

BRISTOL BAY RED KING CRAB STOCK ASSESSMENT IN SPRING 2018

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Executive Summary

1. Stock: red king crab (RKC), *Paralithodes camtschaticus*, in Bristol Bay, Alaska.
2. Catches: The domestic RKC fishery began to expand in the late 1960s and peaked in 1980 with a catch of 129.95 million lbs (58,943 t). The catch declined dramatically in the early 1980s and remained at low levels during the last three decades. Catches during recent years until 2010/11 were among the high catches in last 15 years. The retained catch in 2016/17 was approximately 8.5 million lbs (3,924 t), below the catch in 2015/16 (10 million lbs). The magnitude of bycatch from groundfish trawl and fixed gear fisheries has been stable and small relative to stock abundance during the last 10 years.
3. Stock biomass: Estimated mature biomass increased dramatically in the mid-1970s and decreased precipitously in the early 1980s. Estimated mature crab abundance had increased during 1985-2009 with mature females being about three times more abundant in 2009 than in 1985 and mature males being about two times more abundant in 2009 than in 1985. Estimated mature abundance has steadily declined since 2009.
4. Recruitment: Estimated recruitment was high during 1970s and early 1980s and has generally been low since 1985 (1979 year class). During 1984-2017, only in 1984, 1986, 1995, 1999, 2002 and 2005 were estimated recruitments above the historical average for 1976-2017. Estimated recruitment was extremely low during the last nine years.
5. Management performance:

Status and catch specifications (1,000 t) (scenario 2b):

| Year | MSST | Biomass (MMB) | TAC | Retained Catch | Total Catch | OFL | ABC |
|---------|--------------------|--------------------|------|----------------|-------------|------|------|
| 2013/14 | 12.85 ^A | 27.12 ^A | 3.90 | 3.99 | 4.56 | 7.07 | 6.36 |
| 2014/15 | 13.03 ^B | 27.25 ^B | 4.49 | 4.54 | 5.44 | 6.82 | 6.14 |
| 2015/16 | 12.89 ^C | 27.68 ^C | 4.52 | 4.61 | 5.34 | 6.73 | 6.06 |
| 2016/17 | 12.53 ^D | 25.81 ^D | 3.84 | 3.92 | 4.28 | 6.64 | 5.97 |
| 2017/18 | | 21.31 ^D | | | | 5.60 | 5.04 |

The stock was above MSST in 2016/17 and hence was not overfished. Overfishing did not occur.

Status and catch specifications (million lbs):

| Year | MSST | Biomass (MMB) | TAC | Retained Catch | Total Catch | OFL | ABC |
|---------|-------------------|-------------------|------|----------------|-------------|-------|-------|
| 2013/14 | 28.3 ^A | 59.9 ^A | 8.60 | 8.80 | 10.05 | 15.58 | 14.02 |
| 2014/15 | 28.7 ^B | 60.1 ^B | 9.99 | 10.01 | 11.99 | 15.04 | 13.53 |
| 2015/16 | 28.4 ^C | 61.0 ^C | 9.97 | 10.17 | 11.77 | 14.84 | 13.36 |
| 2016/17 | 27.6 ^D | 56.9 ^D | 8.47 | 8.65 | 9.45 | 14.63 | 13.17 |
| 2017/18 | | 47.0 ^D | | | | 12.35 | 11.11 |

Notes:

A – Calculated from the assessment reviewed by the Crab Plan Team in September 2014

B – Calculated from the assessment reviewed by the Crab Plan Team in September 2015

C – Calculated from the assessment reviewed by the Crab Plan Team in September 2016

D – Calculated from the assessment reviewed by the Crab Plan Team in September 2017

6. Basis for the OFL: All table values are in 1000 t (Scenario 2b):

| Year | Tier | B _{MSY} | Current MMB | B/B _{MSY} (MMB) | F _{OFL} | Years to define B _{MSY} | Natural Mortality |
|---------|------|------------------|-------------|--------------------------|------------------|----------------------------------|-------------------|
| 2013/14 | 3b | 26.4 | 25.0 | 0.95 | 0.27 | 1984-2013 | 0.18 |
| 2014/15 | 3b | 25.7 | 24.7 | 0.96 | 0.28 | 1984-2014 | 0.18 |
| 2015/16 | 3b | 26.1 | 24.7 | 0.95 | 0.27 | 1984-2015 | 0.18 |
| 2016/17 | 3b | 25.8 | 24.0 | 0.93 | 0.27 | 1984-2016 | 0.18 |
| 2017/18 | 3b | 25.1 | 21.3 | 0.85 | 0.24 | 1984-2017 | 0.18 |

Basis for the OFL: All table values are in million lbs:

| Year | Tier | B_{MSY} | Current MMB | B/B_{MSY} (MMB) | F_{OFL} | Years to define B_{MSY} | Natural Mortality |
|---------|------|-----------|----------------|----------------------|-----------|---------------------------------|----------------------|
| 2013/14 | 3b | 58.2 | 55.0 | 0.95 | 0.27 | 1984-2013 | 0.18 |
| 2014/15 | 3b | 56.7 | 54.4 | 0.96 | 0.28 | 1984-2014 | 0.18 |
| 2015/16 | 3b | 57.5 | 54.4 | 0.95 | 0.27 | 1984-2015 | 0.18 |
| 2016/17 | 3b | 56.8 | 52.9 | 0.93 | 0.27 | 1984-2016 | 0.18 |
| 2017/18 | 3b | 55.2 | 47.0 | 0.85 | 0.24 | 1984-2017 | 0.18 |

A. Summary of Major Changes

1. Change to management of the fishery: None.

2. Changes to the input data:

- a. Updated summer trawl survey data and directed pot fishery catch and bycatch data through 2017.
- b. Updated 2016 BSFRF side-by-side trawl survey data in May 2017. In doing so, the total survey biomass decreased from 87,725.1 t initially estimated in September 2016 to 77,815.7 t in the final estimate, about 11.3% reduction. The initial estimate mistakenly included the tows conducted in the recruitment study.
- c. Updated groundfish fisheries bycatch data during 2009-2016 and separated bycatch data by trawl fisheries and fixed gear fisheries.

3. Changes to the assessment methodology:

- a. Two scenarios (scenarios 2bn1 and 2bn2 below) are introduced to deal with observer data without legal crab discard information in the directed pot fishery. The approach to estimate male annual bycatch biomass in the directed pot fishery is detailed in Appendix C.
- b. Five model scenarios are compared in this report (See Section E.3.a for details):

Scenario 2b: the scenario 2b in the SAFE report in September 2017. This scenario assumes that BSFRF survey capture probabilities are 1.0 for all length groups. Under this assumption, NMFS survey selectivities are the products of crab availabilities (equal to BSFRF survey selectivities) and NMFS survey capture probabilities. A survey capture probability for a length group is simply defined as the proportion of the crab in the length group within the area-swept that is caught by the survey net. Also, groundfish fisheries bycatch is separated into trawl fisheries and fixed gear fisheries.

Scenario 2bn1: the same as scenario 2b except that the approach in Appendix C is used to estimate for male annual bycatch biomass since rationalization (2005-present) and total observer male length composition data are used to compute log negative likelihoods in the directed pot fishery. Due to high grading problems in some years since rationalization, two

logistic curves are estimated for retained proportions: one before rationalization (before 2005) and another for 2005-present. See approach 1 in Appendix C.

Scenario 2bn2: the same as scenario 2bn1 except that only one logistic curve is estimated for all years for retained proportions and annual retention adjusted factors are estimated to modify retained proportions for years after 2004. See approach 2 in Appendix C.

Scenario 2b85: the same as Scenario 2b except for starting the model in 1985. This approach is to check NMFS trawl survey catchability estimates without the sharp decline of abundance data in the early 1980s.

Scenario 2c85: the same as Scenario 2c in the draft SAFEE report in May 2017 except for starting the model in 1985 and constant natural mortality of 0.18 for all years.

- c. Analysis of terminal year of recruitment and investigating dynamic B_0 are conducted (see Appendix D).

4. Changes to assessment results:

The population biomass estimates in 2017 are lower than those in 2016. Among the five scenarios, model estimated relative survey biomasses are very similar except slightly lower for scenario 2bn1 during recent years. The absolute population biomass estimates are higher for scenario 2b85 due to lower trawl survey catchability estimate and lower for scenario 2bn1 during last 20 years due to slightly higher estimated trawl survey catchability value. Initial biomass estimates in 1985 for scenarios 2b85 and 2c85 are higher than other three scenarios. The Subtraction approach estimating male discard biomass and combining approach 2 to estimate retained proportions result in population biomass estimates (scenario 2bn2) similar to those from scenario 2b. We recommend scenario 2bn2 for September 2018 assessment because of better fit to the data and better estimating male discard biomass in the directed pot fishery.

B. Responses to SSC and CPT Comments

1. Responses to the most recent two sets of SSC and CPT comments on assessments in general:

CPT and SSC Comments (from January and February 2018)

Conduct a dynamic B_0 analysis and a retrospective analysis of terminal years of recruitment for the CPT meeting of May 2018.

Response: These two analyses are presented in this draft report (see retrospective results and Appendix D).

CPT comments (from January 2018)

“The CPT requested for the May 2018 meeting that assessment authors evaluate the impacts associated with discontinuing the collecting of information on legal retention status by crab observers. In addition, authors were encouraged to evaluate alternative discard calculations and/or suggest alternative methods for the determination of legal male retention status. It was also

suggested that stock assessment authors outline for the CPT how legal not retained information is used or addressed in stock assessments.”

Response: Two approaches to deal with this issue are in Appendix C. The results of these two approaches (scenarios 2bn1 and 2bn2) are presented in this draft report.

2. Responses to the most recent two sets of SSC and CPT comments specific to this assessment:

Response to CPT Comments (from May 2017)

“The CPT recommended the following scenarios be evaluated for the Fall 2017 assessment:

- *Scenario 2a*
- *Scenario 2b*
- *Scenario 2d*

In addition, because the discard biomass time series from the groundfish fixed and trawl gear fisheries are not split by sex, these models should be brought forward using two approaches to Francis (2011) re-weighting of the size compositions: one based on weights calculated as if all the size compositions were sex-specific, and one based on weights calculated from the “extended” size compositions used in the models for the groundfish fixed gear and trawl gear bycatch size compositions. The former approach is based on the expectation of sex-specific changes in mean length, but does not reflect the loss of sex ratio information associated with splitting the size compositions by sex, whereas the latter approach incorporates this information while the weights are based on expectations for changes in size class across the “extended” size composition.”

Response: All nine scenarios in the SAFE report in September 2017 addressed this comment.

Response to CPT Comments (from September 2017)

“Look at the weighting again for this assessment: it is still based on multiplicative lambda’s.”

Response: Corresponding CV values are provided for the lambda values in this SAFE report.

“The difficulties achieving convergence need to be explored: they are unexpected and concerning.”

Response: Yes, it is a concern. At the September 2017 CPT meeting, Jack Turnock mentioned that he had similar problems with the snow crab model. This could be parameter confounding or initial value problems.

“Jittering initial parameter values was not used in this assessment, but may be useful in evaluating convergence issues.”

Response: Agree. We used jittering before and may use it in the future.

“The tensions in the assessment data leading to estimates of NMFS survey Q at 1 need to be identified and approaches to deal with them need to be developed.”

Response: The Q value for scenario C is greater than 1 (about 1.2). It is the parameter transformation that makes Q less than 1.0. There may be several causes to explain this: (1) M and Q are confounded, (2) the sharp decline of abundance in the early 1980s may make estimated Q higher, and (3) few small crab were caught in the survey during the most recent 10 or more years, causing small estimated survey logistic curve values for the small size classes; for a given length, the overall selectivity value (combined catchability and logistic curve value) is Q times logistic curve value, not just Q.

We did several runs to explore Q values: (1) for scenario 2b, estimated Qs are 0.97, 0.95, and 0.93 with base M of 0.18, 0.22 and 0.3; (2) starting the model in 1985 for scenario 2b, resulting in scenario 2b85, Q is estimated as 0.91, which fits the BSFRF survey biomass very well (see the results for scenario 2b85 in this draft SAFE report); (3) starting the model in 1985 for scenario 2c with a fixed M of 0.18, resulting in scenario 2c85, Q is estimated as 0.92.

“The assessment document needs to be updated to reflect changes in the 2016 BSFRF estimate in the main section of text, not just in the Executive Summary.”

Response: This was done in 2017 SAFE report.

“Provide an explanation of why Equation A4 (catch in the directed fishery) is correct (or correct it if it is wrong).”

Response: The equation A4 (below) is correct. It is a simple equation under the assumption of pulse fishing. Total abundance is reduced by natural mortality to the mid-point of the directed pot fishing and then total fishing mortality is applied to the remaining abundance to get catch. For females, it is female bycatch. For males, the retained catch and bycatch are then separated by their selectivity proportions. The Tanner crab fishery and groundfish fisheries are assumed to be pulse fishing and occur after the directed fishery.

$$G_{l,t}^s = (N_{l,t}^s + O_{l,t}^s) e^{-y_t M_t^s} (1 - e^{-F_{l,t}^s})$$

Response to SSC Comments specific to this assessment (from June 2017):

“Five model scenarios were investigated prior to the spring CPT meeting, the results of which suggested relatively minor differences with regard to management quantities. The SSC supports the CPT’s and author’s recommendations regarding model scenarios to bring forward this fall, which include the following: add the 2016 BSFRF data, separate bycatch components, remove the informative prior and reparameterize NMFS survey catchability to exclude values greater than 1.0, as well as alternatives for data weighting within these scenarios.”

Response: Nine scenarios in the SAFE report in September 2017 addressed this comment.

“The SSC noted that only scenarios utilizing Francis weighting methods were proposed for evaluation in the fall. As noted earlier regarding general guidance to the CPT and assessment authors, the SSC encourages stock assessment authors and the CPT to continue to consider alternative approaches, as data weighting is not a ‘one-size-fits-all’ problem. The best method for data weighting will depend on the quality of the data, the time-series length, the conflict among data sources and other factors unique to a specific assessment. Thus, the BBRKC stock assessment author should retain sufficient latitude to use a method appropriate for this particular assessment, noting that internal consistency is more important than blanket consistency across assessments dealing with a variety of unique data configurations and estimation issues. Evaluation of alternative data weighting approaches can be a useful diagnostic tool to better understand conflicts among data sources within the BBRKC assessment.”

Response: Authors wholeheartedly agree with this SSC comment. We used Francis approach in this report and struggled to get scenarios to converge. The effective sample sizes are greatly reduced through the Francis approach. We will search for alternative approaches in the future.

“Also, the SSC encourages the BBRKC author to objectively define the terminal year of recruitment to include in reference point calculations in this assessment. For BBRKC, where all recent recruitment years have been used in the past, dropping one or more years at the end of the time-series might be warranted. A general rule could be based on the variance of the estimated recruitments and/or the youngest ages of crabs sampled by the fishing gear and/or survey gear included in the model.”

Response: This is a very good suggestion. We evaluated it in this draft SAFE report and recommend discarding the recruitment in the terminal year for estimating B_{35%}.

Response to SSC Comments specific to this assessment (from October 2017):

“The SSC reiterates its request from June 2017 for the BBRKC author and CPT to objectively define the terminal year of recruitment to include in reference point calculations in this and other crab assessments, and again requests that the author use the breakpoint analysis applied for Tanner crab to BBRKC to evaluate whether there was a detectable break in production in 2006. The SSC looks forward to the outcomes of a more comprehensive discussion on this topic at the January 2018 CPT meeting.”

Response: Analysis of terminal year of recruitment is included in this draft SAFE report. Based on the results, we recommend not including the recruitment in the most recent year. Breakpoint analysis was done in May 2017, which includes brood years only up to 2005. We will repeat the breakpoint analysis in May 2019 to detect brood year 2006 when we get one more data point.

“This assessment uses the number of lengths measured as a starting point for input sample sizes. The SSC recommends following the approach of other crab and groundfish stocks in using the

number of stations or pots sampled as a better proxy for statistical sample size given the frequently very high correlation among individuals within a single sample.”

Response: Right now for crab stocks, only the AI golden king crab model does not use the number of lengths measured as a starting point for input sample sizes. The golden king crab model uses only directed fishery length composition data, so it is easy for the model to use boat-days for a starting point for effective sample sizes. The Bristol Bay red king crab model includes length composition data from the trawl survey, directed pot fishery, Tanner crab fishery bycatch, groundfish trawl bycatch, and groundfish fixed gear bycatch. It is difficult to find measurement units of sample sizes that are comparable. The number of survey hauls will be almost constant over time, which is difficult to compare with number of pots, or boat-days, or trips. Snow and Tanner crab models have the same problem. Hopefully we can learn from the groundfish stock model approaches and find a better way to deal with sample sizes in the future.

“More research on catchability is needed, including review of existing camera work from BSFRF surveys that may shed light on crab behavior in response to trawl gear. The SSC provided some comments on new research using modifications of the BSFRF Model under the subsection “Crab Bycatch” earlier in this report.”

Response: We agree with these suggestions for needed research. Analysis of camera work from BSFRF surveys will be helpful, especially on the herding effects of BSFRF surveys.

“The CPT suggested that large catches that drove the stock down in the early 1980s could drive the fits, resulting in an estimate of q near 1.0. On this basis, other evaluation of q could include investigating the effect of the period of historical decline (perhaps by down-weighting it) on more recent estimates of catchability, or fitting a research model fit to BBRKC with only data after the stock collapse in the early 1980s.”

“The SSC noted that historical modelling was conducted using relatively simple catch-survey analysis (Collie and Kruse 1998; Can. Spec. Publ. Fish. Aquat. Sci. 125: 73-83). This might provide another tool for exploring why current estimates of catchability are so close to 1.0.”

Response: There may be several causes to explain Q value close to or higher than 1.0: (1) M and Q are confounded, (2) the sharp decline of abundance in the early 1980s may make estimated Q higher and (3) few small crab were caught in the survey during the most recent 10 or more years, causing small estimated survey logistic curve values for the small size classes; for a given length, the overall selectivity value (combined catchability and logistic curve value) is Q times logistic curve value, not just Q .

We did several runs to explore Q values: (1) for scenario 2b, estimated Q s are 0.97, 0.95, and 0.93 with base M of 0.18, 0.22 and 0.3; (2) starting the model in 1985 for scenario 2b, resulting in scenario 2b85, Q is estimated as 0.91, which fits the BSFRF survey biomass very well (see the results for scenario 2b85 in this draft SAFE report); (3) starting the model in 1985 for scenario 2c with a fixed M of 0.18, resulting in scenario 2c85, Q is estimated as 0.92.

The catch-survey analysis (Collie and Kruse 1998; Can. Spec. Publ. Fish. Aquat. Sci. 125: 73-83) is a simple way to explore Q and M relationships. With similar M values as our model, Q is estimated to be 0.95 by Collie and Kruse (1998); however, with a constant M of 0.36, Q is estimated to be 1.01.

“The SSC is also looking forward to continued development of the Gmacs model for BBRKC during 2018.”

Response: We are looking forward to the day of moving over to GMACS too.

C. Introduction

1. Species

Red king crab (RKC), *Paralithodes camtschaticus*, in Bristol Bay, Alaska.

2. General distribution

Red king crab inhabit intertidal waters to depths >200 m of the North Pacific Ocean from British Columbia, Canada, to the Bering Sea, and south to Hokkaido, Japan, and are found in several areas of the Aleutian Islands, eastern Bering Sea, and the Gulf of Alaska.

3. Stock Structure

The State of Alaska divides the Aleutian Islands and eastern Bering Sea into three management registration areas to manage RKC fisheries: Aleutian Islands, Bristol Bay, and Bering Sea (Alaska Department of Fish and Game (ADF&G) 2012). The Bristol Bay area includes all waters north of the latitude of Cape Sarichef (54°36' N lat.), east of 168°00' W long., and south of the latitude of Cape Newenham (58°39' N lat.) and the fishery for RKC in this area is managed separately from fisheries for RKC outside of this area; i.e., the red king crab in the Bristol Bay area are assumed to be a separate stock from red king crab outside of this area. This report summarizes the stock assessment results for the Bristol Bay RKC stock.

4. Life History

Red king crab have a complex life history. Fecundity is a function of female size, ranging from several tens of thousands to a few hundreds of thousands (Haynes 1968; Swiney et al. 2012). The eggs are extruded by females, fertilized in the spring, and held by females for about 11 months (Powell and Nickerson 1965). Fertilized eggs are hatched in the spring, most during April-June (Weber 1967). Primiparous females are bred a few weeks earlier in the season than multiparous females.

Larval duration and juvenile crab growth depend on temperature (Stevens 1990; Stevens and Swiney 2007). Male and female RKC mature at 5–12 years old, depending on stock and temperature (Loher et al. 2001; Stevens 1990) and may live >20 years (Matsuura and Takeshita 1990). Males and females attain a maximum size of 227 and 195 mm carapace length (CL), respectively (Powell and Nickerson 1965). Female maturity is evaluated by the size at which females are observed to carry egg clutches. Male maturity can be defined by multiple criteria

including spermatophore production and size, chelae vs. carapace allometry, and participation in mating *in situ* (reviewed by Webb 2014). For management purposes, females >89 mm CL and males >119 mm CL are assumed to be mature for Bristol Bay RKC. Juvenile RKC molt multiple times per year until age 3 or 4; thereafter, molting continues annually in females for life and in males until maturity. Male molting frequency declines after attaining functional maturity.

5. Fishery

The RKC stock in Bristol Bay, Alaska, supports one of the most valuable fisheries in the United States. A review of the history of the Bristol Bay RKC fishery is provided in Fitch et al. (2012) and Otto (1989). The Japanese fleet started the fishery in the early 1930s, stopped fishing from 1940 to 1952, and resumed the fishery from 1953 until 1974. The Russian fleet fished for RKC from 1959 to 1971. The Japanese fleet employed primarily tanglenets with a very small proportion of catch from trawls and pots. The Russian fleet used only tanglenets. United States trawlers started fishing Bristol Bay RKC in 1947, but the effort and catch declined in the 1950s. The domestic RKC fishery began to expand in the late 1960s and peaked in 1980 with a catch of 129.95 million lbs (58,943 t), worth an estimated \$115.3 million ex-vessel value. The catch declined dramatically in the early 1980s and has remained at low levels during the last two decades (Table 1). After the early 1980s stock collapse, the Bristol Bay RKC fishery took place during a short period in the fall (usually lasting about a week) with the catch quota based on the stock assessment conducted the previous summer (Zheng and Kruse 2002). Beginning with the 2005/2006 season, new regulations associated with fishery rationalization resulted in an increase in the duration of the fishing season (October 15 to January 15). With the implementation of crab rationalization, historical guideline harvest levels (GHL) were changed to a total allowable catch (TAC). Before rationalization, the implementation errors were quite high for some years and total actual catch from 1980 to 2007 was about 6% less than the sum of GHL/TAC over that period.

6. Fisheries Management

King and Tanner crab stocks in the Bering Sea and Aleutian Islands are managed by the State of Alaska through a federal king and Tanner crab fishery management plan (FMP). Under the FMP, management measures are divided into three categories: (1) fixed in the FMP, (2) frameworked in the FMP, and (3) discretion of the State of Alaska. The State of Alaska is responsible for determining and establishing the GHL/TAC under the framework in the FMP.

Harvest strategies for the Bristol Bay RKC fishery have changed over time. Two major management objectives for the fishery are to maintain a healthy stock that ensures reproductive viability and to provide for sustained levels of harvest over the long term (ADF&G 2012). In attempting to meet these objectives, the GHL/TAC is coupled with size-sex-season restrictions. Only males ≥ 6.5 -in carapace width (equivalent to 135-mm carapace length, CL) may be harvested and no fishing is allowed during molting and mating periods (ADF&G 2012). Specification of TAC is based on a harvest rate strategy. Before 1990, harvest rates on legal males were based on population size, abundance of prerecruits to the fishery, and postrecruit abundance, and rates varied from less than 20% to 60% (Schmidt and Pengilly 1990). In 1990, the harvest strategy was modified, and a 20% mature male harvest rate was applied to the abundance of mature-sized (≥ 120 -mm CL) males with a maximum 60% harvest rate cap of legal (≥ 135 -mm CL) males (Pengilly and Schmidt 1995). In addition, a minimum threshold of 8.4 million mature-sized females (≥ 90 -mm

CL) was added to existing management measures to avoid recruitment overfishing (Pengilly and Schmidt 1995). Based on a new assessment model and research findings (Zheng et al. 1995a, 1995b, 1997a, 1997b), the Alaska Board of Fisheries adopted a new harvest strategy in 1996. That strategy had two mature male harvest rates: 10% when effective spawning biomass (ESB) is between 14.5 and 55.0 million lbs and 15% when ESB is at or above 55.0 million lbs (Zheng et al. 1996). The maximum harvest rate cap of legal males was changed from 60% to 50%. A threshold of 14.5 million lbs of ESB was also added. In 1997, a minimum threshold of 4.0 million lbs was established as the minimum GHL for opening the fishery and maintaining fishery manageability when the stock abundance is low. The Board modified the current harvest strategy by adding a mature harvest rate of 12.5% when the ESB is between 34.75 and 55.0 million lbs in 2003 and eliminated the minimum GHL threshold in 2012. The current harvest strategy is illustrated in Figure 1.

D. Data

1. Summary of New Information

The NMFS and BSFRF trawl survey data were updated to include the 2017 survey data.

Catch and biomass data were updated to 2016/17. Groundfish fisheries bycatch data during 2009-2016 were updated and separated into trawl fisheries and fixed gear fisheries bycatches.

Data types and ranges are illustrated in Figure 2.

2. Catch Data

Data on landings of Bristol Bay RKC by length and year and catch per unit effort from 1960 to 1973 were obtained from annual reports of the International North Pacific Fisheries Commission (Hoopes et al. 1972; Jackson 1974; Phinney 1975) and from the ADF&G from 1974 to 2016. Bycatch data are available starting from 1990 and were obtained from the ADF&G observer database and reports (Gaeuman 2013). Sample sizes for catch by length and shell condition are summarized in Table 2. Relatively large samples were taken from the retained catch each year. Sample sizes for trawl bycatch were the annual sums of length frequency samples in the National Marine Fisheries Service (NMFS) database.

(i). Catch Biomass

Retained catch and estimated bycatch biomasses are summarized in Table 1 and illustrated in Figure 2. Retained catch and estimated bycatch from the directed fishery include the general, open-access fishery (prior to rationalization), or the individual fishery quota (IFQ) fishery (after rationalization), as well as the Community Development Quota (CDQ) fishery and the ADF&G cost-recovery harvest. Starting in 1973, the fishery generally occurred during the late summer and fall. Before 1973, a small portion of retained catch in some years was caught from April to June. Because most crab bycatch from the groundfish trawl fisheries occurred during the spring, the years in Table 1 are one year less than those from the NMFS trawl bycatch database to approximate the annual bycatch for reporting years defined as July 1 to June 30; e.g., year 2002 in Table 1 for trawl bycatch corresponds to what is reported for year 2003 in the NMFS database. Catch biomass is shown in Figure 3. Bycatch data for the cost-recovery fishery before 2006 were not available. In this report, pot fisheries include both the

directed fishery and RKC bycatch in the Tanner crab pot fishery and trawl fisheries are groundfish trawl fisheries.

(ii). Catch Size Composition

Retained catch by length and shell condition and bycatch by length, shell condition, and sex were obtained for stock assessments. From 1960 to 1966, only retained catch length compositions from the Japanese fishery were available. Retained catches from the Russian and U.S. fisheries were assumed to have the same length compositions as the Japanese fishery during this period. From 1967 to 1969, the length compositions from the Russian fishery were assumed to be the same as those from the Japanese and U.S. fisheries. After 1969, foreign catch declined sharply and only length compositions from the U.S. fishery were used to distribute catch by length.

(iii). Catch per Unit Effort

Catch per unit effort (CPUE) is defined as the number of retained crab per tan (a unit fishing effort for tanglenets) for the Japanese and Russian tanglenet fisheries and the number of retained crab per potlift for the U.S. fishery (Table 1). Soak time, while an important factor influencing CPUE, is difficult to standardize. Furthermore, complete historical soak time data from the U.S. fishery are not available. Based on the approach of Balsiger (1974), all fishing effort from Japan, Russia, and U.S. were standardized to the Japanese tanglenet from 1960 to 1971, and the CPUE was standardized as crab per tan. Except for the peak-to-crash years of late 1970s and early 1980s the correspondence between U.S. fishery CPUE and area-swept survey abundance is poor (Figure 4). Due to the difficulty in estimating commercial fishing catchability and crab availability to the NMFS annual trawl survey data, commercial CPUE data were not used in the model.

3. NMFS Survey Data

The NMFS has performed annual trawl surveys of the eastern Bering Sea since 1968. Two vessels, each towing an eastern otter trawl with an 83 ft headrope and a 112 ft footrope, conducted this multispecies, crab-groundfish survey during the summer. Stations were sampled in the center of a systematic 20 X 20 nm grid overlaid in an area of $\approx 140,000 \text{ nm}^2$. Since 1972, the trawl survey has covered the full stock distribution except in nearshore waters. The survey in Bristol Bay occurs primarily during late May and June. Tow-by-tow trawl survey data for Bristol Bay RKC during 1975-2017 were provided by NMFS.

Abundance estimates by sex, carapace length, and shell condition were derived from survey data using an area-swept approach (Figures 5a and 5b). Spatial distributions of crab from the standard trawl surveys during recent years are shown in Appendix B. Until the late 1980s, NMFS used a post-stratification approach, but subsequently treated Bristol Bay as a single stratum; the estimates shown for Bristol Bay in Figures 4 and 5 were made without post-stratification. If multiple tows were made for a single station in a given year, the average of the abundances from all tows within that station was used as the estimate of abundance for that station. The new time series since 2015 discards all “hot spot” tows. We used the new area-swept estimates provided by NMFS in 2017.

In addition to standard surveys, NMFS also conducted some surveys after the standard surveys to

better assess mature female abundance. In addition to the standard surveys conducted in early June (late May to early June in 1999 and 2000), a portion of the distribution of Bristol Bay RKC was re-surveyed in 1999, 2000, 2006-2012, and 2017. Resurveys performed in late July, about six weeks after the standard survey, included 31 stations (1999), 23 stations (2000), 31 stations (2006, 1 bad tow and 30 valid tows), 32 stations (2007-2009), 23 stations (2010) and 20 stations (2011 and 2012) with high female density. The resurveys were necessary because a high proportion of mature females had not yet molted or mated when sampled by the standard survey. Differences in area-swept estimates of abundance between the standard surveys and resurveys of these same stations are attributed to survey measurement errors or to seasonal changes in distribution between survey and resurvey. More large females were observed in the resurveys than during the standard surveys in 1999 and 2000 because most mature females had not molted prior to the standard surveys. As in 2006, area-swept estimates of males >89 mm CL, mature males, and legal males within the 32 resurvey stations in 2007 were not significantly different ($P=0.74$, 0.74 and 0.95 ; paired t -test of sample means) between the standard survey and resurvey tows. However, similar to 2006, area-swept estimates of mature females within the 32 resurvey stations in 2007 were significantly different ($P=0.03$; paired t -test) between the standard survey and resurvey tows. Resurvey stations were close to shore during 2010-2012, and mature and legal male abundance estimates were lower for the re-tow than the standard survey. Following the CPT recommendation, we used the standard survey data for male abundance estimates and only the resurvey data, plus the standard survey data outside the resurveyed stations, to assess female abundances during these resurvey years.

4. Bering Sea Fisheries Research Foundation Survey Data

The BSFRF conducted trawl surveys for Bristol Bay RKC in 2007 and 2008 with a small-mesh trawl net and 5-minute tows. The surveys occurred at similar times as the NMFS standard surveys and covered about 97% of the Bristol Bay area. Few Bristol Bay RKC were found outside of the BSFRF survey area. Because of the small mesh size, the BSFRF surveys were expected to catch more of RKC within the swept area. Crab abundances of different size groups were estimated by the kriging method. Mature male abundances were estimated to be 22.331 in 2007 and 19.747 million in 2008 with respective CVs of 0.0634 and 0.0765. BSFRF also conducted a side-by-side survey concurrent with the NMFS trawl survey during 2013-2016 in Bristol Bay. In May 2017, survey biomass and size composition estimates from 2016 BSFRF side-by-side trawl survey data were updated. Total survey biomass decreased from 87,725.1 t initially estimated in September 2016 to 77,815.7 t in the final estimate in May 2017, about 11.3% reduction. The initial estimate mistakenly included the tows conducted in the recruitment study.

E. Analytic Approach

1. History of Modeling Approaches

To reduce annual measurement errors associated with abundance estimates derived from the area-swept method, ADF&G developed a length-based analysis (LBA) in 1994 that incorporates multiple years of data and multiple data sources in the estimation procedure (Zheng et al. 1995a). Annual abundance estimates of the Bristol Bay RKC stock from the LBA have been used to manage the directed crab fishery and to set crab bycatch limits in the groundfish fisheries since

1995 (Figure 1). An alternative LBA (research model) was developed in 2004 to include small size groups for federal overfishing limits. The crab abundance declined sharply during the early 1980s. The LBA estimated natural mortality for different periods of years, whereas the research model estimated additional mortality beyond a basic constant natural mortality during 1976-1993. In this report, we present only the research model that was fit to the data from 1975 to 2017.

2. Model Description

The original LBA model was described in detail by Zheng et al. (1995a, 1995b) and Zheng and Kruse (2002). The model combines multiple sources of survey, catch, and bycatch data using a maximum likelihood approach to estimate abundance, recruitment, selectivities, catches, and bycatch of the commercial pot fisheries and groundfish trawl fisheries. A full model description is provided in Appendix A.

a-f. See appendix A.

g. Critical assumptions of the model:

- i. The base natural mortality is constant over shell condition and length and was estimated assuming a maximum age of 25 and applying the 1% rule (Zheng 2005).
- ii. Survey and fisheries selectivities are a function of length and were constant over shell condition. Selectivities are also a function of sex except for trawl bycatch selectivities, which are the same for both sexes. Two different survey selectivities were estimated: (1) 1975-1981 and (2) 1982-2017, based on modifications to the trawl gear used in the assessment survey.
- iii. Growth is a function of length and is assumed to not change over time for males. For females, growth-per-molt increments as a function of length were estimated for three periods (1975-1982, 1983-1993, and 1994-2017) based on sizes at maturity. Once mature, female red king crab grow with a much smaller growth increment per molt.
- iv. Molting probabilities are an inverse logistic function of length for males. Females molt annually.
- v. Annual fishing seasons for the directed fishery are short.
- vi. The prior of survey catchability (Q) was estimated to be 0.896, based on a trawl experiment by Weinberg et al. (2004) with a standard deviation of 0.025 for some scenarios. Q is assumed to be constant over time and is estimated in the model.
- vii. Males mature at sizes ≥ 120 mm CL. For convenience, female abundance was summarized at sizes ≥ 90 mm CL as an index of mature females.
- viii. Measurement errors were assumed to be normally distributed for length compositions and were log-normally distributed for biomasses.
- h. Changes to the above since previous assessment: see Section A.3. Changes to the assessment methodology.
- i. Outline of methods used to validate the code used to implement the model and whether the code is available: The code is available.

3. Model Selection and Evaluation

a. Alternative model configurations (scenarios):

2b. Scenario 2b is the same as scenario 2b in the SAFE draft report in September 2017. This scenario assumes that BSFRF survey capture probabilities are 1.0 for all length groups. Under this assumption, NMFS survey selectivities are the products of crab availabilities (equal to BSFRF survey selectivities) and NMFS survey capture probabilities. A survey capture probability for a length group is simply defined as the proportion of the crab in the length group within the area-swept that is caught by the survey net. Also, groundfish fisheries bycatch is separated into trawl fisheries and fixed gear fisheries.

Scenario 2b includes:

- (1) Basic $M = 0.18$, with an additional mortality level during 1980-1984 for males and two additional mortality levels (one for 1980-1984 and the other for 1976-1979 and 1985-1993) for females.
- (2) Including BSFRF survey data during 2007-2008 and 2013-2016. The BSFRF survey is treated as an independent survey, and no assumption is made about the capture probabilities of the BSFRF survey. In effect, survey selectivities for both surveys are estimated separately and directly in the model.
- (3) NMFS survey catchability is estimated in the model and is assumed to be constant over time. BSFRF survey catchability is assumed to be 1.0.
- (4) Two levels of molting probabilities for males: one before 1980 and one after 1979, based on survey shell condition data. Each level has two parameters.
- (5) Estimating effective sample size from observed sample sizes. Effective sample sizes are estimated as $\min(0.5 \cdot \text{observed-size}, N)$ for trawl surveys and $\min(0.1 \cdot \text{observed-size}, N)$ for catch and bycatch, where N is the maximum sample size (200 for trawl surveys, 100 for males from the pot fishery and 50 for females from pot fishery and both males and females from the groundfish fisheries). There is a justification for enforcing a maximum limit to effective sample sizes because the number of length measurements is large (Fournier et al. 1998). The effective sample sizes are plotted against the implied effective sample sizes in Figures 6 and 7, where the implied effective sample sizes are estimated as follows:

$$n_y = \sum_l \hat{P}_{y,l}(1-\hat{P}_{y,l}) / \sum_l (P_{y,l} - \hat{P}_{y,l})^2 \quad (1)$$

where $\hat{P}_{y,l}$ and $P_{y,l}$ are estimated and observed size compositions in year y and length group l , respectively.

- (6) Standard survey data for males and NMFS survey retow data (during cold years) for females.
- (7) Estimating initial year length compositions.

For scenario 2b, survey abundances $\hat{N}_{s,y,l}^b$ (BSFRF survey) and $\hat{N}_{s,y,l}^n$ (NMFS survey) by sex s and in year y and length group l are computed as follows:

$$\begin{aligned}\hat{N}_{s,y,l}^b &= N_{s,y,l} s_{s,l}^b, \\ \hat{N}_{s,y,l}^n &= N_{s,y,l} s_{s,l}^n,\end{aligned}\tag{2}$$

where $s_{s,l}^b$ and $s_{s,l}^n$ are survey selectivities for BSFRF and NMFS surveys by sex s and in length group l , respectively, and $N_{s,y,l}$ is the population abundance by sex s and in year y and length group l . BSFRF survey selectivities are computed as

$$s_{s,l}^b = \frac{1}{1 + e^{-\beta_s^b (l - L_{50,s}^b)}},\tag{3}$$

where β and L_{50} are parameters. Survey selectivity for the first length group (67.5 mm) was assumed to be the same for both males and females, so only three parameters (β , L_{50} for females and L_{50} for males) were estimated in the model for each survey. The BSFRF survey catchability is assumed to be 1.0.

Scenario 2b assumes that the BSFRF survey capture probabilities are 1.0 for all length groups. Under this assumption, NMFS survey selectivities are the products of crab availabilities (equal to BSFRF survey selectivities) and NMFS survey capture probabilities (p):

$$s_{s,l}^n = p_{s,l} s_{s,l}^b.\tag{4}$$

Therefore, the model estimates NMFS survey capture probabilities and BSFRF survey selectivities and computes NMFS survey selectivities from these estimates. NMFS survey capture probabilities are computed as

$$p_{s,l} = \frac{Q}{1 + e^{-\beta_s (l - L_{50,s})}},\tag{5}$$

where β and L_{50} are parameters and similar to the survey selectivities, only three parameters (β , L_{50} for females and L_{50} for males) were estimated in the model for each sex. Q is the NMFS survey catchability and is estimated in the model with or without a prior from the double-bag experiment, depending on scenarios.

Since fishing times for both Tanner crab fishery and groundfish fishery are assumed to occur the same time, the fraction separation of fishing mortality rates for both fisheries is used to divide the total fishing mortality rate to individual fisheries, that is, $F_i/F_{\text{tot}} * (1 - \exp(-F_{\text{tot}}))$ for fishery i , and the sum of $F_i = F_{\text{tot}}$.

2bn1: Scenario 2bn1 is the same as scenario 2b except that the approach in Appendix C is used to estimate for male annual bycatch biomass since rationalization (2005-present) and total observer male length composition data are used to compute log negative likelihoods in the directed pot fishery. Due to high grading problems in some years since rationalization, two logistic curves are estimated for retained proportions: one before rationalization (before 2005) and another for 2005-present. See approach 1 in Appendix C.

2bn2: Scenario 2bn2 is the same as scenario 2bn1 except that only one logistic curve is estimated for all years for retained proportions and annual retention adjusted factors are estimated to modify retained proportions for years after 2004. See approach 2 in Appendix C.

2b85: Scenario 2b85 is the same as Scenario 2b except for starting the model in 1985. This approach is to check NMFS trawl survey catchability estimates without the sharp decline of abundance data in the early 1980s.

2c85: Scenario 2c85 is the same as Scenario 2c in the draft SAFEE report in May 2017 except for starting the model in 1985 and constant natural mortality of 0.18 for all years.

- b. Progression of results: See the new results at the beginning of the report.
- c. Evidence of search for balance between realistic and simpler models: NA.
- d. Convergence status/criteria: ADMB default convergence criteria.
- e. Sample sizes for length composition data: observed sample sizes are summarized in Table 2, and estimated implied sample sizes and effective sample sizes are illustrated in Figures 6 and 7.
- f. Credible parameter estimates: All estimated parameters seem to be credible.
- g. Model selection criteria: The likelihood values were used to select among alternatives that could be legitimately compared by that criterion.
- h. Residual analysis: Residual plots are illustrated in figures.
- i. Model evaluation is provided under Results, below.
- j. Jittering: the Stock Synthesis Approach is used to do jittering to find the optimum:

The *Jitter* factor of 0.1 is multiplied by a random normal deviation $rdev=N(0,1)$, to a transformed parameter value based upon the predefined parameter:

$$temp = 0.5 \ rdev \ Jitter \ \ln\left(\frac{P_{max} - P_{min} + 0.0000002}{P_{val} - P_{min} + 0.0000001} - 1\right), \quad (6)$$

with the final jittered starting parameter value backtransformed as:

$$P_{new} = P_{min} + \frac{P_{max} - P_{min}}{1.0 + \exp(-2.0 \ temp)}, \quad (7)$$

where P_{max} and P_{min} are upper and lower bounds of parameters and P_{val} is the estimated parameter value before the jittering. Due to time consuming, the jittering approach is not used in this report.

4. Results

- a. Effective sample sizes and weighting factors.
 - i. For scenario 2b, effective sample sizes are illustrated in Figures 6 and 7.
 - ii. CVs are assumed to be 0.03 for retained catch biomass, and 0.07 for all bycatch biomasses, 0.53 for recruitment variation, and 0.23 for recruitment sex ratio.
 - iii. Initial trawl survey catchability (Q) is estimated to be 0.896 with a standard deviation of 0.025 (CV about 0.03) based on the double-bag experiment results. These values are used as a prior for estimating Q in the model for scenarios 2b, 2bn1, 2bn2, and 2b85.
- b. Tables of estimates.
 - i. Parameter estimates for scenarios 2b, 2bn1, and 2bn2 are summarized in Tables 4 and 5.
 - ii. Abundance and biomass time series are provided in Table 6 for scenarios 2b, 2bn1, and 2bn2.
 - iii. Recruitment time series for scenarios 2b, 2bn1, and 2bn2 are provided in Table 6.
 - iv. Time series of catch biomass is provided in Table 1.

Negative log-likelihood values and parameter estimates are summarized in Tables 4 and 5, respectively. Length-specific fishing mortality is equal to selectivity-at-length times the full fishing mortality. Estimated full pot fishing mortalities for females and full fishing mortalities for groundfish fisheries bycatch were very low due to low bycatch as well as handling mortality rates less than 1.0. Estimated recruits varied greatly from year to year (Table 6). Estimated low selectivities for male pot bycatch, relative to the retained catch, reflected the 20% handling mortality rate (Figure 8). Both selectivities were applied to the same level of full fishing mortality. Estimated selectivities for female pot bycatch were close to 1.0 for all mature females, and the estimated full fishing mortalities for female pot bycatch were lower than for male retained catch and bycatch (Table 5).

- c. Graphs of estimates.
 - i. Selectivities and molting probabilities by length are provided in Figures 8 and 9 for scenarios 2b, 2bn1, 2bn2, 2b85, and 2c85.

One of the most important results is estimated trawl survey selectivity (Figure 8). Survey selectivity affects not only the fitting of the data but also the absolute abundance estimates. Estimated survey selectivities in Figure 8 are generally smaller than the capture probabilities in Figure A1 because survey selectivities include capture probabilities and crab availability. The NMFS survey catchability was estimated to be 0.896 from the trawl experiment, which is higher than that roughly estimated from the

BSFRF surveys (0.854). The reliability of estimated survey selectivities will greatly affect the application of the model to fisheries management. Under- or overestimates of survey selectivities will cause a systematic upward or downward bias of abundance estimates. Information about crab availability to the survey area at survey times will help estimate the survey selectivities.

For all scenarios, estimated molting probabilities during 1975-2017 (Figure 9) were generally lower than those estimated from the 1954-1961 and 1966-1969 tagging data (Balsiger 1974). Lower molting probabilities mean more oldshell crab, possibly due to changes in molting probabilities over time or shell aging errors. Overestimates or underestimates of oldshell crab will result in lower or higher estimates of male molting probabilities.

- ii. Estimated total survey biomass and mature male and female abundances are plotted in Figure 10. Absolute mature male biomasses are illustrated in Figure 11.

Model estimated relative survey biomasses are very similar among the five scenarios except slightly lower for scenario 2bn1 during recent years and fit the survey data quite well. The absolute population biomass estimates are slightly higher for scenario 2b85 due to lower trawl survey catchability estimate and lower for scenario 2bn1 during last 20 years due to slightly higher estimated trawl survey catchability value. Slightly lower estimate of NMFS trawl survey catchability for scenario 2c85 also result in slightly higher absolute biomass estimates during 1985-2010. Initial biomass estimates in 1985 for scenarios 2b85 and 2c85 are higher than other three scenarios. The Subtraction approach estimating male discard biomass and combining approach 2 to estimate retained proportions result in population biomass estimates (scenario 2bn2) similar to those from scenario 2b.

Although the model did not fit the mature crab abundances directly, trends in the mature abundance estimates agree well with observed survey values except in 2014 (Figure 10b). Estimated mature crab abundance increased dramatically in the mid 1970s then decreased precipitously in the early 1980s. Estimated mature crab abundance had increased during 1985-2009 with mature females being about 3 times more abundant in 2009 than in 1985 and mature males being about 2 times more abundant in 2009 than in 1985. Estimated mature abundance has declined since 2009 (Figure 10b). Model estimates of both male and female mature abundances have steadily declined since the late 2000s. Absolute mature male biomasses for all scenarios have a similar trend over time (Figure 11).

The fit to BSFRF survey data and estimated survey selectivities are illustrated in Figures 10c-e.

- iii. Estimated recruitment time series are plotted in Figure 12 for scenarios 2b, 2bn1, and 2bn2.
- iv. Estimated fishing mortality rates are plotted against mature male biomass in Figure 13 for scenarios 2b, 2bn1, and 2bn2.

The average of estimated male recruits from 1984 to 2017 (Figure 12) and mature male biomass per recruit were used to estimate $B_{35\%}$. Alternative periods of 1976-present and 1976-1983 were compared in our report. The full fishing mortalities for the directed pot fishery at the time of fishing were plotted against mature male biomass on Feb. 15 (Figure 13). Estimated fishing mortalities in most years before the current harvest strategy was adopted in 1996 were above $F_{35\%}$ (Figure 13). Under the current harvest strategy, estimated fishing mortalities were at or above the $F_{35\%}$ limits in 1998, 2005, 2007-2009 for scenario 2b but below the $F_{35\%}$ limits in the other post-1995 years. During 1996-2016, estimated fishing mortalities were above the $F_{35\%}$ limits in 1998, 2003, and 2005-2010 for scenario 2bn1. Higher estimated survey selectivities from scenario 2bn1 result in relatively higher fishing mortalities than those with scenarios 2b, 2bn2, 2b85 and 2c85.

For scenario 2b, estimated full pot fishing mortalities ranged from 0.00 to 2.11 during 1975-2016. Estimated values were greater than 0.40 during 1975-1981, 1985-1987, 1993 and 2008 (Table 5, Figure 13). For scenario 2bn1, estimated full pot fishing mortalities ranged from 0.00 to 2.49 during 1975-2016, with estimated values over 0.40 during 1975-1982, 1984-1987, 1990-1991, 1993, 1998, 2005, and 2007-2009 (Figure 13). For scenario 2bn2, estimated full pot fishing mortalities ranged from 0.00 to 2.39 during 1975-2016, with estimated values over 0.40 during 1975-1981, 1984-1987, 1990-1991, 1993, 1998, 2005, and 2007-2008 (Figure 13). Estimated fishing mortalities for pot female and groundfish fisheries bycatches were generally less than 0.06.

- v. Estimated mature male biomass and recruitment are plotted to illustrate their relationships with scenario 2b (Figure 14a). Annual stock productivities are illustrated in Figure 14b.

Stock productivity (recruitment/mature male biomass) was generally lower during the last 20 years (Figure 14b).

