and SSC recommendations that included development of an alternative model with provisional assessment based on survey biomass or some other index of abundance. An alternative 3-stage model was proposed to the CPT in May 2011, but a survey-based approach was requested for the Fall 2011 assessment. In May 2012 the CPT approved a slightly revised and better documented version of the alternative model for assessment. Subsequently, the model developed and used since 2012 was a variant of the previous four-stage SMBKC CSA model and similar in complexity to that described by Collie et al. (2005). Like the earlier model, it considered only male crab $\geq 90 \mathrm{~mm}$ in CL, but combined stages 3 and 4 of the earlier model, resulting in three stages (male size classes) defined by CL measurements of (1) 90-104 mm, (2) 105-119 mm, and (3) $120 \mathrm{~mm}+$ (i.e., 120 mm and above). This consolidation was driven by concern about the accuracy and consistency of shell-condition information, which had been used in distinguishing stages 3 and 4 of the earlier model.
In 2016 the accepted SMBKC assessment model made use of the modeling framework GMACS encompassing a three-stage model structure (Webber et al. 2016). In that assessment, an effort was made to match the 2015 SMBKC stock assessment model to bridge a framework which provided greater flexibility and opportunity to evaluate model assumptions more fully.

## Assessment Methodology

This assessment model again uses the modeling framework GMACS and is detailed in Appendix A.

## Model Selection and Evaluation

Two models are presented with the reference model being the same configuration as approved last year (Palof et al. 2019), one sensitivity is considered which excludes the ADF\&G pot survey data. In addition to
this sensitivty, we evaluated the impacts of adding new data (here just groundfish bycatch) to the reference model. In summary, the following lists the models presented and the naming convention used:

1. 16.0-2019 Model: 2019 accepted model
2. 16.0-2020 Reference Model: updated with 2019/20 groudfish bycatch
3. 16.0a-2020 Reference Model with fixed terminal year recruitment: terminal year recruitment fixed as the average of the last seven years
4. 20.1 - no ADF\&G pot survey data: model 16.0 - excludes ADF\&G pot survey data - abundace and length comps

Note the change in naming convention (per SSC comments). The base model is model 16.0 since that was the year of model development and acceptance.

## Results

## a. Sensitivity to new data

There is no new survey data for the September 2020 model runs, the only additional data is groundfish bycatch data for the $2019 / 20$ crab season. Additionally, the groundfish bycatch data was updated for past years due to some changes in the weights used to estimate crab bycatch in the groundfish fisheries (per. comm. NMFS AKRO). The 2020 reference model is compared here to the 2019 accepted model, which is shown in Figures 6 and 7 with recruitment and spawning biomass shown in Figures 8 and 9, respectively. The 2019 accepted model and the 2020 base model have identical fits to the survey data, as well as identical estimates of SSB and recruitment. This is expected since there are no new influential data in the 2020 model. As has been noted in the past, the reference model still does not capture the recent survey declines in the ADF\&G pot survey, or fit post 2005 trawl survey data points well.

## b. Effective sample sizes and weighting factors

Observed and estimated effective sample sizes are compared in Table 11. Data weighting factors, standard deviation of normalized residuals (SDNRs), and median absolute residual (MAR) are presented in Table 16. Currently the SDNR and MAR are not outputting correctly for the survey data in GMACS. This is on the list to address at the Januaury 2021 modeling workshop. In Sept. 2019 the SDNR for the trawl survey was acceptable at 1.66 in the reference model. Francis (2011) weighting was applied in 2017 but given the relatively few size bins in this assessment, this application was suspended for this assessment.

In Sept. 2019 the SDNRs for the pot surveys showed a similar pattern in each of the scenarios, but are much higher suggesting an inconsistency between the pot survey data and the model structure and other data components. Rather than re-weighting, we chose to retain the values as specified, noting that downweighting these data would effectively exclude the signal from this series. The MAR values for the trawl and pot surveys showed the same pattern among each of the scenarios as the SDNR. The MAR values for the trawl survey and pot survey size compositions were adequate, ranging from 0.60 to 0.68 for the reference case. The SDNRs for the directed pot fishery and other size compositions were similar to previous estimates.

## c. Parameter estimates

Model parameter estimates for each of the GMACS scenarios are summarized in Tables 12, 13, and 14. These parameter estimates are compared in Table 15. Negative log-likelihood values and management measures for each of the model configurations are compared in Tables 4 and 17.

There are differences in parameter estimates among models as reflected in the log-likelihood components and the management quantities. The parameter estimates in the "no pot" scenario differ greatly from the reference model, as expected, due to the removal of recent ADF\&G pot survey data points that pulled the MMB trend downward (Table 15). Also, the size composition residuals are smaller for the trawl survey in the nopot model, presumably because they are allowed to fit these size compositions better due to the removal of the size composition data from the ADF\&G pot survey.

Selectivity estimates for the directed fishery show some variability between models (Figure 10). Estimated recruitment is similar in both models until the mid-2000s when the no pot model (20.1) has consistently higher recruitment, contributing to higher MMB for this model in recent years (Figure 11). Estimated mature male biomass on 15 February also is considerably higher in the no pot model (Figure 13). The no pot model has a better fit to recent years of the NMFS trawl survey data, fitting most of the post-2010 data ranges (fit line encompasses the error bars), compared to the reference model that only fits three of the last 10 years. The improved fit of the trawl survey corresponds to increased MMB estimates in the last 10 years. Not surprisingly this time frame also corresponds to sharp declines in the ADF\&G pot survey abundance estimates that started in the post-2010 data.
Estimated natural mortality in each year $\left(M_{t}\right)$ is presented in Figure 14, showing the mortality event in the late 90s. Estimates of fishing morality, from the reference model (16.0), are shown to assist with the rebuilding and reference point time frame discussions (Figure 26). Fishing mortality can not be ruled out as being an influential factor in the current stock status.

## d. Evaluation of the fit to the data.

The reference model fit to total male ( $\geq 90 \mathrm{~mm}$ CL) trawl survey biomass tends to miss the recent peak around 2010 and fits recent survey data points on the lower end of their error bars (Figures 15). These fits are most likely being pulled down by the recent decline in the ADF\&G pot survey data points, since the no pot model captures more of the error bars for these data points when the NMFS trawl survey data is the only abundance index in the model. However, this model, similar to the additional CV models presenting in May 2020, tend to overfit the recent trawl survey data points (Figure 15).
The reference or base model fit to the pot survey CPUE is similar to past reference models, fitting the overall trends in the data but not capturing some of the high and low points (Figure 16).

For the trawl survey the standardized residuals are more balanced in model 20.1 (no pot), without the ADF\&G pot survey data, especially in recent years. The reference model has a clear residual pattern in the last 15 years, continually under predicting the observed data points (Figure 17). The standardized residuals for the ADF\&G pot survey have similar patterns to past reference model iterations (Figure 18).

Fits to the size compositions for trawl survey, pot survey, and commercial observer data are reasonable but miss the largest size category in some years (Figures 19, 20, and 21) for both scenarios. Representative residual plots of the composition data generally have a poor fit to the three composition data sources (Figures 22, 23 and 24). The model fits to different types of retained and discarded catch values performed as expected given the assumed levels of uncertainty on the input data (Figure 25).

## e. Retrospective and historical analyses

This is the fourth year GMACS has been used for this stock. As such, retrospective patterns and historical analyses of GMACS assessments are limited. However, completion of a retrospective analysis, for the base model, was completed (Figure 28) and is presented in detail in Appendix C.

## f. Uncertainty and sensitivity analyses.

Estimated standard deviations of parameters and selected management measures for the models are summarized for each individual model in Tables 12, 13, 14, and compiled in Table 15. Model estimates of mature
male biomass and OFL in 2020 are presented in Section F.
Uncertainty surrounding the lack of a 2020 trawl survey data point was examined using two approaches and the results are contained in Appendix C. Overall, the authors did not find much additional uncertainty for the reference model due to the lack of a 2020 data point. The current trajectory of the stock (MMB and recruitment) suggests a low status (below $B_{M S Y}$ ) that would not change even with the addition of hypothetical 2020 data point (Approach 3, Appendix C). Appendix C goes into more detail for these analyses and a more thourough discussion of the authors recommendations.

## g. Comparison of alternative model scenarios.

The estimates of mature male biomass (Figure 13) for the no pot model differs from the reference model (16.0) due to the removal of the pot survey abundance and size composition data. This abundance time series contrasts with the NMFS trawl survey and when present tends to lower the scale of the population estimate. This difference is greatest in the last 10 years, recognizing the contrast between these abundance time series and the influence of the ADF\&G pot survey on the current population status.
In summary, the no pot model scenario was provided to explore the sensitivity of this model. Currently, the reference model is still the most appropriate model for settting reference points and model specifications. Research on alternative model specifications that may address the disparities between the trawl and pot survey data are ongoing, as is proposed spatial analyses of these data sets. Additionally, the overfished status of this stock lends itself to maintaining the status quo base model until an appropriate resolution is found to deal with the trawl and pot survey data fit issues. The two reference models presented here, 16.0 and 16.0a, only differ in the estimation of 2019 recruitment. Model 16.0a fixes the 2019 recruitment to be the average of the last seven years of the model, effectively limiting the model's ability to estimate unreasonably high recruitment in the lack of a 2020 data point. However, fixing terminal year recruitment has a minimal effect on the status of the stock, projected MMB, or the resulting OFL for 2020 (Table 4). The recommended model for 2020 would be the reference model (16.0) to maintain consistency for this stock during the rebuilding time frame and with the lack of a 2020 data point for the trawl survey.

## F. Calculation of the OFL and ABC

The overfishing level (OFL) is the total catch associated with the $F_{O F L}$ fishing mortality. The SMBKC stock is currently managed as Tier 4, and only a Tier 4 analysis is presented here. Thus, given stock estimates or suitable proxy values of $B_{M S Y}$ and $F_{M S Y}$, along with two additional parameters $\alpha$ and $\beta, F_{O F L}$ is determined by the control rule

$$
\begin{align*}
& F_{O F L}= \begin{cases}F_{M S Y}, & \text { when } B / B_{M S Y}>1 \\
F_{M S Y} \frac{\left(B / B_{M S Y}-\alpha\right)}{(1-\alpha)}, & \text { when } \beta<B / B_{M S Y} \leq 1\end{cases}  \tag{1}\\
& F_{O F L}<F_{M S Y} \text { with directed fishery } F=0 \text { when } B / B_{M S Y} \leq \beta
\end{align*}
$$

where $B$ is quantified as mature-male biomass (MMB) at mating with time of mating assigned a nominal date of 15 February. Note that as $B$ itself is a function of the fishing mortality $F_{O F L}$ (therefore numerical approximation of $F_{O F L}$ is required). As implemented for this assessment, all calculations proceed according to the model equations given in Appendix A. $F_{O F L}$ is taken to be full-selection fishing mortality in the directed pot fishery and groundfish trawl and fixed-gear fishing mortalities set at their geometric mean values over years for which there are data-based estimates of bycatch-mortality biomass.

The currently recommended Tier 4 convention is to use the full assessment period, currently 1978-2019, to define a $B_{M S Y}$ proxy in terms of average estimated MMB and to set $\gamma=1.0$ with assumed stock natural mortality $M=0.18 \mathrm{yr}^{-1}$ in setting the $F_{M S Y}$ proxy value $\gamma M$. The parameters $\alpha$ and $\beta$ are assigned their default values $\alpha=0.10$ and $\beta=0.25$. The $F_{O F L}$, OFL, ABC, and MMB in 2019 for all scenarios are summarized in Table 4. The currently recommended ABC is $75 \%$ of the OFL (ABC buffer $=25 \%$ ).

Table 4: Comparisons of management measures for the model scenarios. Biomass and OFL are in tons.

| Component | Ref | fixR | nopot |
| :--- | ---: | ---: | ---: |
| $\mathrm{MMB}_{2020}$ | 1060.665 | 1065.996 | 3707.925 |
| $B_{\mathrm{MSY}}$ | 3335.710 | 3391.948 | 3548.160 |
| $M M B / B_{\mathrm{MSY}}$ | 0.337 | 0.334 | 1.171 |
| $F_{\mathrm{OFL}}$ | 0.047 | 0.047 | 0.180 |
| $\mathrm{OFL}_{2020}$ | 50.674 | 48.819 | 618.969 |
| $\mathrm{ABC}_{2020}$ | 38.005 | 36.614 | 464.226 |

## G. Rebuilding Analysis

This stock was declared overfished in fall of 2018 and a rebuilding plan went before the Council for final review in June 2020. The most updated rebuilding plan can be found on the NPFMC website for the June 2020 meeting.

## H. Data Gaps and Research Priorities

The following topics have been listed as areas where more research on SMBKC is needed:

1. Growth increments and molting probabilities as a function of size.
2. Trawl survey catchability and selectivities.
3. Pot survey catchability and selectivities.
4. Temporal changes in spatial distributions near the island.
5. Natural mortality.

## I. Projections and outlook

The outlook for recruitment is pessimistic and the abundance relative to the proxy $B_{M S Y}$ is low. The NMFS survey results in 2019 noted ocean conditions warmer than normal with an absence of a "cold pool" in the region. This could have detrimental effects on the SMBKC stock and should be carefully monitored. Relative to the impact of historical fishing, we again conducted a "dynamic- $B_{0}$ " analysis. This procedure simply projects the population based on estimated recruitment but removes the effect of fishing. For the reference case, this suggests that the impact of fishing has reduced the stock to about $55 \%$ of what it would have been in the absence of fishing (Figure 27, supporting the hypothesis that fishing pressure is not the sole contributer to the decline of this stock in recent years. The other non-fishing contributors to the observed depleted stock trend (ignoring stock-recruit relationship) may reflect variable survival rates due to environmental conditions and also range shifts.