Egg clutch data collected during summer surveys may provide information about mature female reproductive conditions. Although egg clutch data are subject to rating errors as well as sampling errors, data trends over time may be useful. Proportions of empty clutches for newshell mature females >89 mm CL were high in some years before 1990, but have been low since 1990 (Figure 15). The highest proportion of empty clutches (0.2) was in 1986, and primarily involved soft shell females (shell condition 1). Clutch fullness fluctuated annually around average levels during two periods: before 1991 and after 1990 (Figure 15). The average clutch fullness was similar for these two periods (Figure 15). Egg clutch fullness during the last two years is relatively low.

- d. Graphic evaluation of the fit to the data.
 - i. Observed vs. estimated catches are plotted in Figure 16.
 - ii. Model fits to total survey biomass are shown in Figure 10 with a standardized residual plot in Figure 17.

- iii. Model fits to catch and survey proportions by length are illustrated in Figures 18-24 and residual bubble plots are shown in Figures 25-26.

The model (five scenarios) fit the fishery biomass data well and the survey biomass reasonably well (Figures 10 and 16). Because the model estimates annual fishing mortality for directed pot male retained catch, undirected pot male bycatch, pot female bycatch, and trawl bycatch, the deviations of observed and predicted (estimated) fishery biomass are mainly due to size composition differences. Directed pot male discard bycatch was not fit as well as other biomasses because annual fishing mortalities were estimated for total pot male catches and the retained catch has a much smaller CV value than that for the male discard bycatch.

The model also fit the length composition data well (Figures 18-24). The model also fit the length proportions of the pot male bycatch well with two simple linear selectivity functions for scenario 2b (Figure 21). The new approach (Appendix C) also fits the total male length proportions very well for scenarios 2bn1 and 2bn2 (Figure 21).

Modal progressions are tracked well in the trawl survey data, particularly beginning in the mid-1990s (Figures 18 and 19). Cohorts first seen in the trawl survey data in 1975, 1986, 1990, 1995, 1999, 2002 and 2005 can be tracked over time. Some cohorts can be tracked over time in the pot bycatch as well (Figure 21), but the bycatch data did not track the cohorts as well as the survey data. Groundfish trawl bycatch data provide little information to track modal progression (Figures 23 and 24).

Standardized residuals of total survey biomass and proportions of length are plotted to examine their patterns. Residuals were calculated as observed minus predicted and standardized by the estimated standard deviation. Standardized residuals of total survey biomass did not show any consistent patterns (Figure 17). Standardized residuals of proportions of survey males appear to be random over length and year (Figure 25). There is an interesting pattern for residuals of proportions of survey females. Residuals were generally negative for large-sized mature females during 1975-1987 for scenarios 2b, 2bn1 and 2bn2 (Figure 26). Changes in growth over time or increased mortality may cause this pattern. The inadequacy of the model can be corrected by adding parameters to address these factors or with improved growth data.

- e. Retrospective and historic analyses.

Two kinds of retrospective analyses were conducted for this report: (1) the 2017 model (scenario 2b) hindcast results and (2) historical results. The 2017 model results are based on sequentially excluding one-year of data to evaluate the current model performance with fewer data. The historical results are the trajectories of biomass and abundance from previous assessments that capture both new data and changes in methodology over time. Treating the 2017 estimates as the baseline values, we can also evaluate how well the model had done in the past.

- i. Retrospective analysis (retrospective bias in base model or models).

The performance of the 2017 model includes sequentially excluding one-year of data. The model with scenario 2b performed reasonably well during 2011-2016 with a

lower terminal year estimates in 2012 and 2013 and higher estimates in 2011 (Figures 27-28).

ii. Historic analysis (plot of actual estimates from current and previous assessments).

The model first fit the data from 1985 to 2004 in the terminal year of 2004. Thus, sequentially incrementing the terminal year provided 10 historical assessments for comparison with the 2017 assessment model results (Figure 29). The main differences of the 2004 model were weighting factors and effective sample sizes for the likelihood functions. In 2004, the weighting factors were 1,000 for survey biomass, 2,000 for retained catch biomass and 200 for bycatch biomasses. The effective sample sizes were set to be 200 for all proportion data but weighting factors of 5, 2, and 1 were also respectively applied to retained catch proportions, survey proportions and bycatch proportions. Estimates of time series of abundance in 2004 were generally higher than those estimated after 2004 (Figure 29).

In 2005, to improve the fit for retained catch data, the weight for retained catch biomass was increased to 3,000 and the weight for retained catch proportions was increased to 6. All other weights were not changed. In 2006, all weights were re-configured. No weights were used for proportion data, and instead, effective sample sizes were set to 500 for retained catch, 200 for survey data, and 100 for bycatch data. Weights for biomasses were changed to 800 for retained catch, 300 for survey and 50 for bycatch. The weights in 2007 were the same as 2006. Generally, estimates of time series of abundance in 2005 were slightly lower than in 2006 and 2007, and there were few differences between estimates in 2006 and 2007 (Figure 29).

In 2008, estimated coefficients of variation for survey biomass were used to compute likelihood values as suggested by the CPT in 2007. Thus, weights were re-configured to: 500 for retained catch biomass, 50 for survey biomass, and 20 for bycatch biomasses. Effective sample size was lowered to 400 for the retained catch data. These changes were necessary for the estimation to converge and for a relatively good balanced fit to both biomasses and proportion data. Also, sizes at 50% selectivities for all fisheries data were allowed to change annually, subject to a random walk pattern, for all assessments before 2008. The 2008 model does not allow annual changes in any fishery selectivities. Except for higher estimates of abundance during the late 1980s and early 1990s, estimates of time series of abundance in 2008 were generally close to those in 2006 and 2007 (Figure 29).

During 2009-2013, the model was extended to the data through 1968. No weight factors were used for the NMFS survey biomass during 2009-2013 assessments. Since 2013, the model has fitted the data only back to 1975 for consistence of trawl survey data. Two levels of molting probabilities over time were used, shell conditions for males were combined, and length composition data of the BSFRF survey were used as well. In 2014 and 2015, the trawl survey time series were re-estimated and a trawl survey catchability was estimated for some scenarios.

Overall, both historical results (historic analysis) and the 2017 model results (retrospective analysis) performed reasonably well. No great overestimates or underestimates occurred

as was observed in assessments for Pacific halibut (*Hippoglossus stenolepis*) (Parma 1993) and some eastern Bering Sea groundfish stocks (Zheng and Kruse 2002; Ianelli et al. 2003). Since the most recent model was not used to set TAC or overfishing limits until 2009, historical implications for management from the stock assessment errors cannot be evaluated at the current time. However, management implications of the ADF&G stock assessment model were evaluated by Zheng and Kruse (2002).

Ratios of estimated retrospective recruitments to terminal estimates in 2017 as a function of number of years estimated in the model show converging to 1.0 as the number of years increase (Figure 28). Standard deviations of the ratios drop sharply from one year estimated in the model to two years (Figure 28), showing great uncertainty of recruitment estimates for terminal years. Based on these results, we suggest not using recruitment estimates in a terminal year for overfishing/overfished determination.

f. Uncertainty and sensitivity analyses

- i. Estimated standard deviations of parameters are summarized in Table 5 for scenarios 2b, 2bn1 and 2bn2. Estimated standard deviations of mature male biomass are listed in Table 6.
- ii. Probabilities for trawl survey catchability Q are illustrated in Figure 30 for scenarios 2b using the mcmc approach; estimated Q s are generally less than 1.0. Probabilities for mature male biomass and OFL in 2017 are illustrated in Figure 31 for scenario 2b using the mcmc approach. The confidence intervals are quite narrow.
- iii. Sensitivity analysis for handling mortality rate was reported in the SAFE report in May 2010. The baseline handling mortality rate for the directed pot fishery was set at 0.2. A 50% reduction and 100% increase respectively resulted in 0.1 and 0.4 as alternatives. Overall, a higher handling mortality rate resulted in slightly higher estimates of mature abundance, and a lower rate resulted in a minor reduction of estimated mature abundance. Differences of estimated legal abundance and mature male biomass were small among these handling mortality rates.
- iv. Sensitivity of weights. Sensitivity of weights was examined in the SAFE report in May 2010. Weights to biomasses (trawl survey biomass, retained catch biomass, and bycatch biomasses) were reduced to 50% or increased to 200% to examine their sensitivity to abundance estimates. Weights to the penalty terms (recruitment variation and sex ratio) were also reduced or increased. Overall, estimated biomasses were very close under different weights except during the mid-1970s. The variation of estimated biomasses in the mid-1970s was mainly caused by the changes in estimates of additional mortalities in the early 1980s.

g. Comparison of alternative model scenarios

These comparisons, based on the data through 2010, were reported in the SAFE report in May 2011. Estimating length proportions in the initial year (scenario 1a) results in a better fit of survey length compositions at an expense of 36 more parameters than scenario 1. Abundance and biomass estimates with scenario 1a are similar between scenarios. Using only standard

survey data (scenario 1b) results in a poorer fit of survey length compositions and biomass than scenarios using both standard and re-tow data (scenarios 1, 1a, and 1c) and has the lowest likelihood value. Although the likelihood value is higher for using both standard survey and re-tow data for males (scenario 1) than using only standard survey for males (scenario 1c), estimated abundances and biomasses are almost identical. The higher likelihood value for scenario 1 over scenario 1c is due to trawl bycatch length compositions.

In this report (May 2018), five scenarios are compared. Model estimated relative survey biomasses are very similar among the scenarios except slightly lower for scenario 2bn1 during recent years and fit the survey data quite well. The absolute population biomass estimates are slightly higher for scenario 2b85 due to lower trawl survey catchability estimate and lower for scenario 2bn1 during last 20 years due to slightly higher estimated trawl survey catchability value. Slightly lower estimate of NMFS trawl survey catchability for scenario 2c85 also result in slightly higher absolute biomass estimates during 1985-2010. Initial biomass estimates in 1985 for scenarios 2b85 and 2c85 are higher than other three scenarios. Overall, the results for all five scenarios are similar except those impacted by estimates of NMFS trawl survey catchabilities. The Subtraction approach estimating male discard biomass and combining approach 2 to estimate retained proportions result in population biomass estimates (scenario 2bn2) similar to those from scenario 2b. We recommend scenario 2bn2 for September 2018 assessment because of better fit to the data and better estimating male discard biomass in the directed pot fishery.

F. Calculation of the OFL and ABC

1. Bristol Bay RKC is currently placed in Tier 3b (NPFMC 2007).
2. For Tier 3 stocks, estimated biological reference points include $B_{35\%}$ and $F_{35\%}$. Estimated model parameters were used to conduct mature male biomass-per-recruit analysis.
3. Specification of the OFL:

The Tier 3 can be expressed by the following control rule:

$$\begin{aligned}
 \text{a) } \frac{B}{B^*} > 1 & \qquad F_{OFL} = F^* \\
 \text{b) } \beta < \frac{B}{B^*} \leq 1 & \qquad F_{OFL} = F^* \left(\frac{B/B^* - \alpha}{1 - \alpha} \right) \\
 \text{c) } \frac{B}{B^*} \leq \beta & \qquad \text{directed fishery } F = 0 \text{ and } F_{OFL} \leq F^*
 \end{aligned} \tag{1}$$

Where

B = a measure of the productive capacity of the stock such as spawning biomass or fertilized egg production. A proxy of B , MMB estimated at the time of primiparous female mating (February 15) is used as a default in the development of the control rule.

$F^* = F_{35\%}$, a proxy of F_{MSY} , which is a full selection instantaneous F that will produce MSY at the MSY producing biomass,

$B^* = B_{35\%}$, a proxy of B_{MSY} , which is the value of biomass at the MSY producing level,

β = a parameter with restriction that $0 \leq \beta < 1$. A default value of 0.25 is used.

α = a parameter with restriction that $0 \leq \alpha \leq \beta$. A default value of 0.1 is used.

Because trawl bycatch fishing mortality was not related to pot fishing mortality, average trawl bycatch fishing mortality during 2007 to 2016 was used for the per recruit analysis as well as for projections in the next section. Pot female bycatch fishing mortality was set equal to pot male fishing mortality times 0.02, an intermediate level during 1990-2016. Some discards of legal males occurred since the IFQ fishery started in 2005, but the discard rates were much lower during 2007-2013 than in 2005 after the fishing industry minimized discards of legal males. However, due to the high proportion of large oldshell males, the discard rate increased greatly in 2014. The average of retained selectivities and discard male selectivities during 2015-2016 were used to represent current trends for per recruit analysis and projections. Average molting probabilities during 2007-2016 were used for per recruit analysis and projections.

Average recruitments during three periods were used to estimate $B_{35\%}$: 1976-2017, 1984-2017, and 1991-2017 (Figure 12). Estimated $B_{35\%}$ is compared with historical mature male biomass in Figure 13a. We recommend using the average recruitment during 1984-present, corresponding to the 1976/77 regime shift. Note that recruitment period 1984-present has been used since 2011 to set the overfishing limits. Several factors support our recommendation. First, estimated recruitment was lower after 1983 than before 1984, which corresponded to brood years 1978 and later, after the 1976/77 regime shift. Second, high recruitments during the late 1960s and 1970s generally occurred when the spawning stock was primarily located in the southern Bristol Bay, whereas the current spawning stock is mainly in the middle of Bristol Bay. The current flows favor larvae hatched in the southern Bristol Bay (see the section on Ecosystem Considerations for SAFE reports in 2008 and 2009). Finally, stock productivity (recruitment/mature male biomass) was higher before the 1976/1977 regime shift.

If we believe that differences in productivity and other population characteristics before 1978 were caused by fishing, not by the regime shift, then we should use the recruitment from 1976-1983 (corresponding to brood years before 1978) as the baseline to estimate $B_{35\%}$. If we believe that the regime shift during 1976/77 caused the productivity differences, then we should select the recruitments from period 1984-2017 as the baseline.

The control rule is used for stock status determination. If total catch exceeds OFL estimated at B , then “overfishing” occurs. If B equals or declines below $0.5 B_{MSY}$ (i.e., MSST), the stock is “overfished.” If B equals or declines below $\beta^* B_{MSY}$ or β^* a proxy B_{MSY} , then the stock productivity is severely depleted and the fishery is closed.

The estimated probability distribution of MMB in 2017 is illustrated in Figure 30. Based the SSC suggestion in 2011, $ABC = 0.9 * OFL$ is used to estimate ABC.

Status and catch specifications (1,000 t) (scenario 2b):

| Year | MSST | Biomass (MMB) | TAC | Retained Catch | Total Catch | OFL | ABC |
|---------|--------------------|--------------------|------|----------------|-------------|------|------|
| 2013/14 | 12.85 ^A | 27.12 ^A | 3.90 | 3.99 | 4.56 | 7.07 | 6.36 |
| 2014/15 | 13.03 ^B | 27.25 ^B | 4.49 | 4.54 | 5.44 | 6.82 | 6.14 |
| 2015/16 | 12.89 ^C | 27.68 ^C | 4.52 | 4.61 | 5.34 | 6.73 | 6.06 |
| 2016/17 | 12.53 ^D | 25.80 ^D | 3.84 | 3.92 | 4.28 | 6.64 | 5.97 |
| 2017/18 | | 21.31 ^D | | | | 5.60 | 5.04 |

The stock was above MSST in 2016/17 and hence was not overfished. Overfishing did not occur.

Status and catch specifications (million lbs):

| Year | MSST | Biomass (MMB) | TAC | Retained Catch | Total Catch | OFL | ABC |
|---------|-------------------|-------------------|------|----------------|-------------|-------|-------|
| 2013/14 | 28.3 ^A | 59.9 ^A | 8.60 | 8.80 | 10.05 | 15.58 | 14.02 |
| 2014/15 | 28.7 ^B | 60.1 ^B | 9.99 | 10.01 | 11.99 | 15.04 | 13.53 |
| 2015/16 | 28.4 ^C | 61.0 ^C | 9.97 | 10.17 | 11.77 | 14.84 | 13.36 |
| 2016/17 | 27.6 ^D | 56.9 ^D | 8.47 | 8.65 | 9.45 | 14.63 | 13.17 |
| 2017/18 | | 47.0 ^D | | | | 12.35 | 11.11 |

Notes:

- A – Calculated from the assessment reviewed by the Crab Plan Team in September 2014
- B – Calculated from the assessment reviewed by the Crab Plan Team in September 2015
- C – Calculated from the assessment reviewed by the Crab Plan Team in September 2016
- D – Calculated from the assessment reviewed by the Crab Plan Team in September 2017

4. Based on the $B_{35\%}$ estimated from the average male recruitment during 1984-2017, the biological reference points and OFL were estimated in Table 4.
5. Based on the 10% buffer rule used last year, $ABC = 0.9 * OFL$ (Table 4). If $P^*=49\%$ is used, the ABC will be higher.

G. Rebuilding Analyses

NA.

H. Data Gaps and Research Priorities

1. The following data gaps exist for this stock:
 - a. Information about changes in natural mortality in the early 1980s;
 - b. Un-observed trawl bycatch in the early 1980s;
 - c. Natural mortality;

- d. Crab availability to the trawl surveys;
 - e. Juvenile crab abundance;
 - f. Female growth per molt as a function of size and maturity;
 - g. Changes in male molting probability over time.
2. Research priorities:
- a. Estimating natural mortality;
 - b. Estimating crab availability to the trawl surveys;
 - c. Surveying juvenile crab abundance in nearshore;
 - d. Studying environmental factors that affect the survival rates from larvae to recruitment.

I. Projections and Future Outlook

1. Projections

Future population projections primarily depend on future recruitment, but crab recruitment is difficult to predict. Therefore, annual recruitment for the projections was a random selection from estimated recruitments during 1984-2017. Besides recruitment, the other major uncertainty for the projections is estimated abundance in 2017. The 2017 abundance was randomly selected from the estimated normal distribution of the assessment model output for each replicate. Three scenarios of fishing mortality for the directed pot fishery were used in the projections:

- (1) No directed fishery. This was used as a base projection.
- (2) $F_{40\%}$. This fishing mortality creates a buffer between the limits and target levels.
- (3) $F_{35\%}$. This is the maximum fishing mortality allowed under the current overfishing definitions.

Each scenario was replicated 1,000 times and projections made over 10 years beginning in 2017 (Table 7).

As expected, projected mature male biomasses are much higher without the directed fishing mortality than under the other scenarios. At the end of 10 years, projected mature male biomass is above $B_{35\%}$ for all scenarios (Table 7; Figure 32). Projected retained catch for the $F_{35\%}$ scenario is higher than those for the $F_{40\%}$ scenario (Table 7, Figure 33). Due to the poor recruitment in recent years, the projected biomass and retained catch are expected to decline during the next few years.

2. Near Future Outlook

The near future outlook for the Bristol Bay RKC stock is a declining trend. The three recent above-average year classes (hatching years 1990, 1994, and 1997) had entered the legal population by 2006 (Figure 34). Most individuals from the 1997 year class will continue to gain weight to offset loss of the legal biomass to fishing and natural mortalities. The above-average year class (hatching year 2000) with lengths centered around 87.5 mm CL for both males and females in 2006 and with lengths

centered around 112.5-117.5 mm CL for males and around 107.5 mm CL for females in 2008 has largely entered the mature male population in 2009 and the legal population by 2014 (Figure 34). No strong cohorts have been observed in the survey data after this cohort through 2010 (Figure 34). There was a huge tow of juvenile crab of size 45-55 mm in 2011, but these juveniles were not tracked during 2012-2017 surveys. This single tow is unlikely to be an indicator for a strong cohort. The high survey abundance of large males and mature females in 2014 cannot be explained by the survey data during the previous years and were also inconsistent with the 2015-2017 survey results (Figure 34). Due to lack of recruitment, mature and legal crab should continue to decline next year. Current crab abundance is still low relative to the late 1970s, and without favorable environmental conditions, recovery to the high levels of the late 1970s is unlikely.

J. Acknowledgements

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Table 1a. Bristol Bay red king crab annual catch and bycatch mortality biomass (t) from June 1 to May 31. A handling mortality rate of 20% for the directed pot, 25% for the Tanner fishery, and 80% for trawl was assumed to estimate bycatch mortality biomass.

| Year | Retained Catch | | | Pot Bycatch | | Trawl Bycat. | Fixed Bycat. | Tanner Fishery Bycat. | Total Catch |
|------|----------------|---------------|---------|-------------|-------|--------------|--------------|-----------------------|-------------|
| | U.S. | Cost-Recovery | Foreign | Total | Males | | | | |
| 1953 | 1331.3 | | 4705.6 | 6036.9 | | | | | 6036.9 |
| 1954 | 1149.9 | | 3720.4 | 4870.2 | | | | | 4870.2 |
| 1955 | 1029.2 | | 3712.7 | 4741.9 | | | | | 4741.9 |
| 1956 | 973.4 | | 3572.9 | 4546.4 | | | | | 4546.4 |
| 1957 | 339.7 | | 3718.1 | 4057.8 | | | | | 4057.8 |
| 1958 | 3.2 | | 3541.6 | 3544.8 | | | | | 3544.8 |
| 1959 | 0.0 | | 6062.3 | 6062.3 | | | | | 6062.3 |
| 1960 | 272.2 | | 12200.7 | 12472.9 | | | | | 12472.9 |
| 1961 | 193.7 | | 20226.6 | 20420.3 | | | | | 20420.3 |
| 1962 | 30.8 | | 24618.7 | 24649.6 | | | | | 24649.6 |
| 1963 | 296.2 | | 24930.8 | 25227.0 | | | | | 25227.0 |
| 1964 | 373.3 | | 26385.5 | 26758.8 | | | | | 26758.8 |
| 1965 | 648.2 | | 18730.6 | 19378.8 | | | | | 19378.8 |
| 1966 | 452.2 | | 19212.4 | 19664.6 | | | | | 19664.6 |
| 1967 | 1407.0 | | 15257.0 | 16664.1 | | | | | 16664.1 |
| 1968 | 3939.9 | | 12459.7 | 16399.6 | | | | | 16399.6 |
| 1969 | 4718.7 | | 6524.0 | 11242.7 | | | | | 11242.7 |
| 1970 | 3882.3 | | 5889.4 | 9771.7 | | | | | 9771.7 |
| 1971 | 5872.2 | | 2782.3 | 8654.5 | | | | | 8654.5 |
| 1972 | 9863.4 | | 2141.0 | 12004.3 | | | | | 12004.3 |
| 1973 | 12207.8 | | 103.4 | 12311.2 | | | | | 12311.2 |
| 1974 | 19171.7 | | 215.9 | 19387.6 | | | | | 19387.6 |
| 1975 | 23281.2 | | 0 | 23281.2 | | | | | 23281.2 |
| 1976 | 28993.6 | | 0 | 28993.6 | | | 682.8 | | 29676.4 |
| 1977 | 31736.9 | | 0 | 31736.9 | | | 1249.9 | | 32986.8 |
| 1978 | 39743.0 | | 0 | 39743.0 | | | 1320.6 | | 41063.6 |
| 1979 | 48910.0 | | 0 | 48910.0 | | | 1331.9 | | 50241.9 |
| 1980 | 58943.6 | | 0 | 58943.6 | | | 1036.5 | | 59980.1 |
| 1981 | 15236.8 | | 0 | 15236.8 | | | 219.4 | | 15456.2 |
| 1982 | 1361.3 | | 0 | 1361.3 | | | 574.9 | | 1936.2 |
| 1983 | 0.0 | | 0 | 0.0 | | | 420.4 | | 420.4 |
| 1984 | 1897.1 | | 0 | 1897.1 | | | 1094.0 | | 2991.1 |
| 1985 | 1893.8 | | 0 | 1893.8 | | | 390.1 | | 2283.8 |
| 1986 | 5168.2 | | 0 | 5168.2 | | | 200.6 | | 5368.8 |
| 1987 | 5574.2 | | 0 | 5574.2 | | | 186.4 | | 5760.7 |
| 1988 | 3351.1 | | 0 | 3351.1 | | | 597.8 | | 3948.9 |
| 1989 | 4656.0 | | 0 | 4656.0 | | | 174.1 | | 4830.1 |
| 1990 | 9236.2 | 36.6 | 0 | 9272.8 | 526.9 | 651.5 | 247.6 | | 10698.7 |
| 1991 | 7791.8 | 93.4 | 0 | 7885.1 | 407.8 | 75.0 | 316.0 | 1401.8 | 10085.7 |
| 1992 | 3648.2 | 33.6 | 0 | 3681.8 | 552.0 | 418.5 | 335.4 | 244.4 | 5232.2 |
| 1993 | 6635.4 | 24.1 | 0 | 6659.6 | 763.2 | 637.1 | 426.6 | 54.6 | 8541.0 |
| 1994 | 0.0 | 42.3 | 0 | 42.3 | 3.8 | 1.9 | 88.9 | 10.8 | 147.8 |
| 1995 | 0.0 | 36.4 | 0 | 36.4 | 3.3 | 1.6 | 194.2 | 0.0 | 235.5 |
| 1996 | 3812.7 | 49.0 | 0 | 3861.7 | 164.6 | 1.0 | 106.5 | 0.0 | 4133.9 |
| 1997 | 3971.9 | 70.2 | 0 | 4042.1 | 244.7 | 19.6 | 73.4 | 0.0 | 4379.8 |
| 1998 | 6693.8 | 85.4 | 0 | 6779.2 | 959.7 | 864.9 | 159.8 | 0.0 | 8763.7 |
| 1999 | 5293.5 | 84.3 | 0 | 5377.9 | 314.2 | 8.8 | 201.6 | 0.0 | 5902.4 |
| 2000 | 3698.8 | 39.1 | 0 | 3737.9 | 360.8 | 40.5 | 100.4 | 0.0 | 4239.5 |
| 2001 | 3811.5 | 54.6 | 0 | 3866.2 | 417.9 | 173.5 | 164.6 | 0.0 | 4622.1 |

| | | | | | | | | | | |
|------|--------|-------|---|--------|--------|-------|-------|------|------|---------|
| 2002 | 4340.9 | 43.6 | 0 | 4384.5 | 442.7 | 7.3 | 155.1 | | 0.0 | 4989.6 |
| 2003 | 7120.0 | 15.3 | 0 | 7135.3 | 918.9 | 430.4 | 172.3 | | 0.0 | 8656.9 |
| 2004 | 6915.2 | 91.4 | 0 | 7006.7 | 345.5 | 187.0 | 119.6 | | 0.0 | 7658.8 |
| 2005 | 8305.0 | 94.7 | 0 | 8399.7 | 1359.5 | 498.3 | 155.2 | | 0.0 | 10412.8 |
| 2006 | 7005.3 | 137.9 | 0 | 7143.2 | 563.8 | 37.0 | 116.7 | | 3.8 | 7864.4 |
| 2007 | 9237.9 | 66.1 | 0 | 9303.9 | 1001.3 | 186.1 | 138.5 | | 1.8 | 10631.6 |
| 2008 | 9216.1 | 0.0 | 0 | 9216.1 | 1165.5 | 148.4 | 159.5 | | 4.0 | 10693.5 |
| 2009 | 7226.9 | 45.5 | 0 | 7272.5 | 888.1 | 85.2 | 87.2 | 5.0 | 1.6 | 8339.6 |
| 2010 | 6728.5 | 33.0 | 0 | 6761.5 | 797.5 | 122.6 | 78.7 | 2.3 | 0.0 | 7762.6 |
| 2011 | 3553.3 | 53.8 | 0 | 3607.1 | 395.0 | 24.0 | 53.8 | 9.4 | 0.0 | 4089.2 |
| 2012 | 3560.6 | 61.1 | 0 | 3621.7 | 205.2 | 12.3 | 32.4 | 14.9 | 0.0 | 3886.5 |
| 2013 | 3901.1 | 89.9 | 0 | 3991.0 | 310.6 | 99.8 | 61.9 | 39.5 | 28.5 | 4531.1 |
| 2014 | 4530.0 | 8.6 | 0 | 4538.6 | 584.7 | 86.2 | 32.0 | 82.7 | 42.0 | 5366.2 |
| 2015 | 4522.3 | 91.4 | 0 | 4613.7 | 266.1 | 222.9 | 41.7 | 67.9 | 84.2 | 5296.5 |
| 2016 | 3840.4 | 83.4 | 0 | 3923.9 | 237.4 | 87.1 | 21.0 | 14.8 | 0.0 | 4284.2 |

Table 1b. Annual retained catch (millions of crab) and catch per unit effort of the Bristol Bay red king crab fishery.

| Year | Japanese Tanglenet | | Russian Tanglenet | | U.S. Pot/Trawl | | Standardized Crab/tan |
|------|--------------------|----------|-------------------|----------|----------------|--------------|--------------------------|
| | Catch | Crab/tan | Catch | Crab/tan | Catch | Crab/Potlift | |
| 1960 | 1.949 | 15.2 | 1.995 | 10.4 | 0.088 | | 15.8 |
| 1961 | 3.031 | 11.8 | 3.441 | 8.9 | 0.062 | | 12.9 |
| 1962 | 4.951 | 11.3 | 3.019 | 7.2 | 0.010 | | 11.3 |
| 1963 | 5.476 | 8.5 | 3.019 | 5.6 | 0.101 | | 8.6 |
| 1964 | 5.895 | 9.2 | 2.800 | 4.6 | 0.123 | | 8.5 |
| 1965 | 4.216 | 9.3 | 2.226 | 3.6 | 0.223 | | 7.7 |
| 1966 | 4.206 | 9.4 | 2.560 | 4.1 | 0.140 | 52 | 8.1 |
| 1967 | 3.764 | 8.3 | 1.592 | 2.4 | 0.397 | 37 | 6.3 |
| 1968 | 3.853 | 7.5 | 0.549 | 2.3 | 1.278 | 27 | 7.8 |
| 1969 | 2.073 | 7.2 | 0.369 | 1.5 | 1.749 | 18 | 5.6 |
| 1970 | 2.080 | 7.3 | 0.320 | 1.4 | 1.683 | 17 | 5.6 |
| 1971 | 0.886 | 6.7 | 0.265 | 1.3 | 2.405 | 20 | 5.8 |
| 1972 | 0.874 | 6.7 | | | 3.994 | 19 | |
| 1973 | 0.228 | | | | 4.826 | 25 | |
| 1974 | 0.476 | | | | 7.710 | 36 | |
| 1975 | | | | | 8.745 | 43 | |
| 1976 | | | | | 10.603 | 33 | |
| 1977 | | | | | 11.733 | 26 | |
| 1978 | | | | | 14.746 | 36 | |
| 1979 | | | | | 16.809 | 53 | |
| 1980 | | | | | 20.845 | 37 | |
| 1981 | | | | | 5.308 | 10 | |
| 1982 | | | | | 0.541 | 4 | |
| 1983 | | | | | 0.000 | | |
| 1984 | | | | | 0.794 | 7 | |
| 1985 | | | | | 0.796 | 9 | |
| 1986 | | | | | 2.100 | 12 | |
| 1987 | | | | | 2.122 | 10 | |
| 1988 | | | | | 1.236 | 8 | |
| 1989 | | | | | 1.685 | 8 | |
| 1990 | | | | | 3.130 | 12 | |
| 1991 | | | | | 2.661 | 12 | |
| 1992 | | | | | 1.208 | 6 | |
| 1993 | | | | | 2.270 | 9 | |
| 1994 | | | | | 0.015 | | |
| 1995 | | | | | 0.014 | | |
| 1996 | | | | | 1.264 | 16 | |
| 1997 | | | | | 1.338 | 15 | |
| 1998 | | | | | 2.238 | 15 | |
| 1999 | | | | | 1.923 | 12 | |
| 2000 | | | | | 1.272 | 12 | |
| 2001 | | | | | 1.287 | 19 | |
| 2002 | | | | | 1.484 | 20 | |
| 2003 | | | | | 2.510 | 18 | |
| 2004 | | | | | 2.272 | 23 | |
| 2005 | | | | | 2.763 | 30 | |
| 2006 | | | | | 2.477 | 31 | |
| 2007 | | | | | 3.154 | 28 | |
| 2008 | | | | | 3.064 | 22 | |
| 2009 | | | | | 2.553 | 21 | |
| 2010 | | | | | 2.410 | 18 | |
| 2011 | | | | | 1.298 | 28 | |
| 2012 | | | | | 1.176 | 30 | |
| 2013 | | | | | 1.272 | 27 | |
| 2014 | | | | | 1.501 | 26 | |
| 2015 | | | | | 1.527 | 31 | |
| 2016 | | | | | 1.281 | 38 | |

Table 2. Annual sample sizes (>64 mm CL) in numbers of crab for trawl surveys, retained catch and pot and trawl fishery bycatch of Bristol Bay red king crab.

| Year | Trawl Survey | | Retained Catch | Pot Bycatch | | Trawl Bycatch | | Tanner Fishery Bycatch | |
|------|--------------|---------|----------------|-------------|---------|---------------|---------|------------------------|---------|
| | Males | Females | | Males | Females | Males | Females | Males | Females |
| 1975 | 2,943 | 2,139 | 29,570 | | | | | | |
| 1976 | 4,724 | 2,956 | 26,450 | | | 2,327 | 676 | | |
| 1977 | 3,636 | 4,178 | 32,596 | | | 14,014 | 689 | | |
| 1978 | 4,132 | 3,948 | 27,529 | | | 8,983 | 1,456 | | |
| 1979 | 5,807 | 4,663 | 27,900 | | | 7,228 | 2,821 | | |
| 1980 | 2,412 | 1,387 | 34,747 | | | 47,463 | 39,689 | | |
| 1981 | 3,478 | 4,097 | 18,029 | | | 42,172 | 49,634 | | |
| 1982 | 2,063 | 2,051 | 11,466 | | | 84,240 | 47,229 | | |
| 1983 | 1,524 | 944 | 0 | | | 204,464 | 104,910 | | |
| 1984 | 2,679 | 1,942 | 4,404 | | | 357,981 | 147,134 | | |
| 1985 | 792 | 415 | 4,582 | | | 169,767 | 30,693 | | |
| 1986 | 1,962 | 367 | 5,773 | | | 1,199 | 284 | | |
| 1987 | 1,168 | 1,018 | 4,230 | | | 723 | 927 | | |
| 1988 | 1,834 | 546 | 9,833 | | | 437 | 275 | | |
| 1989 | 1,257 | 550 | 32,858 | | | 3,147 | 194 | | |
| 1990 | 858 | 603 | 7,218 | 873 | 699 | 761 | 1,570 | | |
| 1991 | 1,378 | 491 | 36,820 | 1,801 | 375 | 208 | 396 | 885 | 2,198 |
| 1992 | 513 | 360 | 23,552 | 3,248 | 2,389 | 214 | 107 | 280 | 685 |
| 1993 | 1,009 | 534 | 32,777 | 5,803 | 5,942 | | | 232 | 265 |
| 1994 | 443 | 266 | 0 | 0 | 0 | 330 | 247 | | |
| 1995 | 2,154 | 1,718 | 0 | 0 | 0 | 103 | 35 | | |
| 1996 | 835 | 816 | 8,896 | 230 | 11 | 1,025 | 968 | | |
| 1997 | 1,282 | 707 | 15,747 | 4,102 | 906 | 1,202 | 483 | | |
| 1998 | 1,097 | 1,150 | 16,131 | 11,079 | 9,130 | 1,627 | 915 | | |
| 1999 | 764 | 540 | 17,666 | 1,048 | 36 | 2,154 | 858 | | |
| 2000 | 731 | 1,225 | 14,091 | 8,970 | 1,486 | 994 | 671 | | |
| 2001 | 611 | 743 | 12,854 | 9,102 | 4,567 | 4,393 | 2,521 | | |
| 2002 | 1,032 | 896 | 15,932 | 9,943 | 302 | 3,372 | 1,464 | | |
| 2003 | 1,669 | 1,311 | 16,212 | 17,998 | 10,327 | 1,568 | 1,057 | | |
| 2004 | 2,871 | 1,599 | 20,038 | 8,258 | 4,112 | 1,689 | 1,506 | | |
| 2005 | 1,283 | 1,682 | 21,938 | 55,019 | 26,775 | 1,815 | 1,872 | | |
| 2006 | 1,171 | 2,672 | 18,027 | 32,252 | 3,980 | 1,481 | 1,983 | | |
| 2007 | 1,219 | 2,499 | 22,387 | 59,769 | 12,661 | 1,011 | 1,097 | | |
| 2008 | 1,221 | 3,352 | 14,567 | 49,315 | 8,488 | 1,867 | 1,039 | | |
| 2009 | 830 | 1,857 | 16,708 | 52,359 | 6,041 | 1,431 | 848 | | |
| 2010 | 705 | 1,633 | 20,137 | 36,654 | 6,868 | 612 | 837 | | |
| 2011 | 525 | 994 | 10,706 | 20,629 | 1,920 | 563 | 1,068 | | |
| 2012 | 580 | 707 | 8,956 | 7,206 | 561 | 1,507 | 1,751 | | |
| 2013 | 633 | 560 | 10,197 | 13,828 | 6,048 | 4,806 | 4,198 | 218 | 596 |
| 2014 | 1,106 | 1,255 | 9,618 | 13,040 | 1,950 | 1,966 | 2,580 | 256 | 381 |
| 2015 | 600 | 677 | 11,746 | 8,037 | 5,889 | 1,150 | 3,731 | 726 | 2163 |
| 2016 | 374 | 803 | 10,811 | 9,497 | 4,216 | 1,908 | 2,879 | | |
| 2017 | 470 | 558 | | | | | | | |

Table 3. Number of parameters and the list of likelihood components for the model (Scenarios 2b, 2bn1, 2bn2, 2b85, and 2c85).

| Parameter counts | Sce. 2b | 2bn1 | 2bn2 | 2b85 | 2c85 |
|---|----------------|-------------|-------------|-------------|-------------|
| Fixed growth parameters | 9 | 9 | 9 | 9 | 9 |
| Fixed recruitment parameters | 2 | 2 | 2 | 2 | 2 |
| Fixed length-weight relationship parameters | 6 | 6 | 6 | 6 | 6 |
| Fixed mortality parameters | 4 | 4 | 4 | 4 | 4 |
| Fixed survey catchability parameter | 1 | 1 | 1 | 1 | 1 |
| Fixed high grading parameters | 12 | 0 | 0 | 12 | 12 |
| Total number of fixed parameters | 34 | 22 | 22 | 34 | 34 |
| Free survey catchability parameter | 1 | 1 | 1 | 1 | 1 |
| Free growth parameters | 6 | 6 | 6 | 4 | 4 |
| Initial abundance (1975 or 1985) | 1 | 1 | 1 | 1 | 1 |
| Recruitment-distribution parameters | 2 | 2 | 2 | 2 | 2 |
| Mean recruitment parameters | 1 | 1 | 1 | 1 | 1 |
| Male recruitment deviations | 42 | 42 | 42 | 32 | 32 |
| Female recruitment deviations | 42 | 42 | 42 | 32 | 32 |
| Natural and fishing mortality parameters | 4 | 4 | 4 | 1 | 0 |
| Pot male fishing mortality deviations | 43 | 43 | 43 | 33 | 33 |
| Bycatch mortality from the Tanner crab fishery | 11 | 11 | 11 | 11 | 11 |
| Pot female bycatch fishing mortality deviations | 28 | 28 | 28 | 28 | 28 |
| Trawl bycatch fishing mortality deviations | 42 | 42 | 42 | 33 | 33 |
| Fixed gear bycatch fishing mortality deviations | 9 | 9 | 9 | 9 | 9 |
| Initial (1975 or 1985) length compositions | 35 | 35 | 35 | 35 | 35 |
| BSFRF survey extra CV | 1 | 1 | 1 | 1 | 1 |
| Free selectivity parameters | 24 | 25 | 36 | 22 | 22 |
| Total number of free parameters | 292 | 293 | 304 | 246 | 245 |
| Total number of fixed and free parameters | 326 | 327 | 338 | 280 | 279 |

Table 4. Negative log likelihood components and differences for scenarios 2b, 2bn1, 2bn2, 2b85 and 2c85 and some management quantities.

| | Scenario | | | | | | | | |
|-------------------------|----------|---------|---------|---------|---------|-------------|-------------|---------------|---------------|
| | 2b | 2bn1 | 2bn2 | 2b85 | 2c85 | 2b- 2bn1 | 2b- 2bn2 | 2bn1- 2bn2 | 2b85- 2c85 |
| Negative log likelihood | | | | | | | | | |
| R-variation | 87.22 | 87.80 | 86.27 | 48.06 | 54.29 | -0.58 | 0.95 | 1.53 | -6.23 |
| Length-like-retained | -1038.9 | -1047.0 | -1039.5 | -834.8 | -834.4 | 8.09 | 0.62 | -7.47 | -0.40 |
| Length-like-male | -1092.4 | -1224.4 | -1226.3 | -1092.6 | -1094.2 | 132.00 | 133.89 | 1.89 | 1.59 |
| Length-like-discfemale | -812.44 | -812.44 | -811.84 | -795.32 | -796.43 | 0.00 | -0.60 | -0.60 | 1.11 |
| Length-like-survey | -48629 | -48610 | -48632 | -37848 | -37759 | -18.60 | 3.00 | 21.60 | -88.50 |
| Length-like-disctrawl | -3767.1 | -3784.5 | -3785.1 | -2957.1 | -2962.0 | 17.38 | 17.97 | 0.59 | 4.92 |
| Length-like-discfix | -773.41 | -773.29 | -774.05 | -773.38 | -773.13 | -0.12 | 0.64 | 0.76 | -0.25 |
| Length-like-discTanner | -467.04 | -468.17 | -468.51 | -468.44 | -465.25 | 1.13 | 1.47 | 0.34 | -3.20 |
| Length-like-bsfrfsurvey | -645.73 | -645.18 | -647.28 | -644.35 | -641.97 | -0.55 | 1.55 | 2.10 | -2.38 |
| Catchbio_retained | 51.14 | 68.04 | 47.04 | 49.39 | 48.37 | -16.90 | 4.10 | 21.00 | 1.02 |
| Catchbio_discmale | 229.04 | 275.48 | 191.62 | 224.63 | 220.82 | -46.44 | 37.42 | 83.86 | 3.80 |
| Catchbio-discfemale | 0.11 | 0.09 | 0.10 | 0.08 | 0.45 | 0.02 | 0.01 | -0.01 | -0.37 |
| Catchbio-disctrawl | 0.22 | 0.24 | 0.25 | 0.05 | 0.07 | -0.03 | -0.03 | 0.00 | -0.01 |
| Catchbio-discfix | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Catchbio-discTanner | 0.13 | 0.15 | 0.15 | 0.10 | 1.72 | -0.02 | -0.02 | -0.01 | -1.62 |
| Biomass-trawl survey | 103.70 | 105.66 | 108.41 | 51.06 | 53.65 | -1.95 | -4.71 | -2.75 | -2.59 |
| Biomass-bsfrfsurvey | -8.30 | -7.41 | -8.52 | -8.47 | -8.40 | -0.89 | 0.22 | 1.11 | -0.07 |
| Q-trawl survey | 3.84 | 8.60 | 6.22 | 0.22 | 0.00 | -4.76 | -2.38 | 2.38 | 0.22 |
| Others | 18.10 | 17.89 | 26.90 | 14.99 | 14.33 | 0.21 | -8.80 | -9.02 | 0.66 |
| Total | -56741 | -56809 | -56926 | -45034 | -44942 | 68 | 185 | 117 | -92 |
| Free parameters | 292 | 293 | 304 | 246 | 245 | -1 | -12 | -11 | 1 |
| B35%(t) | 25050 | 24283 | 25119 | 23512 | 21985 | 767 | -69 | -836 | 1527 |
| F35% | 0.29 | 0.31 | 0.33 | 0.30 | 0.30 | -0.02 | -0.04 | -0.02 | 0.00 |
| MMB2017(t) | 21312 | 19269 | 22347 | 23089 | 20570 | 2043 | -1035 | -3078 | 2519 |
| OFL2017 | 5599.7 | 4659.0 | 6186.5 | 7448.9 | 6269.2 | 940.7 | -586.8 | -1527.5 | 1179.7 |
| ABC2017(t) | 5039.7 | 4193.1 | 5567.9 | 6704.0 | 5642.3 | 846.6 | -528.1 | -1374.8 | 1061.7 |
| Fofl2017 | 0.24 | 0.24 | 0.29 | 0.29 | 0.28 | 0.00 | -0.05 | -0.05 | 0.01 |
| Q82-17 | 0.97 | 1.00 | 0.98 | 0.91 | 0.92 | -0.03 | -0.01 | 0.02 | -0.01 |

Table 5(2b). Summary of estimated model parameter values and standard deviations and limits for scenario 2b for Bristol Bay red king crab. All values are on a log scale. Male recruit in year t is $\exp(\text{mean}+\text{males}_t)$, and female recruit in year t is $\exp(\text{mean}+\text{males}_t+\text{females}_t)$.

| Year | Recruits | | | | F for Directed Pot Fishery | | | | F for Trawl | |
|---------|----------|-------|--------|-------|----------------------------|-------|----------|-------|-------------|-------|
| | Females | SD | Males | SD | Males | SD | Females | SD | Estimate | SD |
| Mean | 15.796 | 0.023 | 15.796 | 0.023 | -1.680 | 0.041 | 0.012 | 0.001 | -4.621 | 0.070 |
| Limits↑ | 13,18 | | 13,18 | | -3.0,0.0 | | .001,0.1 | | -8.5,-0.5 | |
| Limits↓ | -15,15 | | -15,15 | | -15,2.43 | | -6.0,3.5 | | -10,10 | |
| 1975 | | | | | 0.820 | 0.096 | | | | |
| 1976 | -0.033 | 0.277 | 0.812 | 0.137 | 0.806 | 0.068 | | | 0.201 | 0.111 |
| 1977 | 0.521 | 0.161 | 0.682 | 0.103 | 0.800 | 0.059 | | | 0.710 | 0.107 |
| 1978 | 0.450 | 0.137 | 0.886 | 0.085 | 1.010 | 0.055 | | | 0.781 | 0.106 |
| 1979 | 0.727 | 0.102 | 1.145 | 0.077 | 1.304 | 0.052 | | | 0.949 | 0.106 |
| 1980 | 0.239 | 0.116 | 1.320 | 0.077 | 2.170 | 0.047 | | | 1.663 | 0.106 |
| 1981 | 0.089 | 0.149 | 0.519 | 0.103 | 2.425 | 0.009 | | | 1.203 | 0.107 |
| 1982 | 0.105 | 0.055 | 2.107 | 0.051 | 0.576 | 0.049 | | | 2.378 | 0.107 |
| 1983 | 0.034 | 0.073 | 1.446 | 0.052 | -10.62 | 0.936 | | | 2.036 | 0.104 |
| 1984 | 0.484 | 0.060 | 1.488 | 0.049 | 0.725 | 0.056 | | | 3.089 | 0.104 |
| 1985 | 0.119 | 0.200 | -0.582 | 0.122 | 0.808 | 0.064 | | | 1.935 | 0.106 |
| 1986 | 0.582 | 0.061 | 0.765 | 0.047 | 1.339 | 0.062 | | | 1.000 | 0.107 |
| 1987 | -0.047 | 0.144 | -0.117 | 0.074 | 0.944 | 0.058 | | | 0.585 | 0.106 |
| 1988 | 0.301 | 0.176 | -0.815 | 0.107 | -0.026 | 0.050 | | | 1.367 | 0.102 |
| 1989 | 0.105 | 0.158 | -0.672 | 0.089 | 0.063 | 0.047 | | | -0.073 | 0.102 |
| 1990 | -0.023 | 0.071 | 0.470 | 0.046 | 0.668 | 0.043 | 1.987 | 0.080 | 0.299 | 0.102 |
| 1991 | -0.071 | 0.098 | 0.012 | 0.056 | 0.647 | 0.045 | -0.137 | 0.080 | 0.634 | 0.104 |
| 1992 | -0.584 | 0.427 | -1.744 | 0.171 | 0.132 | 0.047 | 2.167 | 0.081 | 0.687 | 0.103 |
| 1993 | -0.263 | 0.101 | -0.223 | 0.056 | 0.786 | 0.049 | 2.045 | 0.081 | 1.111 | 0.104 |
| 1994 | -0.451 | 0.475 | -2.094 | 0.198 | -4.356 | 0.049 | 1.421 | 0.113 | -0.655 | 0.103 |
| 1995 | 0.021 | 0.041 | 1.349 | 0.036 | -4.707 | 0.046 | 1.541 | 0.119 | -0.058 | 0.102 |
| 1996 | -0.872 | 0.288 | -0.467 | 0.113 | -0.161 | 0.043 | -3.653 | 0.140 | -0.673 | 0.103 |
| 1997 | -0.931 | 0.425 | -1.335 | 0.167 | -0.052 | 0.044 | -1.006 | 0.085 | -1.036 | 0.103 |
| 1998 | -0.330 | 0.128 | -0.078 | 0.068 | 0.646 | 0.044 | 2.052 | 0.078 | -0.172 | 0.102 |
| 1999 | 0.072 | 0.062 | 0.753 | 0.043 | 0.198 | 0.044 | -2.083 | 0.085 | -0.007 | 0.102 |
| 2000 | -0.125 | 0.149 | -0.181 | 0.080 | -0.179 | 0.043 | -0.275 | 0.079 | -0.825 | 0.102 |
| 2001 | 0.642 | 0.191 | -0.844 | 0.138 | -0.164 | 0.043 | 1.092 | 0.078 | -0.394 | 0.101 |
| 2002 | 0.213 | 0.057 | 1.207 | 0.041 | -0.064 | 0.043 | -2.242 | 0.086 | -0.484 | 0.101 |
| 2003 | -0.086 | 0.259 | -0.533 | 0.142 | 0.459 | 0.042 | 1.170 | 0.079 | -0.340 | 0.101 |
| 2004 | -0.230 | 0.161 | 0.200 | 0.082 | 0.315 | 0.043 | 0.378 | 0.079 | -0.707 | 0.101 |
| 2005 | 0.313 | 0.063 | 1.127 | 0.046 | 0.735 | 0.043 | 0.891 | 0.078 | -0.397 | 0.101 |
| 2006 | -0.750 | 0.177 | 0.520 | 0.064 | 0.446 | 0.043 | -1.513 | 0.080 | -0.732 | 0.101 |
| 2007 | -0.290 | 0.161 | -0.046 | 0.082 | 0.763 | 0.044 | -0.284 | 0.079 | -0.558 | 0.102 |
| 2008 | 0.138 | 0.160 | -0.507 | 0.100 | 0.846 | 0.046 | -0.597 | 0.079 | -0.410 | 0.102 |
| 2009 | 0.296 | 0.140 | -0.484 | 0.094 | 0.540 | 0.047 | -0.822 | 0.080 | -1.003 | 0.103 |
| 2010 | 0.050 | 0.100 | 0.108 | 0.063 | 0.395 | 0.048 | -0.284 | 0.080 | -1.201 | 0.104 |
| 2011 | 0.178 | 0.106 | 0.013 | 0.071 | -0.289 | 0.048 | -1.211 | 0.081 | -1.685 | 0.105 |
| 2012 | 0.031 | 0.147 | -0.352 | 0.089 | -0.399 | 0.050 | -1.753 | 0.084 | -2.225 | 0.106 |
| 2013 | -0.478 | 0.193 | -0.541 | 0.090 | -0.225 | 0.052 | 0.187 | 0.080 | -1.547 | 0.106 |
| 2014 | -0.130 | 0.367 | -1.774 | 0.185 | 0.014 | 0.055 | -0.144 | 0.081 | -2.156 | 0.108 |
| 2015 | 0.069 | 0.186 | -1.037 | 0.119 | -0.025 | 0.059 | 0.915 | 0.083 | -1.848 | 0.109 |
| 2016 | 0.236 | 0.179 | -0.964 | 0.124 | -0.110 | 0.063 | 0.160 | 0.085 | -1.442 | 0.110 |
| 2017 | -0.317 | 0.400 | -1.540 | 0.194 | | | | | | |