## J. Acknowledgements

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## Tables

Table 5: Observed proportion of crab by size class during the ADF\&G crab observer pot-lift sampling.
Source: ADF\&G Crab Observer Database.

| Year | Total pot lifts | Pot lifts sampled | Number of crab (90 mm+ CL) | Stage 1 | Stage 2 | Stage 3 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| $1990 / 91$ | 26,264 | 10 | 150 | 0.113 | 0.393 | 0.493 |
| $1991 / 92$ | 37,104 | 125 | 3,393 | 0.133 | 0.177 | 0.690 |
| $1992 / 93$ | 56,630 | 71 | 1,606 | 0.191 | 0.268 | 0.542 |
| $1993 / 94$ | 58,647 | 84 | 2,241 | 0.281 | 0.210 | 0.510 |
| $1994 / 95$ | 60,860 | 203 | 4,735 | 0.294 | 0.271 | 0.434 |
| $1995 / 96$ | 48,560 | 47 | 663 | 0.148 | 0.212 | 0.640 |
| $1996 / 97$ | 91,085 | 96 | 489 | 0.160 | 0.223 | 0.618 |
| $1997 / 98$ | 81,117 | 91,826 | 133 | 3,195 | 0.182 | 0.205 |
| $1998 / 99$ | 135 | 1.322 | 0.193 | 0.216 | 0.513 |  |
| $1999 / 00-2008 / 09$ |  | FISHERY CLOSED |  |  |  |  |
| $2009 / 10$ | 10,484 | 989 | 19,802 | 0.141 | 0.324 | 0.535 |
| $2010 / 11$ | 29,356 | 2,419 | 45,466 | 0.131 | 0.315 | 0.553 |
| $2011 / 12$ | 48,554 | 3,359 | 58,666 | 0.131 | 0.305 | 0.564 |
| $2012 / 13$ | 37,065 | 2,841 | 57,298 | 0.141 | 0.318 | 0.541 |
| $2013 / 14$ |  |  | FISHERY CLOSED |  |  |  |
| $2014 / 15$ | 10,133 | 5,475 | 419 | 9,906 | 0.094 | 0.228 |
| $2015 / 16$ |  |  | 3,248 | 0.115 | 0.252 | 0.639 |
| $2016 / 17-2018 / 19$ |  |  |  |  |  |  |

Table 6: Groundfish SMBKC male bycatch biomass ( t ) estimates. Trawl includes pelagic trawl and nonpelagic trawl types. Source: J. Zheng, ADF\&G, and author estimates based on data from R. Foy, NMFS. Estimates used after 2008/09 are from NMFS Alaska Regional Office.

| Year | Trawl bycatch | Fixed gear bycatch |
| ---: | ---: | ---: |
| 1978 | 0.000 | 0.000 |
| 1979 | 0.000 | 0.000 |
| 1980 | 0.000 | 0.000 |
| 1981 | 0.000 | 0.000 |
| 1982 | 0.000 | 0.000 |
| 1983 | 0.000 | 0.000 |
| 1984 | 0.000 | 0.000 |
| 1985 | 0.000 | 0.000 |
| 1986 | 0.000 | 0.000 |
| 1987 | 0.000 | 0.000 |
| 1988 | 0.000 | 0.000 |
| 1989 | 0.000 | 0.000 |
| 1990 | 0.000 | 0.000 |
| 1991 | 3.538 | 0.045 |
| 1992 | 1.996 | 2.268 |
| 1993 | 1.542 | 0.500 |
| 1994 | 0.318 | 0.091 |
| 1995 | 0.635 | 0.136 |
| 1996 | 0.500 | 0.045 |
| 1997 | 0.500 | 0.181 |
| 1998 | 0.500 | 0.907 |
| 1999 | 0.500 | 1.361 |
| 2000 | 0.500 | 0.500 |
| 2001 | 0.500 | 0.862 |
| 2002 | 0.726 | 0.408 |
| 2003 | 0.998 | 1.134 |
| 2004 | 0.091 | 0.635 |
| 2005 | 0.500 | 0.590 |
| 2006 | 2.812 | 1.451 |
| 2007 | 0.045 | 69.717 |
| 2008 | 0.272 | 6.622 |
| 2009 | 0.638 | 7.522 |
| 2010 | 0.360 | 9.564 |
| 2011 | 0.170 | 0.796 |
| 2012 | 0.011 | 0.739 |
| 2013 | 0.163 | 0.341 |
| 2014 | 0.010 | 0.490 |
| 2015 | 0.010 | 0.711 |
| 2016 | 0.229 | 1.630 |
| 2017 | 0.048 | 5.842 |
| 2018 | 0140 |  |
|  |  |  |
|  |  |  |

Table 7: Fishery characteristics and update. Columns include the 1978/79 to 2015/16 directed St. Matthew Island blue king crab pot fishery. The Guideline Harvest Level (GHL) and Total Allowable Catch (TAC) are in millions of pounds. Harvest includes deadloss. Catch per unit effort (CPUE) in this table is simply the harvest number / pot lifts. The average weight is the harvest weight / harvest number in pounds. The average CL is the average of retained crab in mm from dockside sampling of delivered crab. Source: Fitch et al 2012; ADF\&G Dutch Harbor staff, pers. comm. Note that management (GHL) units are in pounds, for conserving space, conversion to tons is ommitted.

| Year | Dates | GHL/TAC | Harvest |  | Pot lifts | CPUE | avg wt | avg CL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Crab | Pounds |  |  |  |  |
| 1978/79 | 07/15-09/03 |  | 436,126 | 1,984,251 | 43,754 | 10 | 4.5 | 132.2 |
| 1979/80 | 07/15-08/24 |  | 52,966 | 210,819 | 9,877 | 5 | 4.0 | 128.8 |
| 1980/81 | 07/15-09/03 |  |  | CONFID | ENTIAL |  |  |  |
| 1981/82 | 07/15-08/21 |  | 1,045,619 | 4,627,761 | 58,550 | 18 | 4.4 | NA |
| 1982/83 | 08/01-08/16 |  | 1,935,886 | 8,844,789 | 165,618 | 12 | 4.6 | 135.1 |
| 1983/84 | 08/20-09/06 | 8.0 | 1,931,990 | 9,454,323 | 133,944 | 14 | 4.9 | 137.2 |
| 1984/85 | 09/01-09/08 | 2.0-4.0 | 841,017 | 3,764,592 | 73,320 | 11 | 4.5 | 135.5 |
| 1985/86 | 09/01-09/06 | 0.9-1.9 | 436,021 | 2,175,087 | 46,988 | 9 | 5.0 | 139.0 |
| 1986/87 | 09/01-09/06 | 0.2-0.5 | 219,548 | 1,003,162 | 22,073 | 10 | 4.6 | 134.3 |
| 1987/88 | 09/01-09/05 | 0.6-1.3 | 227,447 | 1,039,779 | 28,230 | 8 | 4.6 | 134.1 |
| 1988/89 | 09/01-09/05 | 0.7-1.5 | 280,401 | 1,236,462 | 21,678 | 13 | 4.4 | 133.3 |
| 1989/90 | 09/01-09/04 | 1.7 | 247,641 | 1,166,258 | 30,803 | 8 | 4.7 | 134.6 |
| 1990/91 | 09/01-09/07 | 1.9 | 391,405 | 1,725,349 | 26,264 | 15 | 4.4 | 134.3 |
| 1991/92 | 09/16-09/20 | 3.2 | 726,519 | 3,372,066 | 37,104 | 20 | 4.6 | 134.1 |
| 1992/93 | 09/04-09/07 | 3.1 | 545,222 | 2,475,916 | 56,630 | 10 | 4.5 | 134.1 |
| 1993/94 | 09/15-09/21 | 4.4 | 630,353 | 3,003,089 | 58,647 | 11 | 4.8 | 135.4 |
| 1994/95 | 09/15-09/22 | 3.0 | 827,015 | 3,764,262 | 60,860 | 14 | 4.9 | 133.3 |
| 1995/96 | 09/15-09/20 | 2.4 | 666,905 | 3,166,093 | 48,560 | 14 | 4.7 | 135.0 |
| 1996/97 | 09/15-09/23 | 4.3 | 660,665 | 3,078,959 | 91,085 | 7 | 4.7 | 134.6 |
| 1997/98 | 09/15-09/22 | 5.0 | 939,822 | 4,649,660 | 81,117 | 12 | 4.9 | 139.5 |
| 1998/99 | 09/15-09/26 | 4.0 | 635,370 | 2,968,573 | 91,826 | 7 | 4.7 | 135.8 |
| 1999/00 | 2008/09 |  |  | FISHERY | CLOSED |  |  |  |
| 2009/10 | 10/15-02/01 | 1.17 | 103,376 | 460,859 | 10,697 | 10 | 4.5 | 134.9 |
| 2010/11 | 10/15-02/01 | 1.60 | 298,669 | 1,263,982 | 29,344 | 10 | 4.2 | 129.3 |
| 2011/12 | 10/15-02/01 | 2.54 | 437,862 | 1,881,322 | 48,554 | 9 | 4.3 | 130.0 |
| 2012/13 | 10/15-02/01 | 1.63 | 379,386 | 1,616,054 | 37,065 | 10 | 4.3 | 129.8 |
| 2013/14 |  |  |  | FISHERY | CLOSED |  |  |  |
| 2014/15 | 10/15-02/05 | 0.66 | 69,109 | 308,582 | 10,133 | 7 | 4.5 | 132.3 |
| 2015/16 | 10/19-11/28 | 0.41 | 24,076 | 105,010 | 5,475 | 4 | 4.4 | 132.6 |
| 2016/17 |  |  |  | FISHERY | CLOSED |  |  |  |
| 2017/18 |  |  |  | FISHERY | CLOSED |  |  |  |
| 2018/19 |  |  |  | FISHERY | CLOSED |  |  |  |
| 2019/20 |  |  |  | FISHERY | CLOSED |  |  |  |

Table 8: NMFS EBS trawl-survey area-swept estimates of male crab abundance ( $10^{6}$ crab) and male ( $\geq 90$ mm CL) biomass ( $10^{6} \mathrm{lbs}$ ). Total number of captured male crab $\geq 90 \mathrm{~mm}$ CL is also given. Source: R. Foy, NMFS. The " + " refer to plus group.

| Year | Abundance |  |  |  |  | Biomass |  | Number of crabs |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\begin{gathered} \text { Stage-1 } \\ (90-104 \mathrm{~mm}) \end{gathered}$ | $\begin{gathered} \text { Stage-2 } \\ (105-119 \mathrm{~mm}) \end{gathered}$ | $\begin{gathered} \text { Stage-3 } \\ (120+\mathrm{mm}) \end{gathered}$ | Total | CV | $\begin{gathered} \text { Total } \\ (90+\mathrm{mm} \mathrm{CL}) \end{gathered}$ | CV |  |
| 1978 | 2.213 | 1.991 | 1.521 | 5.726 | 0.411 | 15.064 | 0.394 | 157 |
| 1979 | 3.061 | 2.281 | 1.808 | 7.150 | 0.472 | 17.615 | 0.463 | 178 |
| 1980 | 2.856 | 2.563 | 2.541 | 7.959 | 0.572 | 22.017 | 0.507 | 185 |
| 1981 | 0.483 | 1.213 | 2.263 | 3.960 | 0.368 | 14.443 | 0.402 | 140 |
| 1982 | 1.669 | 2.431 | 5.884 | 9.984 | 0.401 | 35.763 | 0.344 | 271 |
| 1983 | 1.061 | 1.651 | 3.345 | 6.057 | 0.332 | 21.240 | 0.298 | 231 |
| 1984 | 0.435 | 0.497 | 1.452 | 2.383 | 0.175 | 8.976 | 0.179 | 105 |
| 1985 | 0.379 | 0.376 | 1.117 | 1.872 | 0.216 | 6.858 | 0.210 | 93 |
| 1986 | 0.203 | 0.447 | 0.374 | 1.025 | 0.428 | 3.124 | 0.388 | 46 |
| 1987 | 0.325 | 0.631 | 0.715 | 1.671 | 0.302 | 5.024 | 0.291 | 71 |
| 1988 | 0.410 | 0.816 | 0.957 | 2.183 | 0.285 | 6.963 | 0.252 | 81 |
| 1989 | 2.169 | 1.154 | 1.786 | 5.109 | 0.314 | 13.974 | 0.271 | 208 |
| 1990 | 1.053 | 1.031 | 2.338 | 4.422 | 0.302 | 14.837 | 0.274 | 170 |
| 1991 | 1.147 | 1.665 | 2.233 | 5.046 | 0.259 | 15.318 | 0.248 | 197 |
| 1992 | 1.074 | 1.382 | 2.291 | 4.746 | 0.206 | 15.638 | 0.201 | 220 |
| 1993 | 1.521 | 1.828 | 3.276 | 6.626 | 0.185 | 21.051 | 0.169 | 324 |
| 1994 | 0.883 | 1.298 | 2.257 | 4.438 | 0.187 | 14.416 | 0.176 | 211 |
| 1995 | 1.025 | 1.188 | 1.741 | 3.953 | 0.187 | 12.574 | 0.178 | 178 |
| 1996 | 1.238 | 1.891 | 3.064 | 6.193 | 0.263 | 20.746 | 0.241 | 285 |
| 1997 | 1.165 | 2.228 | 3.789 | 7.182 | 0.367 | 24.084 | 0.337 | 296 |
| 1998 | 0.660 | 1.661 | 2.849 | 5.170 | 0.373 | 17.586 | 0.355 | 243 |
| 1998 | 0.223 | 0.222 | 0.558 | 1.003 | 0.192 | 3.515 | 0.182 | 52 |
| 2000 | 0.282 | 0.285 | 0.740 | 1.307 | 0.303 | 4.623 | 0.310 | 61 |
| 2001 | 0.419 | 0.502 | 0.938 | 1.859 | 0.243 | 6.242 | 0.245 | 91 |
| 2002 | 0.111 | 0.230 | 0.640 | 0.981 | 0.311 | 3.820 | 0.320 | 38 |
| 2003 | 0.449 | 0.280 | 0.465 | 1.194 | 0.399 | 3.454 | 0.336 | 65 |
| 2004 | 0.247 | 0.184 | 0.562 | 0.993 | 0.369 | 3.360 | 0.305 | 48 |
| 2005 | 0.319 | 0.310 | 0.501 | 1.130 | 0.403 | 3.620 | 0.371 | 42 |
| 2006 | 0.917 | 0.642 | 1.240 | 2.798 | 0.339 | 8.585 | 0.334 | 126 |
| 2007 | 2.518 | 2.020 | 1.193 | 5.730 | 0.420 | 14.266 | 0.385 | 250 |
| 2008 | 1.352 | 0.801 | 1.457 | 3.609 | 0.289 | 10.261 | 0.284 | 167 |
| 2009 | 1.573 | 2.161 | 1.410 | 5.144 | 0.263 | 13.892 | 0.256 | 251 |
| 2010 | 3.937 | 3.253 | 2.458 | 9.648 | 0.544 | 24.539 | 0.466 | 388 |
| 2011 | 1.800 | 3.255 | 3.207 | 8.263 | 0.587 | 24.099 | 0.558 | 318 |
| 2012 | 0.705 | 1.970 | 1.808 | 4.483 | 0.361 | 13.669 | 0.339 | 193 |
| 2013 | 0.335 | 0.452 | 0.807 | 1.593 | 0.215 | 5.043 | 0.217 | 74 |
| 2014 | 0.723 | 1.627 | 1.809 | 4.160 | 0.503 | 13.292 | 0.449 | 181 |
| 2015 | 0.992 | 1.269 | 1.979 | 4.240 | 0.774 | 12.958 | 0.770 | 153 |
| 2016 | 0.535 | 0.660 | 1.178 | 2.373 | 0.447 | 7.685 | 0.393 | 108 |
| 2017 | 0.091 | 0.323 | 0.663 | 1.077 | 0.657 | 3.955 | 0.600 | 42 |
| 2018 | 0.154 | 0.232 | 0.660 | 1.047 | 0.298 | 3.816 | 0.281 | 62 |
| 2019 | 0.403 | 0.482 | 1.170 | 2.056 | 0.352 | 6.990 | 0.337 | 105 |