Table 5(2b) (continued). Summary of estimated model parameter values and standard deviations and limits for scenario 2b for Bristol Bay red king crab. For initial year length composition deviations, the first 20 length groups are for males and the last 16 length groups are for females.

| Parameter | Value | SD | Limits | Initial Length Composition 1975 | | | |
|--------------------------|--------|-------|-------------|---------------------------------|--------|-------|------------|
| | | | | Length | Value | SD | Limits |
| Mm80-84 | 0.429 | 0.016 | 0.184, 1.0 | 68 | 1.158 | 0.103 | -5, 5 |
| Mf80-84 | 0.797 | 0.021 | 0.276, 1.5 | 73 | 1.188 | 0.090 | -5, 5 |
| Mf76-79,85-93 | 0.097 | 0.006 | 0.0, 0.108 | 78 | 0.528 | 0.108 | -5, 5 |
| log_betaf, females | 0.324 | 0.056 | -0.67, 1.32 | 83 | 0.610 | 0.090 | -5, 5 |
| log_betaf, males | 0.631 | 0.081 | -0.67, 1.32 | 88 | 0.429 | 0.090 | -5, 5 |
| log_betar, females | -0.616 | 0.060 | -1.14, 0.5 | 93 | 0.243 | 0.095 | -5, 5 |
| log_betar, males | -0.604 | 0.051 | -1.14, 0.5 | 98 | 0.254 | 0.094 | -5, 5 |
| Bsfrf_CV | 0.000 | 0.000 | 0.00, 0.40 | 103 | 0.044 | 0.105 | -5, 5 |
| moltp_slope, 75-78 | 0.135 | 0.018 | 0.01, 0.259 | 108 | 0.123 | 0.104 | -5, 5 |
| moltp_slope, 79-17 | 0.099 | 0.004 | 0.01, 0.259 | 113 | 0.255 | 0.101 | -5, 5 |
| log_moltp_L50, 75-78 | 4.974 | 0.011 | 4.445, 5.52 | 118 | 0.056 | 0.119 | -5, 5 |
| log_moltp_L50, 79-17 | 4.949 | 0.004 | 4.445, 5.52 | 123 | 0.100 | 0.123 | -5, 5 |
| log_N75 | 19.953 | 0.031 | 15.0, 22.0 | 128 | 0.019 | 0.138 | -5, 5 |
| log_avg_L50_ret | 4.922 | 0.002 | 4.467, 5.51 | 133 | 0.002 | 0.148 | -5, 5 |
| ret_fish_slope | 0.525 | 0.030 | 0.05, 0.70 | 138 | -0.087 | 0.138 | -5, 5 |
| pot disc.males, ϕ | -0.325 | 0.014 | -0.40, 0.00 | 143 | -0.207 | 0.142 | -5, 5 |
| pot disc.males, κ | 0.004 | 0.000 | 0.0, 0.005 | 148 | -0.395 | 0.154 | -5, 5 |
| pot disc.males, γ | -0.015 | 0.001 | -0.025, 0.0 | 153 | -0.737 | 0.188 | -5, 5 |
| pot disc.fema., slope | 0.174 | 0.060 | 0.05, 0.43 | 158 | -1.277 | 0.262 | -5, 5 |
| log_pot disc.fema., L50 | 4.446 | 0.029 | 4.20, 4.666 | 163 | -1.277 | 0.271 | -5, 5 |
| trawl disc slope | 0.058 | 0.003 | 0.01, 0.20 | 68 | 1.620 | 0.105 | -5, 5 |
| log_trawl disc L50 | 5.113 | 0.047 | 4.50, 5.40 | 73 | 1.517 | 0.102 | -5, 5 |
| log_srv_L50, m, bsfrf | 4.309 | 0.037 | 3.59, 5.48 | 78 | 1.478 | 0.094 | -5, 5 |
| srv_slope, f, bsfrf | 0.039 | 0.007 | 0.01, 0.435 | 83 | 1.312 | 0.093 | -5, 5 |
| log_srv_L50, f, bsfrf | 4.403 | 0.063 | 4.09, 5.54 | 88 | 1.262 | 0.086 | -5, 5 |
| log_srv_L50, m, 75-81 | 4.344 | 0.010 | 4.09, 4.554 | 93 | 0.807 | 0.103 | -5, 5 |
| srv_slope, f, 75-81 | 0.072 | 0.004 | 0.01, 0.303 | 98 | 0.441 | 0.126 | -5, 5 |
| log_srv_L50, f, 75-81 | 4.468 | 0.017 | 4.09, 4.70 | 103 | 0.149 | 0.150 | -5, 5 |
| log_srv_L50, m, 82-17 | 4.403 | 0.084 | 4.09, 5.10 | 108 | -0.007 | 0.157 | -5, 5 |
| srv_slope, f, 82-17 | 0.057 | 0.008 | 0.01, 0.30 | 113 | -0.240 | 0.183 | -5, 5 |
| log_srv_L50, f, 82-17 | 4.302 | 0.051 | 4.09, 4.90 | 118 | -0.839 | 0.290 | -5, 5 |
| TC_slope, females | 0.376 | 0.131 | 0.02, 0.40 | 123 | -0.950 | 0.329 | -5, 5 |
| log_TC_L50, females | 4.534 | 0.014 | 4.24, 4.90 | 128 | -1.261 | 0.440 | -5, 5 |
| TC_slope, males | 0.250 | 0.103 | 0.05, 0.90 | 133 | -2.264 | 1.042 | -5, 5 |
| log_TC_L50, males | 4.570 | 0.019 | 4.25, 5.14 | 138 | -2.373 | 1.252 | -5, 5 |
| Q | 0.965 | 0.021 | 0.59, 1.2 | 143 | NA | NA | |
| log_TC_F, males, 91 | -4.113 | 0.085 | -10.0, 1.00 | Fixed gear bycatch parameters: | | | |
| log_TC_F, males, 92 | -6.086 | 0.086 | -10.0, 1.00 | log_avg | -8.133 | 0.080 | -8.5, -0.5 |
| log_TC_F, males, 93 | -6.804 | 0.088 | -10.0, 1.00 | fmortf_ | -1.355 | 0.111 | -10, 10 |
| log_TC_F, males, 13 | -8.308 | 0.091 | -10.0, 1.00 | fmortf_ | -2.232 | 0.130 | -10, 10 |
| log_TC_F, males, 14 | -7.442 | 0.090 | -10.0, 1.00 | fmortf_ | -0.706 | 0.103 | -10, 10 |
| log_TC_F, males, 15 | -7.024 | 0.091 | -10.0, 1.00 | fmortf_ | -0.166 | 0.100 | -10, 10 |
| log_TC_F, females, 91 | -2.873 | 0.086 | -10.0, 1.00 | fmortf_ | 0.960 | 0.097 | -10, 10 |
| log_TC_F, females, 92 | -4.515 | 0.086 | -10.0, 1.00 | fmortf_ | 1.771 | 0.096 | -10, 10 |
| log_TC_F, females, 93 | -6.395 | 0.087 | -10.0, 1.00 | fmortf_ | 1.408 | 0.097 | -10, 10 |
| log_TC_F, females, 13 | -7.726 | 0.084 | -10.0, 1.00 | fmortf_ | 0.320 | 0.100 | -10, 10 |
| log_TC_F, females, 14 | -7.583 | 0.084 | -10.0, 1.00 | Fix_slo | 0.092 | 0.025 | 0.01, 0.2 |
| log_TC_F, females, 15 | -6.553 | 0.082 | -10.0, 1.00 | log_150 | 4.633 | 0.040 | 4.5, 5.4 |

Table 5(2bn1). Summary of estimated model parameter values and standard deviations and limits for scenario 2bn1 for Bristol Bay red king crab. All values are on a log scale. Male recruit in year t is $\exp(\text{mean}+\text{males}_t)$, and female recruit in year t is $\exp(\text{mean}+\text{males}_t+\text{females}_t)$.

| Year | Recruits | | | | F for Directed Pot Fishery | | | | F for Trawl | |
|---------|----------|--------|---------|--------|----------------------------|--------|----------|--------|-------------|--------|
| | Females | SD | Males | SD | Males | SD | Females | SD | Estimate | SD |
| Mean | 15.758 | 0.023 | 15.758 | 0.023 | -1.516 | 0.026 | 0.012 | 0.001 | -4.433 | 0.068 |
| Limits↑ | 13,18 | | 13,18 | | -3.0,0.0 | | .001,0.1 | | -8.5,-0.5 | |
| Limits↓ | -15,15 | | -15,15 | | -15,2.43 | | -6.0,3.5 | | -10,10 | |
| 1975 | | | | | 0.7544 | 0.0966 | | | | |
| 1976 | -0.0547 | 0.2810 | 0.8458 | 0.1355 | 0.7204 | 0.0652 | | | 0.1462 | 0.1123 |
| 1977 | 0.5198 | 0.1624 | 0.7030 | 0.1030 | 0.7018 | 0.0542 | | | 0.6436 | 0.1078 |
| 1978 | 0.4544 | 0.1376 | 0.8976 | 0.0854 | 0.9065 | 0.0495 | | | 0.7132 | 0.1062 |
| 1979 | 0.7353 | 0.1023 | 1.1532 | 0.0769 | 1.2018 | 0.0463 | | | 0.8866 | 0.1061 |
| 1980 | 0.2308 | 0.1163 | 1.3332 | 0.0764 | 2.0790 | 0.0435 | | | 1.6433 | 0.1058 |
| 1981 | 0.0840 | 0.1500 | 0.5221 | 0.1025 | 2.4250 | 0.0091 | | | 1.2521 | 0.1065 |
| 1982 | 0.1108 | 0.0552 | 2.1077 | 0.0508 | 0.6004 | 0.0509 | | | 2.4106 | 0.1066 |
| 1983 | 0.0318 | 0.0735 | 1.4515 | 0.0515 | -9.9989 | 0.0338 | | | 2.0339 | 0.1043 |
| 1984 | 0.4910 | 0.0605 | 1.4933 | 0.0487 | 0.6842 | 0.0540 | | | 3.0865 | 0.1042 |
| 1985 | 0.1011 | 0.2028 | -0.5769 | 0.1217 | 0.7834 | 0.0615 | | | 1.9434 | 0.1060 |
| 1986 | 0.5936 | 0.0610 | 0.7647 | 0.0463 | 1.3193 | 0.0572 | | | 1.0195 | 0.1074 |
| 1987 | -0.0447 | 0.1451 | -0.1162 | 0.0736 | 0.9234 | 0.0524 | | | 0.5991 | 0.1057 |
| 1988 | 0.2605 | 0.1755 | -0.7916 | 0.1055 | -0.0639 | 0.0440 | | | 1.3584 | 0.1022 |
| 1989 | 0.1189 | 0.1534 | -0.6523 | 0.0884 | 0.0123 | 0.0406 | | | -0.0977 | 0.1023 |
| 1990 | -0.0149 | 0.0712 | 0.4583 | 0.0459 | 0.6220 | 0.0361 | 2.0642 | 0.0816 | 0.2822 | 0.1025 |
| 1991 | -0.0435 | 0.0967 | 0.0032 | 0.0556 | 0.6284 | 0.0383 | -0.5499 | 0.0813 | 0.6341 | 0.1040 |
| 1992 | -0.5983 | 0.4371 | -1.7560 | 0.1710 | 0.1140 | 0.0404 | 2.1739 | 0.0809 | 0.6851 | 0.1034 |
| 1993 | -0.2217 | 0.1018 | -0.2665 | 0.0572 | 0.7745 | 0.0428 | 2.0638 | 0.0818 | 1.1341 | 0.1045 |
| 1994 | -0.4706 | 0.4910 | -2.1384 | 0.2006 | -4.3679 | 0.0428 | 1.4321 | 0.1138 | -0.6542 | 0.1039 |
| 1995 | 0.0077 | 0.0408 | 1.3689 | 0.0354 | -4.7420 | 0.0389 | 1.5546 | 0.1185 | -0.0823 | 0.1025 |
| 1996 | -0.8668 | 0.2786 | -0.4108 | 0.1085 | -0.2085 | 0.0366 | -3.5487 | 0.1366 | -0.7016 | 0.1030 |
| 1997 | -1.0020 | 0.4342 | -1.3063 | 0.1659 | -0.1189 | 0.0369 | -0.2453 | 0.0814 | -1.0640 | 0.1033 |
| 1998 | -0.3340 | 0.1304 | -0.0860 | 0.0691 | 0.7045 | 0.0379 | 1.6136 | 0.0792 | -0.1592 | 0.1019 |
| 1999 | 0.0602 | 0.0632 | 0.7551 | 0.0435 | 0.2027 | 0.0366 | -2.5494 | 0.0884 | 0.0026 | 0.1017 |
| 2000 | -0.1880 | 0.1479 | -0.1147 | 0.0760 | -0.2002 | 0.0359 | 1.1630 | 0.0781 | -0.8349 | 0.1018 |
| 2001 | 0.5475 | 0.1884 | -0.7635 | 0.1316 | -0.2389 | 0.0355 | 0.8222 | 0.0778 | -0.4232 | 0.1014 |
| 2002 | 0.2334 | 0.0567 | 1.1936 | 0.0405 | -0.1118 | 0.0351 | -1.9633 | 0.0841 | -0.5176 | 0.1012 |
| 2003 | -0.0530 | 0.2838 | -0.6605 | 0.1560 | 0.3862 | 0.0345 | 1.0971 | 0.0781 | -0.3709 | 0.1009 |
| 2004 | -0.1876 | 0.1642 | 0.1471 | 0.0849 | 0.3273 | 0.0349 | 0.3370 | 0.0785 | -0.7269 | 0.1013 |
| 2005 | 0.3446 | 0.0648 | 1.0836 | 0.0468 | 0.6648 | 0.0352 | 0.7768 | 0.0783 | -0.4163 | 0.1010 |
| 2006 | -0.6968 | 0.1775 | 0.4739 | 0.0655 | 0.3655 | 0.0353 | -1.3923 | 0.0797 | -0.7552 | 0.1014 |
| 2007 | -0.2151 | 0.1661 | -0.1436 | 0.0877 | 0.7383 | 0.0358 | -0.3672 | 0.0787 | -0.5673 | 0.1017 |
| 2008 | 0.2410 | 0.1630 | -0.6027 | 0.1068 | 0.8675 | 0.0382 | -0.6283 | 0.0791 | -0.3932 | 0.1024 |
| 2009 | 0.3892 | 0.1432 | -0.5727 | 0.1005 | 0.6153 | 0.0395 | -0.8228 | 0.0795 | -0.9664 | 0.1034 |
| 2010 | 0.0387 | 0.0995 | 0.1106 | 0.0623 | 0.4255 | 0.0409 | -0.3551 | 0.0798 | -1.1617 | 0.1041 |
| 2011 | 0.1097 | 0.1012 | 0.1060 | 0.0656 | -0.2259 | 0.0421 | -1.2841 | 0.0820 | -1.6513 | 0.1048 |
| 2012 | -0.0456 | 0.1397 | -0.2326 | 0.0812 | -0.3555 | 0.0434 | -1.8627 | 0.0847 | -2.1948 | 0.1064 |
| 2013 | -0.5319 | 0.1915 | -0.4676 | 0.0862 | -0.1853 | 0.0460 | 0.1382 | 0.0810 | -1.5145 | 0.1061 |
| 2014 | -0.1217 | 0.3816 | -1.8099 | 0.1884 | 0.1521 | 0.0502 | -0.5054 | 0.0826 | -2.1109 | 0.1086 |
| 2015 | 0.0853 | 0.1840 | -1.0271 | 0.1175 | 0.1190 | 0.0559 | 0.6781 | 0.0842 | -1.7991 | 0.1099 |
| 2016 | 0.2305 | 0.1783 | -0.9488 | 0.1226 | -0.0019 | 0.0621 | 0.1600 | 0.0872 | -1.3116 | 0.1114 |
| 2017 | -0.3286 | 0.4055 | -1.5277 | 0.1925 | | | | | | |

Table 5(2bn1) (continued). Summary of estimated model parameter values and standard deviations and limits for scenario 2bn1 for Bristol Bay red king crab. For initial year length composition deviations, the first 20 length groups are for males and the last 16 length groups are for females.

| Parameter | Value | SD | Limits | Initial Length Composition 1975 | | | |
|---------------------------|--------|--------|-------------|---------------------------------|---------|--------|--------|
| | | | | Length | Value | SD | Limits |
| Mm80-84 | 0.4322 | 0.0160 | 0.184, 1.0 | 68 | 1.1581 | 0.1025 | -5, 5 |
| Mf80-84 | 0.7925 | 0.0209 | 0.276, 1.5 | 73 | 1.1860 | 0.0894 | -5, 5 |
| Mf76-79,85-93 | 0.0970 | 0.0066 | 0.0, 0.108 | 78 | 0.5276 | 0.1076 | -5, 5 |
| log_betal, females | 0.3412 | 0.0563 | -0.67, 1.32 | 83 | 0.6109 | 0.0896 | -5, 5 |
| log_betal, males | 0.6177 | 0.0814 | -0.67, 1.32 | 88 | 0.4311 | 0.0895 | -5, 5 |
| log_betar, females | -0.594 | 0.0603 | -1.14, 0.5 | 93 | 0.2471 | 0.0945 | -5, 5 |
| log_betar, males | -0.596 | 0.0513 | -1.14, 0.5 | 98 | 0.2591 | 0.0935 | -5, 5 |
| Bsfrf_CV | 0.0000 | 0.0002 | 0.00, 0.40 | 103 | 0.0514 | 0.1053 | -5, 5 |
| moltp_slope, 75-78 | 0.1342 | 0.0173 | 0.01, 0.259 | 108 | 0.1313 | 0.1037 | -5, 5 |
| moltp_slope, 79-14 | 0.1000 | 0.0042 | 0.01, 0.259 | 113 | 0.2637 | 0.1013 | -5, 5 |
| log_moltp_L50, 75-78 | 4.9756 | 0.0108 | 4.445, 5.52 | 118 | 0.0660 | 0.1187 | -5, 5 |
| log_moltp_L50, 79-14 | 4.9506 | 0.0038 | 4.445, 5.52 | 123 | 0.1105 | 0.1228 | -5, 5 |
| log_N75 | 19.946 | 0.0312 | 15.0, 22.0 | 128 | 0.0305 | 0.1374 | -5, 5 |
| log_avg_L50_total | 4.7800 | 0.0000 | 4.467, 5.51 | 133 | 0.0109 | 0.1470 | -5, 5 |
| sel_fish_slope_total | 0.0992 | 0.0042 | 0.05, 0.70 | 138 | -0.0756 | 0.1382 | -5, 5 |
| log_L50_ret, 75-04 | 4.9206 | 0.0015 | 4.6, 5.1 | 143 | -0.1989 | 0.1422 | -5, 5 |
| sel_fish_slope_ret, 75-04 | 0.4827 | 0.0327 | 0.05, 0.87 | 148 | -0.4177 | 0.1557 | -5, 5 |
| log_L50_ret, 05-16 | 4.9299 | 0.0019 | 4.6, 5.1 | 153 | -0.7841 | 0.1919 | -5, 5 |
| sel_fish_slope_ret, 05-16 | 0.4409 | 0.0487 | 0.05, 0.7 | 158 | -1.3463 | 0.2687 | -5, 5 |
| pot disc.fema., slope | 0.0925 | 0.0166 | 0.05, 0.43 | 163 | -1.3607 | 0.2767 | -5, 5 |
| log_pot disc.fema., L50 | 4.5373 | 0.0359 | 4.20, 4.666 | 68 | 1.6227 | 0.1052 | -5, 5 |
| trawl disc slope | 0.0564 | 0.0025 | 0.01, 0.20 | 73 | 1.5220 | 0.1015 | -5, 5 |
| log_trawl disc L50 | 5.1888 | 0.0704 | 4.50, 5.40 | 78 | 1.4833 | 0.0943 | -5, 5 |
| fixed_disc_slope | 0.0806 | 0.0233 | 3.59, 5.48 | 83 | 1.3181 | 0.0932 | -5, 5 |
| log_fixed_disc L50 | 4.6638 | 0.0550 | 0.01, 0.435 | 88 | 1.2676 | 0.0867 | -5, 5 |
| log_srv_L50, m, bsfrf | 4.2951 | 0.0325 | 3.59, 5.48 | 93 | 0.8144 | 0.1030 | -5, 5 |
| srv_slope, f, bsfrf | 0.0330 | 0.0062 | 0.01, 0.435 | 98 | 0.4506 | 0.1257 | -5, 5 |
| log_srv_L50, f, bsfrf | 4.4113 | 0.0698 | 4.09, 5.54 | 103 | 0.1609 | 0.1505 | -5, 5 |
| log_srv_L50, m, 75-81 | 4.3414 | 0.0101 | 4.09, 4.554 | 108 | 0.0049 | 0.1572 | -5, 5 |
| srv_slope, f, 75-81 | 0.0709 | 0.0040 | 0.01, 0.303 | 113 | -0.2243 | 0.1824 | -5, 5 |
| log_srv_L50, f, 75-81 | 4.4674 | 0.0168 | 4.09, 4.70 | 118 | -0.8258 | 0.2889 | -5, 5 |
| log_srv_L50, m, 82-14 | 4.4805 | 0.0677 | 4.09, 5.10 | 123 | -0.9339 | 0.3276 | -5, 5 |
| srv_slope, f, 82-14 | 0.0593 | 0.0064 | 0.01, 0.30 | 128 | -1.2482 | 0.4390 | -5, 5 |
| log_srv_L50, f, 82-14 | 4.3352 | 0.0388 | 4.09, 4.90 | 133 | -2.2562 | 1.0427 | -5, 5 |
| TC_slope, females | 0.3757 | 0.1299 | 0.02, 0.40 | 138 | -2.3699 | 1.2595 | -5, 5 |
| log_TC_L50, females | 4.5344 | 0.0136 | 4.24, 4.90 | 143 | NA | NA | |
| TC_slope, males | 0.2261 | 0.0863 | 0.05, 0.90 | | | | |
| log_TC_L50, males | 4.5780 | 0.0202 | 4.25, 5.14 | | | | |
| Q | 0.9997 | 0.0203 | 0.592, 1.20 | | | | |

Table 5(2bn1) (continued). Summary of estimated model parameter values and standard deviations and limits for scenario 2bn1 for Bristol Bay red king crab. For initial year length composition deviations, the first 20 length groups are for males and the last 16 length groups are for females.

| Tanner crab fishery F | | | | Fixed gear fishery F | | | |
|-----------------------|--------|-------|--------|----------------------|--------|-------|------------|
| Parameter | Value | SD | Limits | Parameter | Value | SD | Limits |
| log_TC_F, males, 91 | -4.080 | 0.082 | -10, 1 | log avg fmortf | -7.987 | 0.104 | -8.5, -0.5 |
| log_TC_F, males, 92 | -6.052 | 0.084 | -10, 1 | Fmortf, 09 | -1.359 | 0.112 | -10, 10 |
| log_TC_F, males, 93 | -6.766 | 0.085 | -10, 1 | Fmortf, 10 | -2.237 | 0.130 | -10, 10 |
| log_TC_F, males, 13 | -8.280 | 0.090 | -10, 1 | Fmortf, 11 | -0.712 | 0.103 | -10, 10 |
| log_TC_F, males, 14 | -7.414 | 0.088 | -10, 1 | Fmortf, 12 | -0.174 | 0.100 | -10, 10 |
| log_TC_F, males, 15 | -6.996 | 0.090 | -10, 1 | Fmortf, 13 | 0.948 | 0.097 | -10, 10 |
| log_TC_F, females, 91 | -2.848 | 0.085 | -10, 1 | Fmortf, 14 | 1.758 | 0.096 | -10, 10 |
| log_TC_F, females, 92 | -4.490 | 0.085 | -10, 1 | fmortf, 15 | 1.391 | 0.097 | -10, 10 |
| log_TC_F, females, 93 | -6.369 | 0.086 | -10, 1 | Fmortf, 16 | 0.386 | 0.100 | -10, 10 |
| log_TC_F, females, 13 | -7.708 | 0.083 | -10, 1 | | | | |
| log_TC_F, females, 14 | -7.566 | 0.083 | -10, 1 | | | | |
| log_TC_F, females, 15 | -7.566 | 0.083 | -10, 1 | | | | |

Table 5(2bn2). Summary of estimated model parameter values and standard deviations and limits for scenario 2bn2 for Bristol Bay red king crab. All values are on a log scale. Male recruit in year t is $\exp(\text{mean}+\text{males}_t)$, and female recruit in year t is $\exp(\text{mean}+\text{males}_t+\text{females}_t)$.

| Year | Recruits | | | | F for Directed Pot Fishery | | | | F for Trawl | |
|---------|----------|-------|--------|-------|----------------------------|-------|----------|-------|-------------|-------|
| | Females | SD | Males | SD | Males | SD | Females | SD | Estimate | SD |
| Mean | 15.793 | 0.023 | 15.758 | 0.023 | -1.555 | 0.026 | 0.013 | 0.001 | -4.517 | 0.066 |
| Limits↑ | 13,18 | | 13,18 | | -3,0,0 | | .001,0,1 | | -8.5,-0.5 | |
| Limits↓ | -15,15 | | -15,15 | | -15,2,43 | | -6,0,3,5 | | -10,10 | |
| 1975 | | | | | 0.773 | 0.096 | | | | |
| 1976 | -0.044 | 0.283 | 0.808 | 0.137 | 0.744 | 0.065 | | | 0.170 | 0.112 |
| 1977 | 0.525 | 0.162 | 0.680 | 0.103 | 0.728 | 0.054 | | | 0.672 | 0.108 |
| 1978 | 0.454 | 0.137 | 0.879 | 0.085 | 0.934 | 0.049 | | | 0.741 | 0.106 |
| 1979 | 0.730 | 0.102 | 1.142 | 0.077 | 1.226 | 0.046 | | | 0.909 | 0.106 |
| 1980 | 0.227 | 0.116 | 1.326 | 0.077 | 2.102 | 0.043 | | | 1.656 | 0.106 |
| 1981 | 0.077 | 0.149 | 0.514 | 0.102 | 2.425 | 0.009 | | | 1.252 | 0.106 |
| 1982 | 0.101 | 0.055 | 2.093 | 0.051 | 0.597 | 0.051 | | | 2.422 | 0.106 |
| 1983 | 0.026 | 0.073 | 1.423 | 0.051 | -9.999 | 0.035 | | | 2.063 | 0.104 |
| 1984 | 0.485 | 0.060 | 1.452 | 0.049 | 0.725 | 0.054 | | | 3.132 | 0.104 |
| 1985 | 0.095 | 0.201 | -0.615 | 0.120 | 0.841 | 0.062 | | | 1.995 | 0.106 |
| 1986 | 0.588 | 0.061 | 0.719 | 0.046 | 1.373 | 0.058 | | | 1.069 | 0.107 |
| 1987 | -0.048 | 0.145 | -0.161 | 0.073 | 0.973 | 0.053 | | | 0.647 | 0.106 |
| 1988 | 0.251 | 0.174 | -0.827 | 0.104 | -0.017 | 0.044 | | | 1.408 | 0.102 |
| 1989 | 0.114 | 0.152 | -0.686 | 0.087 | 0.061 | 0.041 | | | -0.044 | 0.102 |
| 1990 | -0.017 | 0.071 | 0.413 | 0.046 | 0.681 | 0.036 | 2.034 | 0.082 | 0.343 | 0.103 |
| 1991 | -0.042 | 0.096 | -0.039 | 0.056 | 0.694 | 0.038 | -0.590 | 0.081 | 0.703 | 0.104 |
| 1992 | -0.621 | 0.439 | -1.775 | 0.168 | 0.182 | 0.041 | 2.129 | 0.081 | 0.759 | 0.103 |
| 1993 | -0.219 | 0.101 | -0.303 | 0.057 | 0.847 | 0.043 | 2.012 | 0.082 | 1.212 | 0.105 |
| 1994 | -0.486 | 0.496 | -2.166 | 0.198 | -4.303 | 0.043 | 1.385 | 0.114 | -0.584 | 0.104 |
| 1995 | 0.016 | 0.041 | 1.330 | 0.036 | -4.687 | 0.039 | 1.514 | 0.119 | -0.017 | 0.102 |
| 1996 | -0.819 | 0.271 | -0.441 | 0.108 | -0.152 | 0.037 | -3.588 | 0.137 | -0.637 | 0.103 |
| 1997 | -1.014 | 0.435 | -1.310 | 0.162 | -0.064 | 0.037 | -0.283 | 0.081 | -1.001 | 0.103 |
| 1998 | -0.326 | 0.129 | -0.105 | 0.069 | 0.761 | 0.038 | 1.568 | 0.079 | -0.102 | 0.102 |
| 1999 | 0.068 | 0.063 | 0.735 | 0.044 | 0.250 | 0.037 | -2.589 | 0.088 | 0.049 | 0.102 |
| 2000 | -0.167 | 0.145 | -0.130 | 0.075 | -0.157 | 0.036 | 1.124 | 0.078 | -0.792 | 0.102 |
| 2001 | 0.554 | 0.185 | -0.763 | 0.130 | -0.198 | 0.036 | 0.783 | 0.078 | -0.381 | 0.101 |
| 2002 | 0.229 | 0.056 | 1.214 | 0.041 | -0.074 | 0.036 | -2.005 | 0.084 | -0.478 | 0.101 |
| 2003 | -0.063 | 0.271 | -0.590 | 0.149 | 0.419 | 0.035 | 1.055 | 0.078 | -0.340 | 0.101 |
| 2004 | -0.218 | 0.159 | 0.224 | 0.081 | 0.350 | 0.036 | 0.300 | 0.079 | -0.706 | 0.101 |
| 2005 | 0.300 | 0.064 | 1.166 | 0.046 | 0.671 | 0.039 | 0.747 | 0.080 | -0.414 | 0.101 |
| 2006 | -0.731 | 0.175 | 0.560 | 0.064 | 0.326 | 0.036 | -1.381 | 0.080 | -0.772 | 0.101 |
| 2007 | -0.206 | 0.168 | -0.115 | 0.091 | 0.672 | 0.037 | -0.335 | 0.079 | -0.610 | 0.102 |
| 2008 | 0.248 | 0.166 | -0.569 | 0.110 | 0.750 | 0.041 | -0.551 | 0.080 | -0.472 | 0.103 |
| 2009 | 0.331 | 0.142 | -0.469 | 0.098 | 0.467 | 0.044 | -0.721 | 0.081 | -1.076 | 0.104 |
| 2010 | 0.060 | 0.102 | 0.132 | 0.065 | 0.262 | 0.043 | -0.243 | 0.081 | -1.297 | 0.104 |
| 2011 | 0.197 | 0.106 | 0.058 | 0.072 | -0.310 | 0.046 | -1.256 | 0.083 | -1.794 | 0.105 |
| 2012 | 0.009 | 0.143 | -0.241 | 0.086 | -0.523 | 0.044 | -1.754 | 0.085 | -2.336 | 0.107 |
| 2013 | -0.493 | 0.195 | -0.462 | 0.090 | -0.353 | 0.046 | 0.245 | 0.081 | -1.658 | 0.106 |
| 2014 | -0.132 | 0.387 | -1.756 | 0.189 | 0.138 | 0.054 | -0.554 | 0.085 | -2.267 | 0.109 |
| 2015 | 0.084 | 0.186 | -0.976 | 0.118 | -0.011 | 0.055 | 0.742 | 0.084 | -1.958 | 0.110 |
| 2016 | 0.228 | 0.180 | -0.893 | 0.124 | -0.124 | 0.059 | 0.212 | 0.086 | -1.467 | 0.111 |
| 2017 | -0.350 | 0.415 | -1.473 | 0.194 | | | | | | |

Table 5(2bn2) (continued). Summary of estimated model parameter values and standard deviations and limits for scenario 2bn2 for Bristol Bay red king crab. For initial year length composition deviations, the first 20 length groups are for males and the last 16 length groups are for females.

| Parameter | Value | SD | Limits | Initial Length Composition 1975 | | | |
|-------------------------|--------|-------|-------------|---------------------------------|--------|-------|--------|
| | | | | Length | Value | SD | Limits |
| Mm80-84 | 0.446 | 0.016 | 0.184, 1.0 | 68 | 1.161 | 0.103 | -5, 5 |
| Mf80-84 | 0.801 | 0.021 | 0.276, 1.5 | 73 | 1.193 | 0.089 | -5, 5 |
| Mf76-79,85-93 | 0.096 | 0.007 | 0.0, 0.108 | 78 | 0.532 | 0.108 | -5, 5 |
| log_betal, females | 0.330 | 0.056 | -0.67, 1.32 | 83 | 0.612 | 0.090 | -5, 5 |
| log_betal, males | 0.642 | 0.081 | -0.67, 1.32 | 88 | 0.430 | 0.090 | -5, 5 |
| log_betar, females | -0.600 | 0.060 | -1.14, 0.5 | 93 | 0.244 | 0.095 | -5, 5 |
| log_betar, males | -0.614 | 0.051 | -1.14, 0.5 | 98 | 0.255 | 0.093 | -5, 5 |
| Bsfrf_CV | 0.000 | 0.000 | 0.00, 0.40 | 103 | 0.046 | 0.105 | -5, 5 |
| moltp_slope, 75-78 | 0.136 | 0.018 | 0.01, 0.259 | 108 | 0.125 | 0.104 | -5, 5 |
| moltp_slope, 79-14 | 0.105 | 0.004 | 0.01, 0.259 | 113 | 0.257 | 0.101 | -5, 5 |
| log_moltp_L50, 75-78 | 4.976 | 0.011 | 4.445, 5.52 | 118 | 0.059 | 0.119 | -5, 5 |
| log_moltp_L50, 79-14 | 4.955 | 0.004 | 4.445, 5.52 | 123 | 0.102 | 0.123 | -5, 5 |
| log_N75 | 19.951 | 0.031 | 15.0, 22.0 | 128 | 0.022 | 0.138 | -5, 5 |
| log_avg_L50_total | 4.780 | 0.000 | 4.467, 5.51 | 133 | 0.003 | 0.147 | -5, 5 |
| sel_fish_slope_total | 0.099 | 0.004 | 0.05, 0.70 | 138 | -0.090 | 0.138 | -5, 5 |
| log_L50_ret | 4.919 | 0.001 | 4.6, 5.1 | 143 | -0.202 | 0.142 | -5, 5 |
| sel_fish_slope_ret | 0.521 | 0.031 | 0.05, 0.87 | 148 | -0.410 | 0.155 | -5, 5 |
| pot disc.fema., slope | 0.087 | 0.016 | 0.05, 0.43 | 153 | -0.772 | 0.190 | -5, 5 |
| log_pot disc.fema., L50 | 4.552 | 0.037 | 4.20, 4.666 | 158 | -1.332 | 0.267 | -5, 5 |
| trawl disc slope | 0.058 | 0.003 | 0.01, 0.20 | 163 | -1.345 | 0.275 | -5, 5 |
| log_trawl disc L50 | 5.136 | 0.051 | 4.50, 5.40 | 68 | 1.630 | 0.105 | -5, 5 |
| fixed_disc_slope | 0.093 | 0.024 | 3.59, 5.48 | 73 | 1.526 | 0.102 | -5, 5 |
| log_fixed_disc L50 | 4.636 | 0.039 | 0.01, 0.435 | 78 | 1.486 | 0.094 | -5, 5 |
| log_srv_L50, m, bsfrf | 4.310 | 0.039 | 3.59, 5.48 | 83 | 1.320 | 0.093 | -5, 5 |
| srv_slope, f, bsfrf | 0.037 | 0.006 | 0.01, 0.435 | 88 | 1.269 | 0.087 | -5, 5 |
| log_srv_L50, f, bsfrf | 4.409 | 0.068 | 4.09, 5.54 | 93 | 0.815 | 0.103 | -5, 5 |
| log_srv_L50, m, 75-81 | 4.344 | 0.010 | 4.09, 4.554 | 98 | 0.450 | 0.125 | -5, 5 |
| srv_slope, f, 75-81 | 0.071 | 0.004 | 0.01, 0.303 | 103 | 0.159 | 0.150 | -5, 5 |
| log_srv_L50, f, 75-81 | 4.470 | 0.017 | 4.09, 4.70 | 108 | 0.003 | 0.157 | -5, 5 |
| log_srv_L50, m, 82-14 | 4.436 | 0.075 | 4.09, 5.10 | 113 | -0.229 | 0.182 | -5, 5 |
| srv_slope, f, 82-14 | 0.055 | 0.007 | 0.01, 0.30 | 118 | -0.828 | 0.288 | -5, 5 |
| log_srv_L50, f, 82-14 | 4.323 | 0.051 | 4.09, 4.90 | 123 | -0.938 | 0.327 | -5, 5 |
| TC_slope, females | 0.377 | 0.130 | 0.02, 0.40 | 128 | -1.249 | 0.437 | -5, 5 |
| log_TC_L50, females | 4.535 | 0.014 | 4.24, 4.90 | 133 | -2.253 | 1.033 | -5, 5 |
| TC_slope, males | 0.239 | 0.094 | 0.05, 0.90 | 138 | -2.368 | 1.248 | -5, 5 |
| log_TC_L50, males | 4.576 | 0.019 | 4.25, 5.14 | 143 | NA | NA | |
| Q | 0.984 | 0.020 | 0.592, 1.20 | | | | |

Table 5(2bn2) (continued). Summary of estimated model parameter values and standard deviations and limits for scenario 2bn2 for Bristol Bay red king crab.

| Fixed gear fishery F | | | | | | | |
|---------------------------|--------|--------|--------|----------------|--------|-------|------------|
| Parameter | Value | SD | Limits | Parameter | Value | SD | Limits |
| log_TC_F, males, 91 | -3.987 | 0.084 | -10, 1 | log avg fmortf | -8.150 | 0.079 | -8.5, -0.5 |
| log_TC_F, males, 92 | -5.959 | 0.085 | -10, 1 | Fmortf, 09 | -1.353 | 0.111 | -10, 10 |
| log_TC_F, males, 93 | -6.665 | 0.088 | -10, 1 | Fmortf, 10 | -2.237 | 0.130 | -10, 10 |
| log_TC_F, males, 13 | -8.344 | 0.092 | -10, 1 | Fmortf, 11 | -0.714 | 0.103 | -10, 10 |
| log_TC_F, males, 14 | -7.485 | 0.090 | -10, 1 | Fmortf, 12 | -0.174 | 0.100 | -10, 10 |
| log_TC_F, males, 15 | -7.070 | 0.092 | -10, 1 | Fmortf, 13 | 0.948 | 0.097 | -10, 10 |
| log_TC_F, females, 91 | -2.819 | 0.085 | -10, 1 | Fmortf, 14 | 1.755 | 0.096 | -10, 10 |
| log_TC_F, females, 92 | -4.463 | 0.086 | -10, 1 | fmortf, 15 | 1.389 | 0.097 | -10, 10 |
| log_TC_F, females, 93 | -6.344 | 0.087 | -10, 1 | Fmortf, 16 | 0.386 | 0.100 | -10, 10 |
| log_TC_F, females, 13 | -7.749 | 0.084 | -10, 1 | | | | |
| log_TC_F, females, 14 | -7.611 | 0.084 | -10, 1 | | | | |
| log_TC_F, females, 15 | -6.584 | 0.083 | -10, 1 | | | | |
| Mul_para_ret_proport., 05 | 3.607 | 1.334 | 0, 15 | | | | |
| Mul_para_ret_proport., 06 | 14.999 | 2.676 | 0, 15 | | | | |
| Mul_para_ret_proport., 07 | 14.993 | 27.677 | 0, 15 | | | | |
| Mul_para_ret_proport., 08 | 14.995 | 20.057 | 0, 15 | | | | |
| Mul_para_ret_proport., 09 | 3.240 | 0.949 | 0, 15 | | | | |
| Mul_para_ret_proport., 10 | 2.735 | 0.473 | 0, 15 | | | | |
| Mul_para_ret_proport., 11 | 1.357 | 0.144 | 0, 15 | | | | |
| Mul_para_ret_proport., 12 | 3.362 | 0.538 | 0, 15 | | | | |
| Mul_para_ret_proport., 13 | 2.959 | 0.420 | 0, 15 | | | | |
| Mul_para_ret_proport., 14 | 0.861 | 0.120 | 0, 15 | | | | |
| Mul_para_ret_proport., 15 | 1.894 | 0.199 | 0, 15 | | | | |
| Mul_para_ret_proport., 16 | 1.803 | 0.173 | 0, 15 | | | | |

Table 6(2b). Annual abundance estimates (million crab), mature male biomass (MMB, 1000 t), and total survey biomass (1000t) for red king crab in Bristol Bay estimated by length-based analysis (scenario 2) from 1975-2017. Mature male biomass for year t is on Feb. 15, year $t+1$. Size measurements are mm carapace length.

| Year (t) | Males | | | | Females | Total Recruits | Total Survey Biomass | |
|----------|------------------|----------------|---------------|--------|-----------------|----------------|----------------------|---------------------|
| | Mature (>119 mm) | Legal (>134mm) | MMB (>119 mm) | SD MMB | Mature (>89 mm) | | Model Est. (>64 mm) | Area-Swept (>64 mm) |
| 1975 | 55.605 | 29.131 | 81.728 | 4.958 | 69.633 | | 251.709 | 202.731 |
| 1976 | 60.936 | 35.586 | 91.447 | 4.195 | 105.743 | 32.116 | 288.834 | 331.868 |
| 1977 | 62.265 | 38.209 | 94.265 | 3.524 | 130.640 | 38.466 | 297.283 | 375.661 |
| 1978 | 67.674 | 39.009 | 96.262 | 2.944 | 123.172 | 45.109 | 286.652 | 349.545 |
| 1979 | 63.250 | 40.140 | 80.639 | 2.474 | 105.965 | 69.874 | 261.837 | 167.627 |
| 1980 | 44.669 | 32.562 | 22.332 | 0.840 | 95.596 | 61.586 | 222.496 | 249.322 |
| 1981 | 13.474 | 7.719 | 6.322 | 0.345 | 44.377 | 25.482 | 90.173 | 132.669 |
| 1982 | 6.427 | 2.487 | 6.775 | 0.346 | 20.631 | 125.730 | 44.413 | 143.740 |
| 1983 | 6.056 | 2.634 | 7.834 | 0.350 | 13.545 | 62.622 | 39.407 | 49.320 |
| 1984 | 6.002 | 2.873 | 6.058 | 0.344 | 14.289 | 84.159 | 40.885 | 155.311 |
| 1985 | 7.314 | 2.417 | 10.112 | 0.494 | 13.839 | 8.604 | 34.634 | 34.535 |
| 1986 | 11.990 | 4.636 | 14.881 | 0.738 | 20.318 | 43.413 | 46.376 | 48.158 |
| 1987 | 15.429 | 6.662 | 21.341 | 0.923 | 24.100 | 12.594 | 53.122 | 70.263 |
| 1988 | 16.232 | 9.024 | 27.234 | 1.030 | 28.791 | 7.539 | 57.308 | 55.372 |
| 1989 | 17.644 | 10.848 | 30.898 | 1.086 | 26.250 | 7.806 | 60.651 | 55.941 |
| 1990 | 17.822 | 11.860 | 28.697 | 1.106 | 22.263 | 22.926 | 61.050 | 60.321 |
| 1991 | 14.411 | 10.591 | 23.700 | 1.090 | 20.050 | 14.166 | 55.492 | 85.055 |
| 1992 | 11.308 | 8.451 | 21.535 | 1.048 | 19.747 | 1.973 | 49.400 | 37.687 |
| 1993 | 11.845 | 7.624 | 18.891 | 1.025 | 17.578 | 10.254 | 47.458 | 53.703 |
| 1994 | 11.621 | 6.979 | 24.365 | 1.054 | 14.380 | 1.461 | 41.781 | 32.335 |
| 1995 | 12.099 | 8.796 | 27.144 | 1.029 | 13.930 | 56.431 | 48.278 | 38.396 |
| 1996 | 12.116 | 9.422 | 25.118 | 0.980 | 19.349 | 6.442 | 55.725 | 44.649 |
| 1997 | 11.368 | 8.477 | 23.265 | 0.939 | 28.249 | 2.659 | 59.710 | 85.277 |
| 1998 | 15.626 | 8.184 | 25.495 | 1.019 | 26.369 | 11.524 | 62.804 | 85.176 |
| 1999 | 17.239 | 9.760 | 29.978 | 1.124 | 22.992 | 31.913 | 62.690 | 65.604 |
| 2000 | 15.324 | 11.144 | 29.878 | 1.119 | 25.370 | 11.381 | 65.081 | 68.342 |
| 2001 | 14.292 | 10.661 | 28.736 | 1.078 | 29.521 | 9.038 | 67.670 | 53.188 |
| 2002 | 15.918 | 10.177 | 30.583 | 1.073 | 29.236 | 54.220 | 72.051 | 69.786 |
| 2003 | 16.660 | 10.983 | 29.260 | 1.057 | 34.780 | 8.152 | 76.951 | 116.794 |
| 2004 | 14.880 | 10.434 | 27.220 | 1.016 | 42.196 | 15.880 | 78.602 | 131.910 |
| 2005 | 17.262 | 9.913 | 27.573 | 1.032 | 40.370 | 52.982 | 83.512 | 107.341 |
| 2006 | 17.552 | 10.487 | 29.638 | 1.078 | 44.207 | 17.950 | 86.700 | 95.676 |
| 2007 | 17.028 | 11.076 | 27.006 | 1.095 | 51.198 | 12.102 | 91.822 | 104.841 |
| 2008 | 18.697 | 10.318 | 28.320 | 1.211 | 48.212 | 9.380 | 91.911 | 114.430 |
| 2009 | 19.947 | 11.156 | 32.260 | 1.376 | 43.911 | 10.472 | 89.581 | 91.673 |
| 2010 | 18.964 | 12.422 | 32.564 | 1.466 | 40.493 | 16.563 | 87.419 | 81.642 |
| 2011 | 16.380 | 12.058 | 32.719 | 1.478 | 38.642 | 16.108 | 84.017 | 67.053 |
| 2012 | 14.920 | 11.562 | 31.523 | 1.453 | 38.311 | 10.354 | 83.075 | 61.248 |
| 2013 | 14.622 | 10.799 | 30.376 | 1.454 | 37.431 | 6.835 | 81.440 | 62.410 |
| 2014 | 14.650 | 10.365 | 29.148 | 1.496 | 34.531 | 2.310 | 77.849 | 114.103 |
| 2015 | 13.830 | 9.982 | 27.725 | 1.542 | 30.416 | 5.318 | 72.318 | 64.240 |
| 2016 | 12.499 | 9.447 | 25.804 | 1.564 | 26.513 | 6.266 | 66.256 | 61.231 |
| 2017 | 10.653 | 8.633 | 21.312 | 1.202 | 24.149 | 2.686 | 60.268 | 52.922 |

Table 6(2bn1). Annual abundance estimates (million crab), mature male biomass (MMB, 1000 t), and total survey biomass (1000t) for red king crab in Bristol Bay estimated by length-based analysis (scenario 2bn1) from 1975-2017. Mature male biomass for year t is on Feb. 15, year $t+1$. Size measurements are mm carapace length.

| Year (t) | Males | | | | Females | Total Recruits | Total Survey Biomass | |
|----------|------------------|----------------|---------------|--------|-----------------|----------------|----------------------|---------------------|
| | Mature (>119 mm) | Legal (>134mm) | MMB (>119 mm) | SD MMB | Mature (>89 mm) | | Model Est. (>64 mm) | Area-Swept (>64 mm) |
| 1975 | 54.753 | 28.298 | 79.226 | 4.833 | 69.574 | | 257.743 | 202.731 |
| 1976 | 60.177 | 34.829 | 89.220 | 4.104 | 105.364 | 31.648 | 295.837 | 331.868 |
| 1977 | 61.589 | 37.573 | 92.312 | 3.449 | 129.678 | 37.797 | 304.294 | 375.661 |
| 1978 | 66.977 | 38.458 | 94.431 | 2.883 | 121.886 | 44.094 | 293.194 | 349.545 |
| 1979 | 62.599 | 39.621 | 78.978 | 2.414 | 104.482 | 68.227 | 267.334 | 167.627 |
| 1980 | 44.100 | 32.100 | 21.348 | 0.781 | 93.796 | 59.810 | 226.177 | 249.322 |
| 1981 | 13.159 | 7.476 | 5.619 | 0.285 | 43.265 | 24.553 | 90.657 | 132.669 |
| 1982 | 6.104 | 2.269 | 6.223 | 0.307 | 19.885 | 121.581 | 42.796 | 143.740 |
| 1983 | 5.748 | 2.458 | 7.352 | 0.320 | 13.159 | 60.546 | 38.050 | 49.320 |
| 1984 | 5.701 | 2.718 | 5.601 | 0.315 | 13.922 | 81.821 | 39.650 | 155.311 |
| 1985 | 6.958 | 2.261 | 9.427 | 0.448 | 13.544 | 8.254 | 33.561 | 34.535 |
| 1986 | 11.443 | 4.374 | 13.804 | 0.665 | 19.781 | 42.131 | 45.020 | 48.158 |
| 1987 | 14.698 | 6.255 | 19.853 | 0.822 | 23.415 | 12.152 | 51.461 | 70.263 |
| 1988 | 15.402 | 8.483 | 25.474 | 0.908 | 27.866 | 7.265 | 55.393 | 55.372 |
| 1989 | 16.742 | 10.228 | 28.902 | 0.946 | 25.357 | 7.728 | 58.719 | 55.941 |
| 1990 | 16.884 | 11.178 | 26.511 | 0.950 | 21.463 | 21.905 | 59.097 | 60.321 |
| 1991 | 13.531 | 9.872 | 21.405 | 0.931 | 19.352 | 13.702 | 53.365 | 85.055 |
| 1992 | 10.472 | 7.721 | 19.336 | 0.894 | 19.039 | 1.868 | 47.027 | 37.687 |
| 1993 | 10.987 | 6.941 | 16.708 | 0.868 | 16.977 | 9.628 | 45.086 | 53.703 |
| 1994 | 10.714 | 6.296 | 22.118 | 0.896 | 13.893 | 1.336 | 39.376 | 32.335 |
| 1995 | 11.217 | 8.098 | 24.917 | 0.874 | 13.428 | 55.072 | 45.971 | 38.396 |
| 1996 | 11.233 | 8.729 | 22.893 | 0.834 | 18.724 | 6.572 | 53.495 | 44.649 |
| 1997 | 10.511 | 7.784 | 21.154 | 0.802 | 27.324 | 2.583 | 57.480 | 85.277 |
| 1998 | 14.836 | 7.535 | 22.886 | 0.876 | 25.546 | 10.987 | 60.805 | 85.176 |
| 1999 | 16.306 | 8.969 | 27.423 | 0.976 | 22.485 | 30.616 | 60.384 | 65.604 |
| 2000 | 14.458 | 10.381 | 27.488 | 0.978 | 24.672 | 11.376 | 62.807 | 68.342 |
| 2001 | 13.451 | 9.949 | 26.583 | 0.944 | 28.446 | 8.874 | 65.363 | 53.188 |
| 2002 | 15.085 | 9.532 | 28.406 | 0.940 | 28.194 | 52.091 | 69.928 | 69.786 |
| 2003 | 15.892 | 10.332 | 27.335 | 0.927 | 33.583 | 7.023 | 74.715 | 116.794 |
| 2004 | 14.277 | 9.880 | 25.128 | 0.893 | 40.712 | 14.785 | 76.236 | 131.910 |
| 2005 | 16.367 | 9.288 | 25.725 | 0.905 | 38.834 | 49.727 | 80.272 | 107.341 |
| 2006 | 16.550 | 9.921 | 27.780 | 0.951 | 42.503 | 16.791 | 83.223 | 95.676 |
| 2007 | 15.930 | 10.473 | 24.722 | 0.966 | 49.047 | 10.919 | 87.991 | 104.841 |
| 2008 | 17.236 | 9.531 | 25.209 | 1.064 | 46.211 | 8.679 | 87.416 | 114.430 |
| 2009 | 18.112 | 10.063 | 28.179 | 1.210 | 42.072 | 9.743 | 84.413 | 91.673 |
| 2010 | 16.872 | 11.001 | 28.113 | 1.287 | 38.823 | 15.894 | 81.669 | 81.642 |
| 2011 | 14.338 | 10.541 | 27.916 | 1.296 | 37.065 | 16.415 | 78.336 | 67.053 |
| 2012 | 12.920 | 9.967 | 26.650 | 1.278 | 36.796 | 10.814 | 77.563 | 61.248 |
| 2013 | 12.787 | 9.233 | 25.696 | 1.296 | 36.097 | 6.940 | 76.447 | 62.410 |
| 2014 | 13.182 | 8.933 | 24.623 | 1.364 | 33.406 | 2.153 | 73.580 | 114.103 |
| 2015 | 12.654 | 8.684 | 23.665 | 1.444 | 29.450 | 5.219 | 68.400 | 64.240 |
| 2016 | 11.517 | 8.349 | 22.312 | 1.496 | 25.679 | 6.104 | 62.693 | 61.231 |
| 2017 | 9.768 | 7.722 | 19.269 | 1.063 | 23.402 | 2.605 | 57.078 | 52.922 |

Table 6(2bn2). Annual abundance estimates (million crab), mature male biomass (MMB, 1000 t), and total survey biomass (1000t) for red king crab in Bristol Bay estimated by length-based analysis (scenario 2bn2) from 1975-2017. Mature male biomass for year t is on Feb. 15, year $t+1$. Size measurements are mm carapace length.

| Year (t) | Males | | | | Females | | Total Recruits | Total Survey Biomass | |
|----------|------------------|----------------|---------------|--------|-----------------|---------------------|----------------|----------------------|--|
| | Mature (>119 mm) | Legal (>134mm) | MMB (>119 mm) | SD MMB | Mature (>89 mm) | Model Est. (>64 mm) | | Area-Swept (>64 mm) | |
| 1975 | 54.802 | 28.443 | 79.648 | 4.850 | 69.803 | | 253.779 | 202.731 | |
| 1976 | 60.220 | 34.956 | 89.597 | 4.118 | 105.838 | 31.724 | 291.600 | 331.868 | |
| 1977 | 61.680 | 37.683 | 92.702 | 3.462 | 130.670 | 38.361 | 300.468 | 375.661 | |
| 1978 | 67.184 | 38.593 | 94.978 | 2.897 | 123.041 | 44.807 | 289.972 | 349.545 | |
| 1979 | 62.906 | 39.829 | 79.711 | 2.435 | 105.661 | 69.605 | 264.951 | 167.627 | |
| 1980 | 44.475 | 32.376 | 21.548 | 0.791 | 95.068 | 61.354 | 224.937 | 249.322 | |
| 1981 | 13.220 | 7.574 | 5.723 | 0.288 | 43.591 | 25.127 | 89.746 | 132.669 | |
| 1982 | 6.131 | 2.322 | 6.242 | 0.306 | 19.953 | 123.472 | 42.528 | 143.740 | |
| 1983 | 5.701 | 2.476 | 7.254 | 0.314 | 13.117 | 60.746 | 37.480 | 49.320 | |
| 1984 | 5.589 | 2.689 | 5.425 | 0.304 | 13.811 | 81.018 | 38.687 | 155.311 | |
| 1985 | 6.773 | 2.204 | 9.134 | 0.433 | 13.294 | 8.201 | 32.423 | 34.535 | |
| 1986 | 11.170 | 4.297 | 13.378 | 0.646 | 19.446 | 41.516 | 43.609 | 48.158 | |
| 1987 | 14.336 | 6.145 | 19.279 | 0.798 | 23.036 | 12.018 | 49.887 | 70.263 | |
| 1988 | 15.021 | 8.334 | 24.833 | 0.881 | 27.472 | 7.226 | 53.777 | 55.372 | |
| 1989 | 16.361 | 10.049 | 28.236 | 0.918 | 25.022 | 7.718 | 57.117 | 55.941 | |
| 1990 | 16.530 | 10.986 | 25.808 | 0.920 | 21.210 | 21.658 | 57.527 | 60.321 | |
| 1991 | 13.196 | 9.666 | 20.694 | 0.899 | 19.142 | 13.609 | 51.830 | 85.055 | |
| 1992 | 10.164 | 7.501 | 18.664 | 0.865 | 18.849 | 1.883 | 45.574 | 37.687 | |
| 1993 | 10.701 | 6.735 | 16.070 | 0.843 | 16.831 | 9.623 | 43.736 | 53.703 | |
| 1994 | 10.447 | 6.119 | 21.520 | 0.877 | 13.788 | 1.337 | 38.189 | 32.335 | |
| 1995 | 10.980 | 7.950 | 24.395 | 0.860 | 13.345 | 55.066 | 44.758 | 38.396 | |
| 1996 | 11.028 | 8.593 | 22.432 | 0.822 | 18.667 | 6.699 | 52.236 | 44.649 | |
| 1997 | 10.328 | 7.655 | 20.744 | 0.795 | 27.376 | 2.658 | 56.297 | 85.277 | |
| 1998 | 14.669 | 7.429 | 22.500 | 0.879 | 25.631 | 11.198 | 59.761 | 85.176 | |
| 1999 | 16.169 | 8.919 | 27.128 | 0.991 | 22.594 | 31.216 | 59.523 | 65.604 | |
| 2000 | 14.333 | 10.365 | 27.271 | 0.997 | 24.898 | 11.720 | 62.119 | 68.342 | |
| 2001 | 13.371 | 9.916 | 26.461 | 0.970 | 28.859 | 9.232 | 64.879 | 53.188 | |
| 2002 | 15.109 | 9.525 | 28.458 | 0.979 | 28.707 | 54.947 | 69.805 | 69.786 | |
| 2003 | 16.007 | 10.420 | 27.564 | 0.980 | 34.503 | 7.765 | 75.105 | 116.794 | |
| 2004 | 14.458 | 10.029 | 25.538 | 0.953 | 42.231 | 16.302 | 77.355 | 131.910 | |
| 2005 | 16.881 | 9.506 | 26.656 | 0.985 | 40.470 | 54.482 | 82.407 | 107.341 | |
| 2006 | 17.346 | 10.365 | 29.331 | 1.056 | 44.672 | 18.743 | 86.603 | 95.676 | |
| 2007 | 16.997 | 11.133 | 26.916 | 1.104 | 51.938 | 11.676 | 92.717 | 104.841 | |
| 2008 | 18.984 | 10.428 | 28.748 | 1.261 | 49.056 | 9.333 | 93.415 | 114.430 | |
| 2009 | 20.514 | 11.508 | 33.239 | 1.465 | 44.728 | 10.813 | 91.575 | 91.673 | |
| 2010 | 19.405 | 12.953 | 34.031 | 1.570 | 41.346 | 17.014 | 89.649 | 81.642 | |
| 2011 | 16.726 | 12.649 | 34.151 | 1.585 | 39.564 | 16.980 | 86.717 | 67.053 | |
| 2012 | 15.215 | 12.030 | 32.745 | 1.559 | 39.356 | 11.407 | 86.057 | 61.248 | |
| 2013 | 14.949 | 11.158 | 31.611 | 1.562 | 38.665 | 7.335 | 84.603 | 62.410 | |
| 2014 | 15.118 | 10.756 | 30.418 | 1.609 | 35.831 | 2.342 | 81.318 | 114.103 | |
| 2015 | 14.427 | 10.430 | 28.995 | 1.664 | 31.615 | 5.687 | 75.919 | 64.240 | |
| 2016 | 13.131 | 9.925 | 27.259 | 1.693 | 27.597 | 6.673 | 69.747 | 61.231 | |
| 2017 | 11.249 | 9.162 | 22.347 | 1.153 | 25.189 | 2.825 | 63.703 | 52.922 | |

Table 7(2b). Comparison of projected mature male biomass (1000 t) on Feb. 15, retained catch (1000 t), their 95% limits, and mean fishing mortality with no directed fishery, $F_{40\%}$, and $F_{35\%}$ harvest strategy with $F_{35\%}$ constraint during 2017-2026. Parameter estimates with scenario 2 are used for the projection.

| No Directed Fishery | | | | | | |
|---------------------|--------|---------|---------|-------|---------|---------|
| Year | MMB | 95% LCI | 95% UCI | Catch | 95% LCI | 95% UCI |
| 2017 | 26.310 | 23.719 | 28.757 | 0.000 | 0.000 | 0.000 |
| 2018 | 25.868 | 23.321 | 28.274 | 0.000 | 0.000 | 0.000 |
| 2019 | 25.350 | 22.853 | 27.709 | 0.000 | 0.000 | 0.000 |
| 2020 | 24.805 | 22.432 | 27.316 | 0.000 | 0.000 | 0.000 |
| 2021 | 26.554 | 22.133 | 36.148 | 0.000 | 0.000 | 0.000 |
| 2022 | 30.811 | 22.341 | 49.368 | 0.000 | 0.000 | 0.000 |
| 2023 | 35.911 | 22.448 | 60.877 | 0.000 | 0.000 | 0.000 |
| 2024 | 41.038 | 23.713 | 70.650 | 0.000 | 0.000 | 0.000 |
| 2025 | 45.760 | 25.333 | 78.897 | 0.000 | 0.000 | 0.000 |
| 2026 | 49.955 | 26.680 | 85.860 | 0.000 | 0.000 | 0.000 |

| $F_{40\%}$ | | | | | | |
|------------|--------|---------|---------|-------|---------|---------|
| Year | MMB | 95% LCI | 95% UCI | Catch | 95% LCI | 95% UCI |
| 2017 | 21.981 | 20.158 | 23.665 | 4.465 | 3.672 | 5.252 |
| 2018 | 18.845 | 17.476 | 20.094 | 3.238 | 2.735 | 3.724 |
| 2019 | 16.757 | 15.642 | 17.767 | 2.448 | 2.101 | 2.778 |
| 2020 | 15.364 | 14.354 | 16.580 | 1.968 | 1.713 | 2.252 |
| 2021 | 16.352 | 13.274 | 23.896 | 1.961 | 1.451 | 3.227 |
| 2022 | 19.337 | 12.876 | 33.427 | 2.458 | 1.311 | 4.734 |
| 2023 | 22.430 | 13.092 | 40.377 | 3.243 | 1.285 | 6.517 |
| 2024 | 24.947 | 13.938 | 43.503 | 3.999 | 1.388 | 8.018 |
| 2025 | 26.726 | 15.000 | 47.101 | 4.599 | 1.601 | 8.656 |
| 2026 | 27.928 | 15.453 | 49.508 | 5.002 | 1.802 | 9.333 |

| $F_{35\%}$ | | | | | | |
|------------|--------|---------|---------|-------|---------|---------|
| Year | MMB | 95% LCI | 95% UCI | Catch | 95% LCI | 95% UCI |
| 2017 | 21.339 | 19.616 | 22.926 | 5.127 | 4.231 | 6.014 |
| 2018 | 17.968 | 16.717 | 19.104 | 3.545 | 3.014 | 4.056 |
| 2019 | 15.820 | 14.818 | 16.725 | 2.607 | 2.254 | 2.941 |
| 2020 | 14.444 | 13.521 | 15.592 | 2.065 | 1.806 | 2.355 |
| 2021 | 15.435 | 12.466 | 22.807 | 2.062 | 1.513 | 3.439 |
| 2022 | 18.316 | 12.089 | 31.744 | 2.645 | 1.365 | 5.335 |
| 2023 | 21.187 | 12.329 | 37.890 | 3.535 | 1.345 | 7.308 |
| 2024 | 23.414 | 13.155 | 41.078 | 4.362 | 1.465 | 8.990 |
| 2025 | 24.898 | 14.135 | 43.504 | 4.986 | 1.714 | 9.510 |
| 2026 | 25.839 | 14.579 | 45.380 | 5.379 | 1.901 | 10.226 |

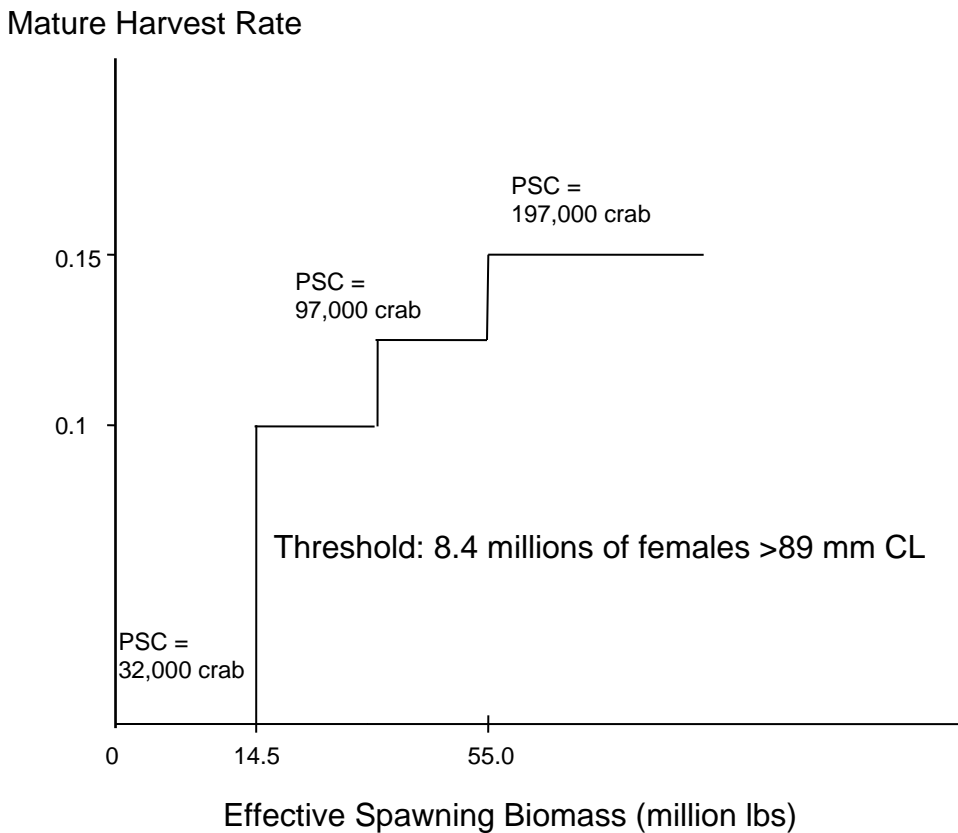


Figure 1. Current harvest rate strategy (line) for the Bristol Bay red king crab fishery and annual prohibited species catch (PSC) limits (numbers of crab) of Bristol Bay red king crab in the groundfish fisheries in zone 1 in the eastern Bering Sea. Harvest rates are based on current-year estimates of effective spawning biomass (ESB), whereas PSC limits apply to previous-year ESB.

Data by type and year

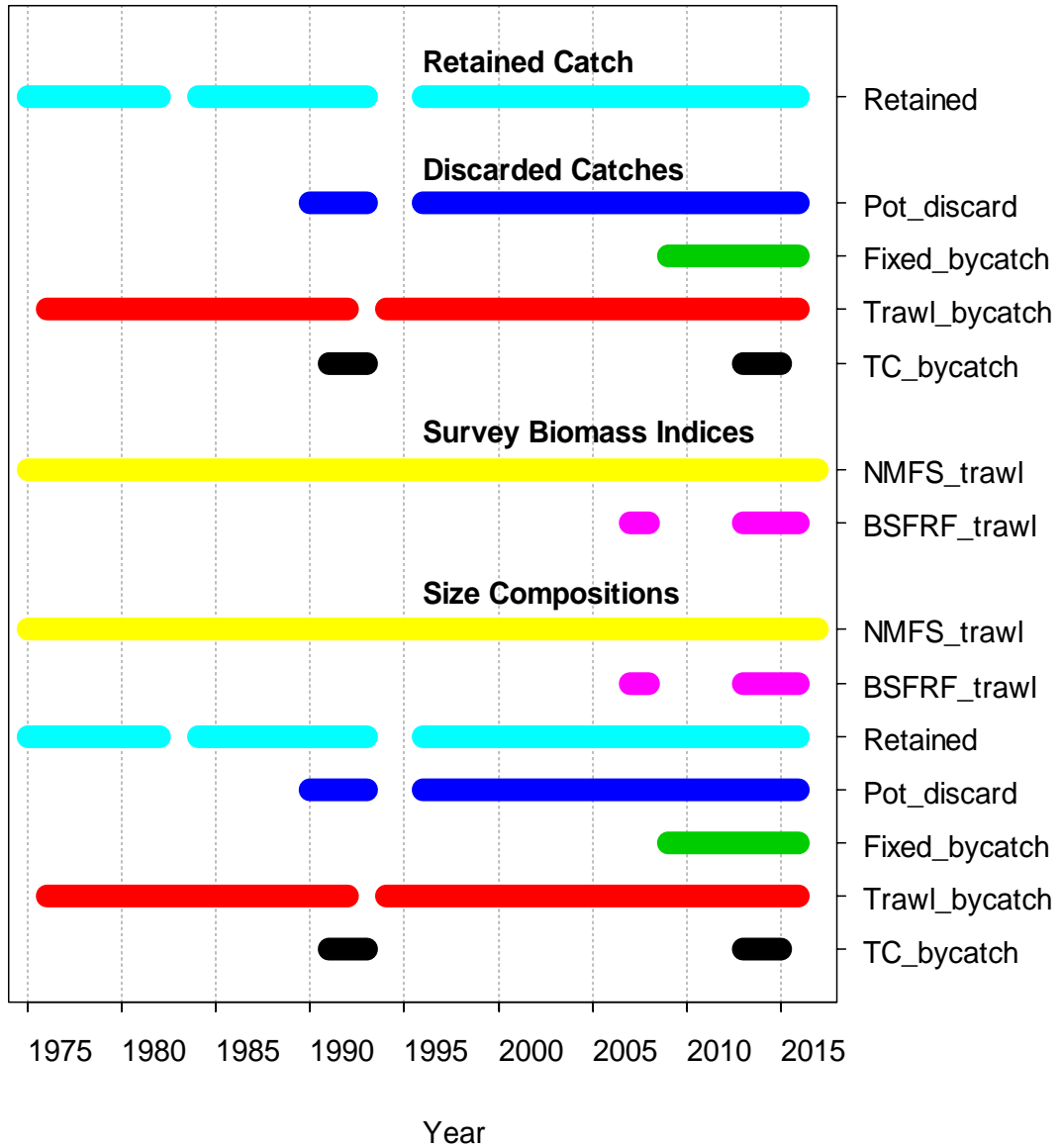


Figure 2. Data types and ranges used for the stock assessment.

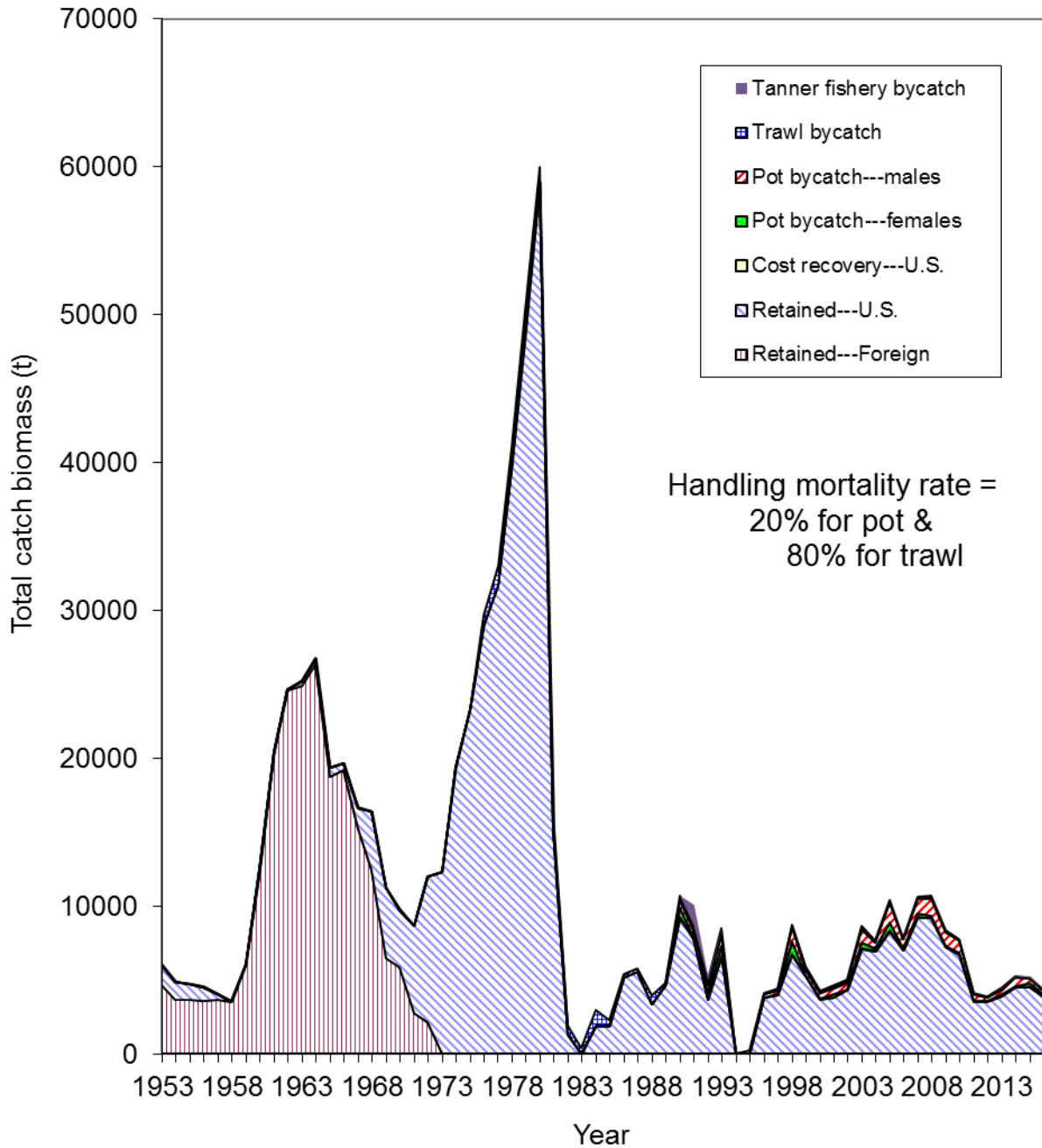


Figure 3. Retained catch biomass and bycatch mortality biomass (t) for Bristol Bay red king crab from 1953 to 2016. Handling mortality rates were assumed to be 0.2 for the directed pot fishery 0.25 for the Tanner crab fishery and 0.8 for the trawl fisheries.

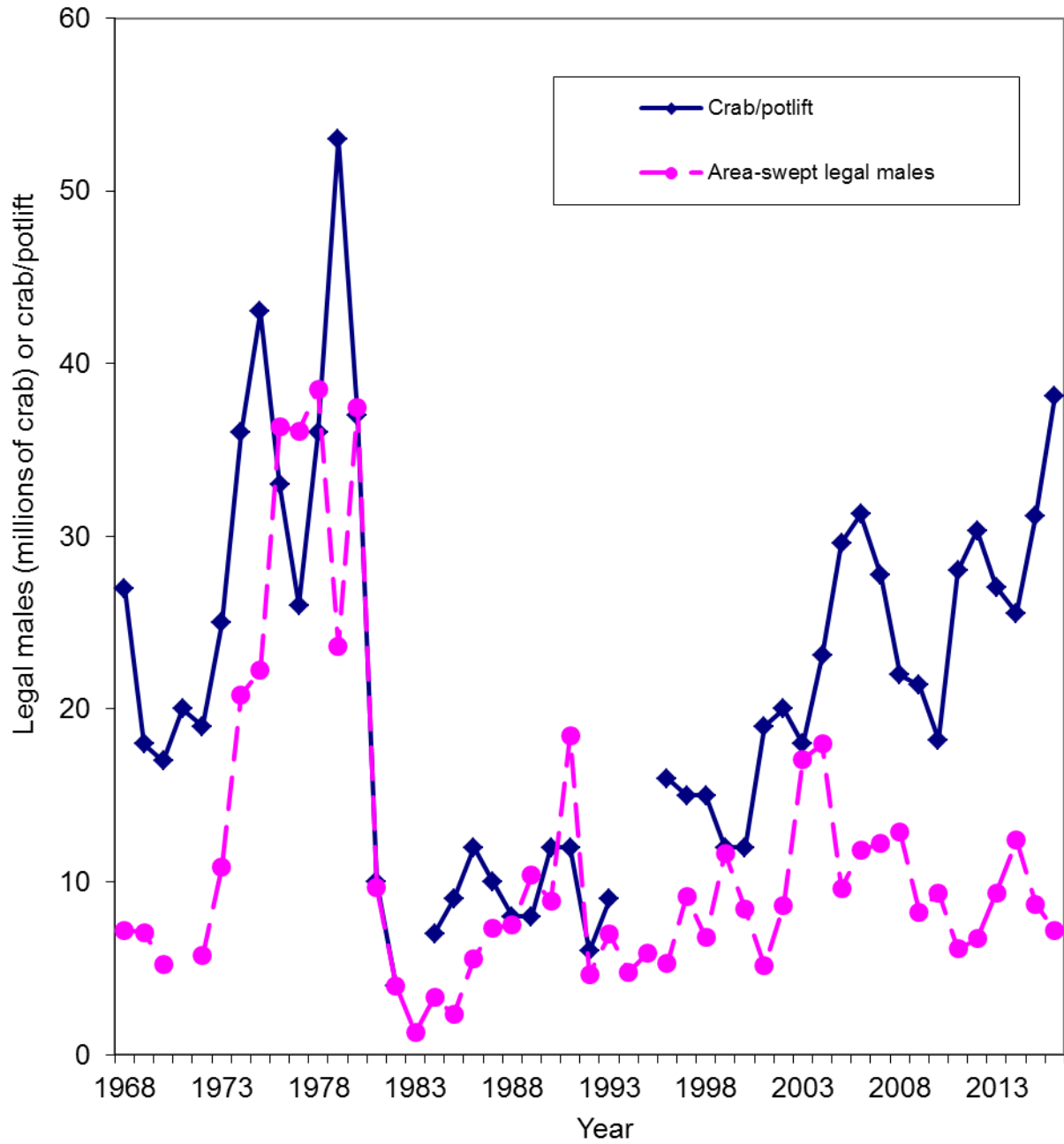


Figure 4. Comparison of survey legal male abundances and catches per unit effort for Bristol Bay red king crab from 1968 to 2016.

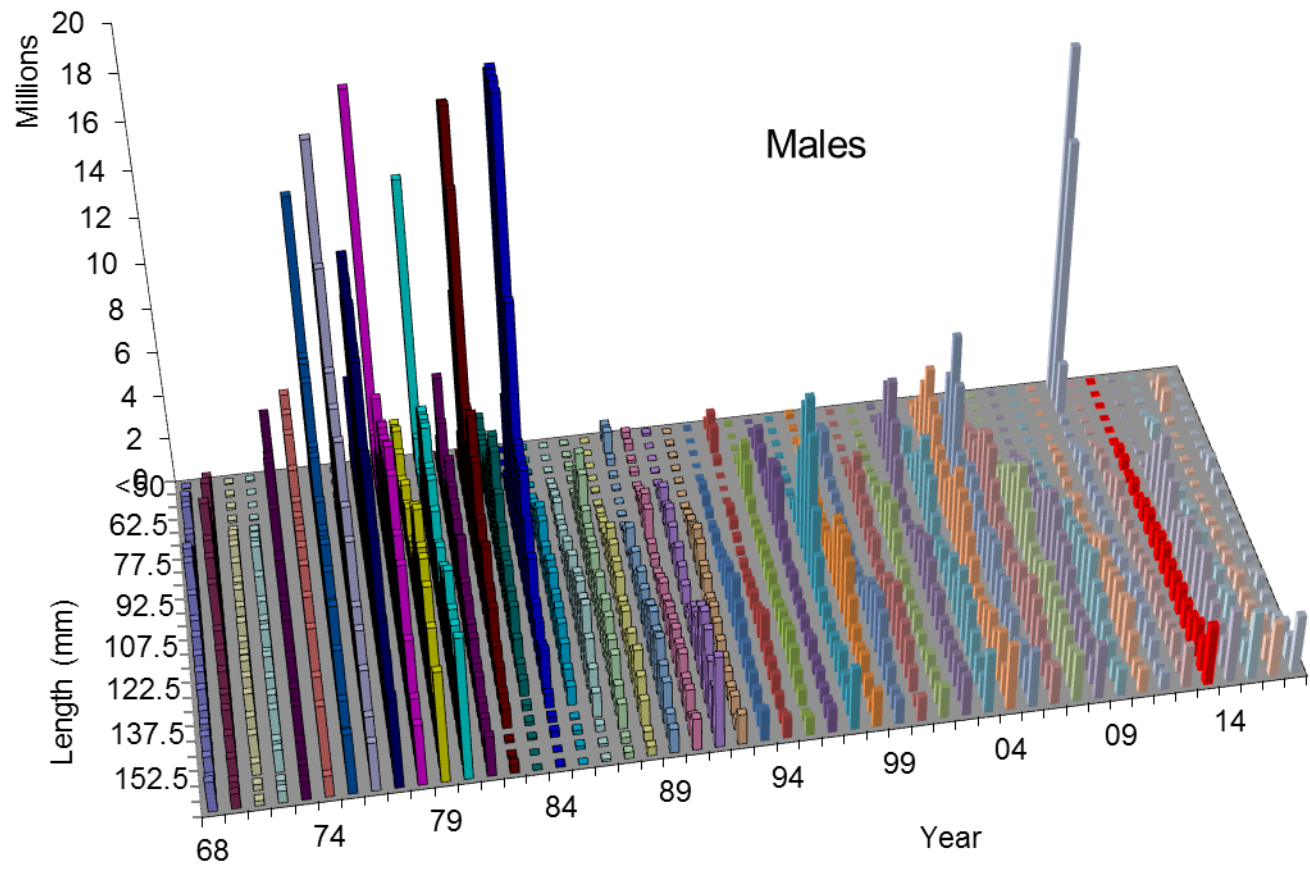


Figure 5a. Survey abundances by 5-mm carapace length bin for male Bristol Bay red king crab from 1968 to 2017.

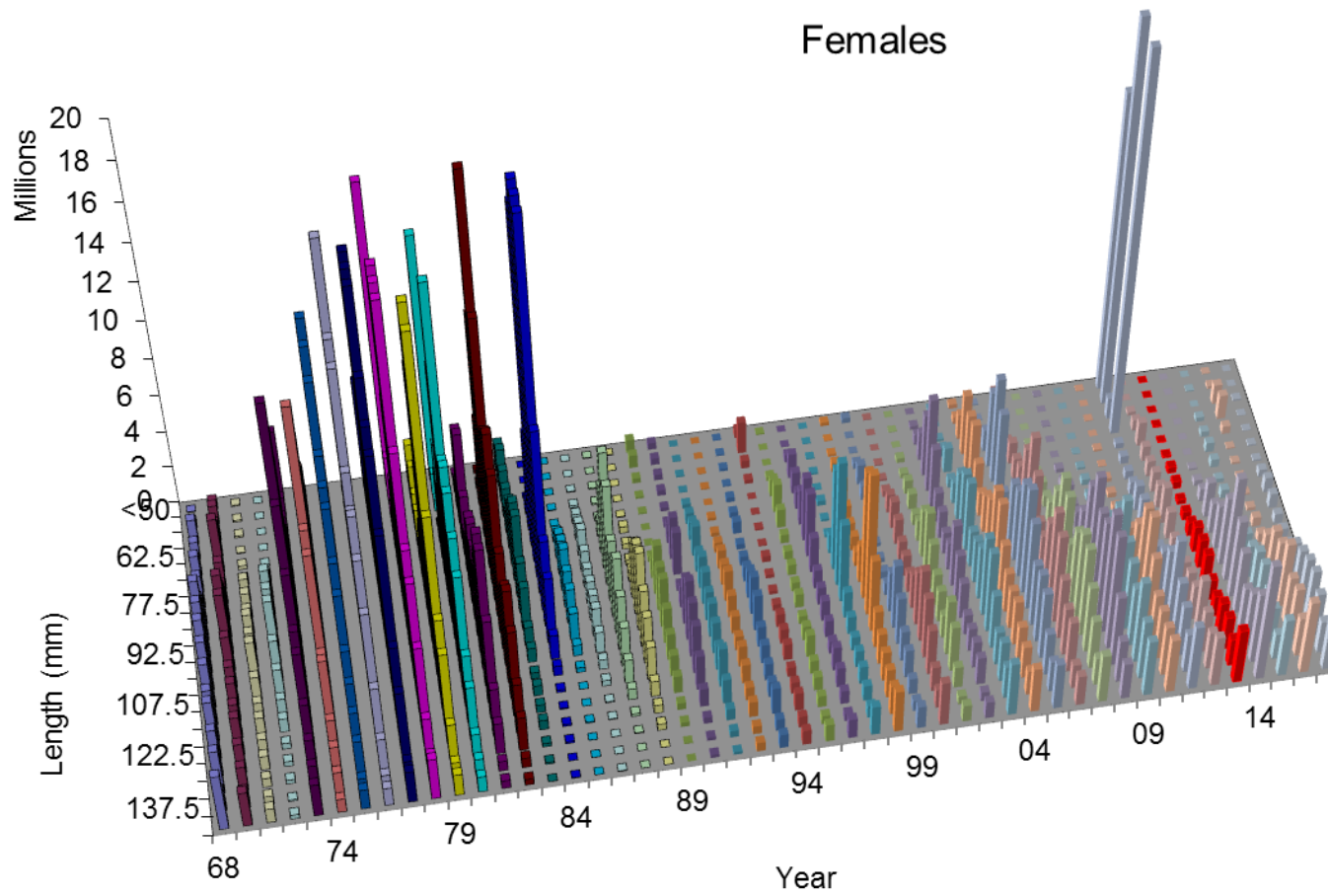


Figure 5b. Survey abundances by 5 mm carapace length bin for female Bristol Bay red king crab from 1968 to 2017.

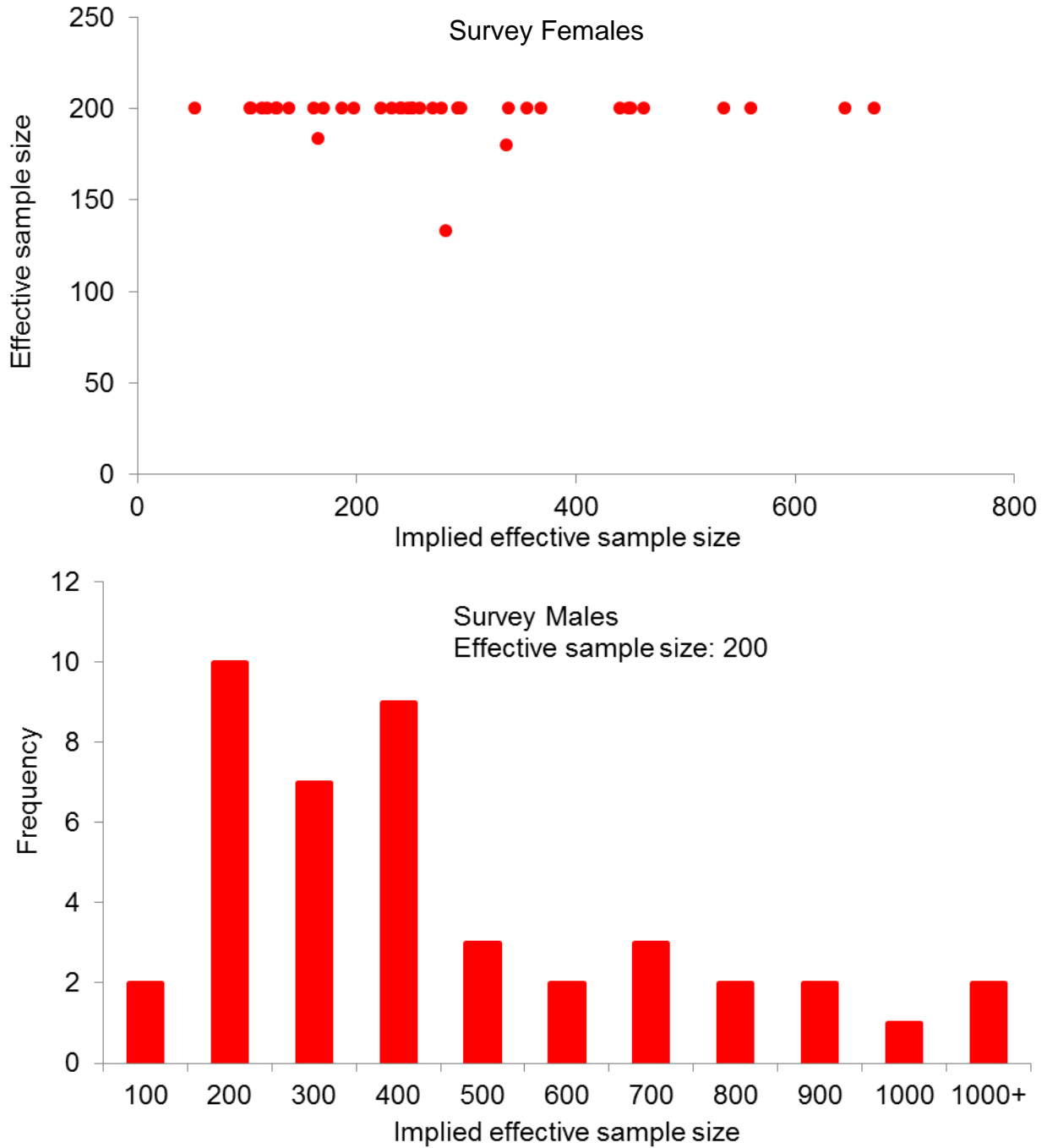


Figure 6. Relationship between implied effective sample sizes (section 3(a)(5)(i)) and effective sample sizes (see effective sample sizes for scenario 2b) for length/sex composition data with scenario 2b: trawl survey data.

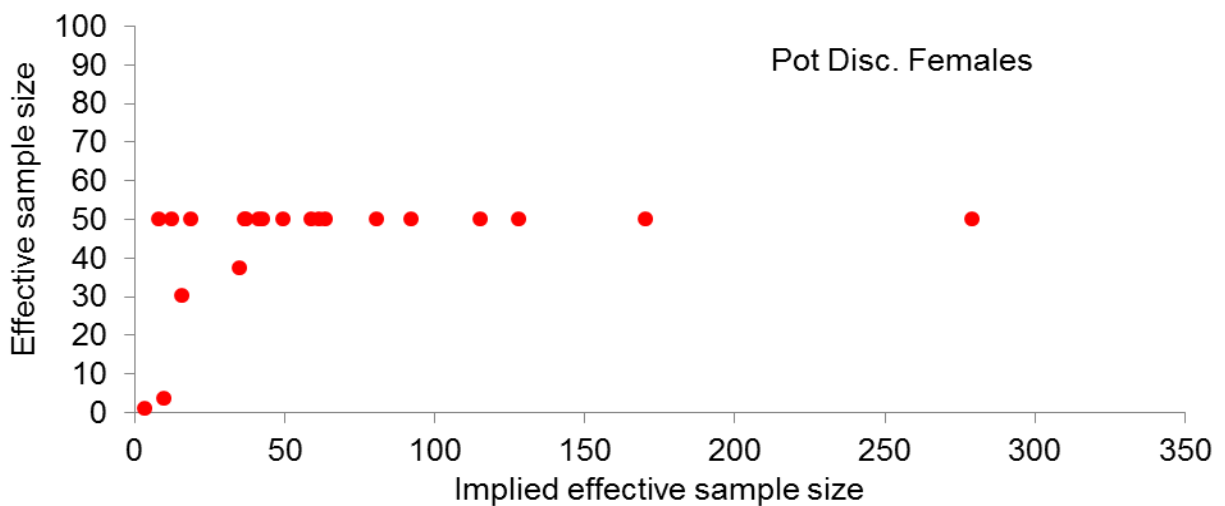
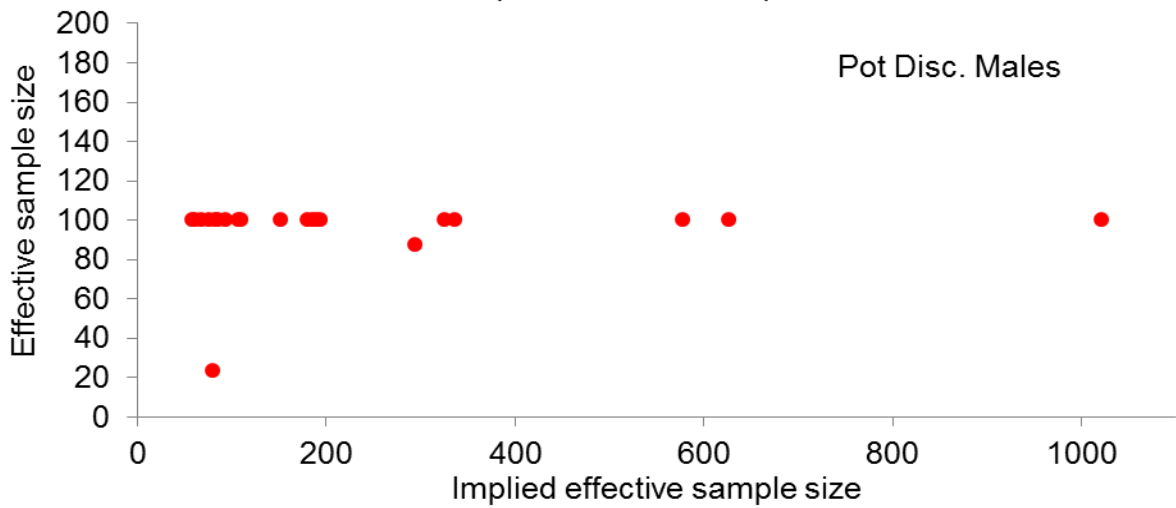
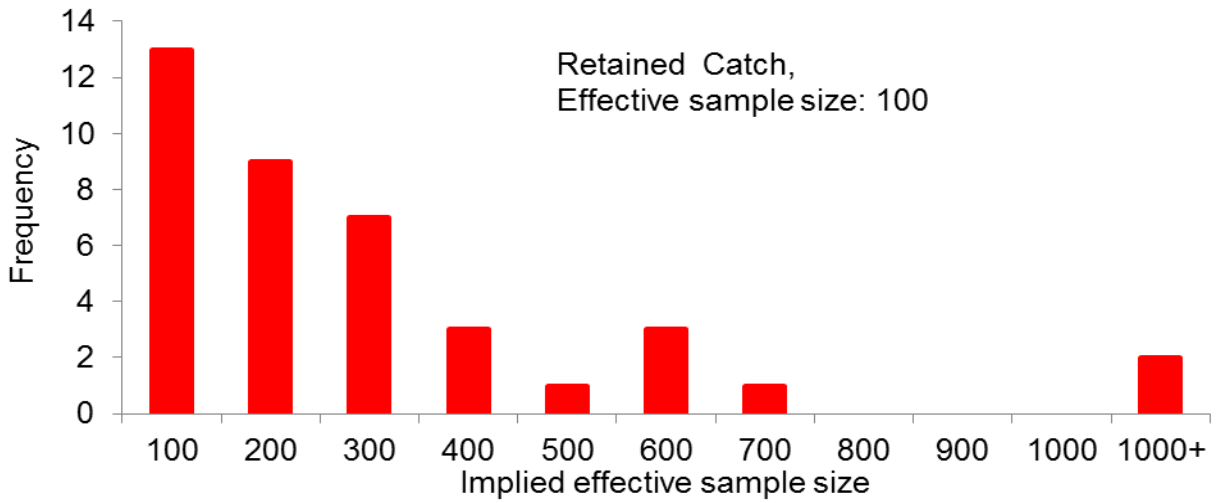


Figure 7. Relationship between implied effective sample sizes (section 3(a)(5)(i)) and effective sample sizes (see effective sample sizes for scenario 2b) for length/sex composition data with scenario 2b: directed pot fishery data.