Table 9: Size-class and total CPUE ( $90+\mathrm{mm}$ CL) with estimated CV and total number of captured crab ( $90+\mathrm{mm}$ CL) from the 96 common stations surveyed during the ADF\&G SMBKC pot surveys. Source: ADF\&G.

| Year | Stage-1 <br> $(90-104 \mathrm{~mm})$ | Stage-2 <br> $(105-119 \mathrm{~mm})$ | Stage-3 <br> $(120+\mathrm{mm})$ | Total CPUE | CV | Number of crabs |
| :---: | :---: | :---: | :---: | ---: | ---: | ---: |
| 1995 | 1.919 | 3.198 | 6.922 | 12.042 | 0.13 | 4624 |
| 1998 | 0.964 | 2.763 | 8.804 | 12.531 | 0.06 | 4812 |
| 2001 | 1.266 | 1.737 | 5.487 | 8.477 | 0.08 | 3255 |
| 2004 | 0.112 | 0.414 | 1.141 | 1.667 | 0.15 | 640 |
| 2007 | 1.086 | 2.721 | 4.836 | 8.643 | 0.09 | 3319 |
| 2010 | 1.326 | 3.276 | 5.607 | 10.209 | 0.13 | 3920 |
| 2013 | 0.878 | 1.398 | 3.367 | 5.643 | 0.19 | 2167 |
| 2015 | 0.198 | 0.682 | 1.924 | 2.805 | 0.18 | 1077 |
| 2016 | 0.198 | 0.456 | 1.724 | 2.378 | 0.19 | 777 |
| 2017 | 0.177 | 0.429 | 1.083 | 1.689 | 0.25 | 643 |
| 2018 | 0.076 | 0.161 | 0.508 | 0.745 | 0.14 | 286 |

Table 10: Mean weight (kg) by stage in used in all of the models (provided as a vector of weights at length each year to GMACS).

| Year | Stage-1 | Stage-2 | Stage-3 |
| ---: | ---: | ---: | ---: |
| 1978 | 0.7 | 1.2 | 1.9 |
| 1979 | 0.7 | 1.2 | 1.7 |
| 1980 | 0.7 | 1.2 | 1.9 |
| 1981 | 0.7 | 1.2 | 1.9 |
| 1982 | 0.7 | 1.2 | 1.9 |
| 1983 | 0.7 | 1.2 | 2.1 |
| 1984 | 0.7 | 1.2 | 1.9 |
| 1985 | 0.7 | 1.2 | 2.1 |
| 1986 | 0.7 | 1.2 | 1.9 |
| 1987 | 0.7 | 1.2 | 1.9 |
| 1988 | 0.7 | 1.2 | 1.9 |
| 1989 | 0.7 | 1.2 | 2.0 |
| 1990 | 0.7 | 1.2 | 1.9 |
| 1991 | 0.7 | 1.2 | 2.0 |
| 1992 | 0.7 | 1.2 | 1.9 |
| 1993 | 0.7 | 1.2 | 2.0 |
| 1994 | 0.7 | 1.2 | 1.9 |
| 1995 | 0.7 | 1.2 | 2.0 |
| 1996 | 0.7 | 1.2 | 2.0 |
| 1997 | 0.7 | 1.2 | 2.1 |
| 1998 | 0.7 | 1.2 | 2.0 |
| 1999 | 0.7 | 1.2 | 1.9 |
| 2000 | 0.7 | 1.2 | 1.9 |
| 2001 | 0.7 | 1.2 | 1.9 |
| 2002 | 0.7 | 1.2 | 1.9 |
| 2003 | 0.7 | 1.2 | 1.9 |
| 2004 | 0.7 | 1.2 | 1.9 |
| 2005 | 0.7 | 1.2 | 1.9 |
| 2006 | 0.7 | 1.2 | 1.9 |
| 2007 | 0.7 | 1.2 | 1.9 |
| 2008 | 0.7 | 1.2 | 1.9 |
| 2009 | 0.7 | 1.2 | 1.9 |
| 2010 | 0.7 | 1.2 | 1.8 |
| 2011 | 0.7 | 1.2 | 1.8 |
| 2012 | 0.7 | 1.2 | 1.8 |
| 2013 | 0.7 | 1.2 | 1.9 |
| 2014 | 0.7 | 1.2 | 1.9 |
| 2015 | 0.7 | 1.2 | 1.9 |
| 2016 | 0.7 | 1.2 | 1.9 |
| 2017 | 0.7 | 1.2 | 1.9 |
| 2018 | 0.7 | 1.2 | 1.9 |
| 2019 | 0.7 | 1.2 | 1.9 |
|  |  |  |  |

Table 11: Observed and input sample sizes for observer data from the directed pot fishery, the NMFS trawl survey, and the ADF\&G pot survey.

| Year | Number measured |  |  | Input sample sizes |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Observer pot | NMFS trawl | ADF\&G pot | Observer pot | NMFS trawl | ADF\&G pot |
| 1978 |  | 157 |  |  | 50 |  |
| 1979 |  | 178 |  |  | 50 |  |
| 1980 |  | 185 |  |  | 50 |  |
| 1981 |  | 140 |  |  | 50 |  |
| 1982 |  | 271 |  |  | 50 |  |
| 1983 |  | 231 |  |  | 50 |  |
| 1984 |  | 105 |  |  | 50 |  |
| 1985 |  | 93 |  |  | 46.5 |  |
| 1986 |  | 46 |  |  | 23 |  |
| 1987 |  | 71 |  |  | 35.5 |  |
| 1988 |  | 81 |  |  | 40.5 |  |
| 1989 |  | 208 |  |  | 50 |  |
| 1990 | 150 | 170 |  | 15 | 50 |  |
| 1991 | 3393 | 197 |  | 25 | 50 |  |
| 1992 | 1606 | 220 |  | 25 | 50 |  |
| 1993 | 2241 | 324 |  | 25 | 50 |  |
| 1994 | 4735 | 211 |  | 25 | 50 |  |
| 1995 | 663 | 178 | 4624 | 25 | 50 | 100 |
| 1996 | 489 | 285 |  | 25 | 50 |  |
| 1997 | 3195 | 296 |  | 25 | 50 |  |
| 1998 | 1323 | 243 | 4812 | 25 | 50 | 100 |
| 1999 |  | 52 |  |  | 26 |  |
| 2000 |  | 61 |  |  | 30.5 |  |
| 2001 |  | 91 | 3255 |  | 45.5 | 100 |
| 2002 |  | 38 |  |  | 19 |  |
| 2003 |  | 65 |  |  | 32.5 |  |
| 2004 |  | 48 | 640 |  | 24 | 100 |
| 2005 |  | 42 |  |  | 21 |  |
| 2006 |  | 126 |  |  | 50 |  |
| 2007 |  | 250 | 3319 |  | 50 | 100 |
| 2008 |  | 167 |  |  | 50 |  |
| 2009 | 19802 | 251 |  | 50 | 50 |  |
| 2010 | 45466 | 388 | 3920 | 50 | 50 | 100 |
| 2011 | 58667 | 318 |  | 50 | 50 |  |
| 2012 | 57282 | 193 |  | 50 | 50 |  |
| 2013 |  | 74 | 2167 |  | 37 | 100 |
| 2014 | 9906 | 181 |  | 50 | 50 |  |
| 2015 | 3248 | 153 | 1077 | 50 | 50 | 100 |
| 2016 |  | 108 | 777 |  | 50 | 100 |
| 2017 |  | 42 | 643 |  | 21 | 100 |
| 2018 |  | 62 | 286 |  | 31 | 100 |
| 2019 |  | 105 |  |  | 50 |  |

Table 12: Model parameter estimates, selected derived quantities, and their standard deviations (SD) for the reference (16.0) model.

| Parameter | Estimate | SD |
| :--- | ---: | ---: |
| Natural mortality deviation in 1998/99 $\left(\delta_{1998}^{M}\right)$ | 1.573 | 0.138 |
| $\log (\bar{R})$ | 13.899 | 0.200 |
| $\log \left(n_{1}^{0}\right)$ | 14.950 | 0.175 |
| $\log \left(n_{2}^{0}\right)$ | 14.509 | 0.211 |
| $\log \left(n_{3}^{0}\right)$ | 14.326 | 0.207 |
| $q_{p o t}$ | 3.838 | 0.253 |
| $\log \left(\bar{F}^{\text {df }}\right)$ | -2.125 | 0.052 |
| $\log \left(\bar{F}^{\text {tb }}\right)$ | -9.470 | 0.073 |
| $\log \left(\bar{F}^{\text {fb }}\right)$ | -8.093 | 0.073 |
| $\log$ Stage-1 directed pot selectivity 1978-2008 | -0.819 | 0.179 |
| $\log$ Stage-2 directed pot selectivity 1978-2008 | -0.452 | 0.129 |
| $\log$ Stage-1 directed pot selectivity 2009-2017 | -0.483 | 0.162 |
| $\log$ Stage-2 directed pot selectivity $2009-2017$ | -0.000 | 0.000 |
| $\log$ Stage-1 NMFS trawl selectivity | -0.320 | 0.066 |
| $\log$ Stage-2 NMFS trawl selectivity | -0.000 | 0.000 |
| $\log$ Stage-1 ADF\&G pot selectivity | -0.725 | 0.126 |
| $\log$ Stage-2 ADF\&G pot selectivity | -0.000 | 0.000 |
| $F_{\text {OFL }}$ | 0.040 | 0.007 |
| OFL | 50.674 | 17.412 |

Table 13: Model parameter estimates, selected derived quantities, and their standard deviations (SD) for the reference model with fixed terminal year recruitment 'fixR' (16.0a).

| Parameter | Estimate | SD |
| :--- | ---: | ---: |
| Natural mortality deviation in 1998/99 $\left(\delta_{1998}^{M}\right)$ | 1.573 | 0.138 |
| $\log (\bar{R})$ | 13.870 | 0.198 |
| $\log \left(n_{1}^{0}\right)$ | 14.950 | 0.175 |
| $\log \left(n_{2}^{0}\right)$ | 14.508 | 0.211 |
| $\log \left(n_{3}^{0}\right)$ | 14.326 | 0.207 |
| $q_{p o t}$ | 3.833 | 0.253 |
| $\log \left(\bar{F}^{\text {df }}\right)$ | -2.126 | 0.052 |
| $\log \left(\bar{F}^{\text {tb }}\right)$ | -9.472 | 0.073 |
| $\log \left(\bar{F}^{\text {fb }}\right)$ | -8.094 | 0.073 |
| $\log$ Stage-1 directed pot selectivity 1978-2008 | -0.820 | 0.179 |
| $\log$ Stage-2 directed pot selectivity 1978-2008 | -0.452 | 0.129 |
| $\log$ Stage-1 directed pot selectivity 2009-2017 | -0.484 | 0.162 |
| $\log$ Stage-2 directed pot selectivity 2009-2017 | -0.000 | 0.000 |
| $\log$ Stage-1 NMFS trawl selectivity | -0.320 | 0.066 |
| $\log$ Stage-2 NMFS trawl selectivity | -0.000 | 0.000 |
| $\log$ Stage-1 ADF\&G pot selectivity | -0.727 | 0.125 |
| $\log$ Stage-2 ADF\&G pot selectivity | -0.000 | 0.000 |
| $F_{\text {OFL }}$ | 0.047 | 0.007 |
| OFL | 48.819 | 9.115 |

Table 14: Model parameter estimates, selected derived quantities, and their standard deviations (SD) for the 'no pot' (20.1) model.

| Parameter | Estimate | SD |
| :--- | ---: | ---: |
| Natural mortality deviation in 1998/99 $\left(\delta_{1998}^{M}\right)$ | 1.829 | 0.235 |
| $\log (\bar{R})$ | 14.225 | 0.203 |
| $\log \left(n_{1}^{0}\right)$ | 14.945 | 0.174 |
| $\log \left(n_{2}^{0}\right)$ | 14.459 | 0.211 |
| $\log \left(n_{3}^{0}\right)$ | 14.290 | 0.205 |
| $\log \left(\bar{F}^{\text {df }}\right)$ | -2.319 | 0.056 |
| $\log \left(\bar{F}^{\mathrm{tb}}\right)$ | -9.716 | 0.079 |
| $\log \left(\bar{F}^{\mathrm{fb}}\right)$ | -8.341 | 0.079 |
| $\log$ Stage-1 directed pot selectivity 1978-2008 | -0.817 | 0.178 |
| $\log$ Stage-2 directed pot selectivity 1978-2008 | -0.482 | 0.133 |
| $\log$ Stage-1 directed pot selectivity 2009-2017 | -0.982 | 0.182 |
| $\log$ Stage-2 directed pot selectivity 2009-2017 | -0.000 | 0.000 |
| $\log$ Stage-1 NMFS trawl selectivity | -0.376 | 0.062 |
| $\log$ Stage-2 NMFS trawl selectivity | -0.000 | 0.000 |
| $F_{\text {OFL }}$ | 0.047 | 0.000 |
| OFL | 618.969 | 144.208 |

Table 15: Comparisons of parameter estimates for the model scenarios.