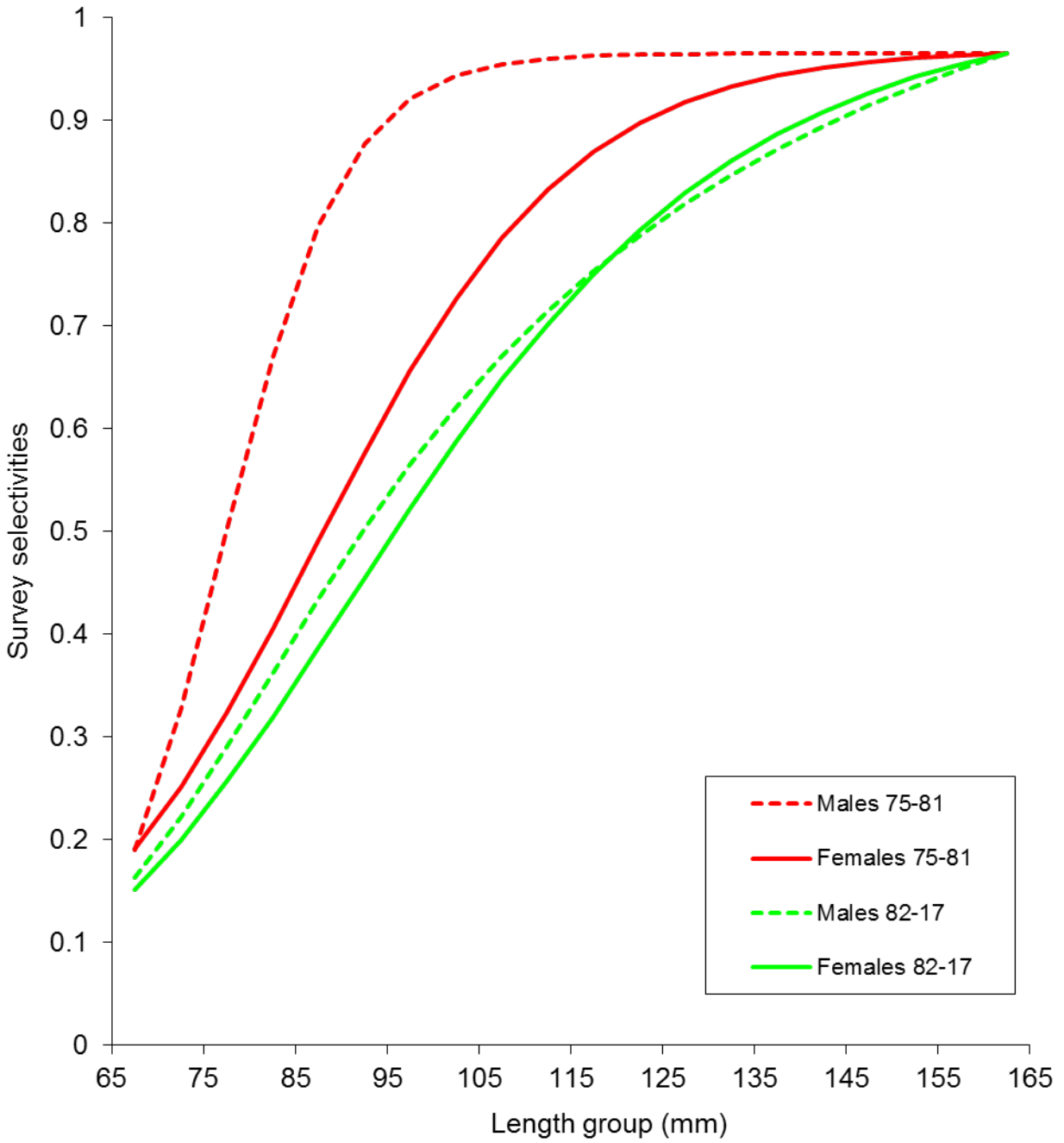


Figure 8a(2b). Estimated trawl survey selectivities under scenario 2b. Pot, fixed gear and trawl handling mortality rates were assumed to be 0.2 0.5 and 0.8, respectively.

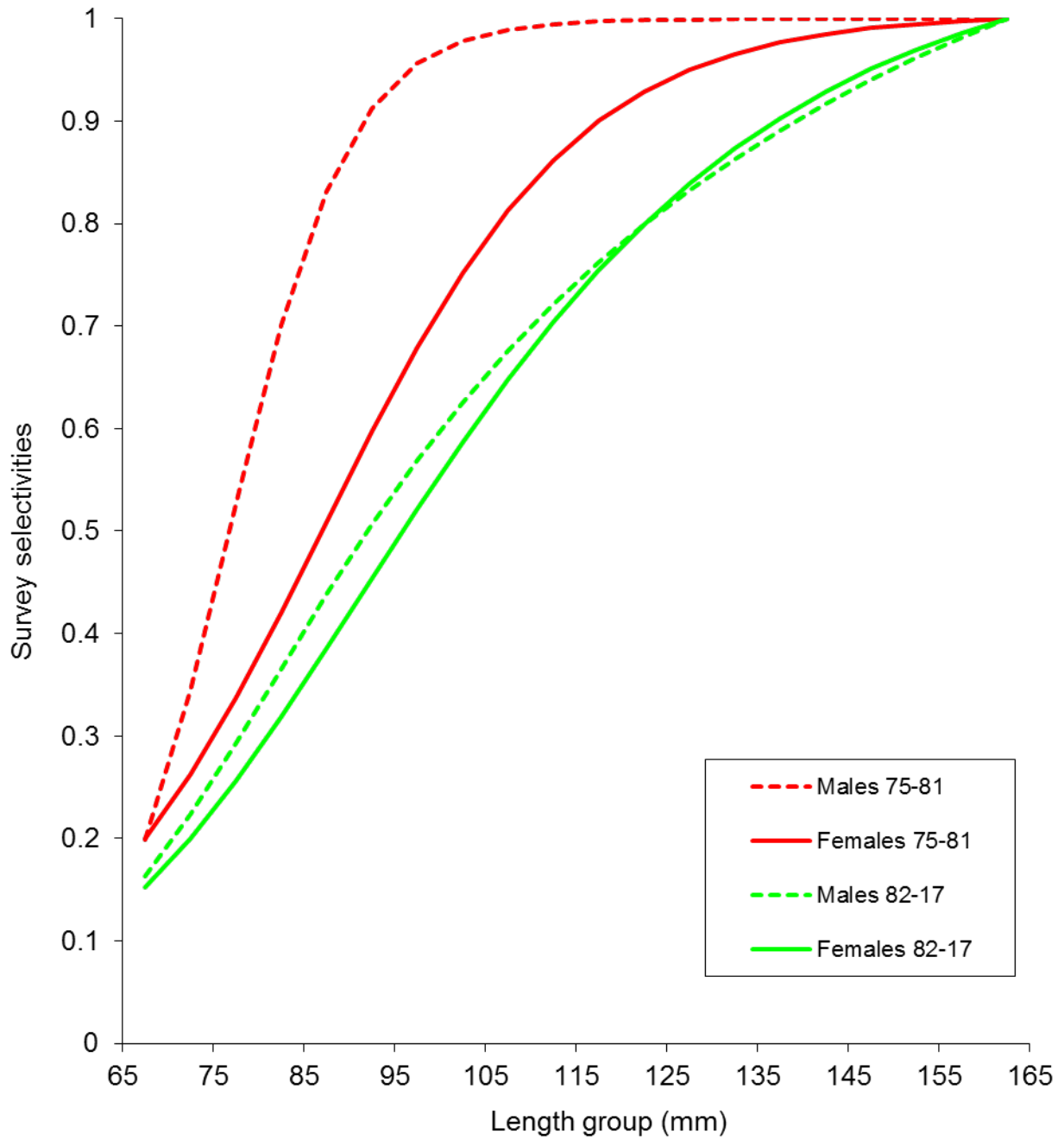


Figure 8a(2bn1). Estimated trawl survey selectivities under scenario 2bn1. Pot, fixed gear and trawl handling mortality rates were assumed to be 0.2, 0.5 and 0.8, respectively.

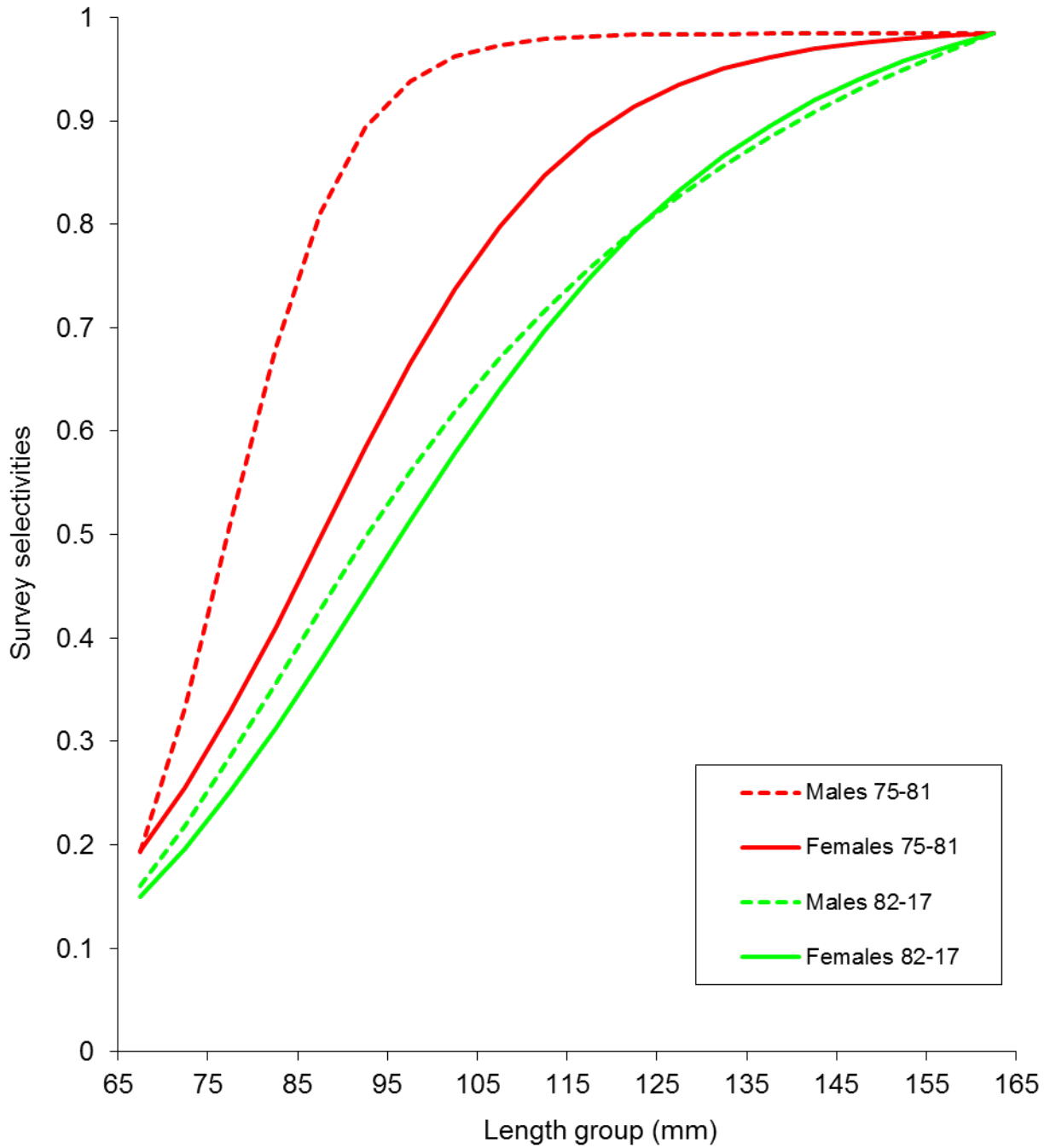


Figure 8a(2bn2). Estimated trawl survey selectivities under scenario 2bn2. Pot, fixed gear and trawl handling mortality rates were assumed to be 0.2, 0.5 and 0.8, respectively.

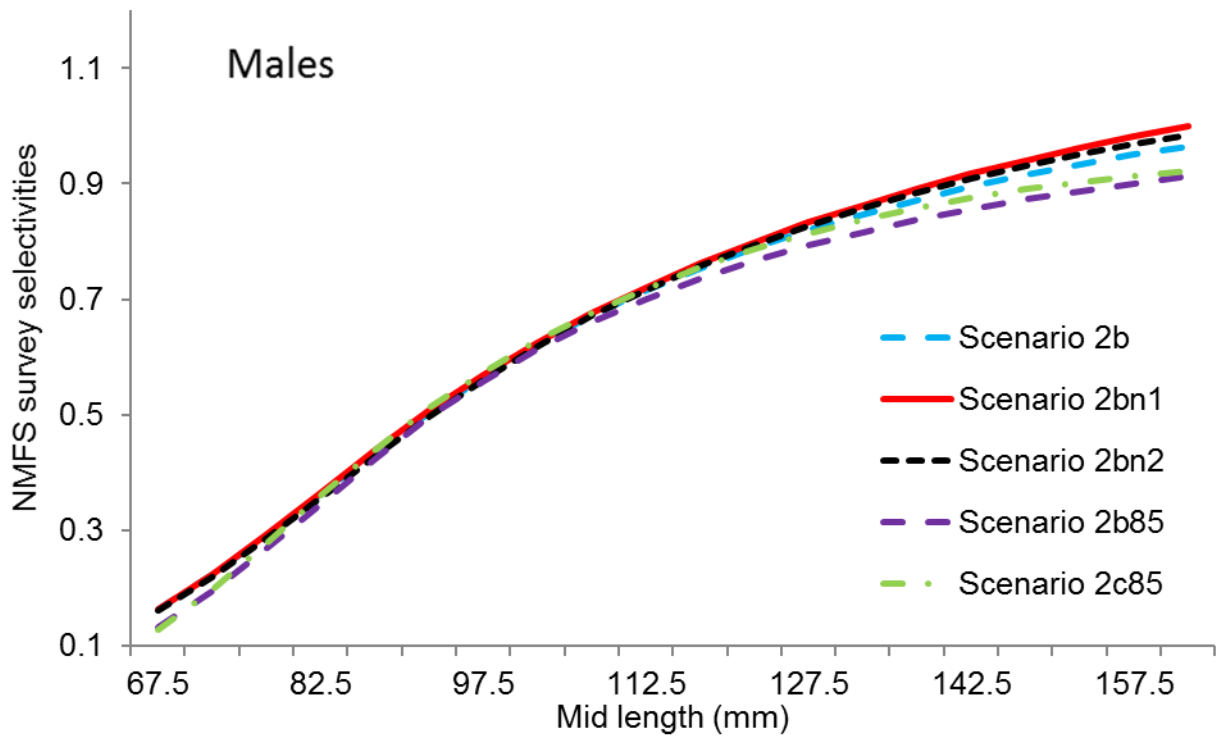
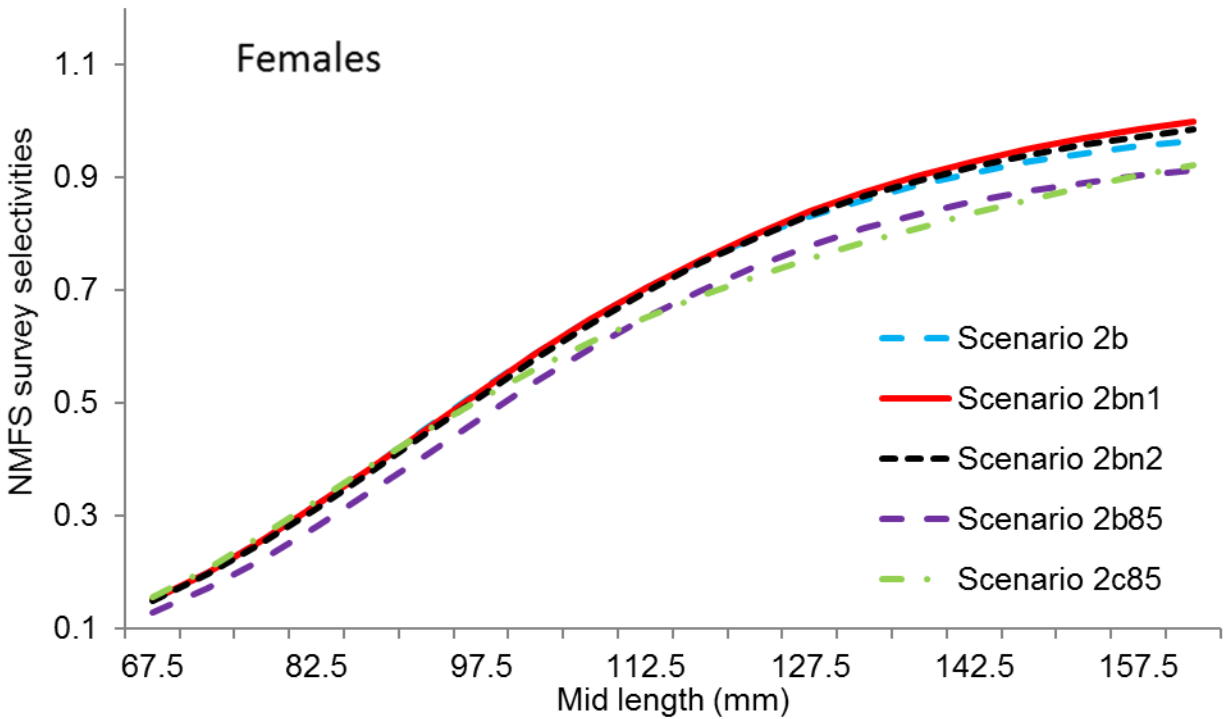


Figure 8b. Comparisons of estimated NMFS trawl survey selectivities for period 1982-2017 under scenarios 2b, 2bn1, 2bn2, 2b85 and 2c85. Pot, fixed gear and trawl handling mortality rates were assumed to be 0.2, 0.5 and 0.8, respectively.

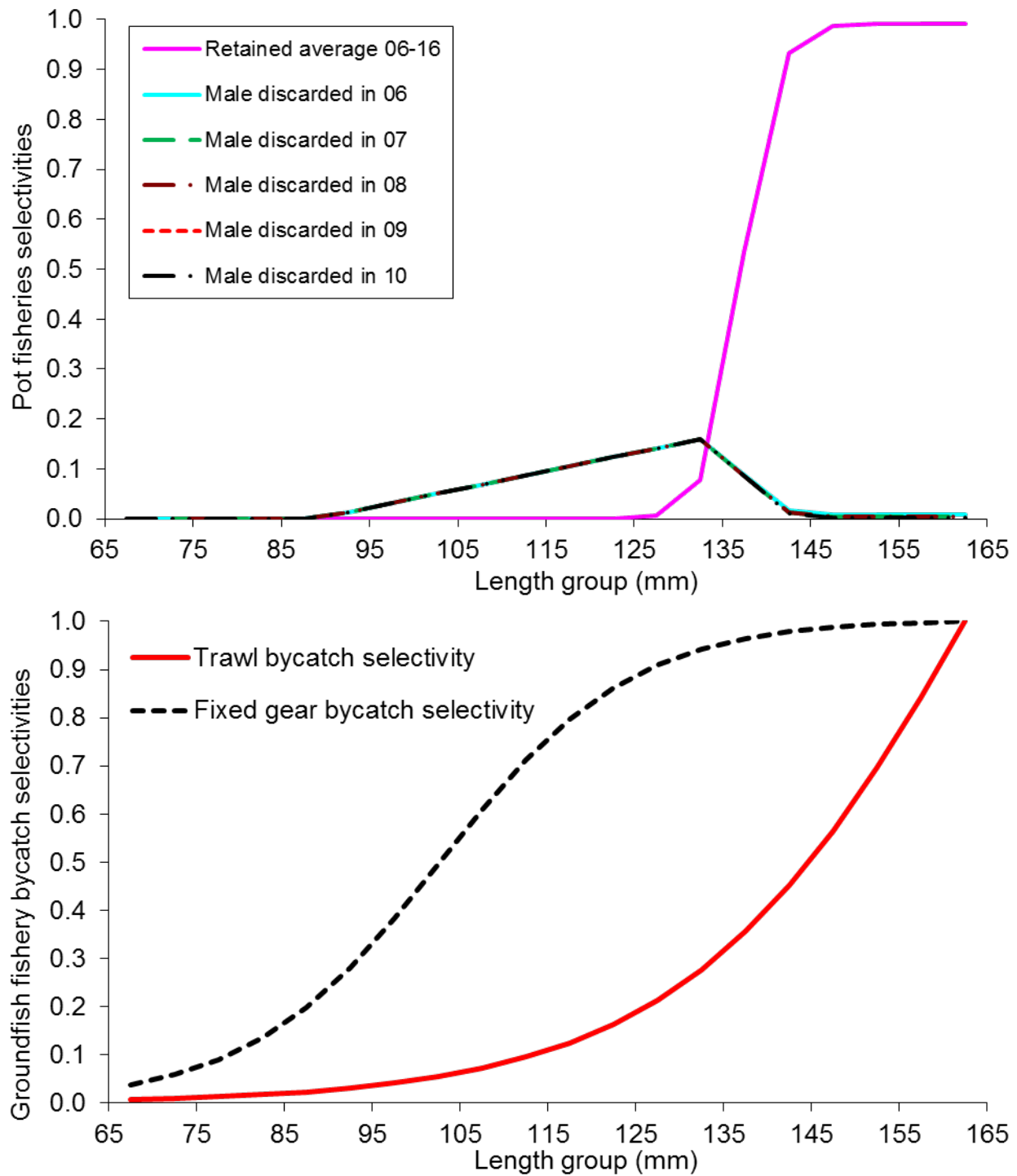


Figure 8c. Estimated pot fishery selectivities and groundfish trawl bycatch selectivities under scenario 2b. Pot, fixed gear and trawl handling mortality rates were assumed to be 0.2, 0.5 and 0.8, respectively.

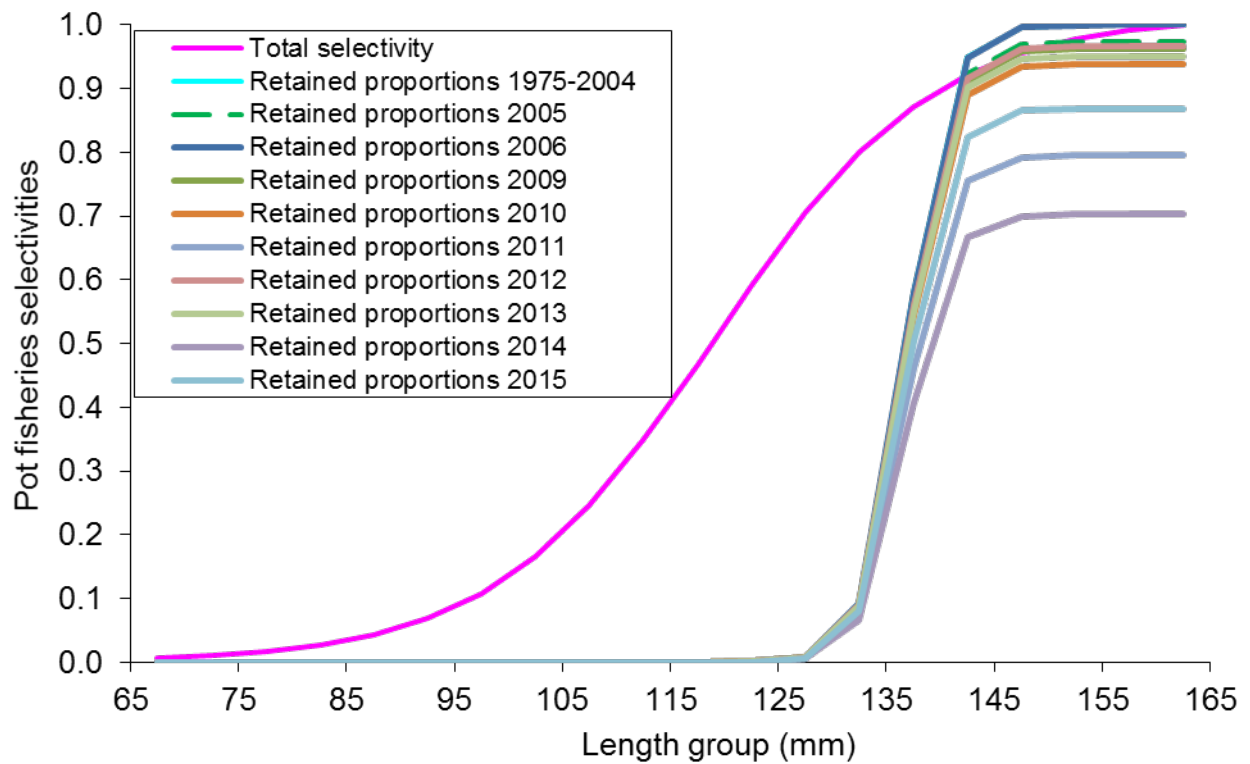
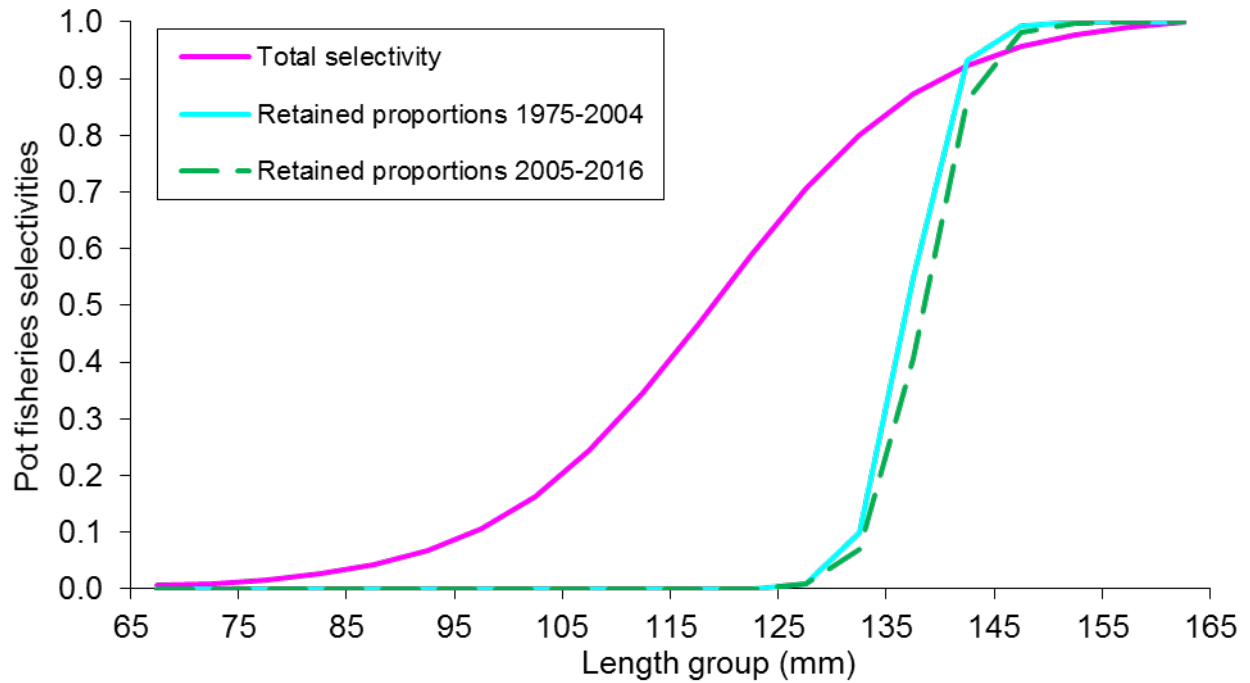


Figure 8d. Estimated pot fishery selectivities and retained proportions under scenarios 2bn1 (upper panel) and 2bn2 (lower panel). Pot, fixed gear and trawl handling mortality rates were assumed to be 0.2, 0.5 and 0.8, respectively.

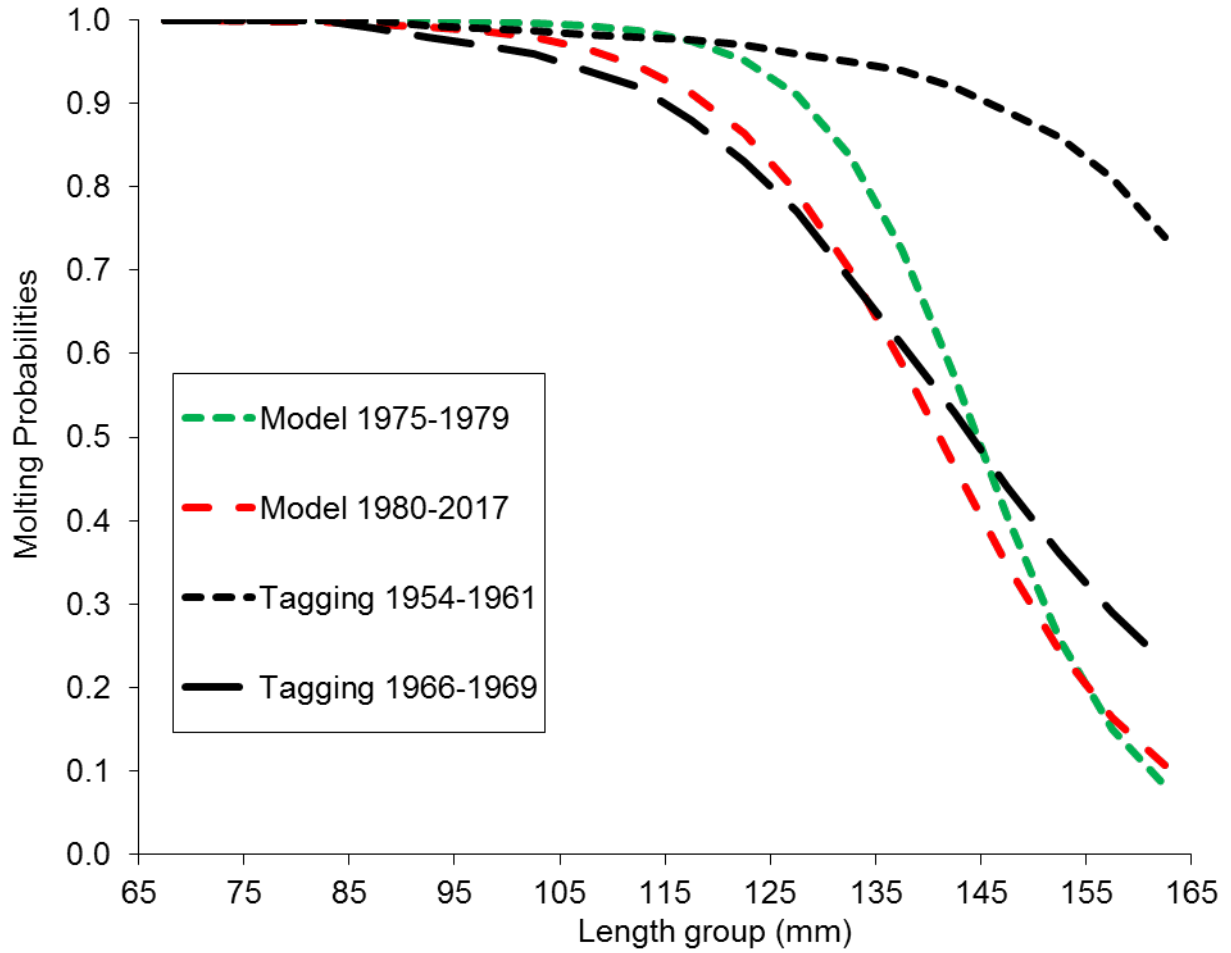


Figure 9(2b). Comparison of estimated probabilities of molting of male red king crab in Bristol Bay for different periods. Molting probabilities for periods 1954-1961 and 1966-1969 were estimated by Balsiger (1974) from tagging data. Molting probabilities for 1975-2017 were estimated with a length-based model with pot handling mortality rate of 0.2 under scenario 2b.

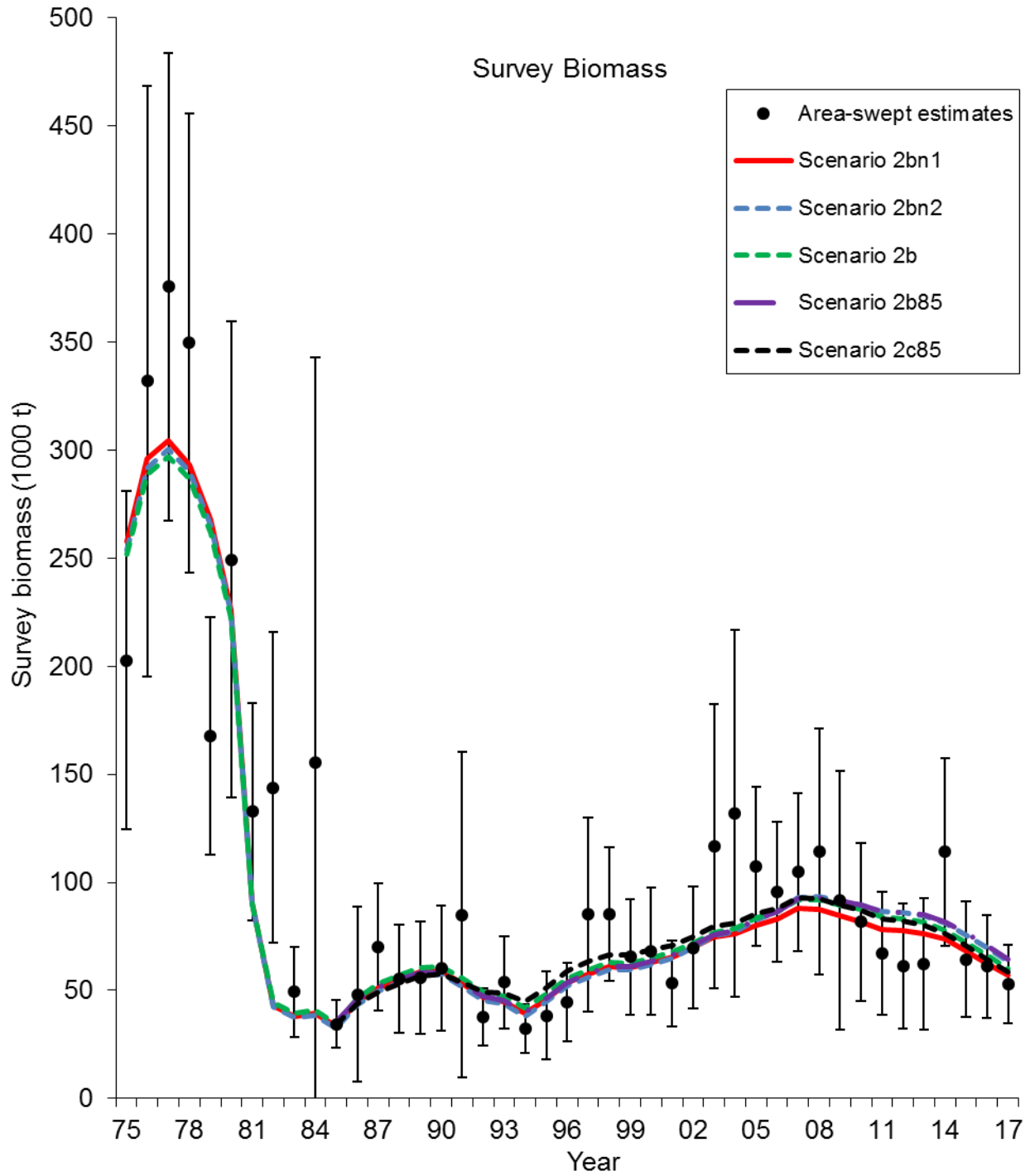


Figure 10a. Comparisons of area-swept estimates of total survey biomass and model prediction for model estimates in 2017 under scenarios 2b, 2bn1, 2bn2, 2b85, and 2c85. Pot, fixed gear, and trawl handling mortality rates were assumed to be 0.2, 0.5 and 0.8, respectively. The error bars are plus and minus 2 standard deviations.

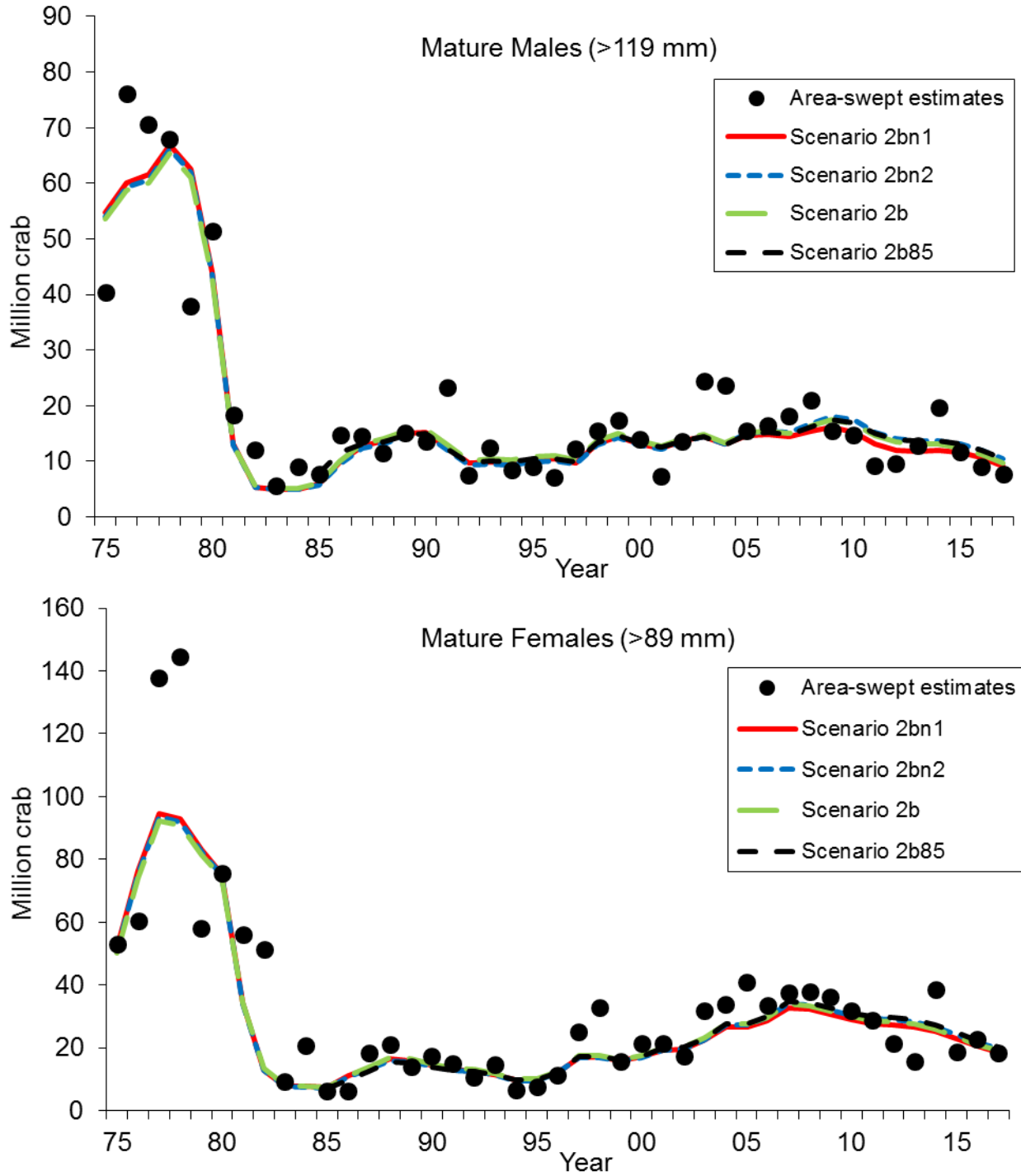


Figure 10b. Comparisons of area-swept estimates of male (>119 mm) and female (>89 mm) abundance and model prediction for model estimates in 2017 under scenarios 2b, 2bn1, 2bn2, and 2b85. Pot, fixed gear and trawl handling mortality rates were assumed to be 0.2, 0.5 and 0.8, respectively.

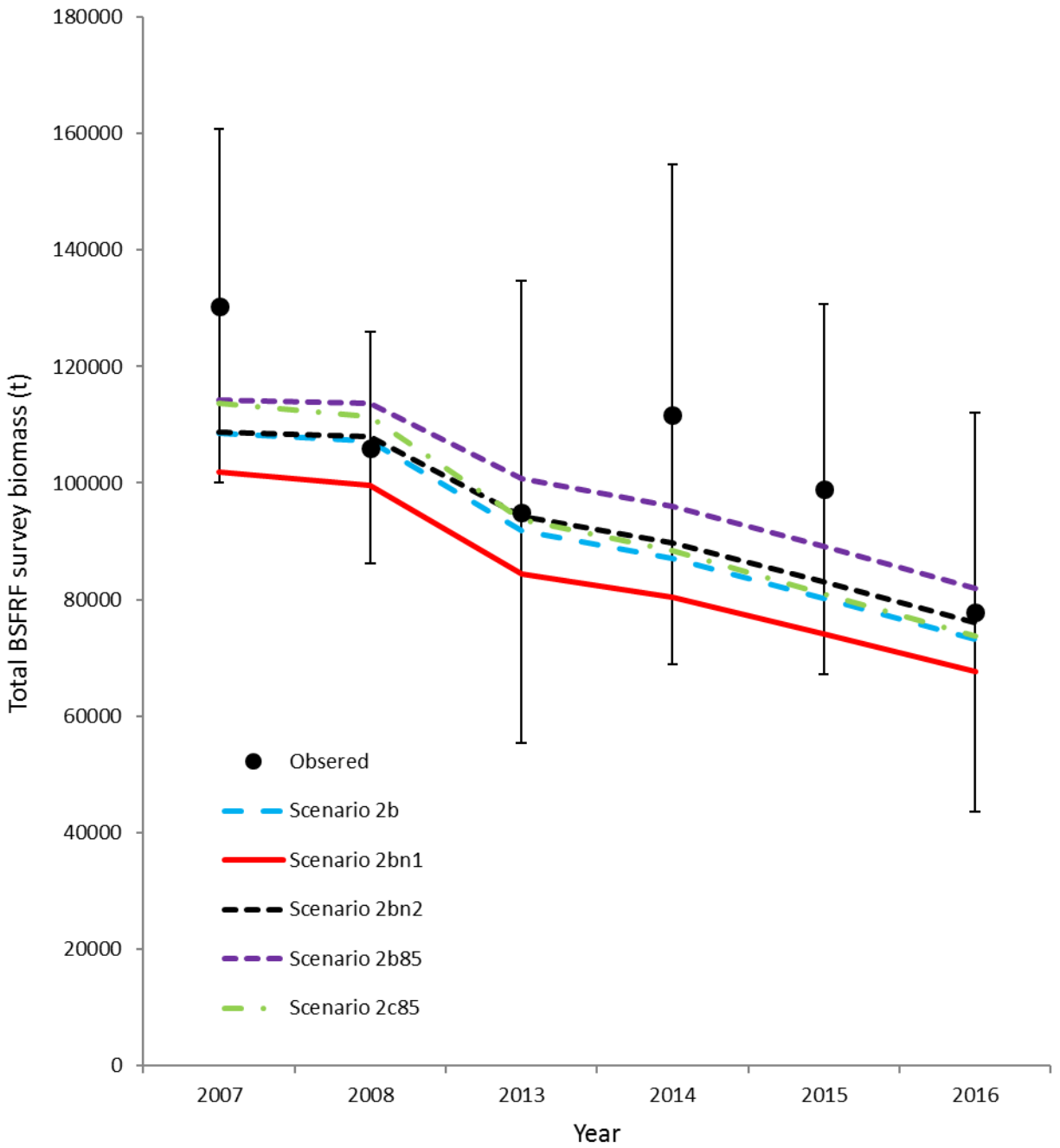


Figure 10c. Comparisons of total survey biomass estimates by the BSFRF survey and the model for model estimates in 2017 (scenarios 2b, 2bn1, 2bn2, 2b85, and 2c85). The error bars are plus and minus 2 standard deviations of scenario 2b.

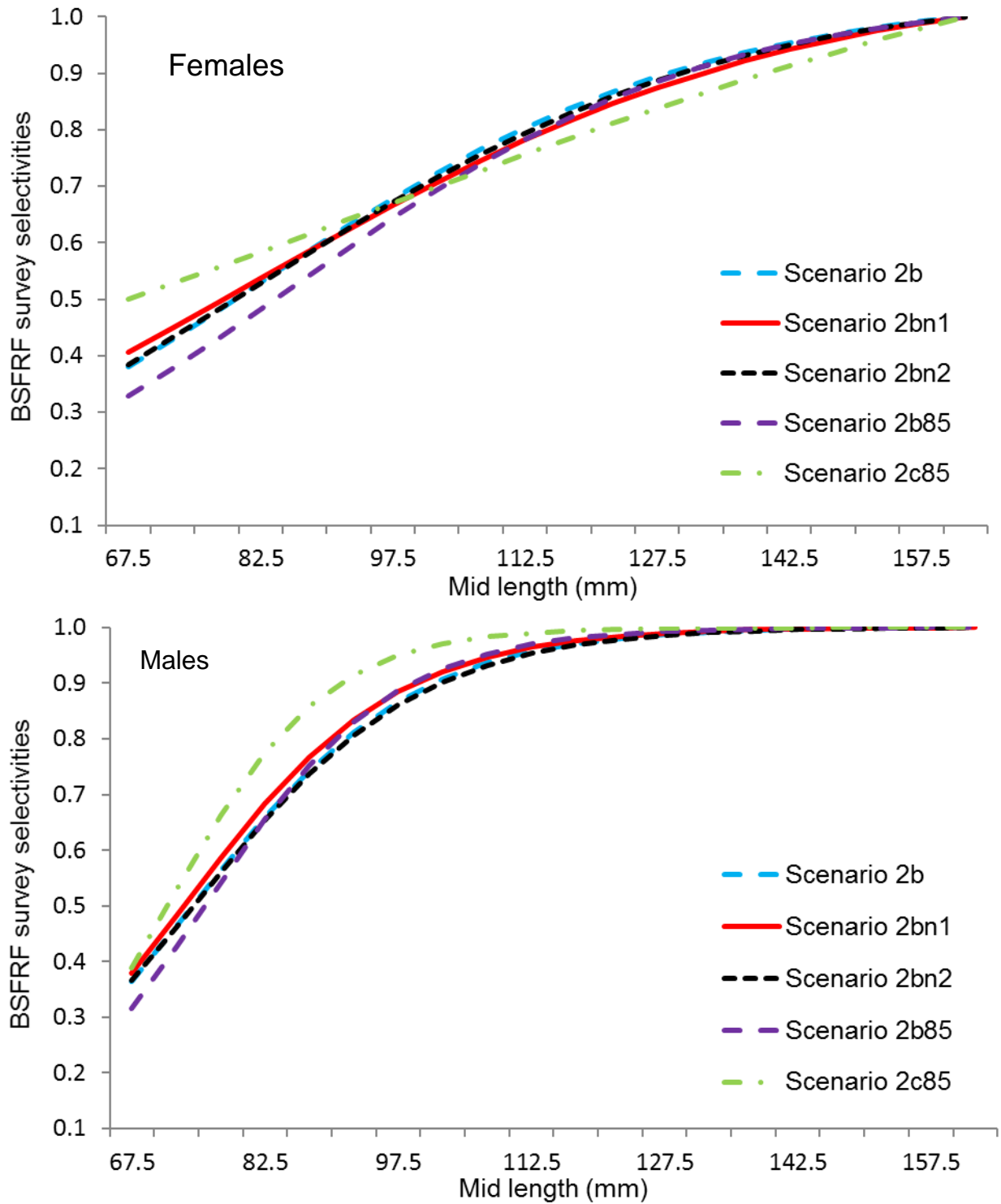


Figure 10d. Comparisons of estimated BSFRF survey selectivities with scenarios 2b, 2bn1, 2bn21, 2b85, and 2c85. The catchability is assumed to be 1.0.

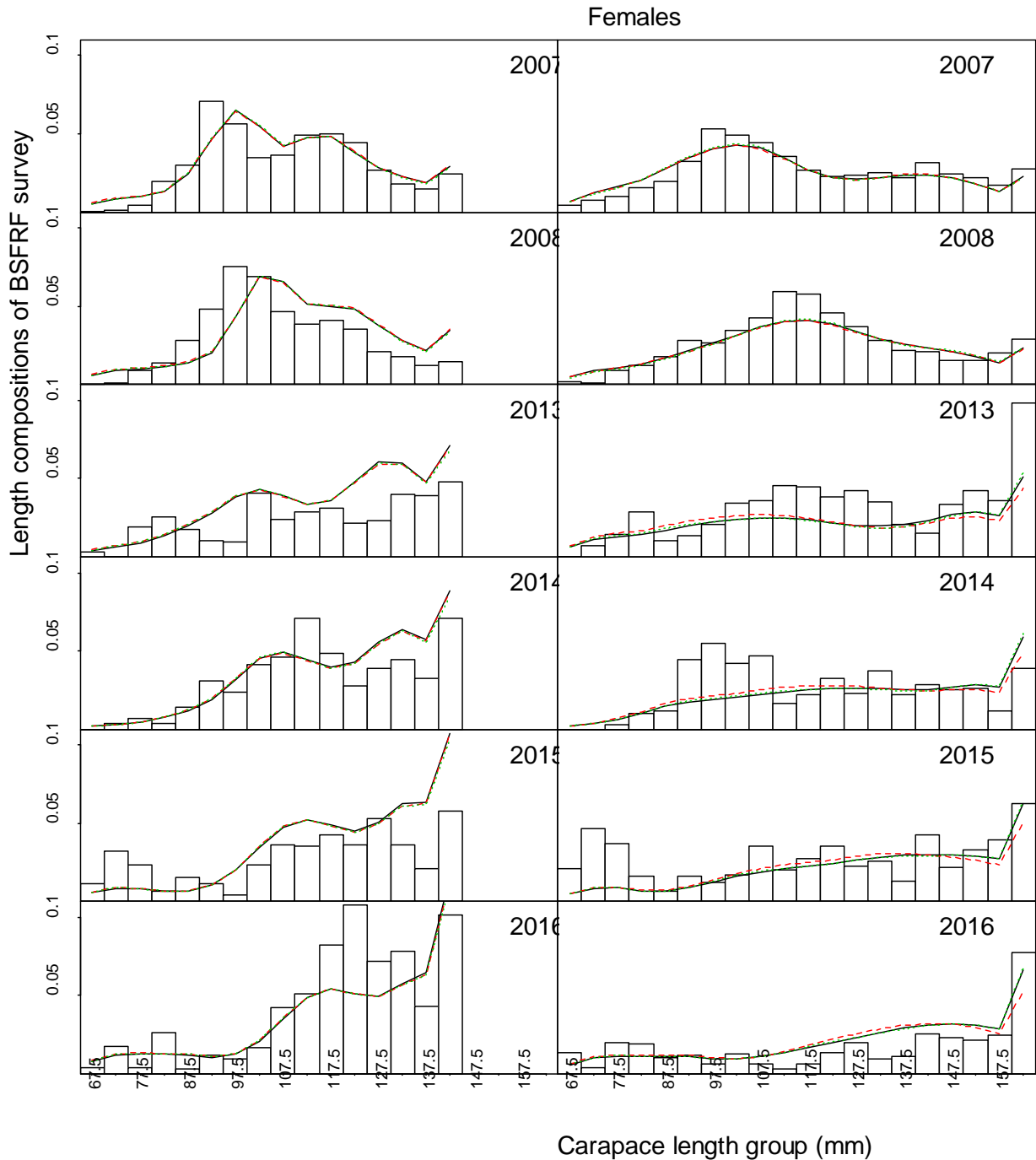


Figure 10e(2b, 2bn1 & 2bn2). Comparisons of length compositions by the BSFRF survey and the model estimates during 2007-2008 and 2013-2016 with scenarios 2b (solid black), 2bn1 (dashed red), and 2bn2 (green lines).

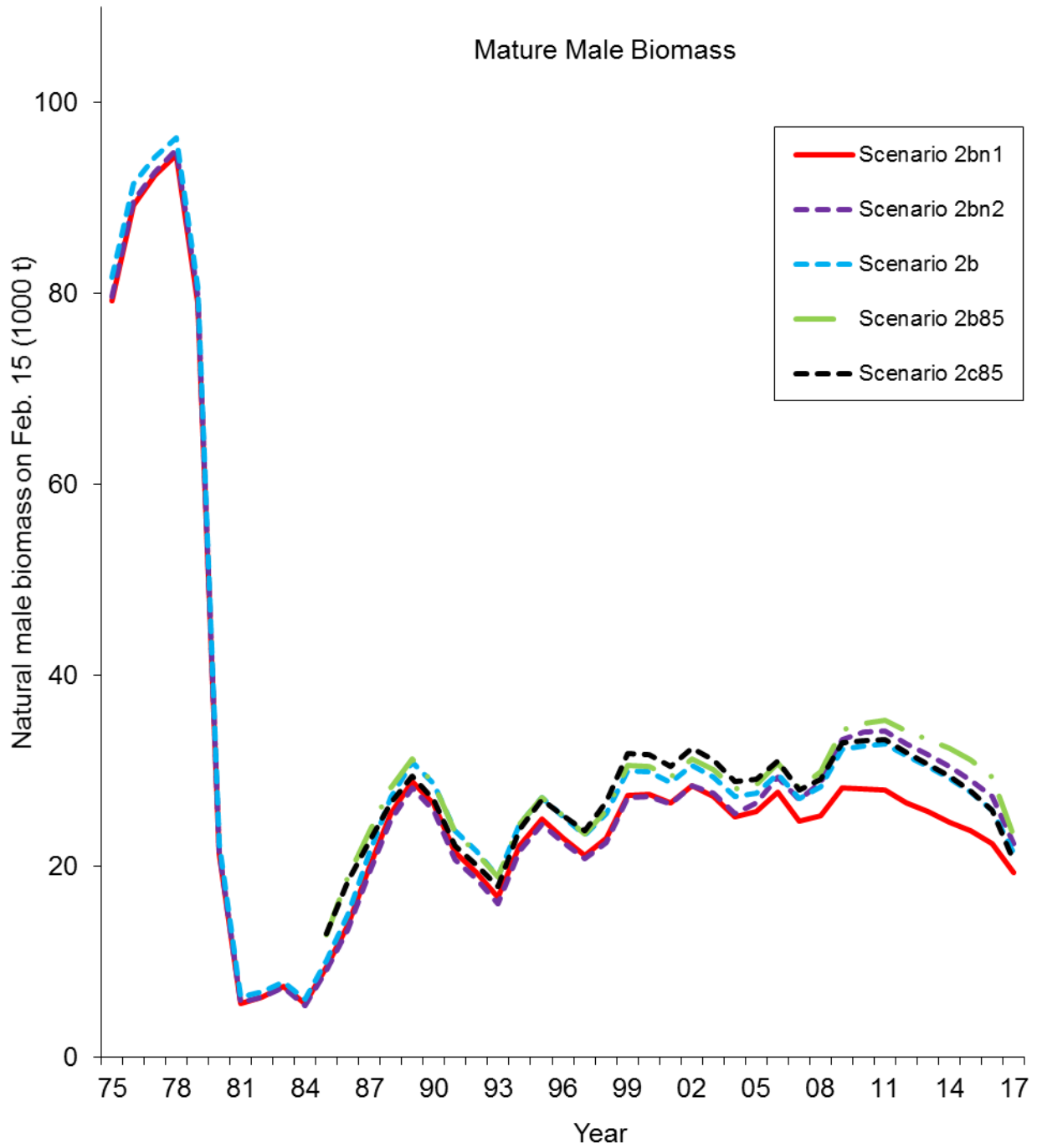


Figure 11. Estimated absolute mature male biomasses during 1975-2017 for scenarios 2b, 2bn1, 2bn2, 2b85, and 2c85.

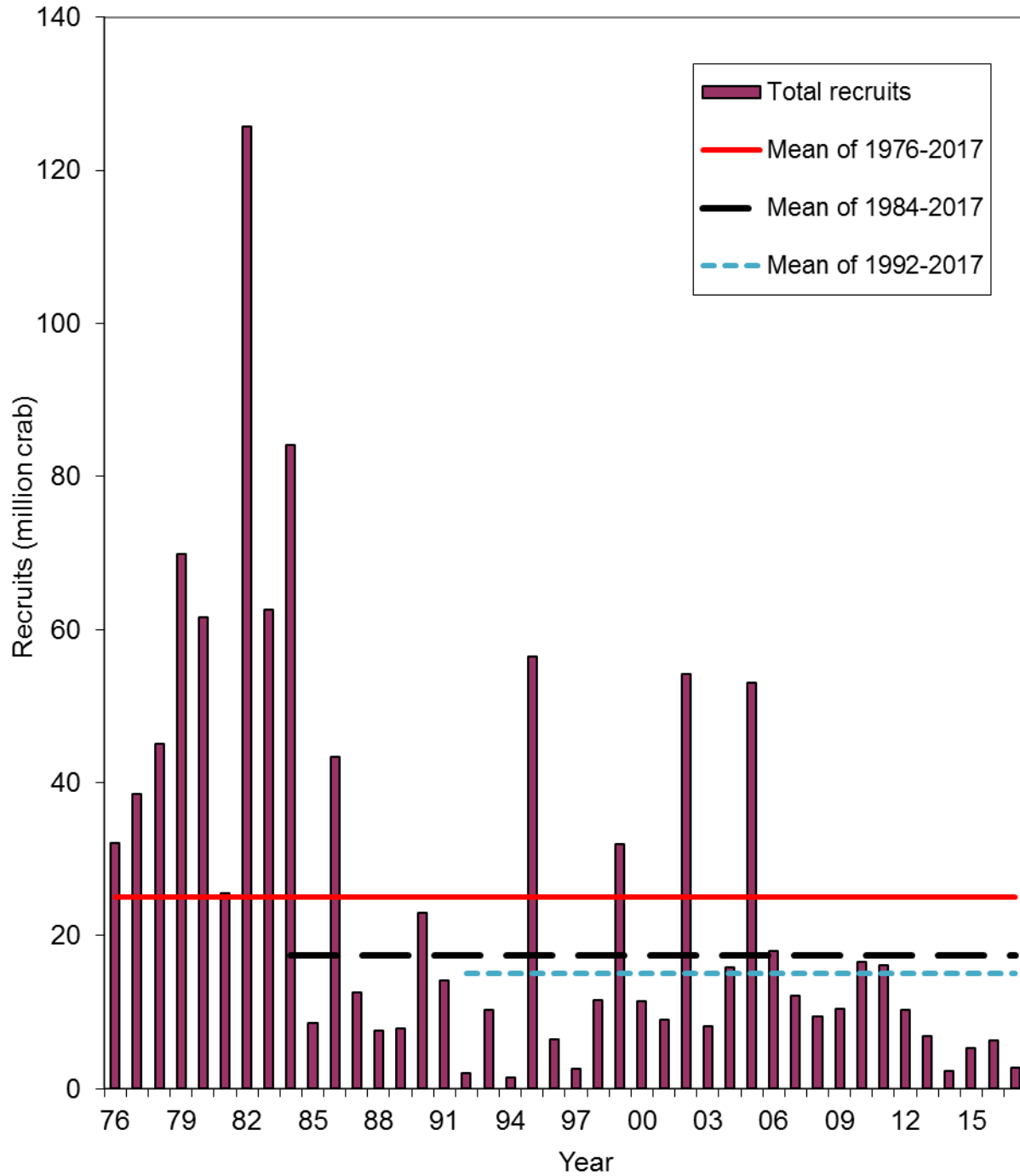


Figure 12(2b). Estimated recruitment time series during 1976-2017 with scenario 2b. Mean male recruits during 1984-2016 was used to estimate $B_{35\%}$.

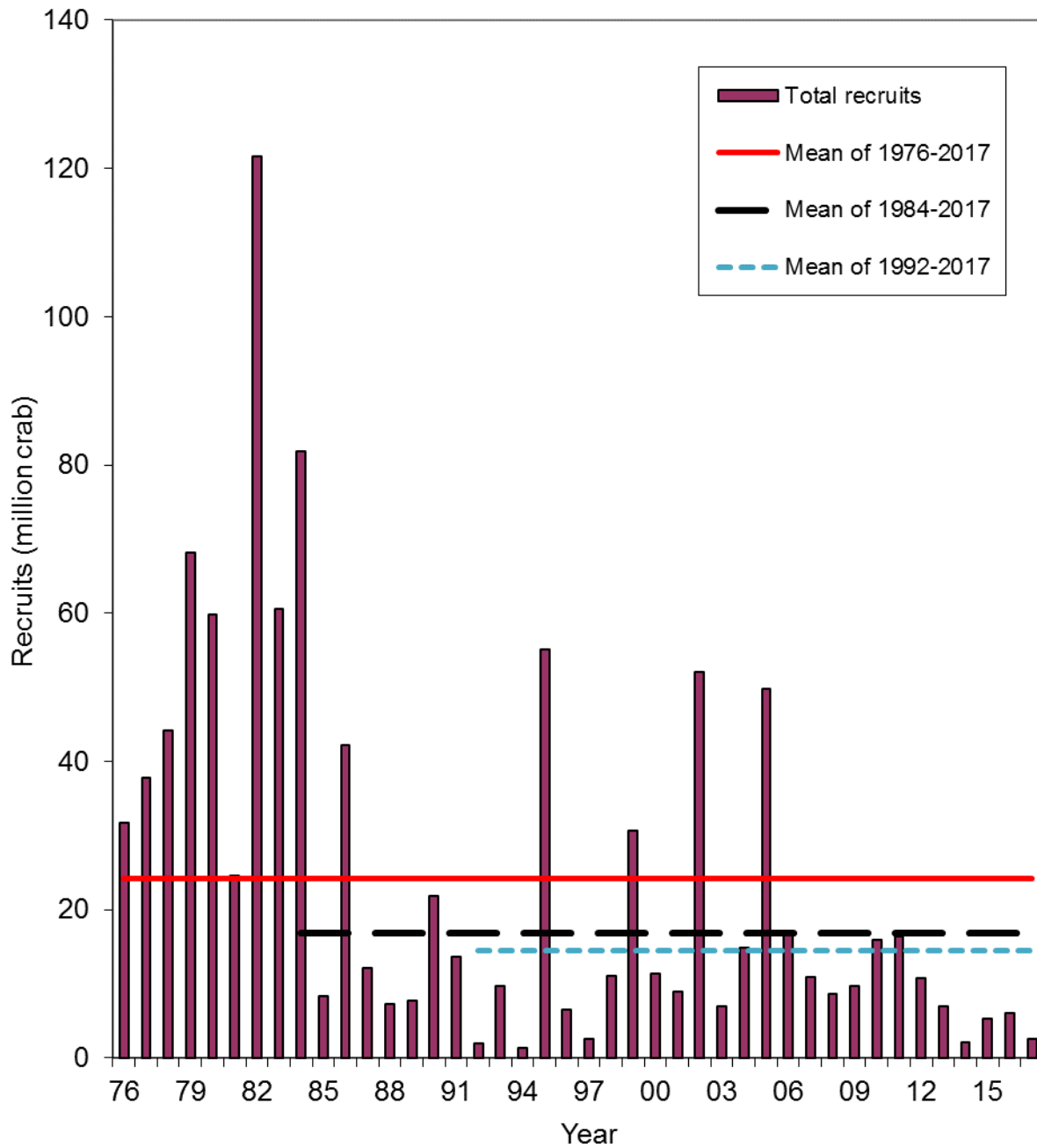


Figure 12(2bn1). Estimated recruitment time series during 1976-2017 with scenario 2bn1. Mean male recruits during 1984-2017 was used to estimate B35%.

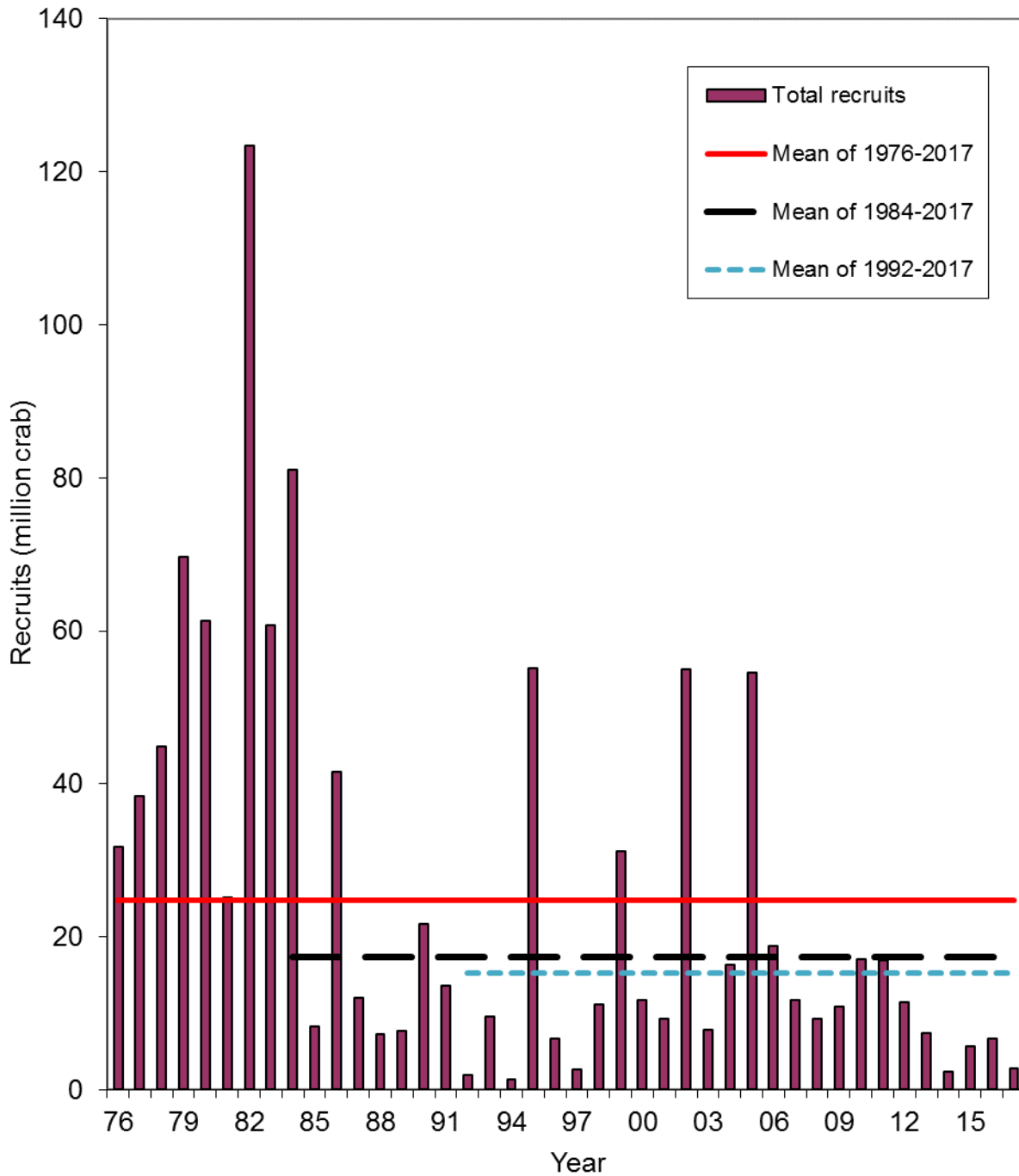


Figure 12(2bn2). Estimated recruitment time series during 1976-2017 with scenario 2bn2. Mean male recruits during 1984-2017 was used to estimate B35%.

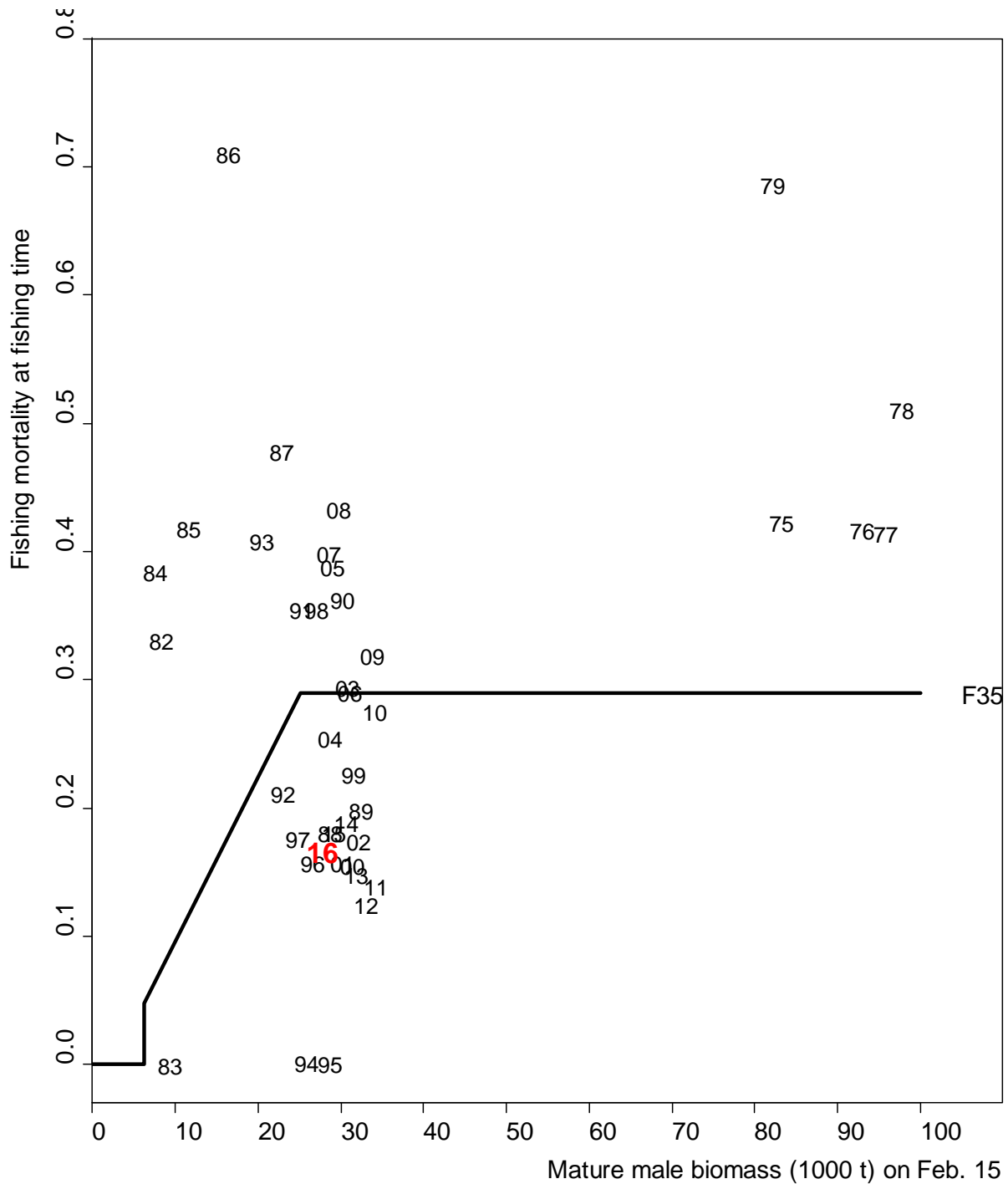


Figure 13(2b). Relationships between full fishing mortalities for the directed pot fishery and mature male biomass on Feb. 15 during 1975-2016 under scenario 2b. Average of recruitment from 1984 to 2017 was used to estimate B_{MSY} . Pot, fixed gear and trawl handling mortality rates were assumed to be 0.2, 0.5 and 0.8, respectively.

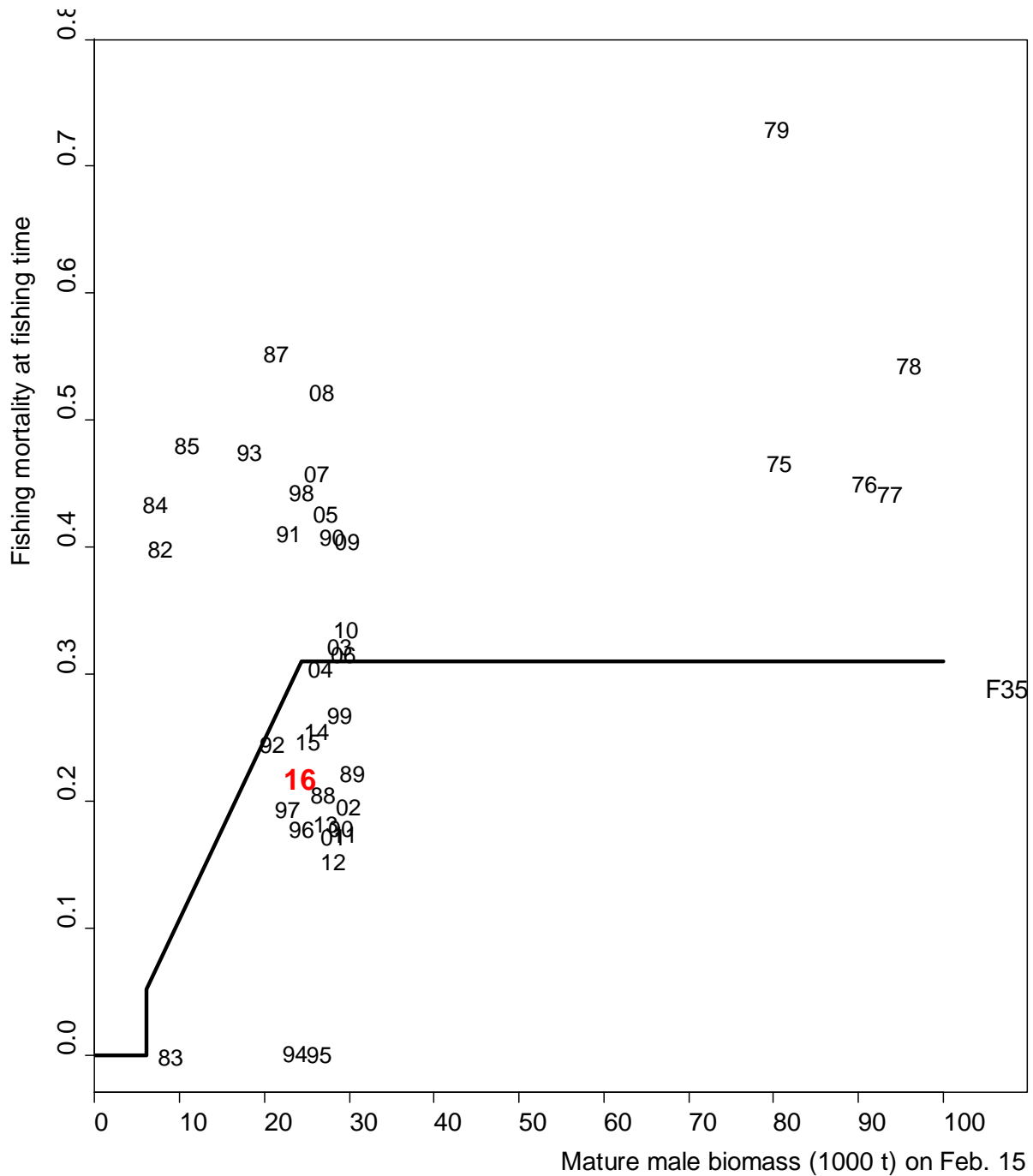


Figure 13(2bn1). Relationships between full fishing mortalities for the directed pot fishery and mature male biomass on Feb. 15 during 1975-2016 under scenario 2bn1. Average of recruitment from 1984 to 2017 was used to estimate BMSY. Pot, fixed gear and trawl handling mortality rates were assumed to be 0.2, 0.5 and 0.8, respectively.

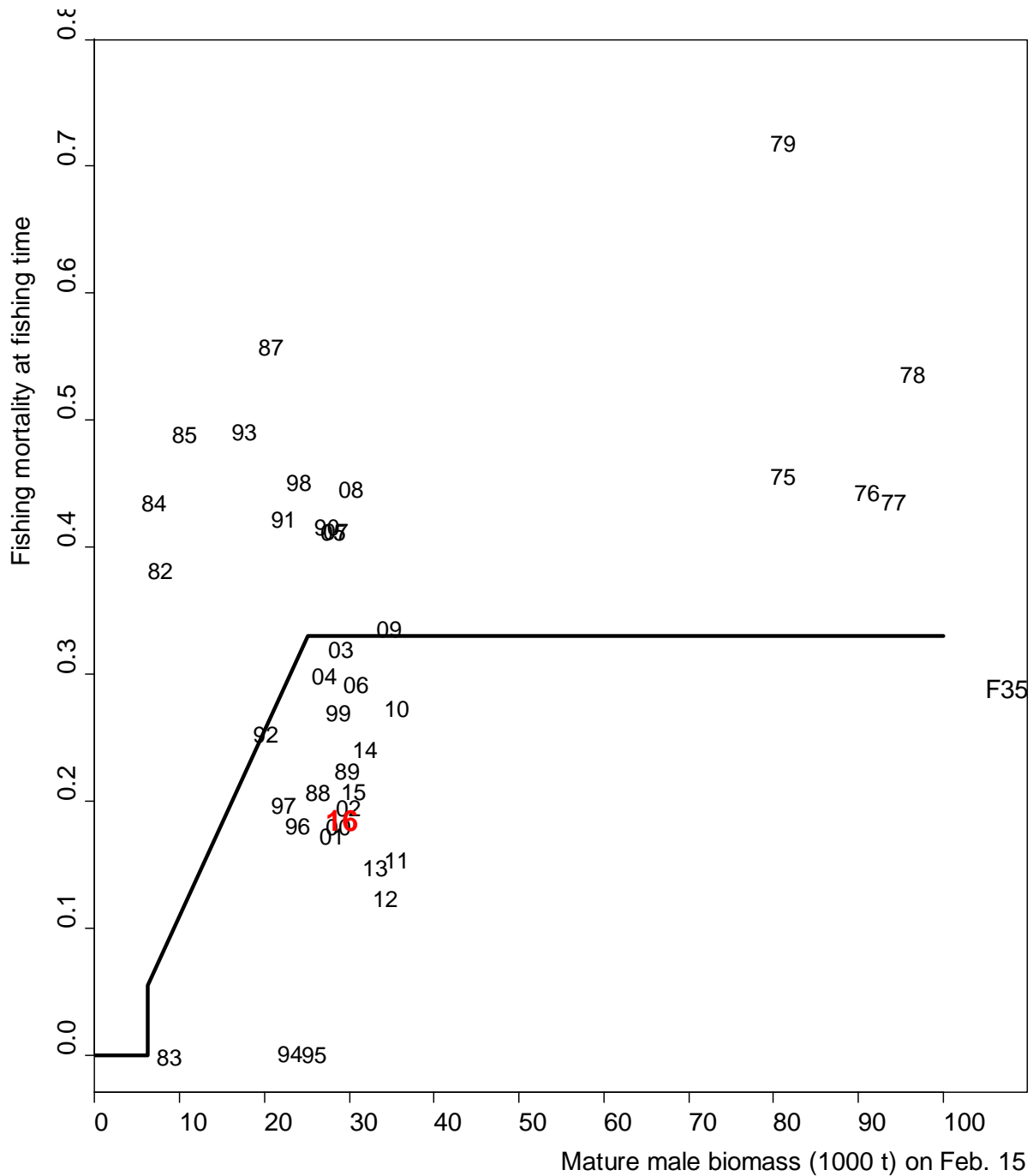


Figure 13(2bn2). Relationships between full fishing mortalities for the directed pot fishery and mature male biomass on Feb. 15 during 1975-2016 under scenario 2bn2. Average of recruitment from 1984 to 2017 was used to estimate BMSY. Pot, fixed gear and trawl handling mortality rates were assumed to be 0.2, 0.5 and 0.8, respectively.

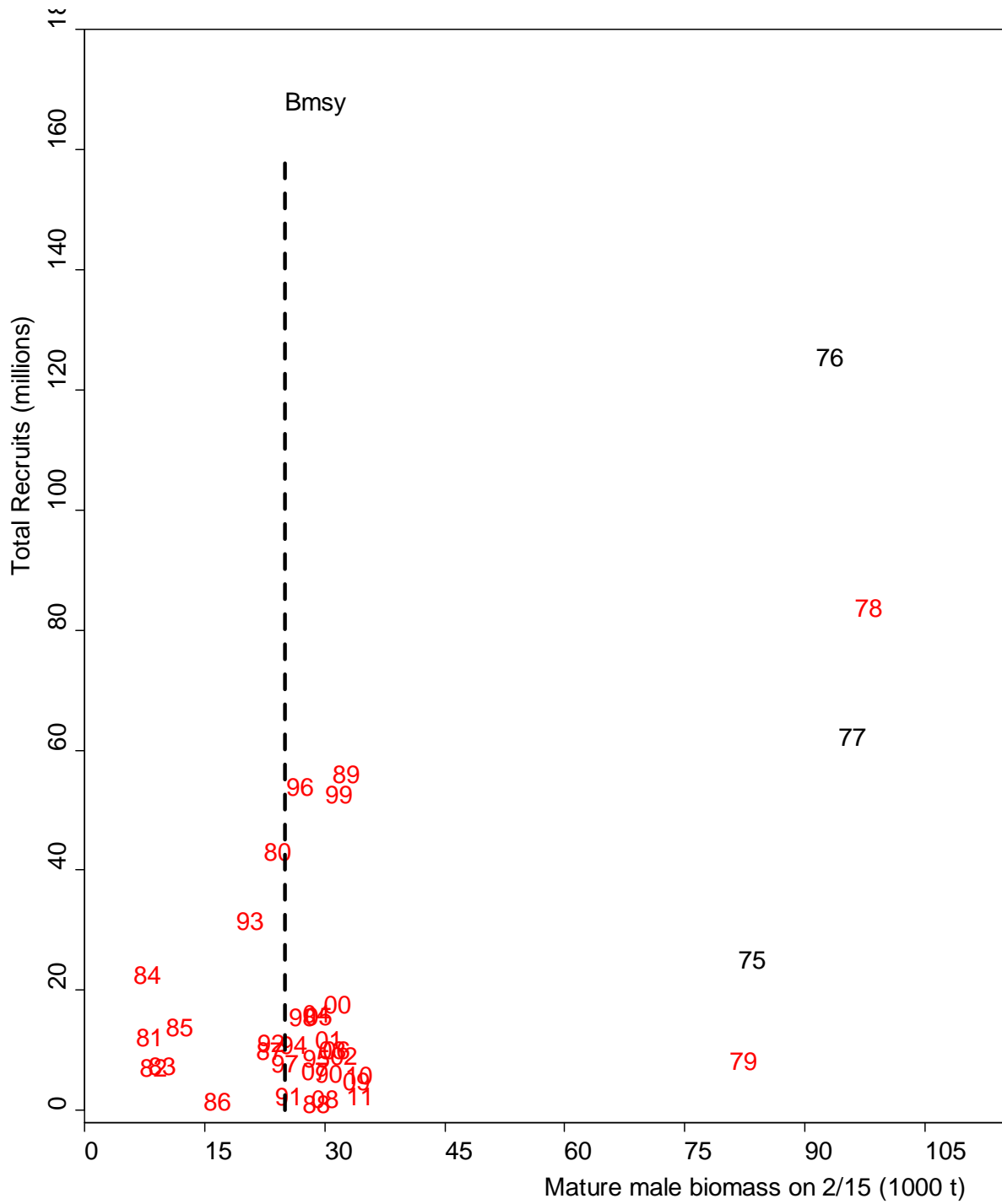


Figure 14a. Relationships between mature male biomass on Feb. 15 and total recruits at age 5 (i.e., 6-year time lag) for Bristol Bay red king crab with pot handling mortality rate of 0.2 under scenario 2b. Numerical labels are years of mating, and the vertical dotted line is the estimated $B_{35\%}$ based on the mean recruitment level during 1984 to 2017.

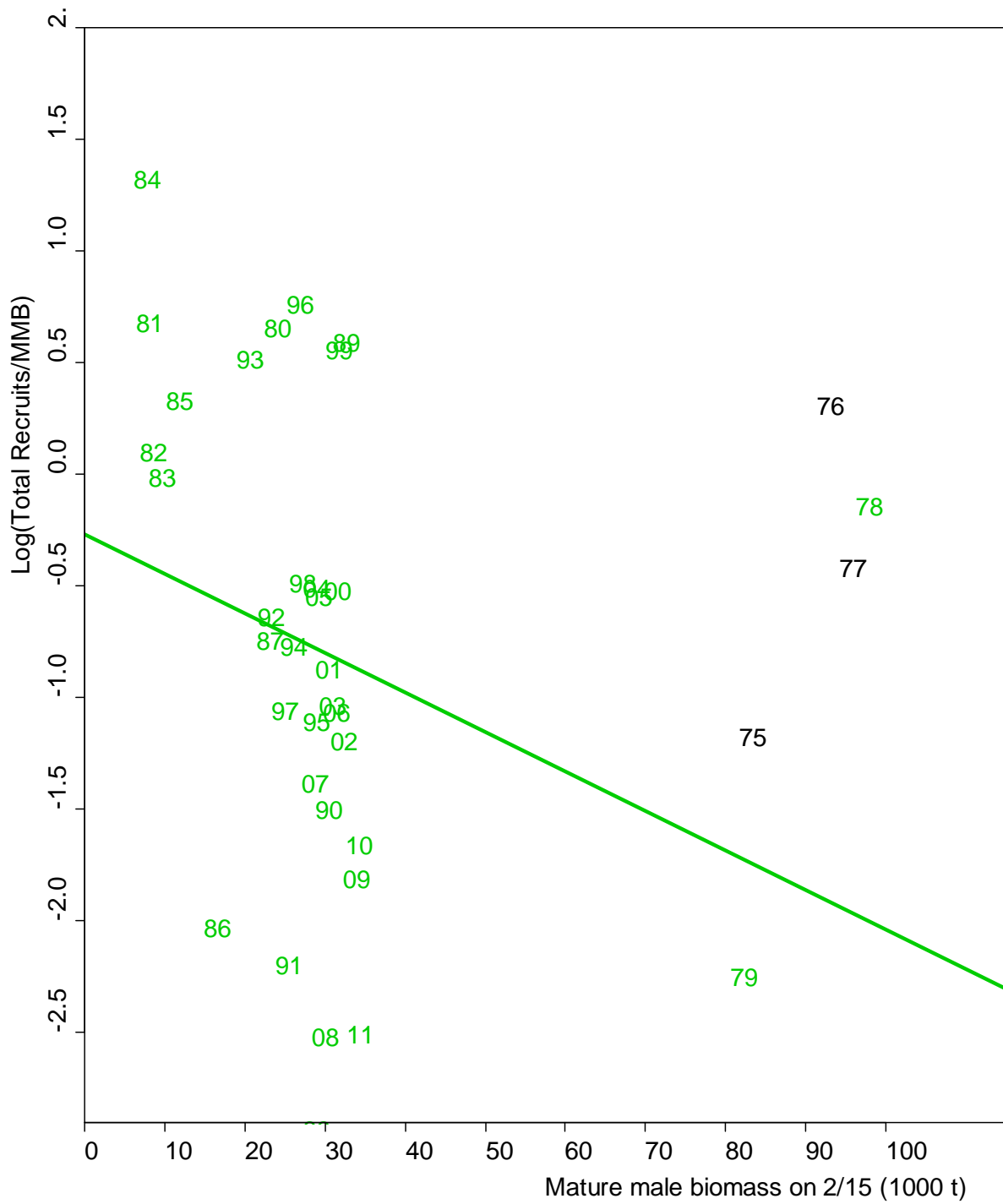


Figure 14b. Relationships between log recruitment per mature male biomass and mature male biomass on Feb. 15 for Bristol Bay red king crab with pot handling mortality rate of 0.2 under scenario 2b. Numerical labels are years of mating, and the line is the regression line for data of 1978-2011.

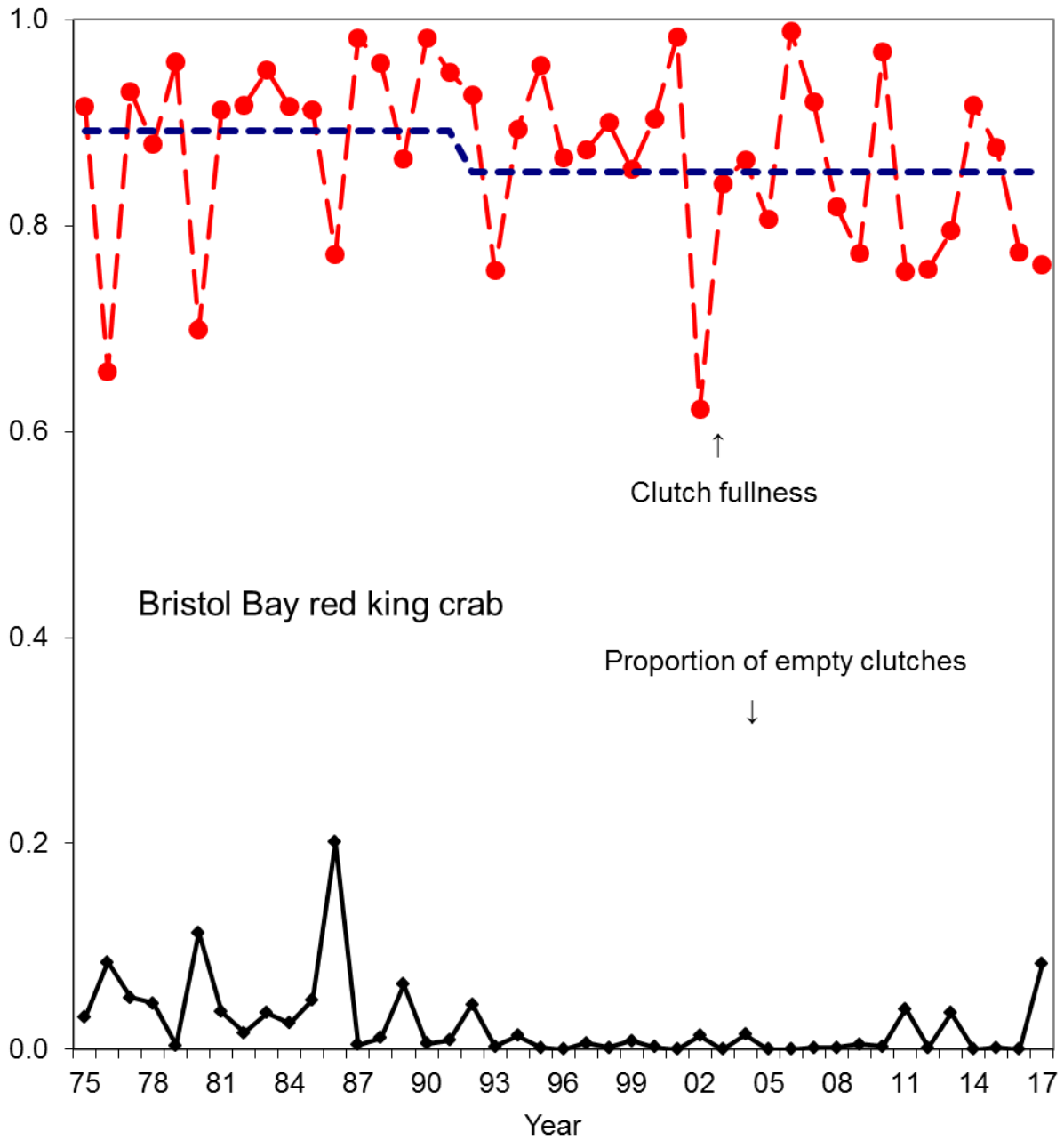


Figure 15. Average clutch fullness and proportion of empty clutches of newshell (shell conditions 1 and 2) mature female crab >89 mm CL from 1975 to 2017 from survey data. Oldshell females were excluded.

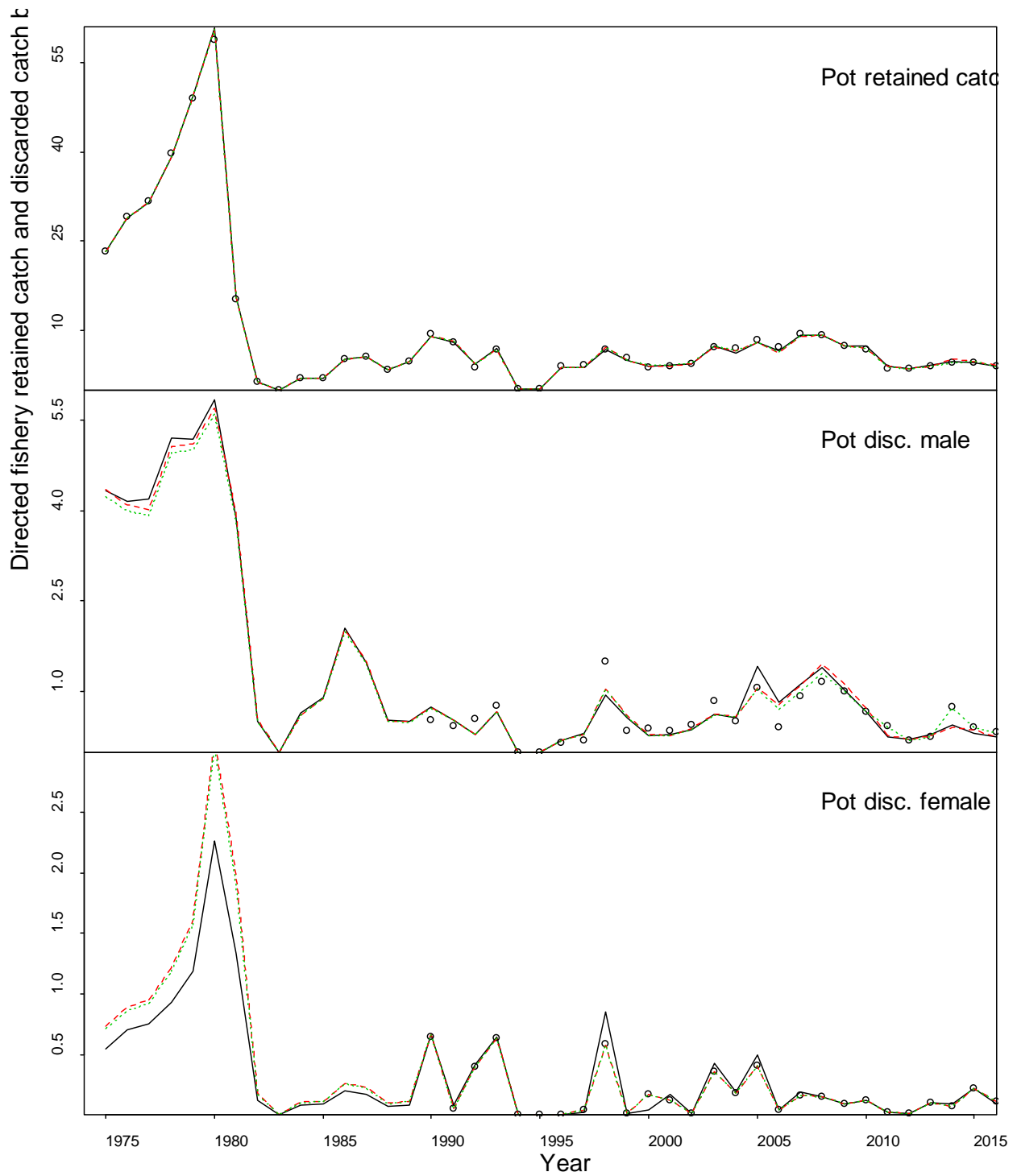


Figure 16a. Observed and predicted catch mortality biomass under scenarios 2b(solid black), 2bn1 (dashed red), and 2bn2 (green lines). Mortality biomass is equal to caught biomass times a handling mortality rate. Pot handling mortality rate is 0.2.

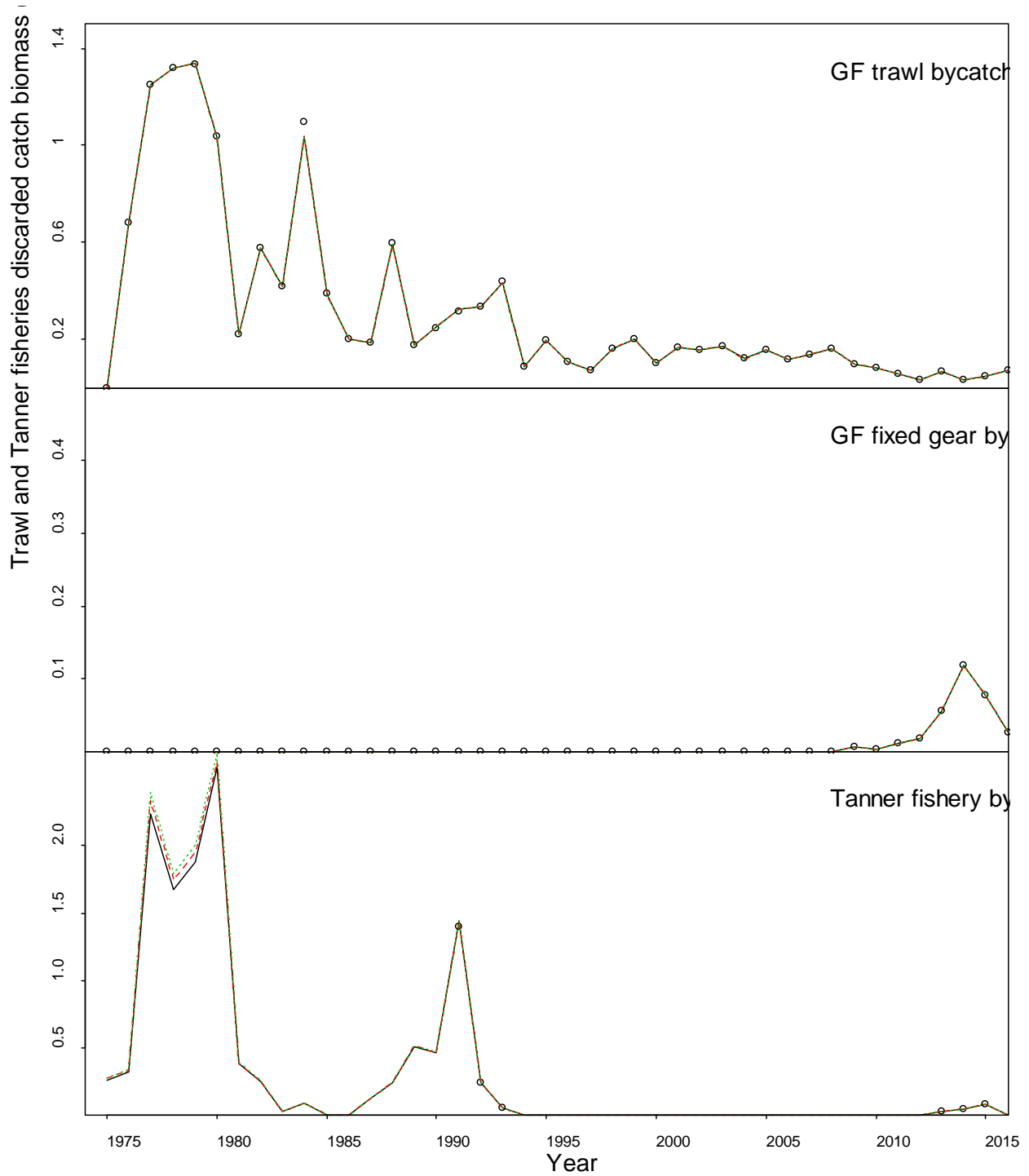


Figure 16b. Observed and predicted bycatch mortality biomass from groundfish fisheries and the Tanner crab fishery under scenarios 2b(solid black), 2bn1(dashed red), and 2bn2(green lines). Mortality biomass is equal to caught biomass times a handling mortality rate. Trawl handling mortality rate is 0.8, fixed gear handling mortality rate is 0.5, and Tanner crab pot handling mortality is 0.25. Trawl bycatch biomass was 0 before 1976.