| Parameter | Ref | fixR | nopot |
| :--- | ---: | ---: | ---: |
| Natural mortality deviation in 1998/99 $\left(\delta_{1998}^{M}\right)$ | 1.573 | 1.573 | 1.829 |
| $\log (\bar{R})$ | 13.899 | 13.870 | 14.225 |
| $\log \left(n_{1}^{0}\right)$ | 14.950 | 14.950 | 14.945 |
| $\log \left(n_{2}^{0}\right)$ | 14.509 | 14.508 | 14.459 |
| $\log \left(n_{3}^{0}\right)$ | 14.326 | 14.326 | 14.290 |
| $q_{p o t}$ | 3.838 | 3.833 | - |
| $\log \left(\bar{F}^{\text {df }}\right)$ | -2.125 | -2.126 | -2.319 |
| $\log \left(\bar{F}^{\text {tb }}\right)$ | -9.470 | -9.472 | -9.716 |
| $\log \left(\bar{F}^{\text {fb }}\right)$ | -8.093 | -8.094 | -8.341 |
| $\log$ Stage-1 directed pot selectivity 1978-2008 | -0.819 | -0.820 | -0.817 |
| $\log$ Stage-2 directed pot selectivity 1978-2008 | -0.452 | -0.452 | -0.482 |
| $\log$ Stage-1 directed pot selectivity 2009-2017 | -0.483 | -0.484 | -0.982 |
| $\log$ Stage-2 directed pot selectivity 2009-2017 | -0.000 | -0.000 | -0.000 |
| $\log$ Stage-1 NMFS trawl selectivity | -0.320 | -0.320 | -0.376 |
| $\log$ Stage-2 NMFS trawl selectivity | -0.000 | -0.000 | -0.000 |
| $\log$ Stage-1 ADF\&G pot selectivity | -0.725 | -0.727 | - |
| $\log$ Stage-2 ADF\&G pot selectivity | -0.000 | -0.000 | - |
| $F_{\text {OFL }}$ | 0.047 | 0.047 | 0.180 |
| OFL | 50.674 | 48.819 | 618.969 |

Table 16: Comparisons of data weights, SDNR and MAR (standard deviation of normalized residuals and median absolute residual) values for the model scenarios.

| Component | Ref | fixR | nopot |
| :--- | :---: | :---: | ---: |
| NMFS trawl survey weight | 1.00 | 1.00 | 1.00 |
| ADF\&G pot survey weight | 1.00 | 1.00 |  |
| Directed pot LF weight | 1.00 | 1.00 | 1.00 |
| NMFS trawl survey LF weight | 1.00 | 1.00 | 1.00 |
| ADF\&G pot survey LF weight | 1.00 | 1.00 |  |
| SDNR NMFS trawl survey | 0.00 | 0.00 | 0.00 |
| SDNR ADF\&G pot survey | 0.00 | 0.00 |  |
| SDNR directed pot LF | 0.70 | 0.70 | 0.77 |
| SDNR NMFS trawl survey LF | 1.30 | 1.30 | 1.23 |
| SDNR ADF\&G pot survey LF | 0.95 | 0.95 |  |
| MAR NMFS trawl survey | 0.00 | 0.00 | 0.00 |
| MAR ADF\&G pot survey | 0.00 | 0.00 |  |
| MAR directed pot LF | 0.52 | 0.52 | 0.46 |
| MAR NMFS trawl survey LF | 0.60 | 0.60 | 0.78 |
| MAR ADF\&G pot survey LF | 0.68 | 0.68 |  |

Table 17: Comparisons of negative log-likelihood values for the selected model scenarios. It is important to note that comparisons among models may be limited since the number of parameters between models changes (e.g., nopot model).

| Component | Ref | fixR | nopot |
| :--- | ---: | ---: | ---: |
| Pot Retained Catch | -68.50 | -68.51 | -56.27 |
| Pot Discarded Catch | 4.89 | 4.89 | 6.29 |
| Trawl bycatch Discarded Catch | -7.99 | -7.99 | 6.11 |
| Fixed bycatch Discarded Catch | -7.95 | -7.95 | 4.84 |
| NMFS Trawl Survey | 8.84 | 8.62 | -4.42 |
| ADF\&G Pot Survey CPUE | 84.62 | 84.93 |  |
| Directed Pot LF | -103.99 | -103.99 | -102.34 |
| NMFS Trawl LF | -252.91 | -252.93 | -256.22 |
| ADF\&G Pot LF | -91.02 | -91.05 |  |
| Recruitment deviations | 59.56 | 60.01 | 59.37 |
| F penalty | 9.66 | 9.66 | 9.66 |
| M penalty | 6.46 | 6.46 | 6.45 |
| Prior | 13.71 | 13.71 | 12.11 |
| Total | -344.61 | -344.12 | -314.40 |
| Total estimated parameters | 147.00 | 146.00 | 144.00 |

Table 18: Population abundances $(\boldsymbol{n})$ by crab stage in numbers of crab at the time of the survey and mature male biomass (MMB) in tons on 15 February for the model configuration used in 2019.

| Year | $n_{1}$ | $n_{2}$ | $n_{3}$ | MMB | CV MMB |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 1978 | 3109715 | 2000299 | 1666848 | 4550 | 0.178 |
| 1979 | 4376763 | 2355384 | 2282776 | 6433 | 0.124 |
| 1980 | 3779544 | 3257707 | 3463738 | 10256 | 0.083 |
| 1981 | 1439955 | 3221560 | 4866873 | 10705 | 0.062 |
| 1982 | 1618361 | 1833987 | 4894696 | 7604 | 0.072 |
| 1983 | 811849 | 1447417 | 3468928 | 4537 | 0.099 |
| 1984 | 662337 | 858825 | 1983059 | 3022 | 0.124 |
| 1985 | 928011 | 622498 | 1406806 | 2656 | 0.144 |
| 1986 | 1366392 | 705833 | 1186990 | 2600 | 0.140 |
| 1987 | 1330701 | 989214 | 1278483 | 3074 | 0.129 |
| 1988 | 1241066 | 1061590 | 1484711 | 3360 | 0.126 |
| 1989 | 2898487 | 1033510 | 1638093 | 3849 | 0.121 |
| 1990 | 1877184 | 1956744 | 1939926 | 4970 | 0.094 |
| 1991 | 1938968 | 1673531 | 2420850 | 4992 | 0.095 |
| 1992 | 2099715 | 1593816 | 2382018 | 5175 | 0.085 |
| 1993 | 2372747 | 1673953 | 2494925 | 5427 | 0.077 |
| 1994 | 1608587 | 1844929 | 2573586 | 5200 | 0.070 |
| 1995 | 1749039 | 1461936 | 2471794 | 5073 | 0.073 |
| 1996 | 1780265 | 1429663 | 2364609 | 4775 | 0.075 |
| 1997 | 912655 | 1434576 | 2265018 | 4155 | 0.094 |
| 1998 | 603985 | 936010 | 1844896 | 2740 | 0.110 |
| 1999 | 369997 | 310550 | 711971 | 1680 | 0.102 |
| 2000 | 408474 | 312747 | 786233 | 1822 | 0.084 |
| 2001 | 372448 | 335395 | 853220 | 1973 | 0.076 |
| 2002 | 129931 | 322415 | 917072 | 2077 | 0.070 |
| 2003 | 290682 | 180441 | 940677 | 1961 | 0.071 |
| 2004 | 187364 | 224669 | 903940 | 1943 | 0.071 |
| 2005 | 468821 | 180737 | 886078 | 1860 | 0.072 |
| 2006 | 702839 | 325974 | 875801 | 2003 | 0.072 |
| 2007 | 403315 | 506459 | 961977 | 2337 | 0.069 |
| 2008 | 835694 | 391131 | 1082101 | 2461 | 0.060 |
| 2009 | 682211 | 603380 | 1179630 | 2497 | 0.054 |
| 2010 | 624238 | 577600 | 1251605 | 2110 | 0.057 |
| 2011 | 496132 | 520319 | 1099028 | 1528 | 0.070 |
| 2012 | 228196 | 415162 | 788179 | 998 | 0.108 |
| 2013 | 251502 | 235864 | 506691 | 1158 | 0.097 |
| 2014 | 204364 | 220853 | 566085 | 1090 | 0.103 |
| 2015 | 162705 | 185039 | 537244 | 1070 | 0.105 |
| 2016 | 169495 | 152401 | 534064 | 1116 | 0.102 |
| 2017 | 131331 | 146586 | 538681 | 1116 | 0.101 |
| 2018 | 141883 | 122799 | 535054 | 1085 | 0.100 |
| 2019 | 250747 | 121140 | 521618 | 1022 | 0.103 |
|  |  |  |  |  |  |

Table 19: Population abundances ( $\boldsymbol{n}$ ) by crab stage in numbers of crab at the time of the survey (1 July, season 1) and mature male biomass (MMB) in tons on 15 February for the 2020 reference model.

| Year | $n_{1}$ | $n_{2}$ | $n_{3}$ | MMB | CV MMB |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 1978 | 3151217 | 2048032 | 1704813 | 4676 | 0.176 |
| 1979 | 4405644 | 2394327 | 2341979 | 6576 | 0.122 |
| 1980 | 3774514 | 3287008 | 3535569 | 10427 | 0.083 |
| 1981 | 1435061 | 3228410 | 4941160 | 10851 | 0.062 |
| 1982 | 1622665 | 1833539 | 4959495 | 7725 | 0.072 |
| 1983 | 826815 | 1449709 | 3522402 | 4646 | 0.099 |
| 1984 | 673504 | 867978 | 2029459 | 3119 | 0.123 |
| 1985 | 940551 | 631919 | 1451162 | 2759 | 0.143 |
| 1986 | 1398609 | 716293 | 1230084 | 2694 | 0.139 |
| 1987 | 1351732 | 1011045 | 1322901 | 3183 | 0.127 |
| 1988 | 1256200 | 1080852 | 1534825 | 3474 | 0.123 |
| 1989 | 2919885 | 1048636 | 1691144 | 3969 | 0.119 |
| 1990 | 1888479 | 1974231 | 1993985 | 5088 | 0.093 |
| 1991 | 1953255 | 1686052 | 2476052 | 5111 | 0.094 |
| 1992 | 2112699 | 1606335 | 2435840 | 5290 | 0.085 |
| 1993 | 2392964 | 1685630 | 2547439 | 5543 | 0.077 |
| 1994 | 1638537 | 1860336 | 2625259 | 5314 | 0.070 |
| 1995 | 1766633 | 1483754 | 2525427 | 5201 | 0.073 |
| 1996 | 1804613 | 1446768 | 2421768 | 4904 | 0.075 |
| 1997 | 941521 | 1454055 | 2323563 | 4296 | 0.094 |
| 1998 | 618296 | 958642 | 1906137 | 2860 | 0.109 |
| 1999 | 381326 | 315898 | 737767 | 1735 | 0.102 |
| 2000 | 421648 | 320952 | 811560 | 1879 | 0.084 |
| 2001 | 383990 | 345593 | 879772 | 2034 | 0.076 |
| 2002 | 134380 | 332345 | 945496 | 2142 | 0.071 |
| 2003 | 302039 | 186255 | 969851 | 2022 | 0.072 |
| 2004 | 191454 | 233042 | 932326 | 2006 | 0.072 |
| 2005 | 479484 | 185831 | 914401 | 1919 | 0.072 |
| 2006 | 718464 | 333716 | 903047 | 2062 | 0.072 |
| 2007 | 409910 | 517899 | 990132 | 2402 | 0.069 |
| 2008 | 844891 | 398703 | 1112005 | 2526 | 0.061 |
| 2009 | 692584 | 611117 | 1209302 | 2557 | 0.055 |
| 2010 | 634017 | 586098 | 1281337 | 2168 | 0.058 |
| 2011 | 509421 | 528796 | 1129162 | 1588 | 0.072 |
| 2012 | 239665 | 425751 | 819051 | 1062 | 0.109 |
| 2013 | 264030 | 246289 | 539320 | 1227 | 0.098 |
| 2014 | 216047 | 231419 | 599794 | 1160 | 0.104 |
| 2015 | 171673 | 195187 | 571890 | 1140 | 0.106 |
| 2016 | 178308 | 160859 | 568985 | 1187 | 0.103 |
| 2017 | 138175 | 154391 | 572956 | 1186 | 0.101 |
| 2018 | 147990 | 129272 | 568274 | 1151 | 0.101 |
| 2019 | 262671 | 126752 | 553209 | 1081 | 0.103 |
|  |  |  |  |  |  |

## Figures



Figure 1: Distribution of blue king crab (Paralithodes platypus) in the Gulf of Alaska, Bering Sea, and Aleutian Islands waters (shown in blue).


Figure 2: Blue king crab Registration Area Q (Bering Sea)

Data by type and year


Figure 3: Data extent for the SMBKC assessment.


Figure 4: Trawl and pot-survey stations used in the SMBKC stock assessment.


Figure 5: Catches (in numbers) of male blue king crab > 90mm CL from the 2011-2019 NMFS trawl-survey at the 56 stations used to assess the SMBKC stock.


Figure 6: Fits to NMFS area-swept trawl estimates of total $(>90 \mathrm{~mm})$ male survey biomass for the reference model only ( 16.0 ref for 2020 and 16.02019 accepted model). Error bars are plus and minus 2 standard deviations.


Figure 7: Comparisons of fits to CPUE from the ADFG pot surveys for the reference model 16.0 reference model in 2019 and 2020. Error bars are plus and minus 2 standard deviations.


Figure 8: Reference model estimated recruitment (2019 and 2020) for comparison from 1978-2018, does not show recent recruitment, i.e. 2019.


Figure 9: Sensitivity of new data in 2020 on estimated mature male biomass (MMB); 1978-2020.


Figure 10: Comparisons of the estimated stage-1 and stage-2 selectivities for the different model scenarios (the stage-3 selectivities are all fixed at 1). Estimated selectivities are shown for the directed pot fishery, the trawl bycatch fishery, the fixed bycatch fishery, the NMFS trawl survey, and the ADFG pot survey. Two selectivity periods are estimated in the directed pot fishery, from 1978-2008 and 2009-2019.

## Recruitment model scenarios



Figure 11: Estimated recruitment 1979-2019 comparing model alternatives. The solid horizontal lines in the background represent the estimate of the average recruitment parameter $(\bar{R})$ in each model scenario. Note the high uncertainty in recruitment in both the ref and the nopot model due to the lack of 2020 data.


Figure 12: Estimated recruitment 1979-2019 comparing ref model (16.0) and model with fixed recruitment in the terminal year (16.0a). The solid horizontal lines in the background represent the estimate of the average recruitment parameter $(\bar{R})$ in each model scenario.


Figure 13: Comparisons of estimated mature male biomass (MMB) time series on 15 February during 19782020 for each of the model scenarios.


Model
$\rightarrow-\quad$ model 16.0 (2020)
-- model 16.0a (fix R ter)
-- model 20.1 (no pot)

Figure 14: Time-varying natural mortality $\left(M_{t}\right)$. Estimated pulse period occurs in 1998/99 (i.e. $M_{1998}$ ).


Figure 15: Comparisons of area-swept estimates of total ( $90+\mathrm{mm}$ CL) male survey biomass (tons) and model predictions for the model scenarios. The error bars are plus and minus 2 standard deviations.


Figure 16: Comparisons of total ( $90+\mathrm{mm}$ CL) male pot survey CPUEs and model predictions for the model scenarios. The error bars are plus and minus 2 standard deviations.


Figure 17: Standardized residuals for area-swept estimates of total male survey biomass for the model scenarios.


Figure 18: Standardized residuals for total male pot survey CPUEs for each of the GMACS model scenarios.


Figure 19: Observed and model estimated size-frequencies of SMBKC by year retained in the directed pot fishery for the model scenarios.


Figure 20: Observed and model estimated size-frequencies of discarded male SMBKC by year in the NMFS trawl survey for the model scenarios.


Figure 21: Observed and model estimated size-frequencies of discarded SMBKC by year in the ADFG pot survey for the model scenarios.


Figure 22: Bubble plots of residuals by stage and year for the all the size composition data sets (ADFG pot survey, NMFS trawl survey, and the directed pot fishery) for SMBKC in the 'reference' model (16.0).


Figure 23: Bubble plots of residuals by stage and year for the all the size composition data sets (NMFS trawl survey, and the directed pot fishery) for SMBKC in the 'fixR' model (16.0a).


Figure 24: Bubble plots of residuals by stage and year for the all the size composition data sets (NMFS trawl survey, and the directed pot fishery) for SMBKC in the 'no pot' model (20.1).


Figure 25: Comparison of observed and model predicted retained catch and bycatches in each of the GMACS models. Note that difference in units between each of the panels, some panels are expressed in numbers of crab, some as biomass (tons).






Year

Figure 26: Fishing mortality estimates from the reference model (16.0) for directed and bycatch fleets


Figure 27: Comparison of mature male biomass relative to the dynamic B zero value, (15 February, 19782019) for each of the model scenarios.


Figure 28: Retrospective pattern in mature male biomass (MMB ( t ) ) for the reference (base) model (16.0), Mohn's rho $=-0.346$

## Appendix A: SMBKC Model Description

## 1. Introduction

The GMACS model has been specified to account only for male crab $\geq 90 \mathrm{~mm}$ in carapace length (CL). These are partitioned into three stages (size- classes) determined by CL measurements of (1) 90-104 mm, (2) 105-119 mm, and (3) $120+\mathrm{mm}$. For management of the St. Matthew Island blue king crab (SMBKC) fishery, 120 mm CL is used as the proxy value for the legal measurement of 5.5 inch carapace width (CW), whereas 105 mm CL is the management proxy for mature-male size (state regulation 5 AAC 34.917 (d)). Accordingly, within the model only stage-3 crab are retained in the directed fishery, and stage- 2 and stage- 3 crab together comprise the collection of mature males. Some justification for the 105 mm value is presented in Pengilly and Schmidt (1995), who used it in developing the current regulatory SMBKC harvest strategy. The term "recruit" here designates recruits to the model, i.e., annual new stage- 1 crab, rather than recruits to the fishery. The following description of model structure reflects the GMACS base model configuration.

## 2. Model Population Dynamics

Within the model, the beginning of the crab year is assumed contemporaneous with the NMFS trawl survey, nominally assigned a date of 1 July. Although the timing of the fishery is different each year, MMB is estimated at 15 February, which is the reference date for calculation of federal management biomass quantities. To accommodate this, each model year is split into 5 seasons $(t)$ and a proportion of the natural mortality $\left(\tau_{t}\right)$, scaled relative to the portions of the year, is applied in each of these seasons where $\sum_{t=1}^{t=5} \tau_{t}=1$. Each model year consists of the following processes with time-breaks denoted here by "Seasons." However, it is important to note that actual seasons are survey-to-fishery, fishery-to Feb 15, and Feb 15 to July 1. The following breakdown accounts for events and fishing mortality treatments:

1. Season 1 (survey period)

- Beginning of the SMBKC fishing year (1 July)
- $\tau_{1}=0$
- Surveys

2. Season 2 (natural mortality until pulse fishery)

- $\tau_{2}$ ranges from 0.05 to 0.44 depending on the time of year the fishery begins each year (i.e., a higher value indicates the fishery begins later in the year; see Table reftab:smbkc-fishery)

3. Season 3 (pulse fishery)

- $\tau_{3}=0$
- fishing mortality applied

4. Season 4 (natural mortality until spawning)

- $\tau_{4}=0.63-\sum_{i=1}^{i=4} \tau_{i}$
- Calculate MMB (15 February)

5. Season 5 (natural mortality and somatic growth through to June 30th)

- $\tau_{5}=0.37$
- Growth and molting
- Recruitment (all to stage-1)

The proportion of natural mortality $\left(\tau_{t}\right)$ applied during each season in the model is provided in Table 20. The beginning of the year ( 1 July) to the date that MMB is measured ( 15 February) is $63 \%$ of the year. Therefore $63 \%$ of the natural mortality must be applied before the MMB is calculated. Because the timing of the fishery is different each year, $\tau_{2}$ varies and thus $\tau_{4}$ varies also.
With boldface lower-case letters indicating vector quantities we designate the vector of stage abundances during season $t$ and year $y$ as

$$
\begin{equation*}
\boldsymbol{n}_{t, y}=n_{l, t, y}=\left[n_{1, t, y}, n_{2, t, y}, n_{3, t, y}\right]^{\top} \tag{2}
\end{equation*}
$$

The number of new crab, or recruits, of each stage entering the model each season $t$ and year $y$ is represented as the vector $\boldsymbol{r}_{t, y}$. The SMBKC formulation of GMACS specifies recruitment to stage- 1 only during season $t=5$, thus the recruitment size distribution is

$$
\begin{equation*}
\phi_{l}=[1,0,0]^{\top}, \tag{3}
\end{equation*}
$$

and the recruitment is

$$
\boldsymbol{r}_{t, y}= \begin{cases}0 & \text { for } \quad t<5  \tag{4}\\ \bar{R} \phi_{l} \delta_{y}^{R} & \text { for } \quad t=5\end{cases}
$$

where $\bar{R}$ is the average annual recruitment and $\delta_{y}^{R}$ are the recruitment deviations each year $y$

$$
\begin{equation*}
\delta_{y}^{R} \sim \mathcal{N}\left(0, \sigma_{R}^{2}\right) . \tag{5}
\end{equation*}
$$

Using boldface upper-case letters to indicate a matrix, we describe the size transition matrix $\boldsymbol{G}$ as

$$
\boldsymbol{G}=\left[\begin{array}{ccc}
1-\pi_{12}-\pi_{13} & \pi_{12} & \pi_{13}  \tag{6}\\
0 & 1-\pi_{23} & \pi_{23} \\
0 & 0 & 1
\end{array}\right]
$$

with $\pi_{j k}$ equal to the proportion of stage- $j$ crab that molt and grow into stage- $k$ within a season or year.
The natural mortality each season $t$ and year $y$ is

$$
\begin{equation*}
M_{t, y}=\bar{M} \tau_{t}+\delta_{y}^{M} \text { where } \delta_{y}^{M} \sim \mathcal{N}\left(0, \sigma_{M}^{2}\right) \tag{7}
\end{equation*}
$$

Fishing mortality by year $y$ and season $t$ is denoted $F_{t, y}$ and calculated as

$$
\begin{equation*}
F_{t, y}=F_{t, y}^{\mathrm{df}}+F_{t, y}^{\mathrm{tb}}+F_{t, y}^{\mathrm{fb}} \tag{8}
\end{equation*}
$$

where $F_{t, y}^{\mathrm{df}}$ is the fishing mortality associated with the directed fishery, $F_{t, y}^{\mathrm{tb}}$ is the fishing mortality associated with the trawl bycatch fishery, $F_{t, y}^{\mathrm{fb}}$ is the fishing mortality associated with the fixed bycatch fishery. Each of these are derived as

$$
\begin{array}{lll}
F_{t, y}^{\mathrm{df}}=\bar{F}^{\mathrm{df}}+\delta_{t, y}^{\mathrm{df}} & \text { where } & \delta_{t, y}^{\mathrm{df}} \sim \mathcal{N}\left(0, \sigma_{\mathrm{df}}^{2}\right), \\
F_{t, y}^{\mathrm{tb}}=\bar{F}^{\mathrm{tb}}+\delta_{t, y}^{\mathrm{tb}} & \text { where } & \delta_{t, y}^{\mathrm{df}} \sim \mathcal{N}\left(0, \sigma_{\mathrm{tb}}^{2}\right), \\
F_{t, y}^{\mathrm{fb}}=\bar{F}^{\mathrm{fb}}+\delta_{t, y}^{\mathrm{fb}} & \text { where } & \delta_{t, y}^{\mathrm{df}} \sim \mathcal{N}\left(0, \sigma_{\mathrm{fb}}^{2}\right), \tag{9}
\end{array}
$$

where $\delta_{t, y}^{\mathrm{df}}, \delta_{t, y}^{\mathrm{tb}}$, and $\delta_{t, y}^{\mathrm{fb}}$ are the fishing mortality deviations for each of the fisheries, each season $t$ during each year $y, \bar{F}^{\text {df }}, \bar{F}^{\text {tb }}$, and $\bar{F}^{\text {fb }}$ are the average fishing mortalities for each fishery. The total mortality $Z_{l, t, y}$ represents the combination of natural mortality $M_{t, y}$ and fishing mortality $F_{t, y}$ during season $t$ and year $y$

$$
\begin{equation*}
\boldsymbol{Z}_{t, y}=Z_{l, t, y}=M_{t, y}+F_{t, y} \tag{10}
\end{equation*}
$$

The survival matrix $\boldsymbol{S}_{t, y}$ during season $t$ and year $y$ is

$$
\boldsymbol{S}_{t, y}=\left[\begin{array}{ccc}
1-e^{-Z_{1, t, y}} & 0 & 0  \tag{11}\\
0 & 1-e^{-Z_{2, t, y}} & 0 \\
0 & 0 & 1-e^{-Z_{3, t, y}}
\end{array}\right]
$$

The basic population dynamics underlying GMACS can thus be described as

$$
\begin{array}{lr}
\boldsymbol{n}_{t+1, y}=\boldsymbol{S}_{t, y} \boldsymbol{n}_{t, y}, & \text { if } t<5 \\
\boldsymbol{n}_{t, y+1}=\boldsymbol{G} \boldsymbol{S}_{t, y} \boldsymbol{n}_{t, y}+\boldsymbol{r}_{t, y} & \text { if } t=5 .
\end{array}
$$

## 3. Model Data

Data inputs used in model estimation are listed in Table 21.

## 4. Model Parameters

Table 22 lists fixed (externally determined) parameters used in model computations. In all scenarios, the stage-transition matrix is

$$
\boldsymbol{G}=\left[\begin{array}{ccc}
0.2 & 0.7 & 0.1  \tag{13}\\
0 & 0.4 & 0.6 \\
0 & 0 & 1
\end{array}\right]
$$

which is the combination of the growth matrix and molting probabilities.
Estimated parameters are listed in Table 23 and include an estimated natural mortality deviation parameter in 1998/99 ( $\delta_{1998}^{M}$ ) assuming an anomalous mortality event in that year, as hypothesized by Zheng and Kruse (2002), with natural mortality otherwise fixed at $0.18 \mathrm{yr}^{-1}$.

## 5. Model Objective Function and Weighting Scheme

The objective function consists of the sum of several "negative log-likelihood" terms characterizing the hypothesized error structure of the principal data inputs (Table 17). A lognormal distribution is assumed to characterize the catch data and is modelled as

$$
\begin{align*}
\sigma_{t, y}^{\text {catch }} & =\sqrt{\log \left(1+\left(C V_{t, y}^{\text {catch }}\right)^{2}\right)}  \tag{14}\\
\delta_{t, y}^{\text {catch }} & =\mathcal{N}\left(0,\left(\sigma_{t, y}^{\text {catch }}\right)^{2}\right) \tag{15}
\end{align*}
$$

where $\delta_{t, y}^{c a t c h}$ is the residual catch. The relative abudance data is also assumed to be lognormally distributed

$$
\begin{align*}
& \sigma_{t, y}^{\mathrm{I}}=\frac{1}{\lambda} \sqrt{\log \left(1+\left(C V_{t, y}^{\mathrm{I}}\right)^{2}\right)}  \tag{16}\\
& \delta_{t, y}^{\mathrm{I}}=\log \left(I^{\mathrm{obs}} / I^{\mathrm{pred}}\right) / \sigma_{t, y}^{\mathrm{I}}+0.5 \sigma_{t, y}^{\mathrm{I}} \tag{17}
\end{align*}
$$

and the likelihood is

$$
\begin{equation*}
\sum \log \left(\delta_{t, y}^{\mathrm{I}}\right)+\sum 0.5\left(\sigma_{t, y}^{\mathrm{I}}\right)^{2} \tag{18}
\end{equation*}
$$

GMACS calculates standard deviation of the normalised residual (SDNR) values and median of the absolute residual (MAR) values for all abundance indices and size compositions to help the user come up with resonable likelihood weights. For an abundance data set to be well fitted, the SDNR should not be much greater than 1 (a value much less than 1 , which means that the data set is fitted better than was expected, is not a cause for concern). What is meant by "much greater than 1 " depends on $m$ (the number of years in the data set). Francis (2011) suggests upper limits of $1.54,1.37$, and 1.26 for $m=5,10$, and 20, respectively. Although an SDNR not much greater than 1 is a necessary condition for a good fit, it is not sufficient. It is important to plot the observed and expected abundances to ensure that the fit is good.
GMACS also calculates Francis weights for each of the size composition data sets supplied (Francis 2011). If the user wishes to use the Francis iterative re-weighting method, first the weights applied to the abundance indices should be adjusted by trial and error until the SDNR (and/or MAR) are adequte. Then the Francis weights supplied by GMACS should be used as the new likelihood weights for each of the size composition data sets the next time the model is run. The user can then iteratively adjust the abudance index and size composition weights until adequate SDNR (and/or MAR) values are achieved, given the Francis weights.

## 6. Estimation

The model was implemented using the software AD Model Builder (Fournier et al. 2012), with parameter estimation by minimization of the model objective function using automatic differentiation. Parameter estimates and standard deviations provided in this document are AD Model Builder reported values assuming maximum likelihood theory asymptotics.