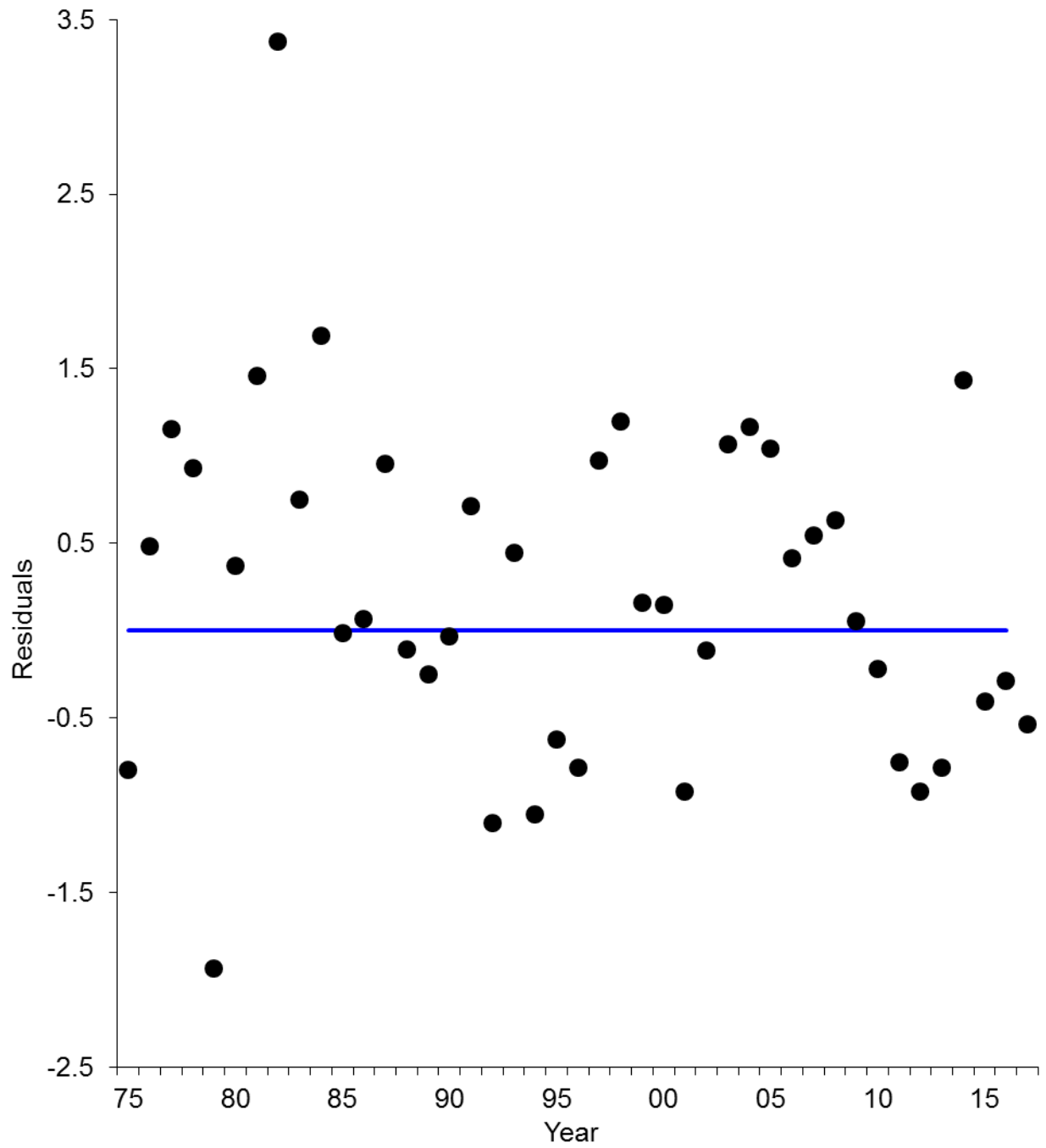


Figure 17(2b). Standardized residuals of total survey biomass under scenario 2b. Pot, fixed gear and trawl handling mortality rates were assumed to be 0.2, 0.5 and 0.8, respectively.

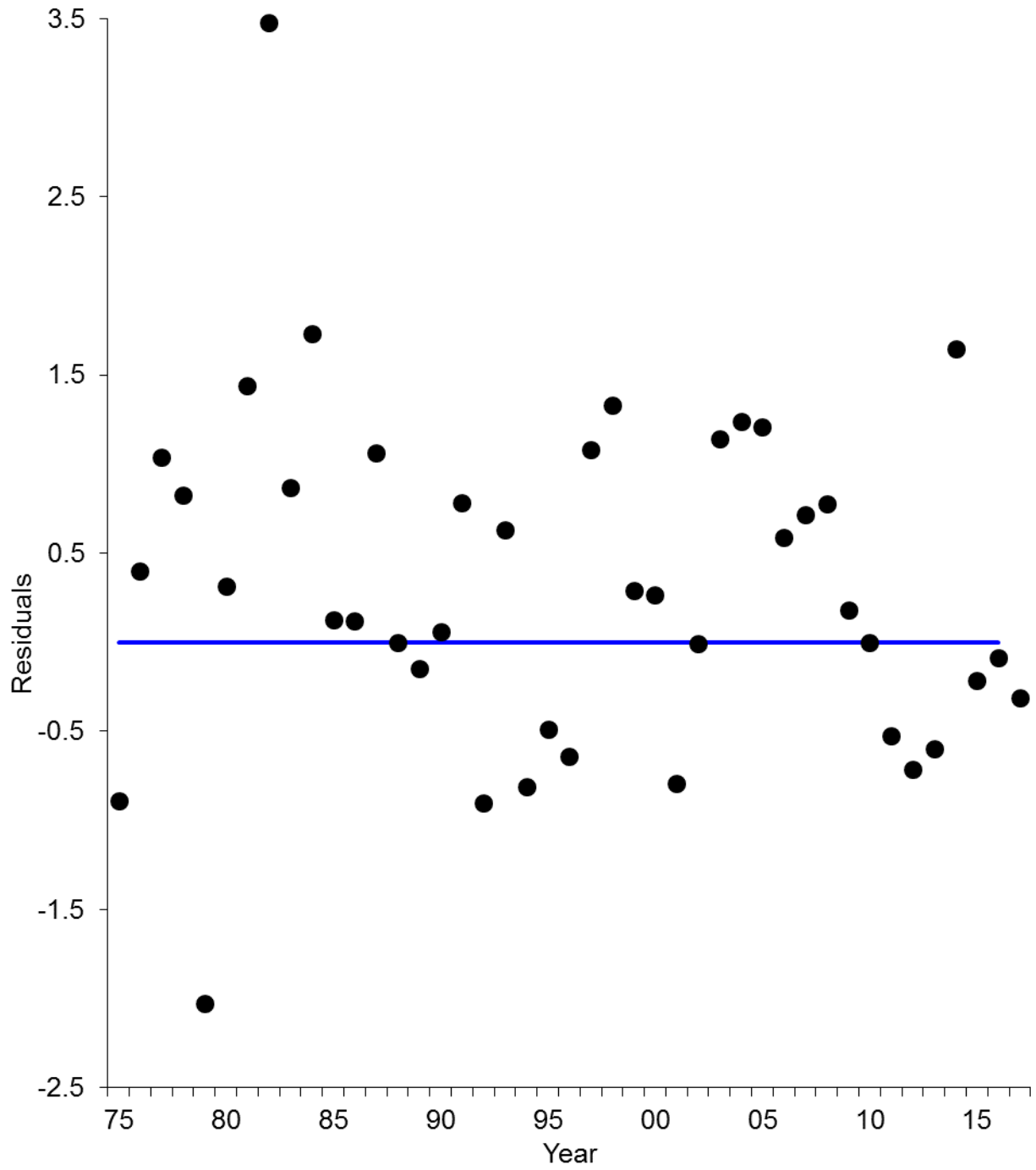


Figure 17(2bn1). Standardized residuals of total survey biomass under scenario 2bn1. Pot, fixed gear and trawl handling mortality rates were assumed to be 0.2, 0.5 and 0.8, respectively.

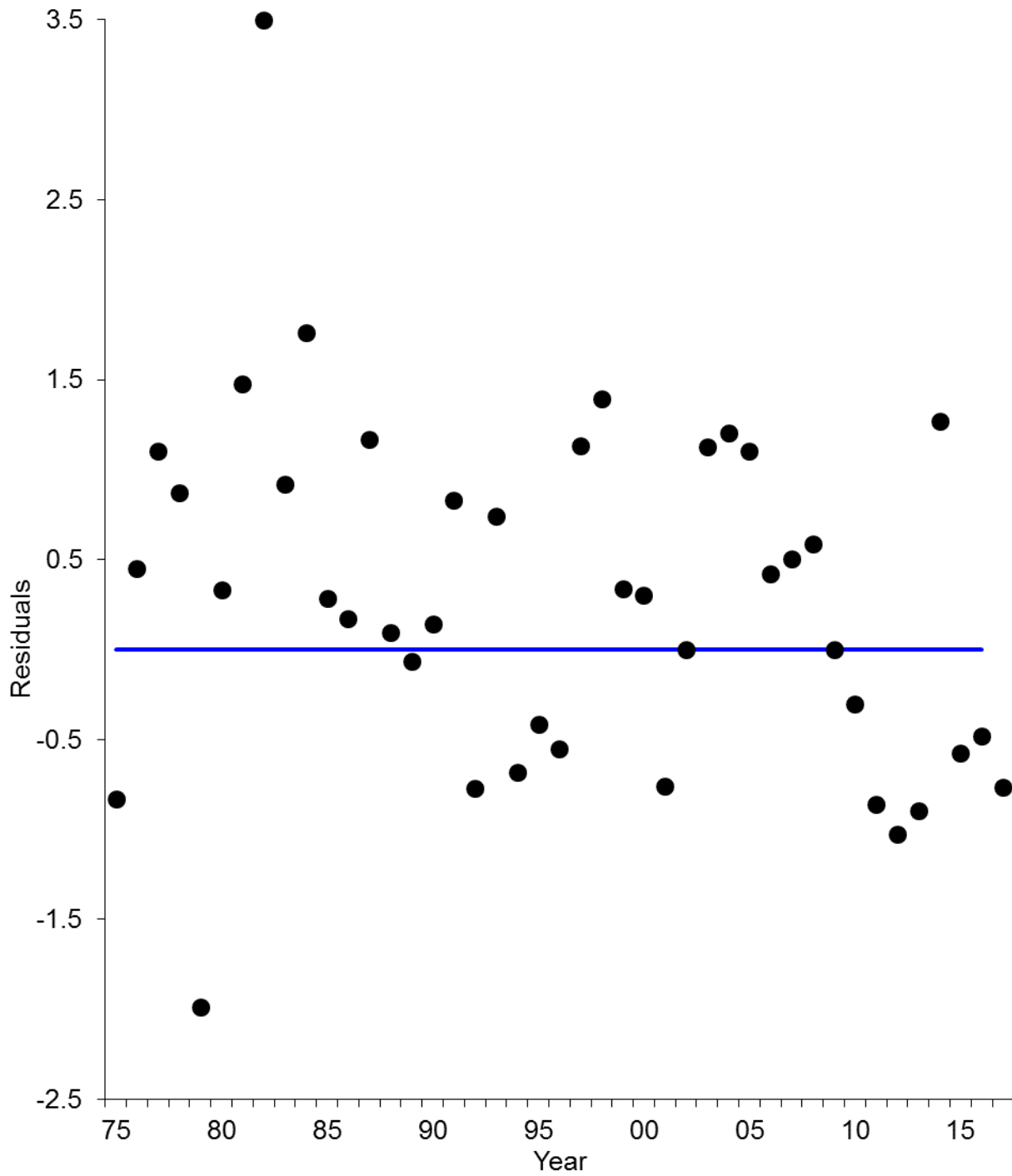


Figure 17(2bn2). Standardized residuals of total survey biomass under scenario 2bn2. Pot, fixed gear and trawl handling mortality rates were assumed to be 0.2, 0.5 and 0.8, respectively.

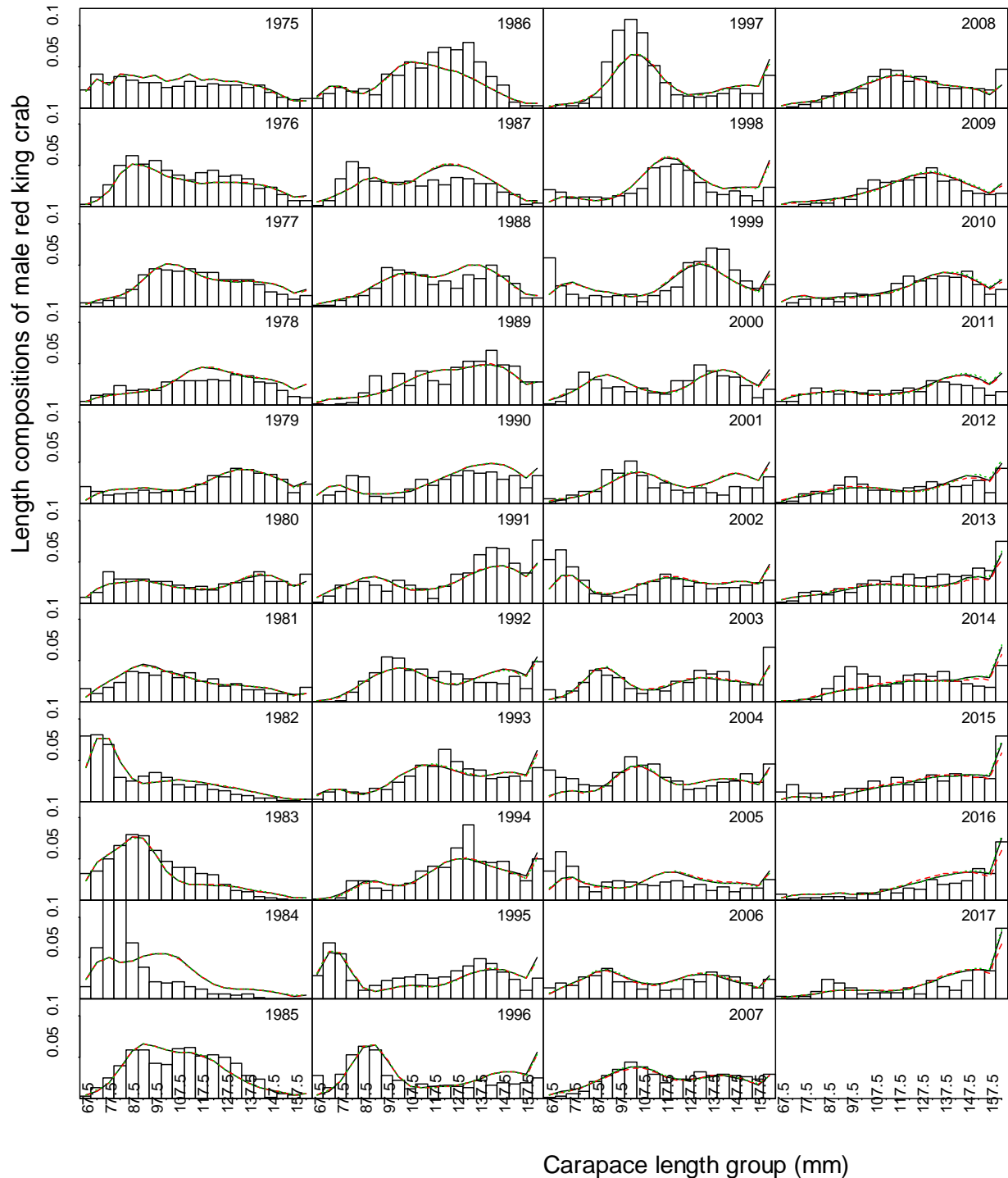


Figure 18(2b, 2bn1 & 2bn2). Comparison of area-swept and model estimated survey length frequencies of Bristol Bay male red king crab by year under scenarios 2b(solid black), 2bn1(dashed red), and 2bn2(green lines). Pot, fixed gear and trawl handling mortality rates were assumed to be 0.2, 0.5 and 0.8.

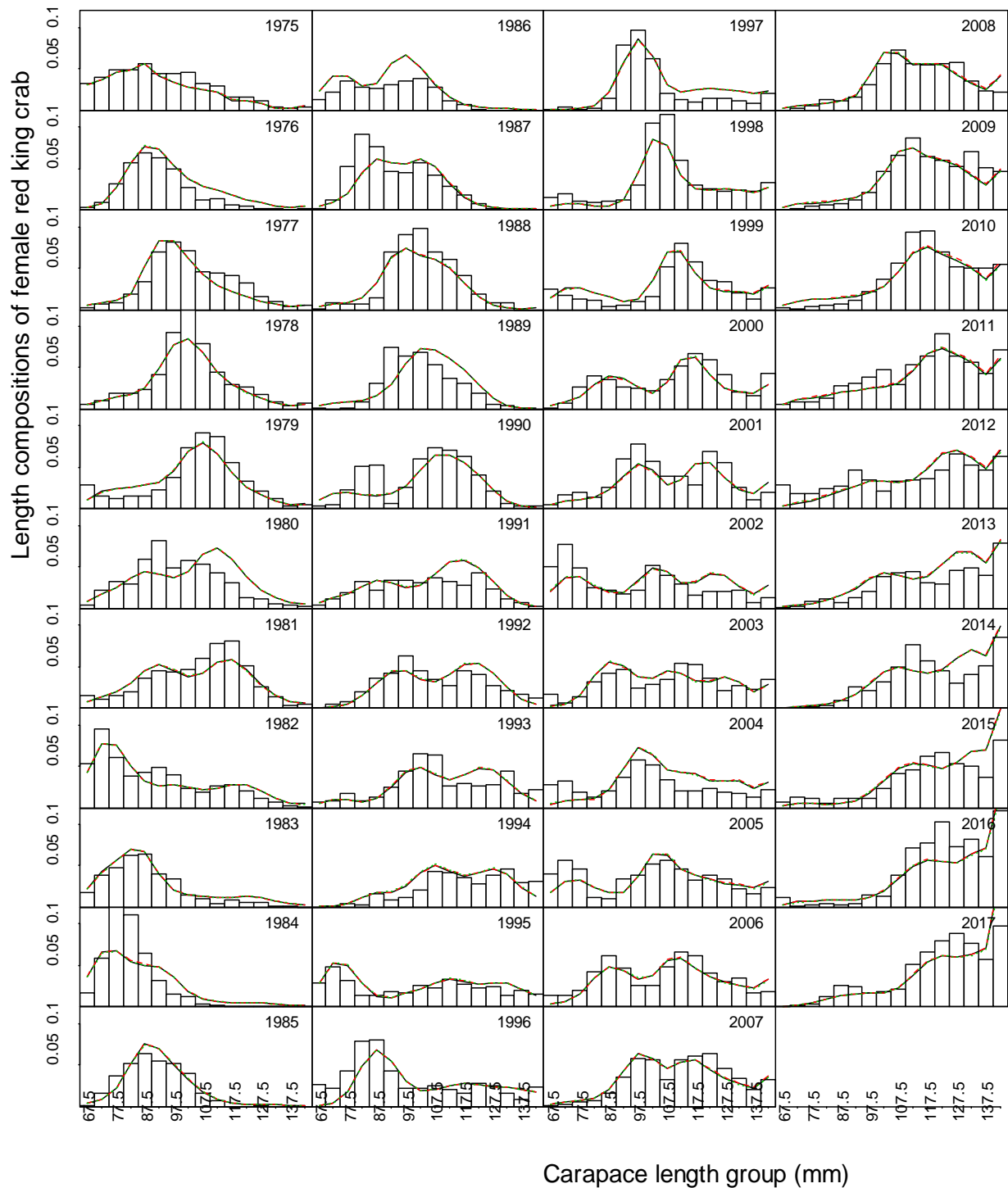


Figure 19(2b, 2bn1 & 2bn2). Comparison of area-swept and model estimated survey length frequencies of Bristol Bay female red king crab by year under scenarios 2b(solid black), 2bn1(dashed red), and 2bn2 (green lines). Pot, fixed gear and trawl handling mortality rates were assumed to be 0.2, 0.5 and 0.8.

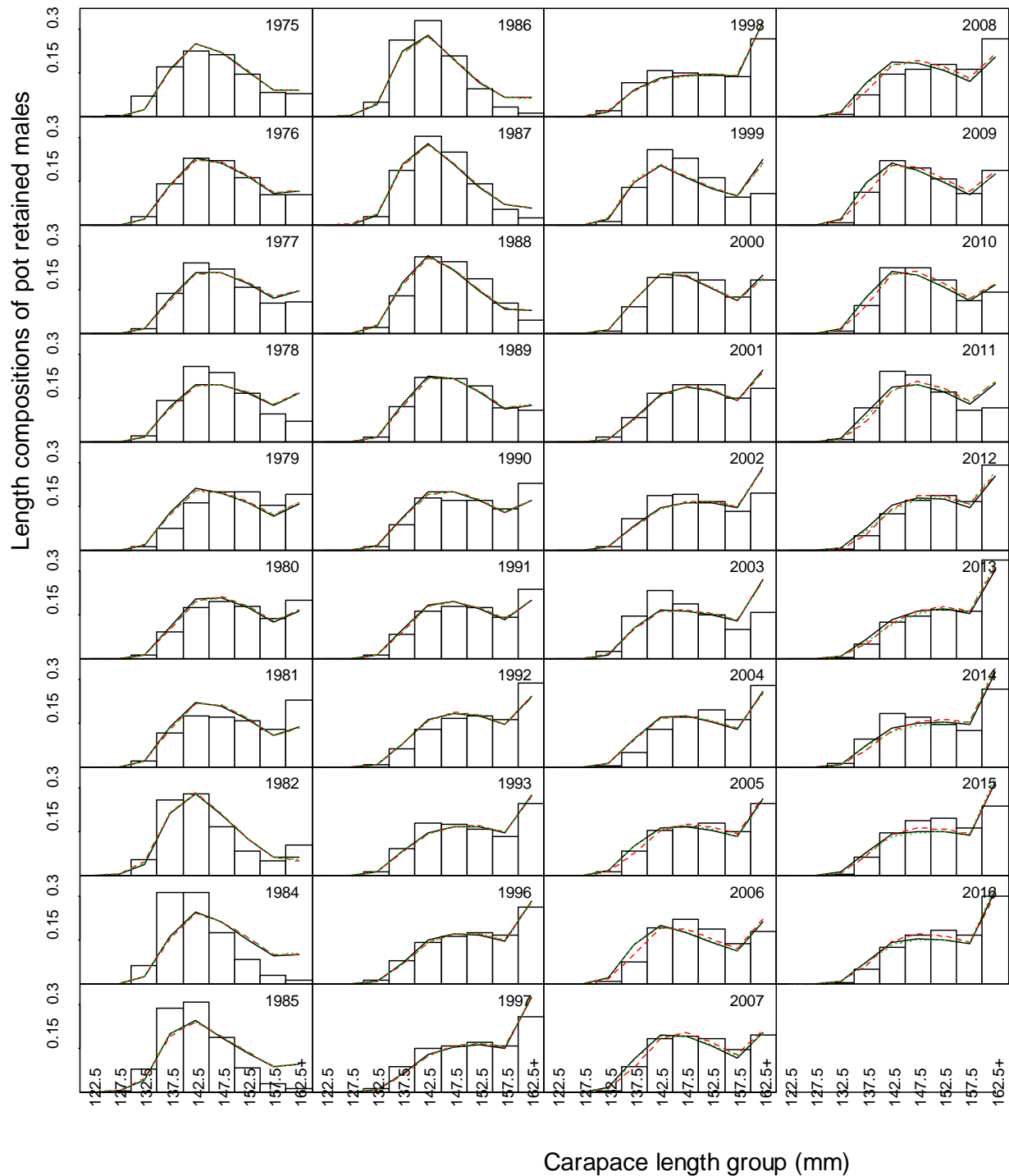


Figure 20(2b, 2bn1 & 2bn2). Comparison of observed and model estimated retained length frequencies of Bristol Bay male red king crab by year in the directed pot fishery under scenarios 2b(solid black), 2bn1(dashed red), and 2bn2(green lines). Pot, fixed gear and trawl handling mortality rates were assumed to be 0.2, 0.5 and 0.8.

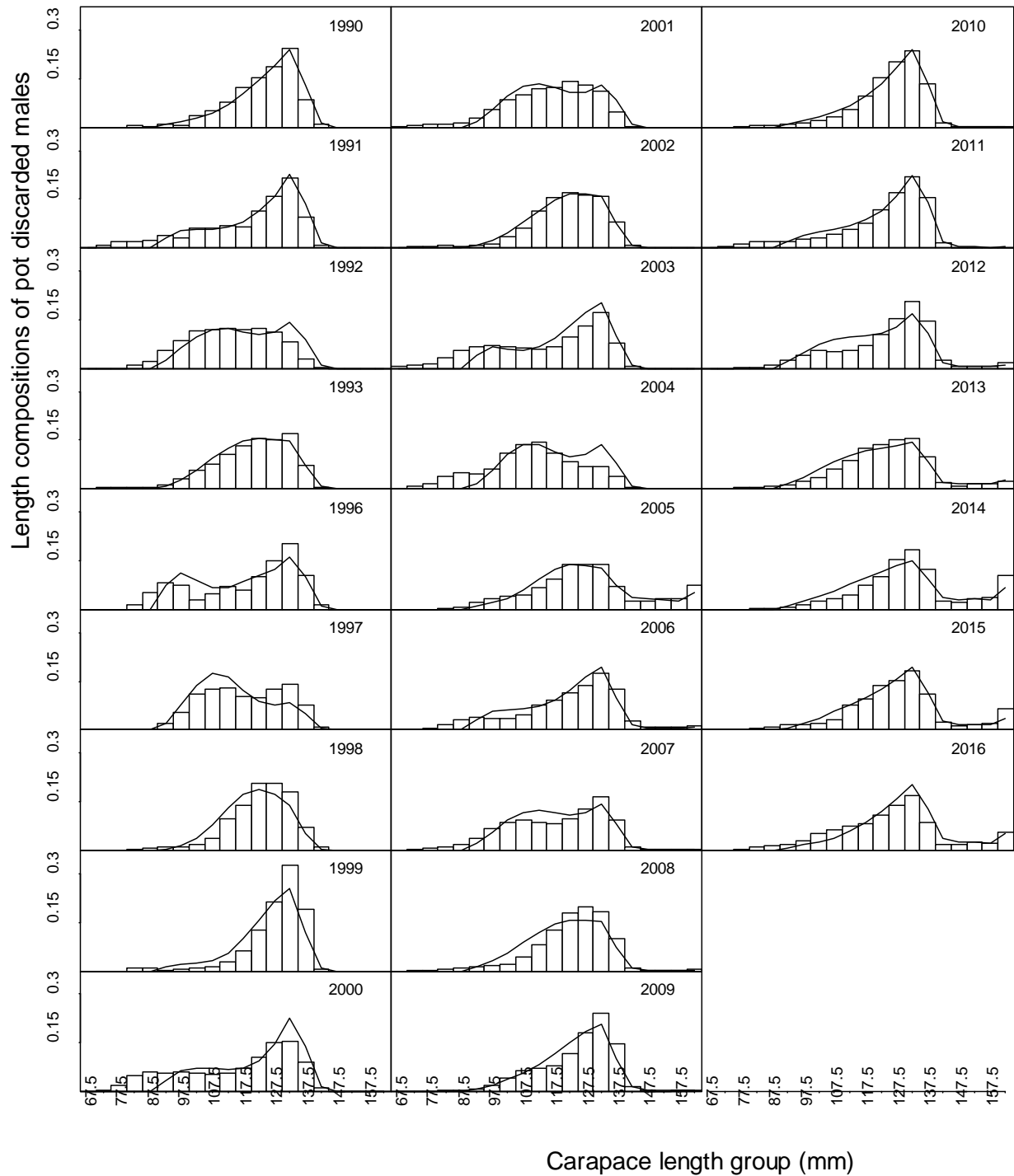


Figure 21(2b). Comparison of observer and model estimated discarded length frequencies of Bristol Bay male red king crab by year in the directed pot fishery under scenario 2b(solid black). Pot, fixed gear and trawl handling mortality rates were assumed to be 0.2, 0.5 and 0.8.

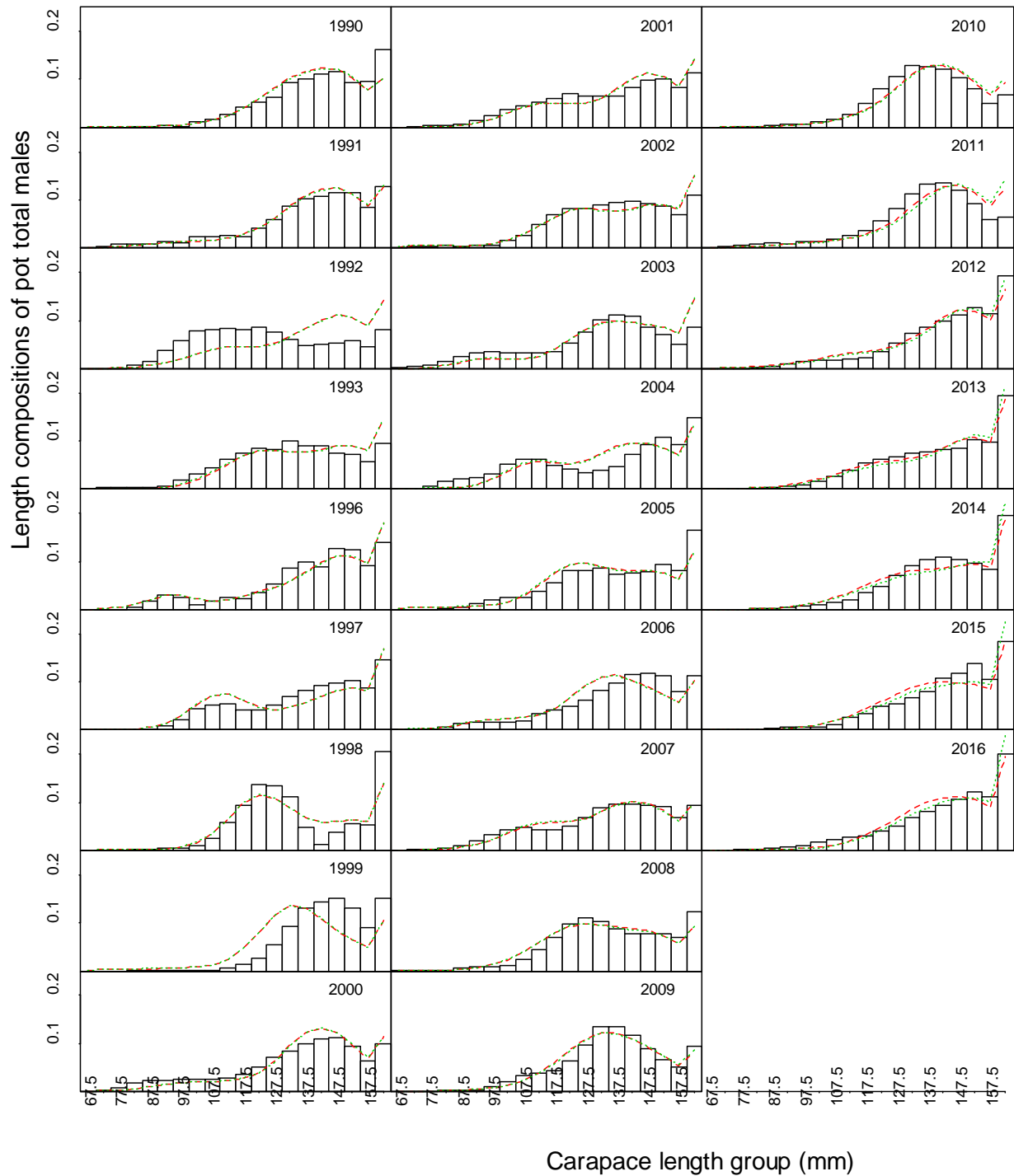


Figure 21(2bn1 & 2bn2). Comparison of observer and model estimated total length frequencies of Bristol Bay male red king crab by year in the directed pot fishery under scenarios 2bn1(dashed red) and 2bn2(green lines). Pot, fixed gear and trawl handling mortality rates were assumed to be 0.2, 0.5 and 0.8.

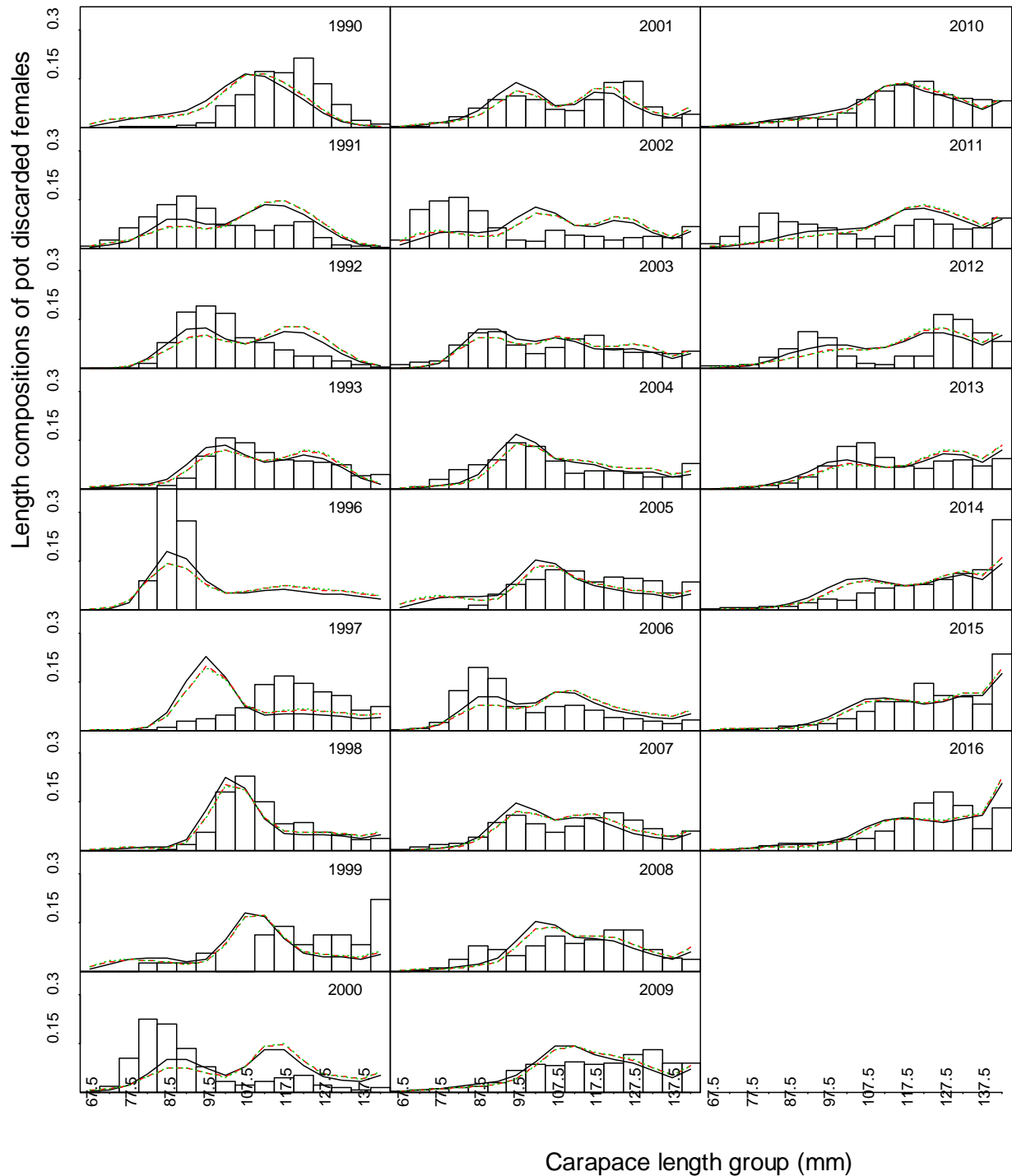


Figure 22(2b, 2bn1 & 2bn2). Comparison of observer and model estimated discarded length frequencies of Bristol Bay female red king crab by year in the directed pot fishery under scenarios 2b(solid black), 2bn1(dashed red), and 2bn2(green lines). Pot, fixed gear and trawl handling mortality rates were assumed to be 0.2, 0.5 and 0.8.

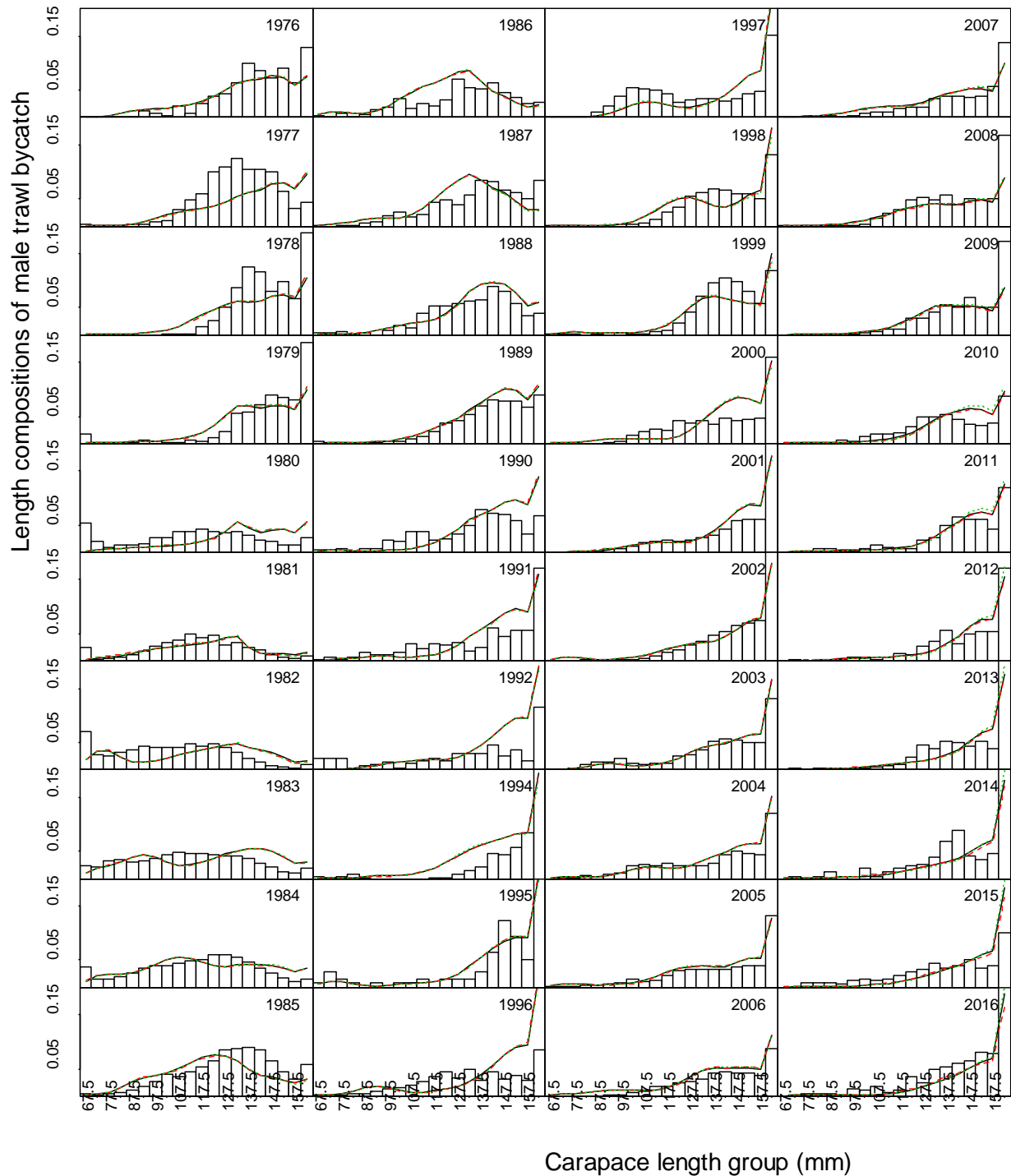


Figure 23(2b, 2bn1 & 2bn2). Comparison of observer and model estimated discarded length frequencies of Bristol Bay male red king crab by year in the groundfish trawl fisheries under scenarios 2b(solid black), 2bn1(dashed red), and 2bn2(green lines). Pot, fixed gear and trawl handling mortality rates were assumed to be 0.2, 0.5 and 0.8.

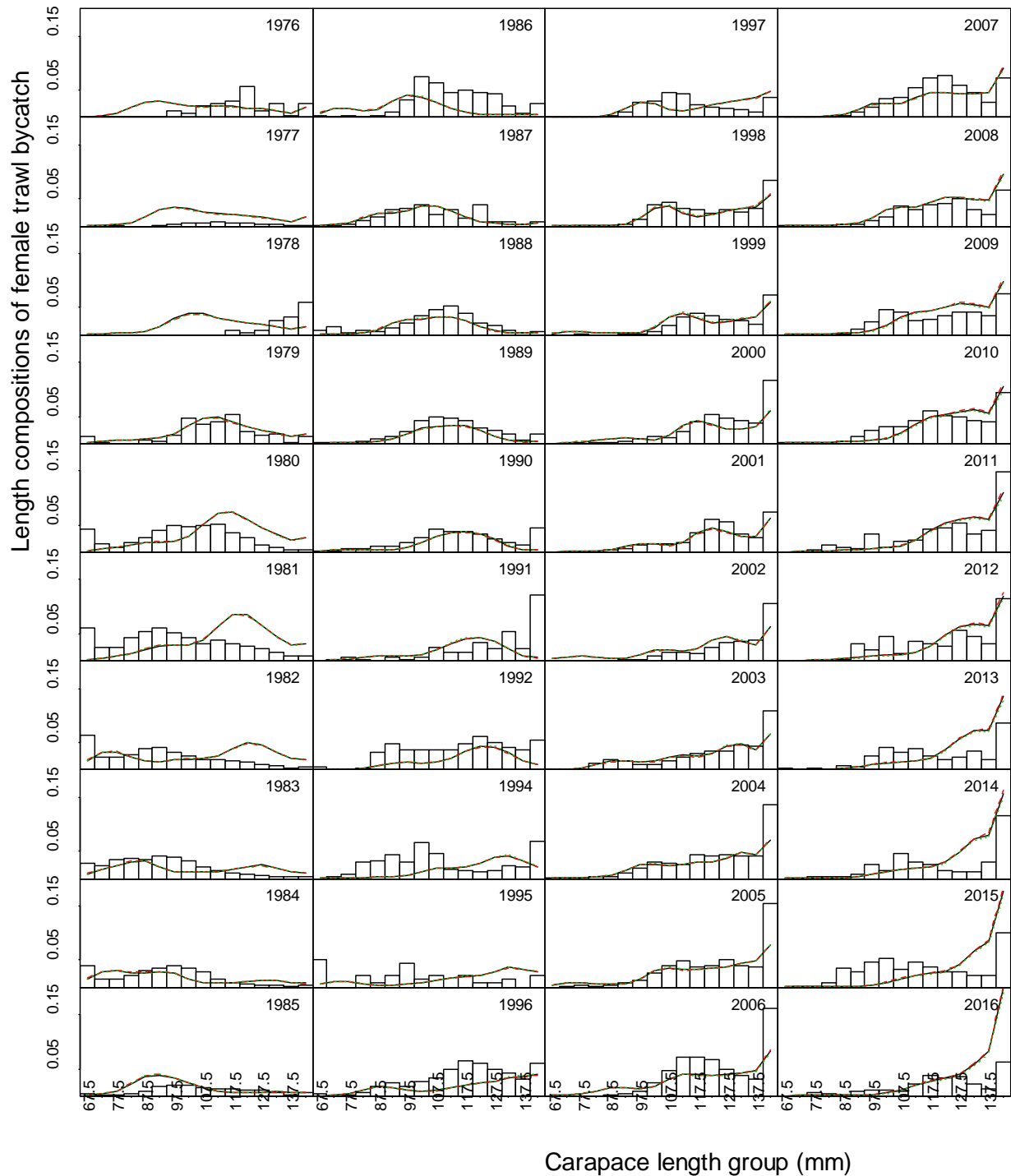


Figure 23(2b, 2bn1 & 2bn2). Comparison of observer and model estimated discarded length frequencies of Bristol Bay female red king crab by year in the groundfish trawl fisheries under scenarios 2b(solid black), 2bn1(dashed red), and 2bn2(green lines). Pot, fixed gear and trawl handling mortality rates were assumed to be 0.2, 0.5 and 0.8.

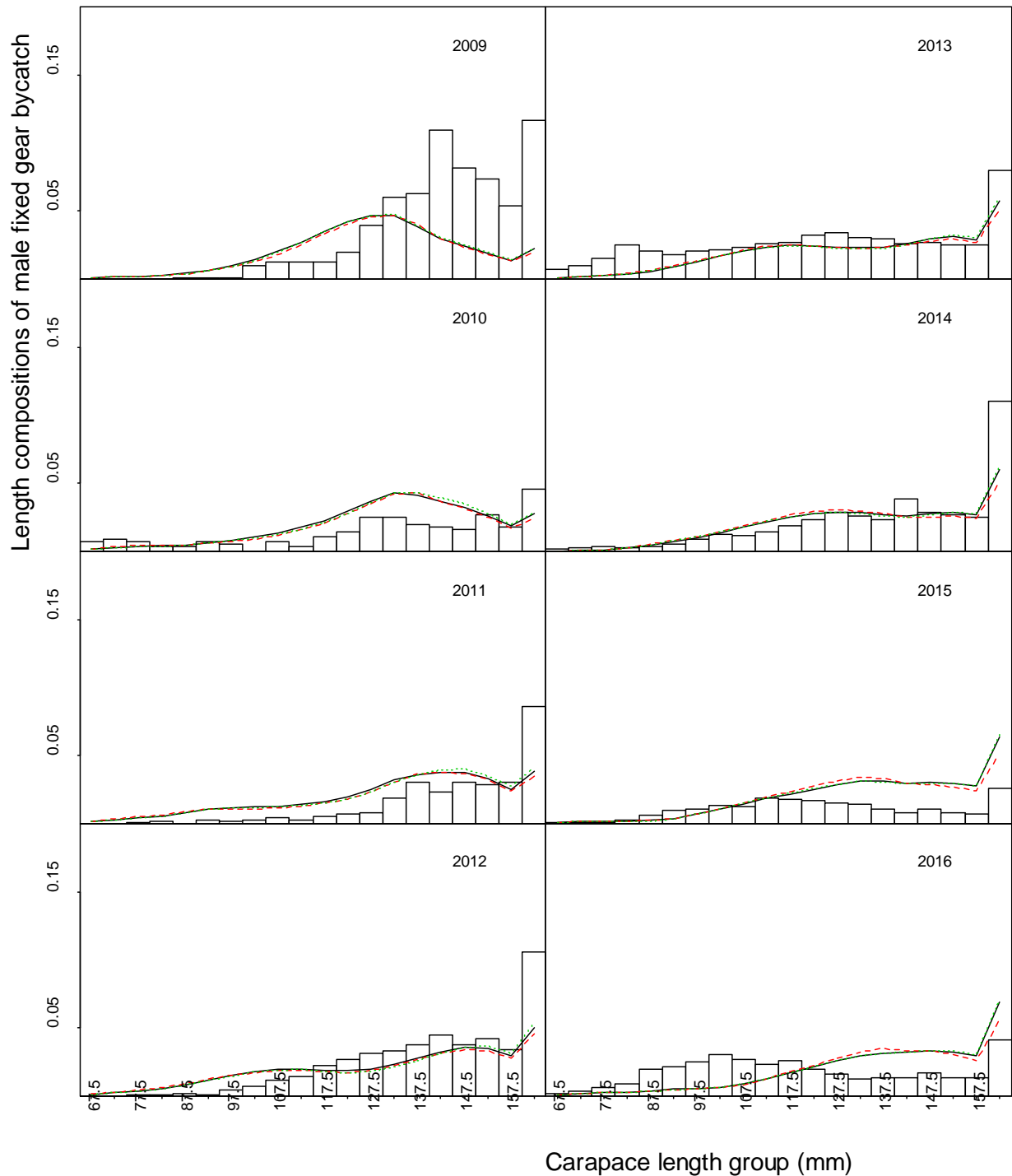


Figure 24(2b, 2bn1 & 2bn2). Comparison of observer and model estimated discarded length frequencies of Bristol Bay male red king crab by year in the groundfish fixed gear fisheries under scenarios 2b(solid black), 2bn1(dashed red), and 2bn2(green lines). Pot, fixed gear and trawl handling mortality rates were assumed to be 0.2, 0.5 and 0.8.