Table 20: Proportion of the natural mortality $\left(\tau_{t}\right)$ that is applied during each season $(t)$ in the model.

| Year | Season 1 | Season 2 | Season 3 | Season 4 | Season 5 |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 1978 | 0.00 | 0.07 | 0.00 | 0.56 | 0.37 |
| 1979 | 0.00 | 0.06 | 0.00 | 0.57 | 0.37 |
| 1980 | 0.00 | 0.07 | 0.00 | 0.56 | 0.37 |
| 1981 | 0.00 | 0.05 | 0.00 | 0.58 | 0.37 |
| 1982 | 0.00 | 0.07 | 0.00 | 0.56 | 0.37 |
| 1983 | 0.00 | 0.12 | 0.00 | 0.51 | 0.37 |
| 1984 | 0.00 | 0.10 | 0.00 | 0.53 | 0.37 |
| 1985 | 0.00 | 0.14 | 0.00 | 0.49 | 0.37 |
| 1986 | 0.00 | 0.14 | 0.00 | 0.49 | 0.37 |
| 1987 | 0.00 | 0.14 | 0.00 | 0.49 | 0.37 |
| 1988 | 0.00 | 0.14 | 0.00 | 0.49 | 0.37 |
| 1989 | 0.00 | 0.14 | 0.00 | 0.49 | 0.37 |
| 1990 | 0.00 | 0.14 | 0.00 | 0.49 | 0.37 |
| 1991 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 1992 | 0.00 | 0.14 | 0.00 | 0.49 | 0.37 |
| 1993 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 1994 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 1995 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 1996 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 1997 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 1998 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 1999 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 2000 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 2001 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 2002 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 2003 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 2004 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 2005 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 2006 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 2007 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 2008 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 2009 | 0.00 | 0.44 | 0.00 | 0.19 | 0.37 |
| 2010 | 0.00 | 0.44 | 0.00 | 0.19 | 0.37 |
| 2011 | 0.00 | 0.44 | 0.00 | 0.19 | 0.37 |
| 2012 | 0.00 | 0.44 | 0.00 | 0.19 | 0.37 |
| 2013 | 0.00 | 0.44 | 0.00 | 0.19 | 0.37 |
| 2014 | 0.00 | 0.44 | 0.00 | 0.19 | 0.37 |
| 2015 | 0.00 | 0.44 | 0.00 | 0.19 | 0.37 |
| 2016 | 0.00 | 0.44 | 0.00 | 0.19 | 0.37 |
| 2017 | 0.00 | 0.44 | 0.00 | 0.19 | 0.37 |
| 2018 | 0.00 | 0.44 | 0.00 | 0.19 | 0.37 |
| 2019 | 0.00 | 0.44 | 0.00 | 0.19 | 0.37 |
|  |  |  |  |  |  |

Table 21: Data inputs used in model estimation.

| Data | Years | Source |
| :--- | :--- | :--- |
| Directed pot-fishery retained-catch number <br> (not biomass) | $1978 / 79-1998 / 99$ <br> $2009 / 10-2015 / 16$ | Fish tickets <br> (fishery closed 1999/00-2008/09 <br> and 2016/17-2018/19) |
| Groundfish trawl bycatch biomass | $1992 / 93-2018 / 19$ | NMFS groundfish observer program |
| Groundfish fixed-gear bycatch biomass | $1992 / 93-2018 / 19$ | NMFS groundfish observer program |
| NMFS trawl-survey biomass index <br> (area-swept estimate) and CV | $1978-2019$ | NMFS EBS trawl survey |
| ADF\&G pot-survey abundance index <br> (CPUE) and CV | $1995-2018$ | ADF\&G SMBKC pot survey |

Table 22: Fixed model parameters for all scenarios.

| Parameter | Symbol | Value | Source/rationale |
| :---: | :---: | :---: | :---: |
| Trawl-survey catchability | $q$ | 1.0 | Default |
| Natural mortality | M | $0.18 \mathrm{yr}^{-1}$ | NPFMC (2007) |
| Size transition matrix | G | Equation 13 | Otto and Cummiskey (1990) |
| Stage-1 and stage-2 mean weights | $w_{1}, w_{2}$ | $0.7,1.2 \mathrm{~kg}$ | Length-weight equation (B. Foy, NMFS) applied to stage midpoints |
| Stage-3 mean weight | $w_{3, y}$ | Depends on year | Fishery reported average retained weight from fish tickets, or its average, and mean weights of legal males |
| Recruitment SD | $\sigma_{R}$ | 1.2 | High value |
| Natural mortality SD | $\sigma_{M}$ | 10.0 | High value (basically free parameter) |
| Directed fishery handling mortality |  | 0.2 | 2010 Crab SAFE |
| Groundfish trawl handling mortality |  | 0.8 | 2010 Crab SAFE |
| Groundfish fixed-gear handling mortality |  | 0.5 | 2010 Crab SAFE |

Table 23: The lower bound (LB), upper bound (UB), initial value, prior, and estimation phase for each estimated model parameter.

| Parameter | LB | Initial value | UB | Prior | Phase |
| :--- | ---: | ---: | ---: | :--- | ---: |
| Average recruitment $\log (R)$ | -7 | 10.0 | 20 | Uniform $(-7,20)$ | 1 |
| Stage-1 initial numbers $\log \left(n_{1}^{0}\right)$ | 5 | 14.5 | 20 | Uniform $(5,20)$ | 1 |
| Stage-2 initial numbers $\log \left(n_{2}^{0}\right)$ | 5 | 14.0 | 20 | Uniform $(5,20)$ | 1 |
| Stage-3 initial numbers $\log \left(n_{3}^{0}\right)$ | 5 | 13.5 | 20 | Uniform $(5,20)$ | 1 |
| ADF\&G pot survey catchability $q$ | 0 | 3.0 | 5 | Uniform $(0,5)$ | 1 |
| Stage-1 directed fishery selectivity 1978-2008 | 0 | 0.4 | 1 | Uniform $(0,1)$ | 3 |
| Stage-2 directed fishery selectivity 1978-2008 | 0 | 0.7 | 1 | Uniform $(0,1)$ | 3 |
| Stage-1 directed fishery selectivity 2009-2017 | 0 | 0.4 | 1 | Uniform $(0,1)$ | 3 |
| Stage-2 directed fishery selectivity 2009-2017 | 0 | 0.7 | 1 | Uniform $(0,1)$ | 3 |
| Stage-1 NMFS trawl survey selectivity | 0 | 0.4 | 1 | Uniform $(0,1)$ | 4 |
| Stage-2 NMFS trawl survey selectivity | 0 | 0.7 | 1 | Uniform $(0,1)$ | 4 |
| Stage-1 ADF\&G pot survey selectivity | 0 | 0.4 | 1 | Uniform $(0,1)$ | 4 |
| Stage-2 ADF\&G pot survey selectivity | 0 | 0.7 | 1 | Uniform $(0,1)$ | 4 |
| Natural mortality deviation during 1998 $\delta_{1998}^{M}$ | -3 | 0.0 | 3 | Normal $\left(0, \sigma_{M}^{2}\right)$ | 4 |
| Recruitment deviations $\delta_{y}^{R}$ | -7 | 0.0 | 7 | Normal $\left(0, \sigma_{R}^{2}\right)$ | 3 |
| Average directed fishery fishing mortality $\bar{F}^{\text {df }}$ | - | 0.2 | - | - | 1 |
| Average trawl bycatch fishing mortality $\bar{F}^{\mathrm{tb}}$ | - | 0.001 | - | - | 1 |
| Average fixed gear bycatch fishing mortality $\bar{F}^{\mathrm{fb}}$ | - | 0.001 | - | - | 1 |

## Appendix B. Data files for the reference model (16.0)

## The reference model (16.0) data file for 2020





$\left.\begin{array}{lllllllllllllll}1996 & 3 & 1 & 1 & 0 & 0 & 0 & 25 & 0.1595 & 0.2229 & 0.6176 \\ 1997 & 3 & 1 & 1 & 0 & 0 & 0 & 25 & 0.1818 & 0.2053 & 0.6128 \\ 1998 & 3 & 1 & 1 & 0 & 0 & 0 & 25 & 0.1927 & 0.2162 & 0.5911 \\ 2009 & 3 & 1 & 1 & 0 & 0 & 0 & 50 & 0.1413 & 0.3235 & 0.5352 \\ 2010 & 3 & 1 & 1 & 0 & 0 & 0 & 50 & 0.1314 & 0.3152 & 0.5534 \\ 2011 & 3 & 1 & 1 & 0 & 0 & 0 & 50 & 0.1314 & 0.3051 & 0.5636 \\ 2012 & 3 & 1 & 1 & 0 & 0 & 0 & 50 & 0.1417 & 0.3178 & 0.5406 \\ 2014 & 3 & 1 & 1 & 0 & 0 & 0 & 50 & 0.0939 & 0.2275 & 0.6786 \\ 2015 & 3 & 1 & 1 & 0 & 0 & 0 & 50 & 0.1148 & 0.2518 & 0.6333\end{array}\right]$ nno fishery so not updated

```
# MidPoint Sex Increment CV
# 97.5 1 14.1 0.2197
#112.5 1 14.1 0.2197
#127.5 1 14.1
# 97.5 1 13.8 0.2197
# 112.5 1 14.1 0.2197
# 127.5 1 14.4 0.2197
## eof
9999
```


## The reference model (16.0) control file for 2020







```
\#\# OTHER CONTROLS
```



```
1978 \# First rec_dev
2019 \# last rec_dev (updated annually)
\# Estimated rec_dev phase
\# Estimated sex_ratio
\# initial sex-ratio
\# Estimated rec_ini phase
\# VERBOSE FLAG ( \(0=\) off, \(1=\) on, 2 = objective func)
\# Initial conditions ( \(0=\) Unfished, 1 = Steady-state fished, 2 = Free parameters)
\# Lambda (proportion of mature male biomass for SPR reference points)
\# Stock-Recruit-Relationship ( \(0=\) None, \(1=\) Beverton-Holt)
\# Maximum phase (stop the estimation after this phase).
\# Maximum number of function calls
```



```
\#\# EMPHASIS FACTORS (CATCH)
\#\# ====================================================================================14 \#\#
\#Ret_POT Disc_POT Disc_trawl Disc_fixed
    \(\begin{array}{llll}1 & 1 & 1\end{array}\)
\#\# ==================================================================================14
\#\# EMPHASIS FACTORS (Priors)
```



```
\(\begin{array}{rrrrrcrl}\text { \# Log_fdevs } & \text { meanF } & \text { Mdevs } & \text { Rec_devs } \text { Initial_devs Fst_dif_dev Mean_sex-Ratio } & \\ 10000 & 1 & 1 & 1 & 0 & 0 & 1 & \text { \#(10000) }\end{array}\)
\#\# EOF
9999
```


# Appendix C. Assessing uncertainty in model output due to lack of terminal year survey data for St. Matthew blue king crab (SMBKC) 

## Introduction

NMFS trawl surveys during the summer of 2020 were cancelled due to logistic difficulties caused by the global pandemic COVID-19. Therefore, the crab assessment authors met to discuss approaches to address the potential of additional uncertainty in the current year models - specifically the projected mature male biomass and associated reference points. The objective of these approaches/simulations was to provide the crab plan team (CPT) and the scientific and statistical committee (SSC) a range of potential additional uncertainty that could be applied to the buffers used on the OFL calculations to produce an appropriate ABC for the 2020/21 crab season.

## Objectives

1. Can we characterize the additional uncertainty in the current years estimates due to the lack of terminal year survey data? If so, what does it look like?
2. Is the model uncertainty characterized in objective \#1 currently included in the ABC buffer applied to this stock or do we need to apply additional uncertainty measures?

## Approaches

## Approach 1 (and 2): retrospective patterns with and without terminal survey data

Retrospective analysis are typically performed on models to characterize the tendencies of a model to over or under estimate current trends in biomass, recruitment, etc. Retrospective patterns are described as a clear tendency for a model to either over or under estimate. Approach 1 compares the output of retrospective models with the terminal year of survey data and ones where the terminal year of trawl survey data are removed (both abundance and size composition data). Approach 2 was to do this for the last year's model - 2019 - which is included in the analysis.

A number of key model outputs were compared for these retrospective runs. These include: average recruitment, $B_{m s y}$, status of the stock, terminal year MMB, and reference point calculations (OFL).

## Results

Retrospective analysis of the base model show a retrospective pattern that tends to overestimate mature male biomass (MMB) in the terminal year (Figure 1 and 2). Using a peel of the last 5 years estimates of MMB the estimated Mohn's $\rho$ is -0.346 , which suggests a retrospecive pattern in the MMB estimates for the base model. Since 2018 the MMB estimates have been relatively stable, however, they are the lowest in the model history and reflect a time of overfished declaration for the stock.

In general, models that lacked the terminal year of survey data performed similarly to models with the survey data for each model end year (Figure 3). In cases where the model outputs differed the model without the terminal year of survey data tended to have results similar to the previous years model. For the last 5 years of retrospective model runs the models with and without the terminal year of survey data performed very similarly. These results support the hypothesis that for SMBKC in the last few years no additional uncertainty is present in the mmb estimates with the lack of the terminal year survey data (Figure 4).

Figures 5 through 10 display the small differences between these model runs in each model end year. There are some small differences in the model with and without the terminal year of survey data, but most of these exist around between 2013 and 2015 where the population was transitioning from healthy levels to overfished. This is most evident in the terminal MMB, $F_{O F L}$, and OFL comparisons for 2013 (Figures 6, 9, and 10).

Hypothetically if the uncertainty about the quantities of interest increased due to the lack of a terminal year of survey data the resuling average CVs for the quantities would be larger in runs without the terminal year of survey data. Table 2 summarises the average CVs over all years for the "normal" retrospective runs and those without the terminal year of survey data. There are small differences in the average CVs, with those in the "missing survey" retrospective runs being slightly larger on average, but this difference is small and does not suggest increase uncertainty in the "missing survey" runs.

The average percent difference between these quantities was approximately $1 \%$ overall and was the highest in OFL comparisons at an average difference of $4 \%$ (Table 1). Most differences were small and even unnoticeable in years where the population trajectory was similar to the previous year. The underlying model processes (growth, mortality, selectivity, etc.) drive the current year's model estimates without the presence of new abundance or size data, and the uncertainty about these processes has not increased with the lack of one year of survey data.
Based on this analysis the author does not recommend additional uncertainty in the ABC buffer for SMBKC for the 2020 base model.

## Approach 3: encompassing expected variability

This approach was designed to run models with "fake" 2020 data to determine how much a data point in 2020 could have potential influenced the model outcome. The same key model outputs were compared in this approach as in approach 1.

This approach evaluates the impact of different hypothetical 2020 survey outcomes, and is based on a SSC recommendation in its June minutes. Using the NMFS trawl survey time series fit in the proposed base or reference model the multiplicative residuals were calculated (predicted survey fit/observed survey data point) for each year. The 25 th and 75 th percentiles of the multiplicative residual distribution were obtained, which would represent a typical low and high value for the survey (Martin Dorn per comm.).

A predicted survey value was obtained for 2020 by running the base model with a hypothetical survey value with a very high CV (100), so that the model did not attempt to fit the observation. For SMBKC the hypothetical survey value was an average of the last 4 years of the survey to best estimate the hypothetical 2020 data point even though the CV for this data point was large. Once the base model was fit with this hypothetical data point the resulting estimate for the 2020 survey was used to complete two additional model runs. These runs multiplied the predicted 2020 survey data point by the 25 th and 75 th percentiles of the multiplicative residuals to simulate a "low" and "high" survey data point. The CV for these runs was set equal to the median survey CV. These two runs were evaluated along side the 2020 base model to determine the sensitivity of model output and management quantities on the 2020 survey data point.

## Results

Overall, the model output and management quantities did not differ much between the base and the low and high hypothetical survey data runs for 2020 (Figure 11 and Table 3).

The estimated mature male biomass trend was the same, with little difference evident when viewing the entire time series (Figure 12). A detailed view of the last 10 years is provided for the MMB estimates in order to view the small difference in the three model estimates. The trends are all similar, with the only difference being the scale of the MMB estimate in the last 7 years (Figure 13). In reference to the base model the "high" run increased the MMB by a very small amount, where the "low" run decreased the MMB trend by about twice as much. All model estimates were very similar and within the typical range of uncertainty
of the base model (Figure 14). Based on this analysis the author does not recommend additional uncertainty in the ABC buffer for SMBKC for the 2020 base model.

## Recommendations on uncertainty

The analysis performed in this appendix, including the general retrospective analysis, suggest that no additional uncertainty is neccessary for SMBKC. Any additional variability in the model estimates from not having a survey data point in 2020 would like produce a small change in the calculated 2020 OFL. The current buffer of $20 \%$ includes the expected uncertainty in the model output that is observed in the retrospective analysis, adding to this uncertainty does not appear neccessary at this time.

The current status of the stock is still overfished, and the directed fishery is closed. The only harvest for this stock comes from bycatch in the groundfish and other crab fisheries which occurs at very low levels. While increasing the buffer on the ABC would not impact these fisheries, it also does not appear neccessary to keep the bycatch numbers well below the projected ABC.

## Figures



Figure 1: Retrospective run estimates of mature male biomass (mmb) for the SMBKC reference model (16.0) for the last 10 years.


Figure 2: Retrospective run estimates of mature male biomass (mmb) for the SMBKC reference model (16.0) for the last 10 years, only showing the last 20 years for a detailed view.


Figure 3: Retrospective run estimates of mature male biomass (mmb) for the SMBKC reference model (16.0) including models that eliminated the terminal year survey data.


Figure 4: Retrospective run estimates of mature male biomass ( mmb ) for the SMBKC reference model (16.0) including models that eliminated the terminal year survey data for the last 5 model years. Highlighting the last 20 years for a more detailed view.


Figure 5: Comparison of average recruitment model estimates from 'normal' retrospective runs and those without the terminal year survey data.


Figure 6: Comparison of Bmsy model estimates from 'normal' retrospective runs and those without the terminal year survey data.

Status ( $\mathrm{B}_{\mathrm{pr}} / \mathrm{B}_{\mathrm{MSY}}$ )


Figure 7: Comparison of the model estimate of 'status' (B/Bmsy) from 'normal' retrospective runs and those without the terminal year survey data.


Figure 8: Comparison of the model estimate of terminal year mmb from 'normal' retrospective runs and those without the terminal year survey data.


Figure 9: Comparison of the model estimate of fof from 'normal' retrospective runs and those without the terminal year survey data.

## OFL



Figure 10: Comparison of the model estimate of OFL from 'normal' retrospective runs and those without the terminal year survey data.

Approach 3 - high, low, 2020 base


Figure 11: Model output and reference points from approach 3. Comparing the 2020 base model with a model that has a high 'fake' 2020 survey data point and one that has a low 'fake' survey data point.


Model

- model 16.0 (2020 base)
- 2020 base - App 3 low
- 2020 base - App 3 high

Figure 12: Mature male biomass estimates from approach 3. Comparing the 2020 base model with a model that has a high 'fake' 2020 survey data point and one that has a low 'fake' survey data point.


Figure 13: Mature male biomass estimates from approach 3. Comparing the 2020 base model with a model that has a high 'fake' 2020 survey data point and one that has a low 'fake' survey data point, only showing the last 10 years for detail on model differentiation.


Figure 14: Mature male biomass estimates with associated variability from approach 3. Comparing the 2020 base model with a model that has a high 'fake' 2020 survey data point and one that has a low 'fake' survey data point, only showing the last 20 years for detail on model differentiation.

## Tables

Table 1: Comparisons of the percent difference in parameter estimates for the retrospective models with and without the terminal year of survey data.

| Year | AvgR | Bmsy | Terminal MMB | Status | Fofl | OFL |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| 2010 | -3.921 | -0.606 | -0.582 | 0.024 | 0.000 | -1.692 |
| 2011 | -1.980 | -0.117 | -5.674 | -5.564 | 0.000 | -3.183 |
| 2012 | 1.410 | 0.835 | 0.863 | 0.027 | 0.000 | 3.898 |
| 2013 | 9.199 | 3.471 | 30.491 | 26.113 | 30.537 | 72.124 |
| 2014 | -0.399 | -0.208 | -5.101 | -4.903 | -5.563 | -7.861 |
| 2015 | -2.176 | 0.037 | -1.912 | -1.948 | -2.345 | -3.588 |
| 2016 | -2.469 | -0.256 | -3.270 | -3.021 | -3.816 | -6.579 |
| 2017 | 0.602 | 0.125 | -0.364 | -0.488 | -0.713 | -0.419 |
| 2018 | -1.882 | -0.630 | -4.642 | -4.038 | -6.091 | -10.343 |
| 2019 | 0.501 | -1.927 | -4.270 | -2.389 | -3.722 | -2.330 |
| RMS | 3.479 | 1.318 | 10.214 | 8.787 | 10.173 | 23.368 |

Table 2: Average CV over all years (2010-2019) for normal retrospective runs and those missing the terminal year of survey data.

| Type | CV-Bmsy | CV-OFL | CV-status | CV-temrinal-SSB |
| :--- | ---: | ---: | ---: | ---: |
| retro | 4.32 | 20.19 | 11.12 | 11.77 |
| missing-survey | 4.36 | 21.42 | 11.71 | 12.51 |

Table 3: Comparisons of the percent difference in parameter estimates for the low and high models in approach 3 compared to the 2020 base model (16.0).

| Variable | Diff-Ltobase | Diff-Htobase |
| :--- | ---: | ---: |
| avgR | -2.176 | 2.020 |
| Bmsy | -0.291 | 0.156 |
| Terminal-MMB | -2.746 | 1.226 |
| Status | -2.463 | 1.068 |
| F-ofl | -3.586 | 1.477 |
| OFL | -7.261 | 6.303 |

# Appendix D. Ecosystem and Socioeconomic Profile of the Saint Matthew Blue King Crab Stock 

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September 2020


With Contributions from:
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## Executive Summary

National initiative and NPFMC recommendations suggest a high priority for conducting an ecosystem and socioeconomic profile (ESP) for Saint Matthew blue king crab (SMBKC) due to the stock's current overfished status and poor recruitment in recent years. Scores for stock assessment prioritization, habitat prioritization, climate vulnerability assessment, and data classification analysis were moderate to high. Furthermore, in 2018 when the stock was declared overfished, the Crab Plan Team requested an evaluation of ecosystem factors to inform the stock rebuilding plan.

We follow the standardized template for conducting an ESP and present results of applying the ESP process through a metric and subsequent indicator assessment. We use information from a variety of data streams available for the SMBKC stock. Analysis of the ecosystem and socioeconomic processes for SMBKC by life history stage along with information from the literature identified a suite of indicators for testing and continued monitoring within the ESP. Results of the metric and indicator assessment are summarized below as ecosystem and socioeconomic considerations that can be used for evaluating concerns in the main stock assessment.

Please refer to the last full ESP document for further information regarding the ecosystem and socioeconomic linkages for this stock (Fedewa et al., 2019, available online within the SMBKC SAFE, Appendix E, pp. 99-120 at: https://meetings.npfmc.org/CommentReview/DownloadFile?p=6ffde3ce-67be-4139-b165-cbff9062da06.pdf\&fileName=C4\ 6\ SMBKC\ SAFE\ 2019.pdf).

## Summary of Changes in Assessment Inputs

## Changes in the Metric or Indicator Data

The 2020 SMBKC ESP update includes a suite of new ecosystem indicators that were developed from remote sensing data and Bering 10K ROMS model output hindcasts. The suite of socioeconomic indicators for SMBKC remain unchanged due to the continued closure of the fishery while the stock rebuilds.

## Changes in the Indicator Analysis

We have included the addition of a Stage 2 Importance Test in the Indicator Analysis section of the 2020 SMBKC ESP update. Results from the analysis are outlined below.

## Summary of Results

Important ecosystem and socioeconomic processes that may identify dominant pressures on the SMBKC stock were reviewed in the last full ESP document. We updated the suite of ecosystem indicators for SMBKC using these mechanistic linkages or hypothesized relationships. Specifically, the addition of spring bottom temperature, wind stress and chlorophyll $a$ indicators likely represent environmental conditions and prey availability for BKC early life stages. Please reference the 2019 full SMBKC ESP document for complete descriptions of indicators that occurred in the last full ESP. Any changes in methodology for indicators developed in 2019 are outlined below, as well as full descriptions for new indicators.

## Indicator Suite

## Ecosystem Indicators:

1.) Physical Indicators

- Cold Pool Index: Due to the cancelation of the 2020 EBS summer bottom trawl survey, the cold pool index was calculated from ROMS model output as the fraction of the EBS
survey area with bottom waters less than $2^{\circ} \mathrm{C}$ on July 1 of each year (Kearney et al., 2020).
- Summer Bottom Temperature: Due to the cancelation of the 2020 EBS summer bottom trawl survey, June-July bottom temperatures were averaged within the SMBKC management area from ROMS model output (Kearney et al., 2020).
- Spring Bottom Temperature: Average of Feb-March bottom temperatures within the SMBKC management boundary from ROMS model output (Kearney et al., 2020).
- Corrosivity Index: Percent of the SMBKC management area containing an average bottom aragonite saturation state of < from Feb-April (D. Pilcher, pers. commun., 2020)
- Chlorophyll $a$ Biomass: April-June average chlorophyll-a biomass within the St. Matthew region of the Bering Sea; calculated with 8-day composite data from MODIS satellites (J. Nielsen, pers. commun., 2020)
- Wind Stress: June ocean surface wind stress within the SMBKC management boundary. Product of NOAA blended winds and MetOp ASCAP sensors from multiple satellites (Zhang et al., 2006, NOAA/NESDIS, CoastWatch)
2.) Biological Indicators
- Pacific Cod Biomass: Pacific cod comprise the majority of total biomass in the Benthic Predator Biomass indicator developed for the 2019 full ESP document. As such, we refined a predation indicator to solely include pacific cod biomass within the SMBKC management area.
- Benthic Invert Biomass
- SMBKC Recruit Biomass (Palof, pers. commun, 2020)

Socioeconomic Indicators:
1.) Fishery Performance Indicators

- CPUE (mean no. of crabs per potlift): Fishing effort efficiency, as measured by estimated mean number of retained SMBKC per potlift.
- Total Potlifts: Fishing effort, as measured by estimated number of crab pots lifted by vessels during the SMBKC fishery.
- Vessels active in fishery: Annual count of crab vessels that delivered commercial landings of SMBKC to processors.
- SMBKC male bycatch biomass: Incidental bycatch biomass estimates of male BBRKC (tons) in trawl and fixed gear fisheries
2.) Economic Indicators
- TAC Utilization (\%): Percentage of the annual SMBKC TAC (GHL prior to 2005) that was harvested by active vessels, including deadloss discarded at landing.
- SMBKC ex-vessel revenue share (\% of total exvessel revenue): SMBKC ex-vessel revenue share as percentage of total calendar year ex-vessel revenue from all commercial landings in Alaska fisheries, mean value over all vessels active in SMBKC during the respective year.
- Ex-vessel price per pound: commercial value per unit (pound) of SMBKC landings (as adjusted by CFEC to account for post-season adjustments to ex-vessel settlements), measured as weighted average value over all ex-vessel sales reported.
3.) Community Indicators
- Processors active in fishery: Total number of crab processors that purchased landings of SMBKC from delivering vessels during the calendar year. This provides an indicator of the level of participation of buyers in the market for SMBKC landings.
- Local Quotient of SMBKC landed catch in Saint Paul: Ex-vessel value share of SMBKC landings to communities on St. Paul Island, as percentage of total value of commercial landings to St. Paul processors from all commercial Alaska fisheries, as aggregate
percentage over all landings during the respective year. St Paul represents the principal port of landing for the SMBKC fishery during the post-rationalization period, representing from $78 \%$ to $100 \%$ of all purchased landings in the fishery. The local quotient (LQ) represents the share of community landings attributed to SMBKC in relation to revenue from all other species landed in the community during years when the fishery was opened.


## Indicator Analysis

We provide an update to the list and time-series of ecosystem and socioeconomic indicators (Tables 1-2, Figures 1-2) and then report the results of the first and second stage statistical tests for the indicator analysis with the inclusion of current-year data. The third stage has not yet been completed, and will require more indicator development and review of the ESP modeling applications.

## Stage 1: Traffic Light Test

The first stage of the indicator analysis is a simple assessment of the most recent year relative value and a traffic-light evaluation of the most current year where available (Tables 1-2). Details of the analysis can be found in the 2019 full ESP document.

Current year trends suggest relatively average environmental conditions for the SMBKC stock in 2020, although SMBKC recruit biomass is still well below the long-term average (Figure 1). While summer bottom temperatures in the St. Matthew management area were $1-2^{\circ} \mathrm{C}$ below 2018-2019 temperatures, the region still experienced warmer than average conditions relative to the long-term mean. However, a larger fraction of bottom waters were $<2^{\circ} \mathrm{C}$ in 2020 compared to previous years. The addition of a corrosivity indicator suggests that SMBKC are exposed to significant interannual variability in the aragonite saturation state of bottom waters. All stations within the SMBKC management area contained undersaturated bottom waters ( $\Omega$ arag < 1) in spring 2020 which suggests potential consequences for shell formation following the spring molt, as well as reduced condition and survival of embryos and larval stages.

Chlorophyll $a$ biomass was above the long-term average in 2020, suggesting a more intense spring bloom and good first-feeding conditions for BKC larvae. Likewise, June wind speeds around St. Matthew Island were near-average in 2020 and on a downward trend since 2015, which may promote increased larval encounter rates with diatom prey. Current-year data for benthic invertebrate and Pacific cod biomass indicators were not available due to the cancellation of the EBS bottom trawl survey. Benthic invertebrate biomass has remained high since the late 1980's (possibly coinciding with a 1989 regime shift in the North Pacific), while Pacific cod biomass has been on a downward trend after reaching an all-time high in 2016.

With the exception of SMBKC male bycatch, all socioeconomic indicators in Table 2 are derived from SMBKC fishery data reported from the most recent open season (2015/16), and thus are not updated in this report. Bycatch of SMBKC in the groundfish fisheries during 2019 was near the lower bound of the historical range, and was slightly reduced from 2018.

## Stage 2: Importance Test

Bayesian adaptive sampling (BAS) was used for the second stage statistical test to quantify the association between hypothesized predictors and SMBKC mature male biomass (MMB), and to assess the strength of support for each hypothesis. BAS explores model space, or the full range of candidate combinations of predictor variables, to calculate marginal inclusion probabilities for each predictor, model weights for each combination of predictors, and generate Bayesian model averaged predictions for
outcomes (Clyde et al., 2011). In this second test, the full set of indicators is first winnowed to the predictors that could directly relate to MMB, and have consistent temporal data coverage. We then provide the mean relationship between each predictor variable and $\log$ MMB over time (Figure 3a), with error bars describing the uncertainty ( 1 standard deviation) in each estimated effect and the marginal inclusion probabilities for each predictor variable (Figure 3b). A higher probability indicates that the variable is a better candidate predictor of SMBKC MMB. The highest ranked predictor variables ( $\geq 0.25$ inclusion probability) were: SMBKC recruit biomass, summer bottom temperatures, and benthic invertebrate biomass. Unfortunately, due to the nature of the BAS model only being able to fit years with complete observations for each covariate, the final subset of covariates was quite small and creates a significant data gap. Despite this shortcoming, predictive performance of the BAS model appears to generally capture SMBKC MMB trends across the time series (Figure 3d).

## Ecosystem Considerations

- Despite repeated fishery closures, SMBKC mature male biomass and recruitment estimates remain below-average following a 1989 regime shift in the Bering Sea, suggesting that environmental factors may be impeding recruitment success and stock recovery.
- Highly specific thermal optimums and habitat requirements of SMBKC likely limit mobility in response to warmer than average bottom temperatures and shifting predator distributions in the Bering Sea.
- Large catches of Pacific cod in the St. Matthew Island management boundary in 2016 preceded declines in BKC mature male biomass, recruitment, and the overfished declaration in 2018.
- Trend modeling for SMBKC ecosystem indicators revealed near-average conditions for SMBKC in 2020, although persistent, corrosive bottom waters surrounding St. Matthew Island suggest potential impacts on shell formation, growth and survival of BKC.


## Socioeconomic Considerations

- Vessel engagement in the SMBKC fishery as measured by annual counts of active vessels during years that the fishery has opened, has declined relative to the pre-rationalization period reflecting consolidation of the crab fleet following rationalization.
- In the most recent open seasons, the active fleet has been reduced to 3-4 vessels, with TAC utilization also declining to $26 \%$ during the 2015/16 season.
- Ex-vessel revenue share and the Local Quotient for Saint Paul both reached high values during 2010, concurrent with a peak in ex-vessel price; large declines in both metrics over the subsequent open seasons, despite relatively high ex-vessel prices during the next four open SMBKC seasons indicate that both vessels and processors active during those years have shifted into other fisheries.


## Data Gaps and Future Research Priorities

Additional data on BKC life history characteristics (i.e. growth-per-molt data and molting probabilities) as well as estimates for natural mortality would aide in a better understanding of stage-specific vulnerabilities for the metric panel. In addition, process-based studies are necessary in order to identify links between larval survival, recruitment and environmental factors. Examining larval drift patterns and spatial distributions of mature BKC around St. Matthew Island in relation to habitat characteristics will help to inform essential fish habitat models and support the future development of a larval retention indicator. Developing an EFH habitat indicator for SMBKC should also be prioritized, as metric assessment results highlighted several vulnerabilities related to habitat. Furthermore, given the prevalence of corrosive bottom water conditions in the SMBKC management area, continued research efforts should focus on the potential impacts of ocean acidification on BKC physiology and the role pH levels may play in determining habitat use and spatial distributions of the stock.

In most socioeconomic dimensions, SMBKC fishery is relatively data rich in many respects. In the context of the ESP, however, the intermittent nature of the fishery and reliance on fishery-dependent socioeconomic data limits the available socioeconomic information to years when the fishery has opened. This complicates the depiction and/or interpretation of long-term averages for most socioeconomic indicators and suggests the need for development of indicators that are informative of social and economic factors relevant to the purposes of the ESP, but function on a continuous basis, including during years when the fishery is closed. Potential examples include estimation of current value of PSMFC QS assets, calculation of revenue share metrics for SMBKC processors and vessels identified with the SMBKC fishery on the basis of more continuous association than participation in the fishery during a particular year. Substantial improvements over the indicators reported above are feasible, however, are largely dependent on further development of clear objectives for the inclusion of social and economic indicators within the ESP framework.

## Responses to SSC and Plan Team Comments on ESPs in General

"Regarding ESPs in general, the SSC recommends development of a method to aggregate indices into a score that could be estimated over time and compared to stock history. One potential pathway forward may be to normalize and use an unweighted sum of all the indicators where all time series overlap, or just assign +1 or -1 to each indicator so that a neutral environment would be zero." (SSC, February 2020, pg. 7)

A presentation on a scoring option for the indicator suite was provided in the ESP Model Workshop in March 2020. The score used a simple $+1,0$, and -1 assignment to the indicator based on whether the current year was above, within, or below 1 standard deviation from the mean for the time series. Sablefish and GOA pollock were provided as case studies and scores were calculated historically for the past 15 years. The score timeline trajectory was also evaluated with respect to the general ecosystem and socioeconomic considerations provided in the ESP documents. We plan to provide this score in next year's ESPs for SMBKC and hope for feedback on the method.

## Responses to SSC and Plan Team Comments Specific to this ESP

"The SSC is very pleased to see the Ecosystem and Socioeconomic Profile for SMBKC. The conceptual model was appreciated especially by those that are less familiar with crab life history characteristics. The introduction of some new ecosystem indicators was a good start. It was noted that the stock showed a high vulnerability to ocean acidification (OA), so if there is a way to index OA in the ESP that might be a good addition." (SSC, Oct, 2019, pg. 12)
In response to this recommendation, we updated the 2020 SMBKC ecosystem indicator suite to include a Corrosivity Index developed from Bering 10K ROMS output. This index, representing the percent of SMBKC management area containing low pH bottom waters undersaturated in aragonite, will provide the means to highlight vulnerabilities across BKC life stages to acidified conditions.
"The SMBKC ESP provides a tool to track, for the first time, the socioeconomic context of a fishery that has not successfully provided for the continuous, sustained participation of fishing communities over time. The SSC recommends that the ESP be augmented to track indices of community engagement and dependency, by community or aggregations of communities, across the relevant vessel and processing sectors and, for the years following rationalization, quota share ownership by community by share type. Where data confidentiality constraints dictate, the analysts should consider the use of regional as well as local quotient indicators." (SSC, Oct, 2019, pg. 12)

This recommendation has not been accomplished in this update. AFSC is currently developing a dedicated annual report to accompany the Crab and Groundfish Economic SAFE reports, focused on providing comprehensive analysis and monitoring of community participation and engagement in groundfish and crab fisheries. The Annual Community Engagement and Participation Overview
(ACEPO) will provide detailed, community-level metrics of fishery participation, including income and employment, and ownership of vessel, plant, permit and quota share assets. Development of methods and indices for effectively capturing these and other dimensions of management effects on communities is currently concentrated on producing the ACEPO report. It is expected that this will provide the basis for identifying reduced-form indicators of community effects that will be suitable for incorporation in ESPs in the future.

## Acknowledgements

We would like to thank all contributors and stock assessment authors for their timely response to requests and questions regarding data, report summaries, and manuscripts. We also thank all attendees and presenters at ESP Data workshops (May 2019 and March 2020) for their valuable insight on the development of the BBRKC ESP and future indicator development. Lastly, we thank the Crab Plan Team, North Pacific Fisheries Management Council, and AFSC for supporting the development of this report and future reports.

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Table 1. First stage ecosystem indicator analysis for St. Matthew blue king crab (SMBK), including indicator title and short description. The most recent year relative value (greater than (+), less than (-) or within 1 standard deviation $(\cdot)$ of long-term mean) of the time series is provided. Fill color is based on a traffic light evaluation for SMBKC of the current year conditions relative to 1 standard deviation of the longterm mean $($ white $=$ average, blue $=$ good, red $=$ poor, no fill $=$ no current year data) .

| Title | Description | Recent |
| :---: | :---: | :---: |
| Cold Pool Index | Fraction of the EBS BT survey area with bottom water less than $2^{\circ} \mathrm{C}$ on 1 July of each year from Bering10K ROMS model output hindcasts | $\bullet$ |
| Summer Bottom Temperature | Average of June-July bottom temperatures ( ${ }^{\circ} \mathrm{C}$ ) within the SMBKC management boundary from the Bering 10K ROMS model output hindcasts | $\bullet$ |
| Corrosivity Index | Percent of the SMBKC management area containing an average bottom aragonite saturation state of $<1$ from FebApril | + |
| Spring Bottom Temperature | Average of Feb-March bottom temperatures $\left({ }^{\circ} \mathrm{C}\right)$ within the SMBKC management boundary from the Bering 10K ROMS model output hindcasts | $\bullet$ |
| Wind Stress | June ocean surface wind stress within the SMBKC management boundary. Product of NOAA blended winds and MetOp ASCAP sensors from multiple satellites | $\bullet$ |
| Chlorophyll-a Biomass | April-June average chlorophyll-a biomass within the St. Matthew region; calculated with 8-day composite data from MODIS satellites | $\bullet$ |
| Pacific cod biomass | Biomass ( $1,000 \mathrm{t}$ ) of Pacific cod within the SMBKC management boundary on the EBS bottom trawl survey | $\bullet$ |
| Benthic invertebrate biomass | Combined biomass $(1,000 \mathrm{t})$ of benthic invertebrates within the SMBKC management boundary on the EBS bottom trawl survey | + |
| SMBKC Prerecruit Biomass | Model estimates for SMBKC recruitment. Includes male crab ( $90-104 \mathrm{~mm} \mathrm{CL}$ ) that will likely enter the fishery the following year. | $\bullet$ |

Table 2. First stage socioeconomic indicator analysis for St. Matthew blue king crab (SMBK), including indicator title and short description. The most recent year relative value (greater than ( + ), less than $(-)$ or within 1 standard deviation $(\cdot)$ of long-term mean) of the time series is provided. Fill color is based on a traffic light evaluation for SMBKC of the current year conditions relative to 1 standard deviation of the longterm mean $($ white $=$ average, blue $=$ good, red $=$ poor, no fill $=$ no current year data ).

| Title | Description | Recent |
| :---: | :---: | :---: |
| Vessels active in fishery | Annual count of crab vessels that delivered commercial landings of SMBKC to processors ${ }^{1}$ | $\bullet$ |
| TAC Utilization | Percentage of the annual SMBKC TAC (GHL prior to 2005) that was harvested by active vessels, including deadloss discarded at landing. | $\bullet$ |
| Total Potlifts | Fishing effort, as measured by estimated number of crab pots lifted by vessels during the SMBKC fishery | + |
| CPUE | Fishing effort efficiency, as measured by estimated mean number of retained SMBKC per potlift | $\bullet$ |
| Ex-vessel price per pound | Commercial value per unit (pound) of SMBKC landings (as adjusted by CFEC to account for post-season adjustments to ex-vessel settlements), measured as weighted average value over all ex-vessel sales reported. | $\bullet$ |
| SMBKC ex-vessel revenue share | SMBKC ex-vessel revenue share as percentage of total calendar year ex-vessel revenue from all commercial landings in Alaska fisheries, mean value over all vessels active in SMBKC during the respective year. | $\bullet$ |
| Processors active in fishery | Total number of crab processors that purchased landings of SMBKC from delivering vessels during the calendar year. | - |
| Local Quotient of SMBKC landed catch in St. Paul | Ex-vessel value share of SMBKC landings to communities on St. Paul Island, as percentage of total value of commercial landings to St. Paul processors from all commercial Alaska fisheries, aggregate percentage over all landings during the respective year. | $\bullet$ |
| SMBKC Male <br> Bycatch in Groundfish Fishery | Incidental bycatch biomass estimates of male SMBKC (tons) in trawl and fixed gear fisheries | $\bullet$ |

[^3]

St. Matthew Summer Bottom Temperature



St Matthew Spring Bottom Temperature



Chlorophyll-a Biomass


Figure 1. Selected ecosystem indicators for SMBKC with time series ranging from 1980-2020. Upper and lower dotted horizontal lines are $90^{\text {th }}$ and $10^{\text {th }}$ percentiles of time series. Dashed horizontal line is mean of time series. Light green shaded area represents most recent year data for traffic light analysis.



Figure 1. (cont.) Selected ecosystem indicators for SMBKC with time series ranging from 1980-2020. Upper and lower dotted horizontal lines are $90^{\text {th }}$ and $10^{\text {th }}$ percentiles of time series. Dashed horizontal line is mean of time series. Light green shaded area represents most recent year data for traffic light analysis.



Figure 2. Selected socioeconomic indicators for SMBKC with time series ranging from 1980-2019. Upper and lower dotted horizontal lines are $90^{\text {th }}$ and $10^{\text {th }}$ percentiles of time series. Dashed horizontal line is mean of time series. Light green shaded area represents most recent year data for traffic light analysis.



TAC Utilization (\%)


Figure 2. (cont.) Selected socioeconomic indicators for SMBKC with time series ranging from 1980 2019. Upper and lower dotted horizontal lines are $90^{\text {th }}$ and $10^{\text {th }}$ percentiles of time series. Dashed horizontal line is mean of time series. Light green shaded area represents most recent year data for traffic light analysis.


Figure 3. Bayesian adaptive sampling output showing the mean relationship and uncertainty ( $\pm 1 \mathrm{SD}$ ) with log-transformed St. Matthew blue king crab mature male biomass: a) the estimated effect and b) marginal inclusion probabilities for each predictor variable of the subsetted covariate ecosystem indicator dataset. Output also includes model c) predicted fit ( $1: 1$ line) and $d$ ) average fit across the MMB time series.


[^0]:    ${ }^{1} 1983 / 84$ refers to a fishing year that extends from 1 July 1983 to 30 June 1984.

[^1]:    ${ }^{2}$ NOAA grant Bering Sea Crab Research II, NA16FN2621, 1997.

[^2]:    ${ }^{3}$ D. Pengilly, ADF\&G, pers. comm.

[^3]:    ${ }^{1}$ Includes crab catcher/processors that harvested and processed SMBKC catch on-board.