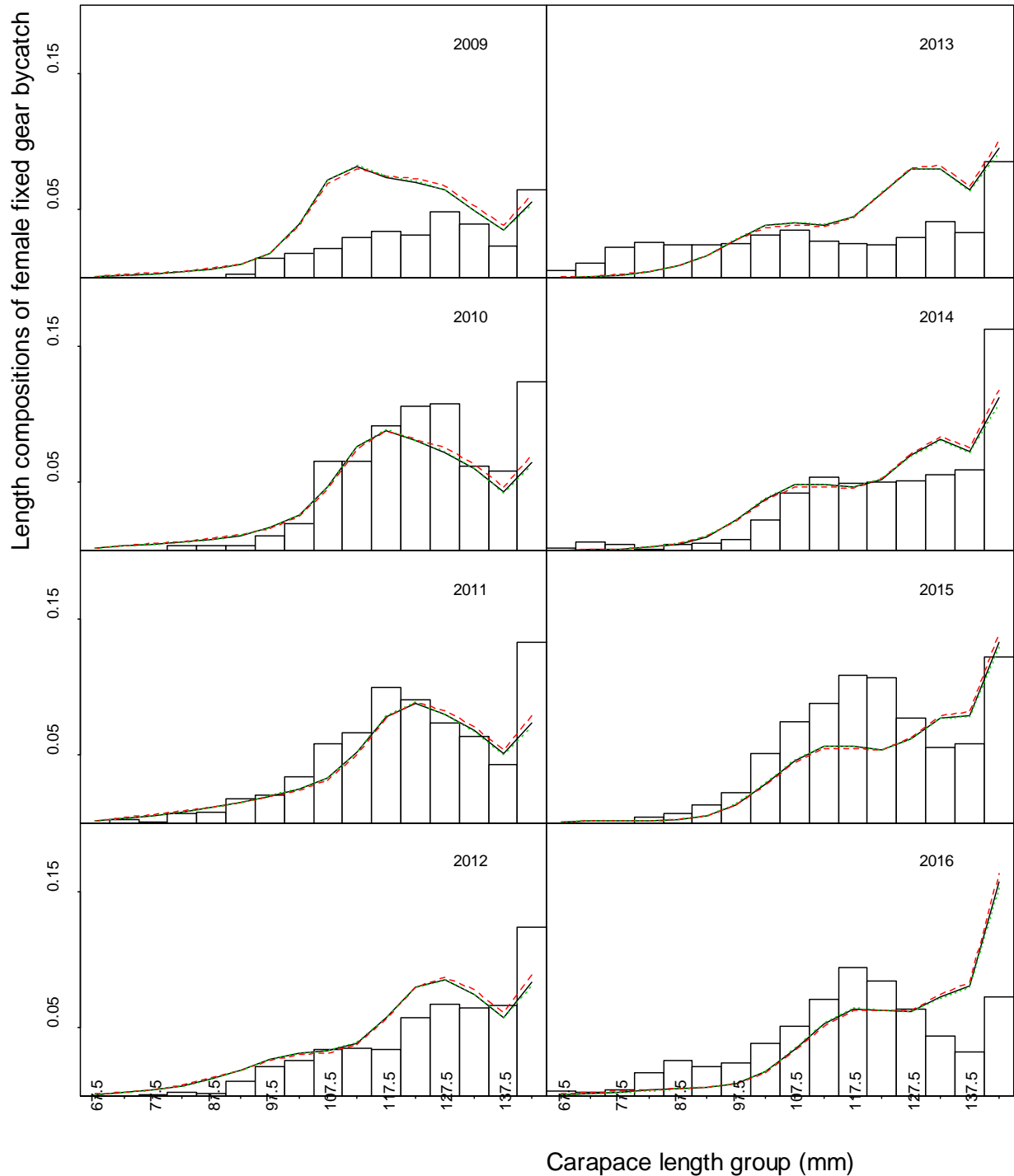


Figure 24(2b, 2bn1 & 2bn2). Comparison of observer and model estimated discarded length frequencies of Bristol Bay female red king crab by year in the groundfish fixed gear fisheries under scenarios 2b(solid black), 2bn1(dashed red), and 2bn2(green lines). Pot, fixed gear and trawl handling mortality rates were assumed to be 0.2, 0.5 and 0.8.

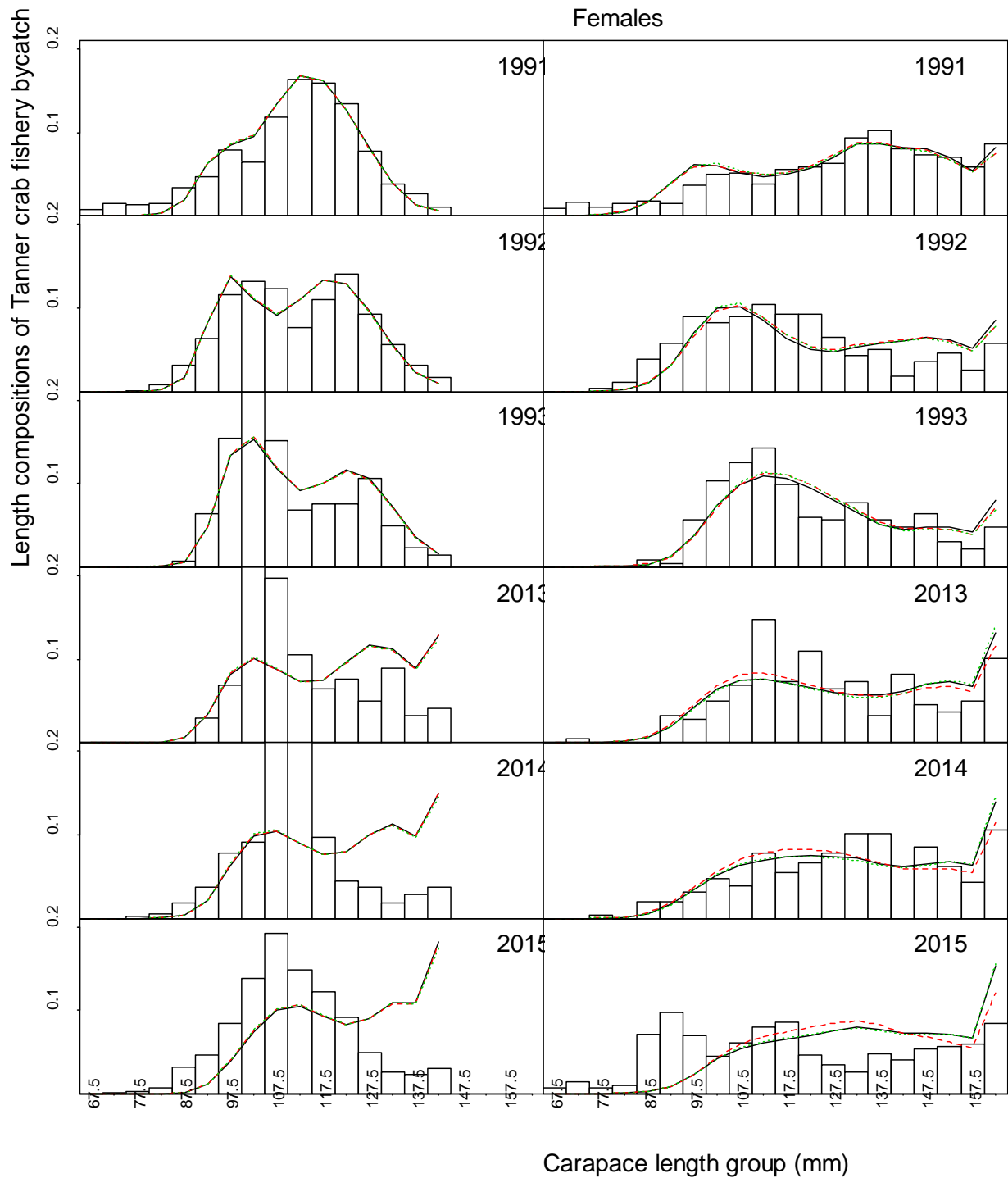


Figure 24(2b, 2bn1 & 2bn2). Comparison of observer and model estimated discarded length frequencies of Bristol Bay red king crab by year in the Tanner crab fishery under scenarios 2b(solid black), 2bn1(dashed red), and 2bn2(green lines). Pot, fixed gear and trawl handling mortality rates were assumed to be 0.2, 0.5 and 0.8.

Scenario 2b, Trawl Survey Males

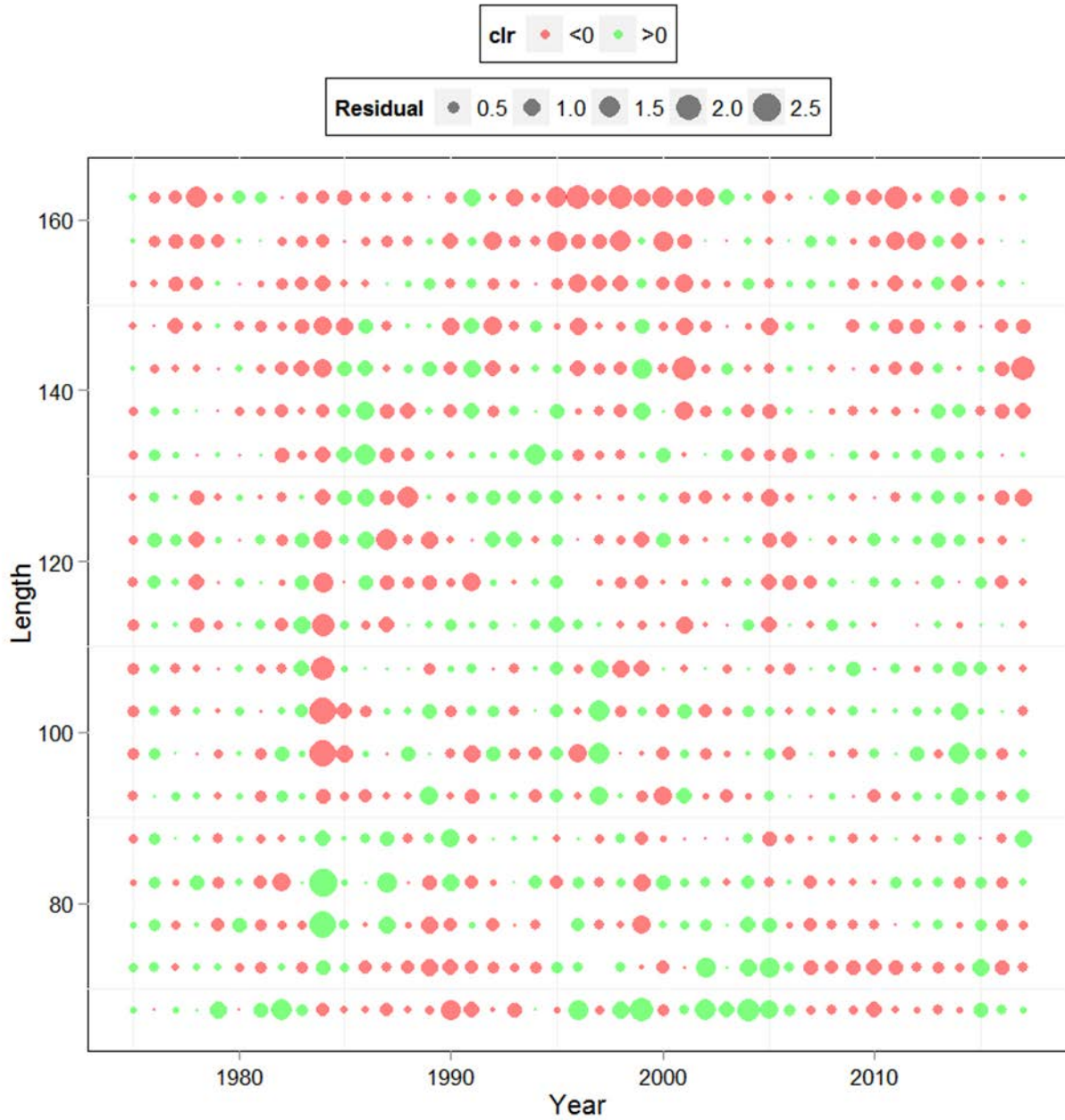


Figure 25(2b). Standardized residuals of proportions of survey male red king crab by year and carapace length (mm) under scenario 2b. Green circles are positive residuals, and red circles are negative residuals. Pot, fixed gear and trawl handling mortality rates were assumed to be 0.2, 0.5 and 0.8, respectively.

Scenario 2bn1, Trawl Survey Males

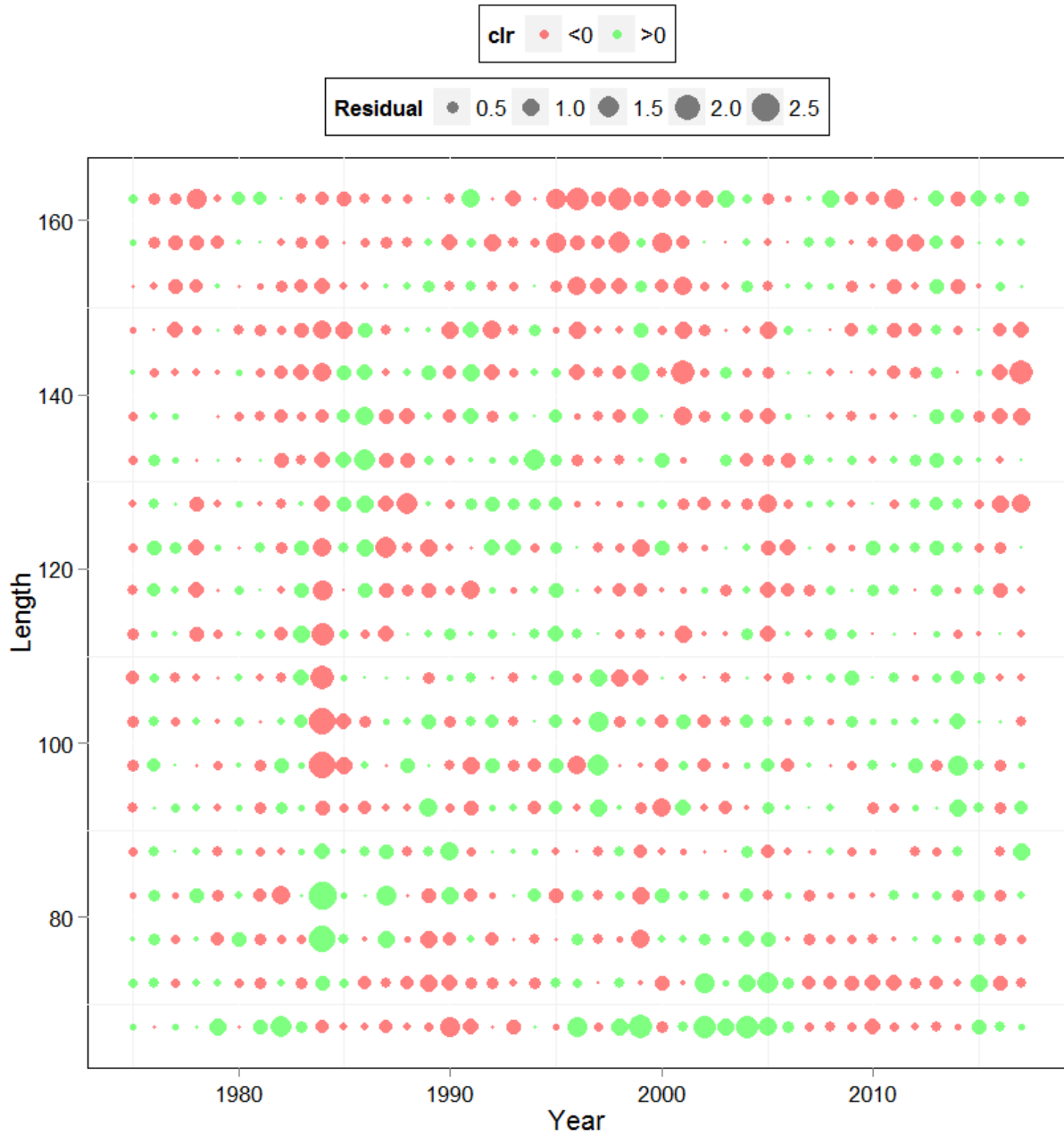


Figure 25(2bn1). Standardized residuals of proportions of survey male red king crab by year and carapace length (mm) under scenario 2bn1. Green circles are positive residuals, and red circles are negative residuals. Pot, fixed gear and trawl handling mortality rates were assumed to be 0.2, 0.5 and 0.8, respectively.

Scenario 2bn2, Trawl Survey Males

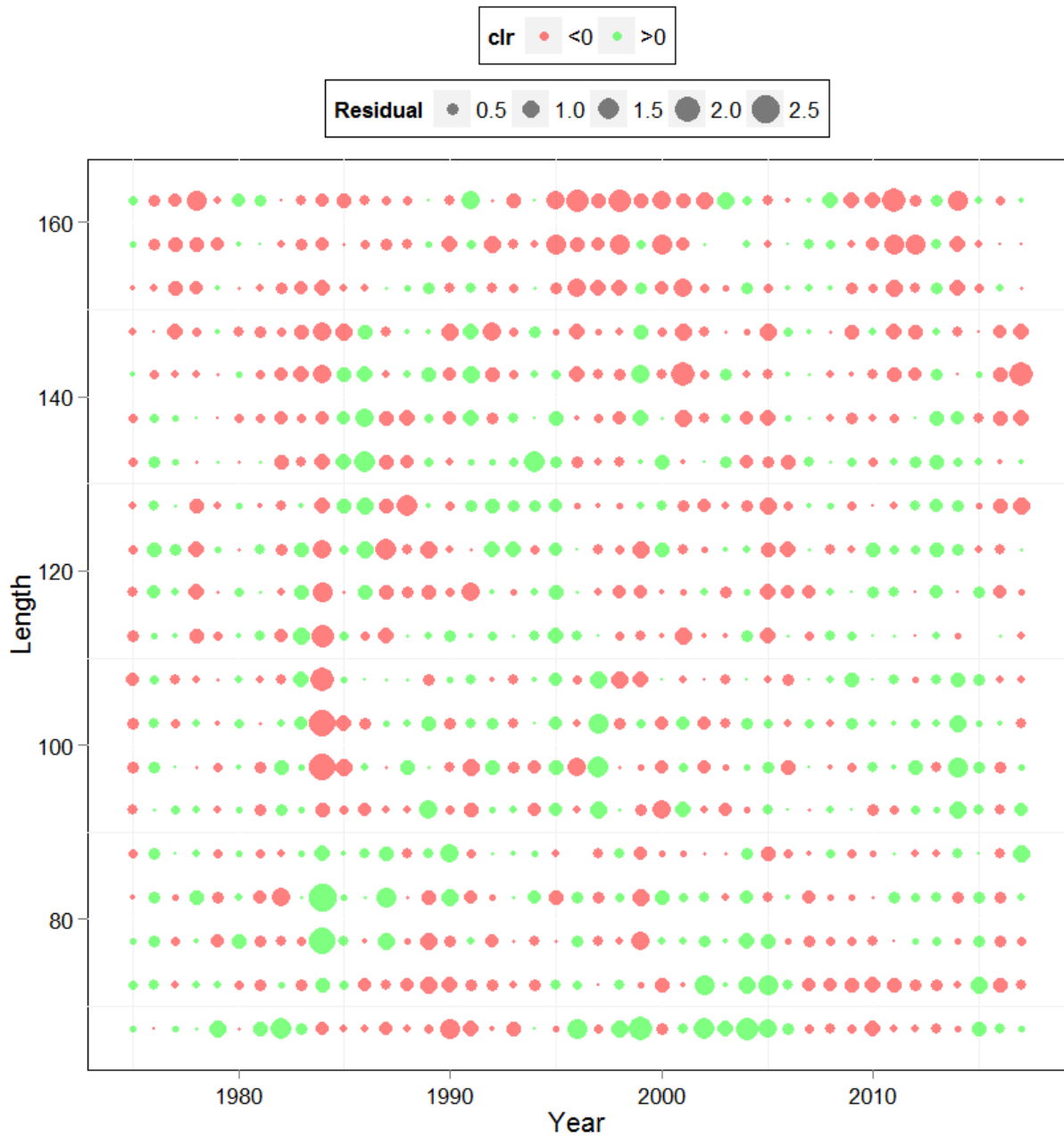


Figure 25(2bn2). Standardized residuals of proportions of survey male red king crab by year and carapace length (mm) under scenario 2bn2. Green circles are positive residuals, and red circles are negative residuals. Pot, fixed gear and trawl handling mortality rates were assumed to be 0.2, 0.5 and 0.8, respectively.

Scenario 2b, Trawl Survey Females

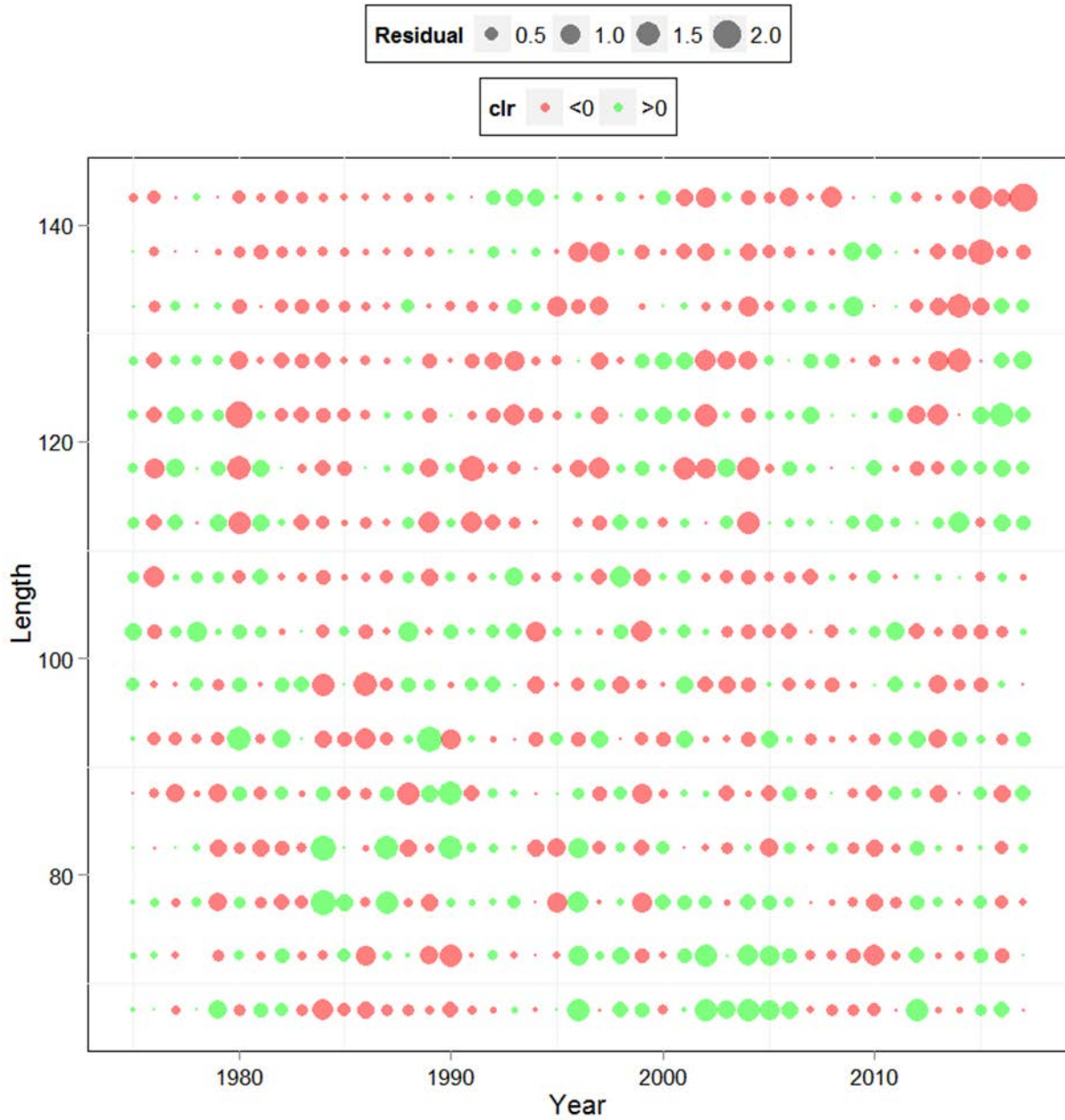


Figure 26(2b). Standardized residuals of proportions of survey female red king crab by year and carapace length (mm) under scenario 2b. Green circles are positive residuals, and red circles are negative residuals. Pot, fixed gear and trawl handling mortality rates were assumed to be 0.2, 0.5 and 0.8, respectively.

Scenario 2bn1, Trawl Survey Females

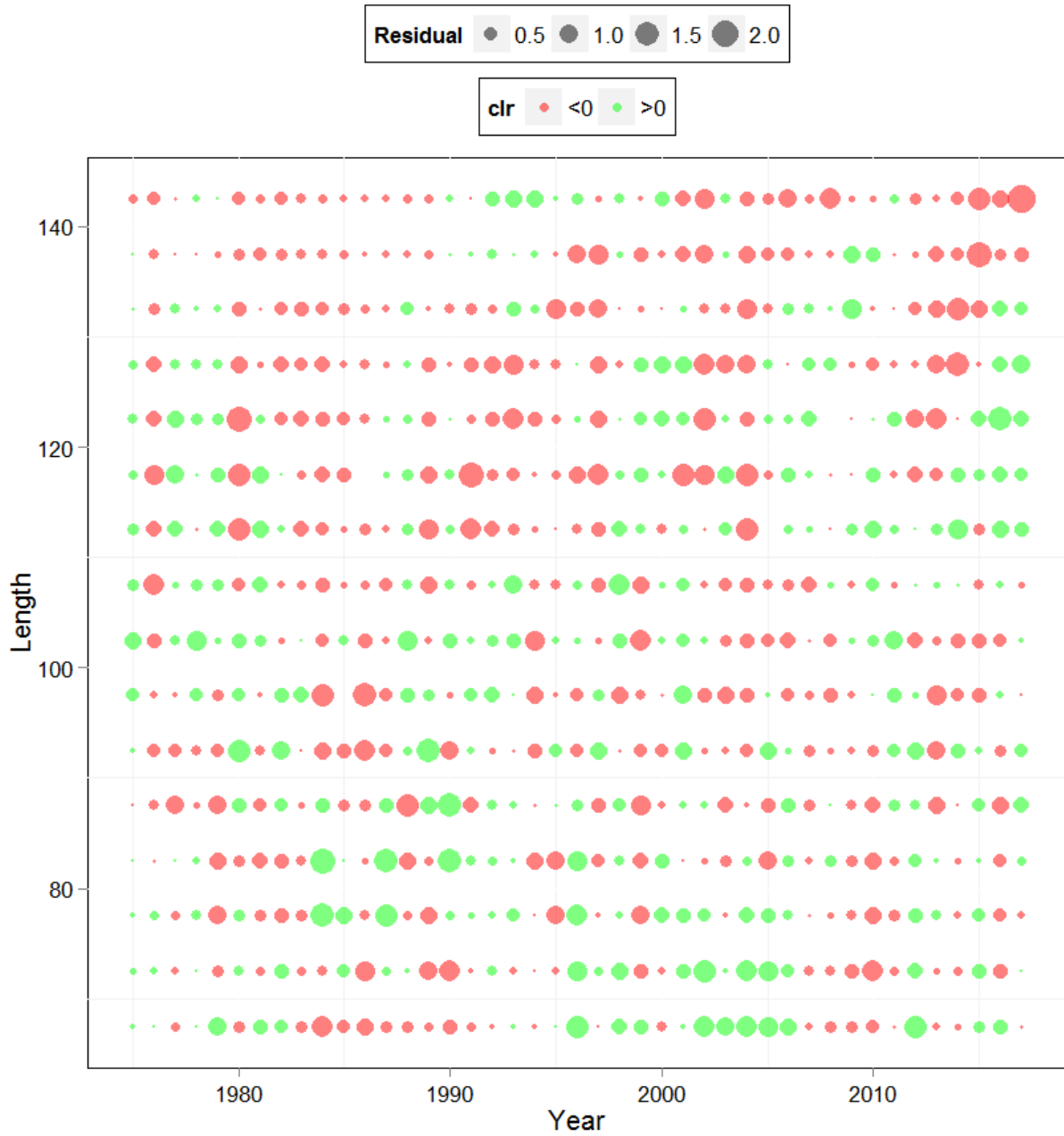


Figure 26(2bn1). Standardized residuals of proportions of survey female red king crab by year and carapace length (mm) under scenario 2bn1. Green circles are positive residuals, and red circles are negative residuals. Pot, fixed gear and trawl handling mortality rates were assumed to be 0.2, 0.5 and 0.8, respectively.

Scenario 2bn2, Trawl Survey Females

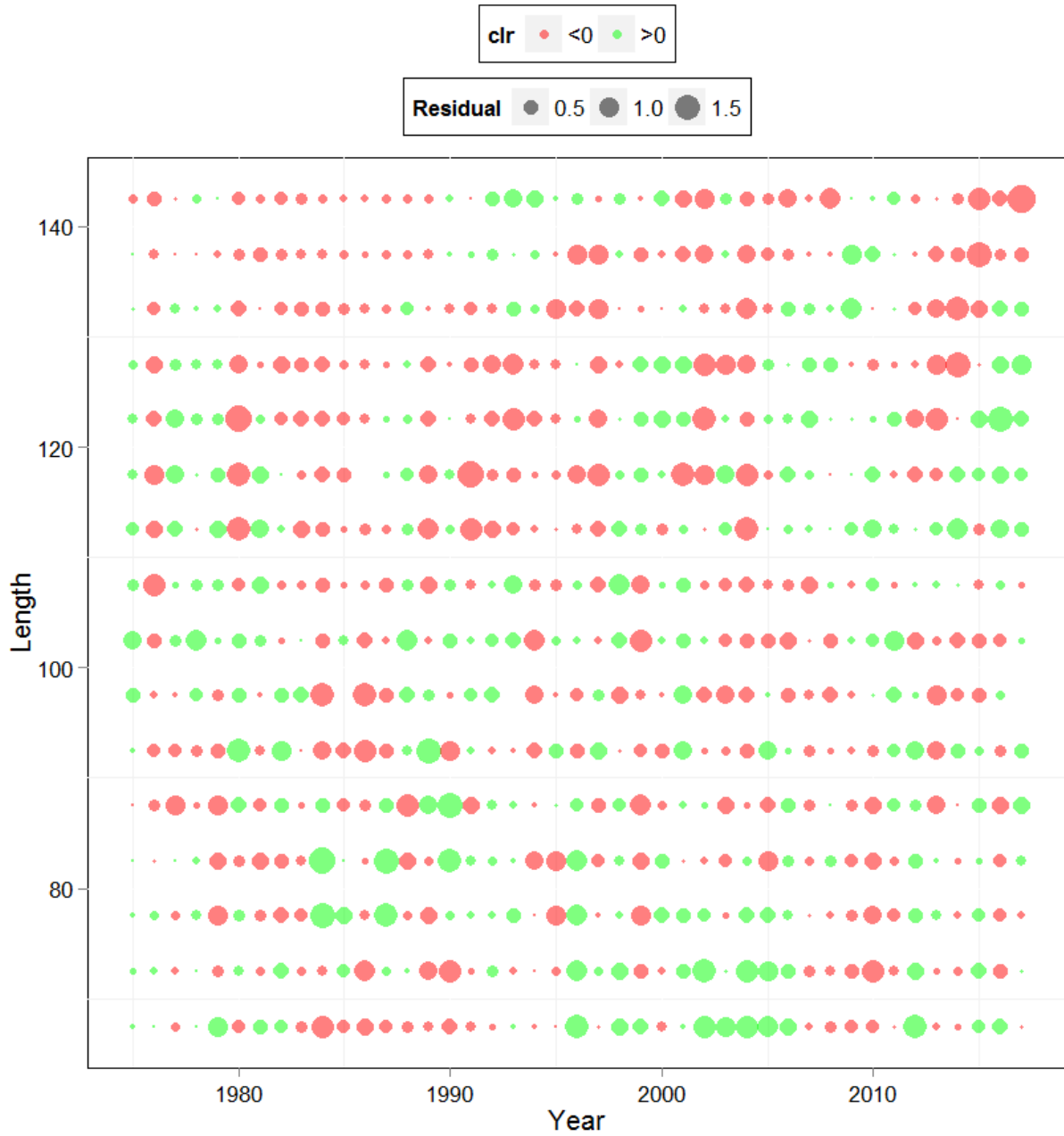


Figure 26(2bn2). Standardized residuals of proportions of survey female red king crab by year and carapace length (mm) under scenario 2dn2. Green circles are positive residuals, and red circles are negative residuals. Pot, fixed gear and trawl handling mortality rates were assumed to be 0.2, 0.5 and 0.8, respectively.

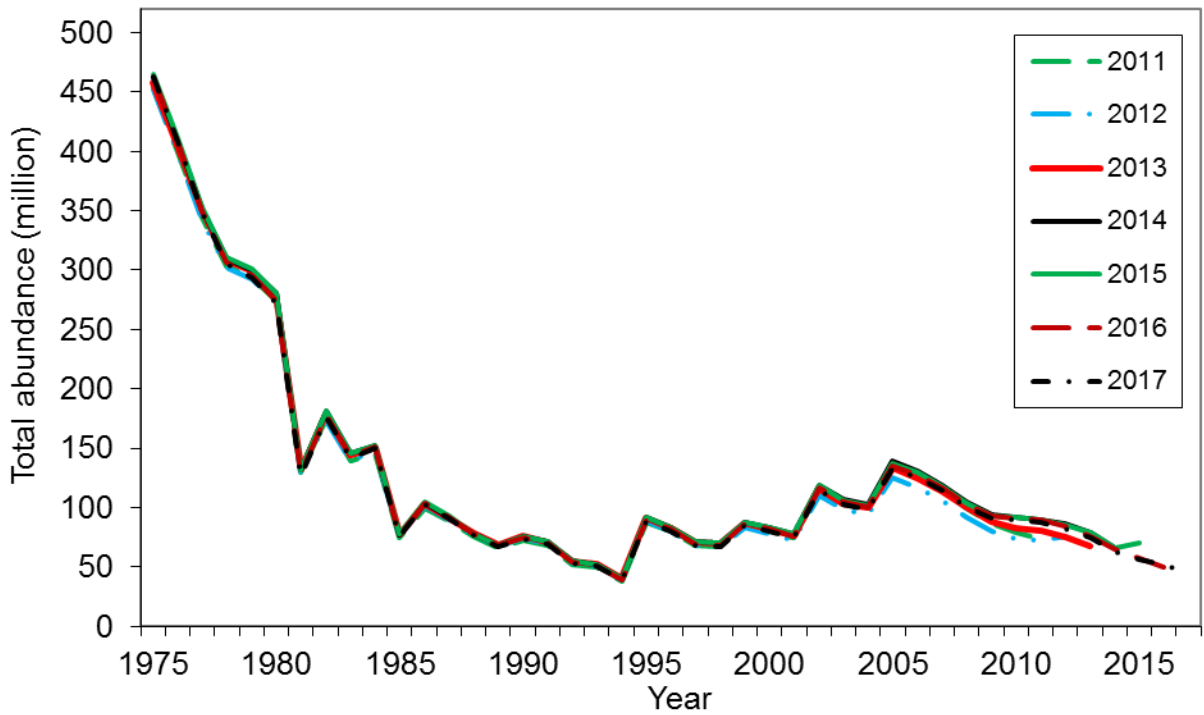
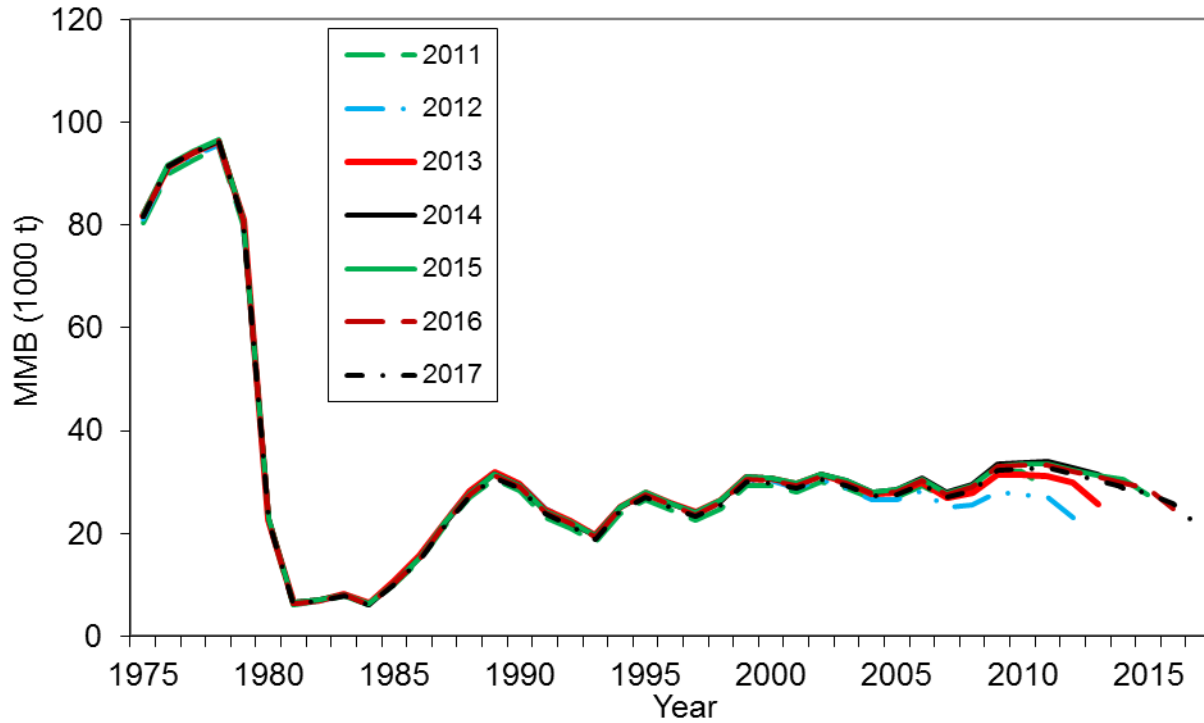


Figure 27. Comparison of hindcast estimates of mature male biomass on Feb. 15 (top) and total abundance (bottom) of Bristol Bay red king crab from 1975 to 2017 made with terminal years 2011-2017 with scenario 2b. These are results of the 2017 model. Legend shows the terminal year. Pot, fixed gear and trawl handling mortality rates were assumed to be 0.2, 0.5 and 0.8, respectively.

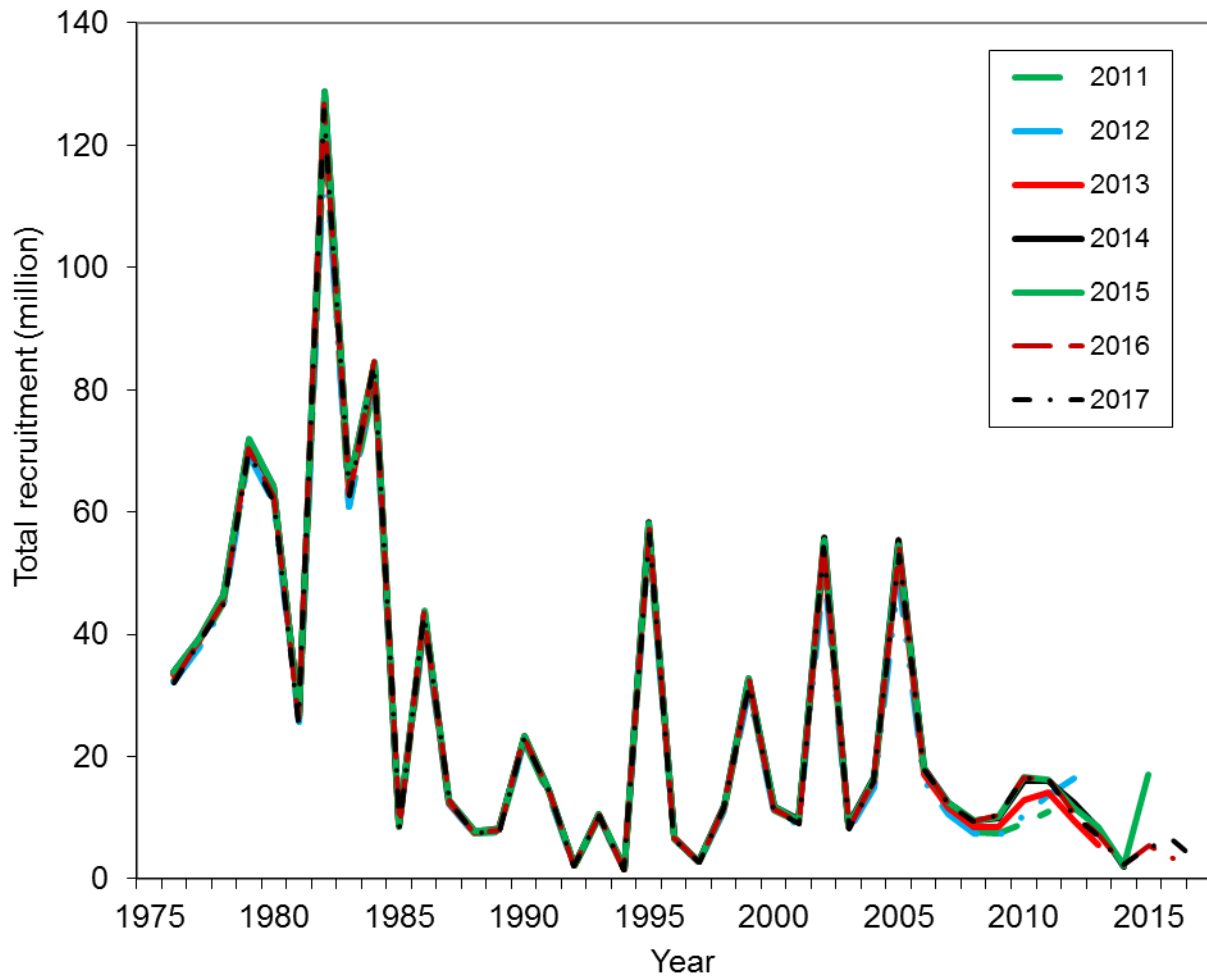


Figure 28a. Comparison of hindcast estimates of total recruitment for scenario 2b of Bristol Bay red king crab from 1976 to 2017 made with terminal years 2011-2017. These are results of the 2017 model. Legend shows the terminal year. Pot, fixed gear and trawl handling mortality rates were assumed to be 0.2, 0.5 and 0.8, respectively.

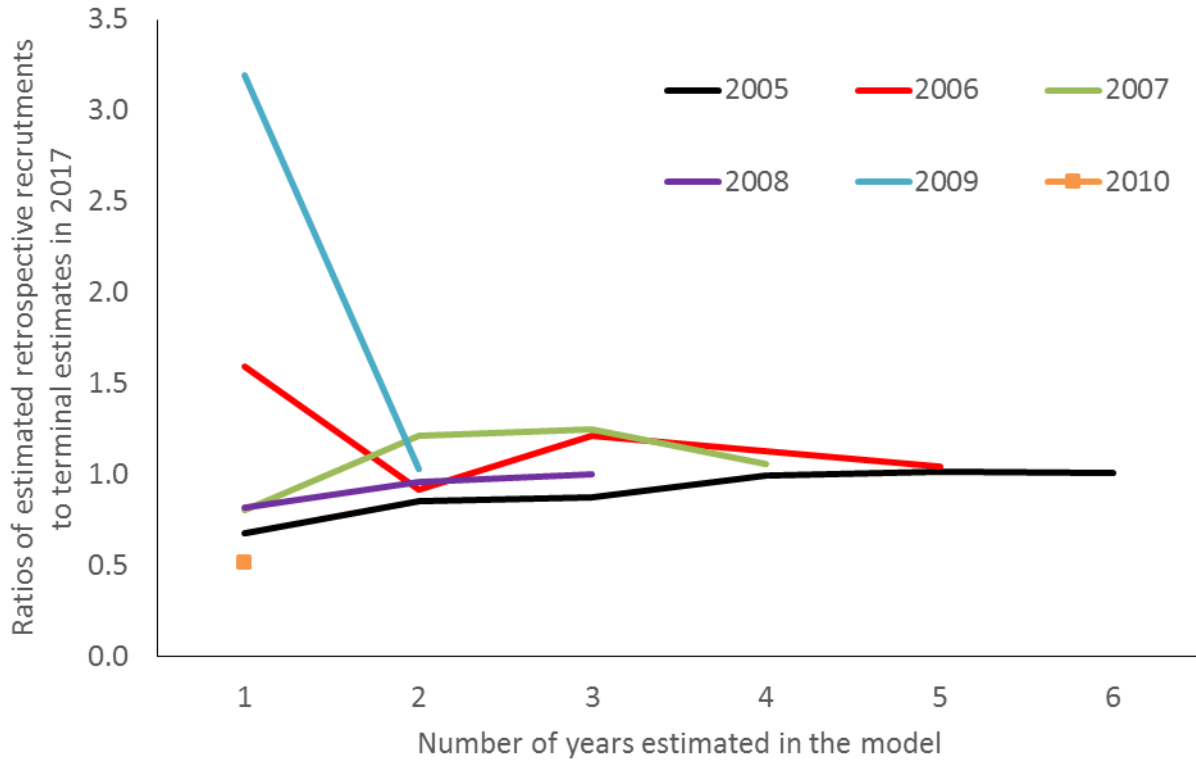


Figure 28b. Evaluation of Bristol Bay red king crab retrospective errors on recruitment estimates as a function of the number of years in the model for scenario 2b.

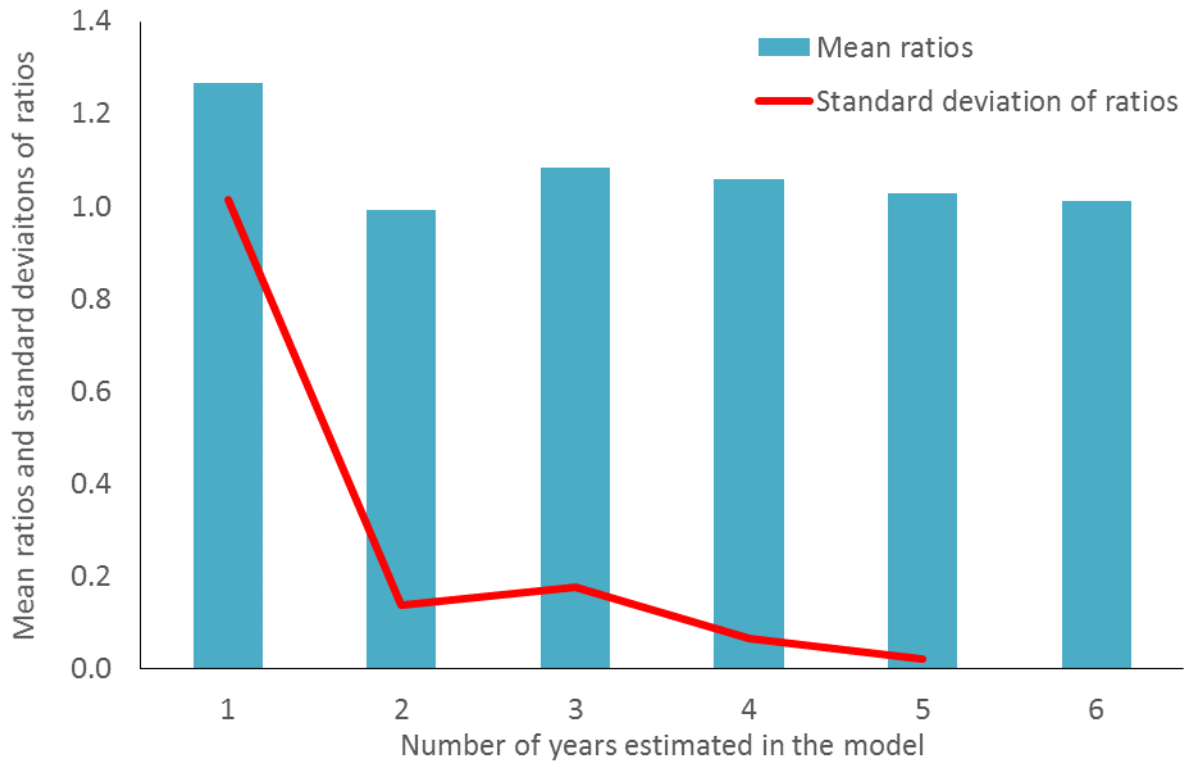


Figure 28c. Mean ratios of retrospective estimates of recruitments to those estimated in the most recent year (2017) and standard deviations of the ratios as a function of the number of years in the model for scenario 2b.

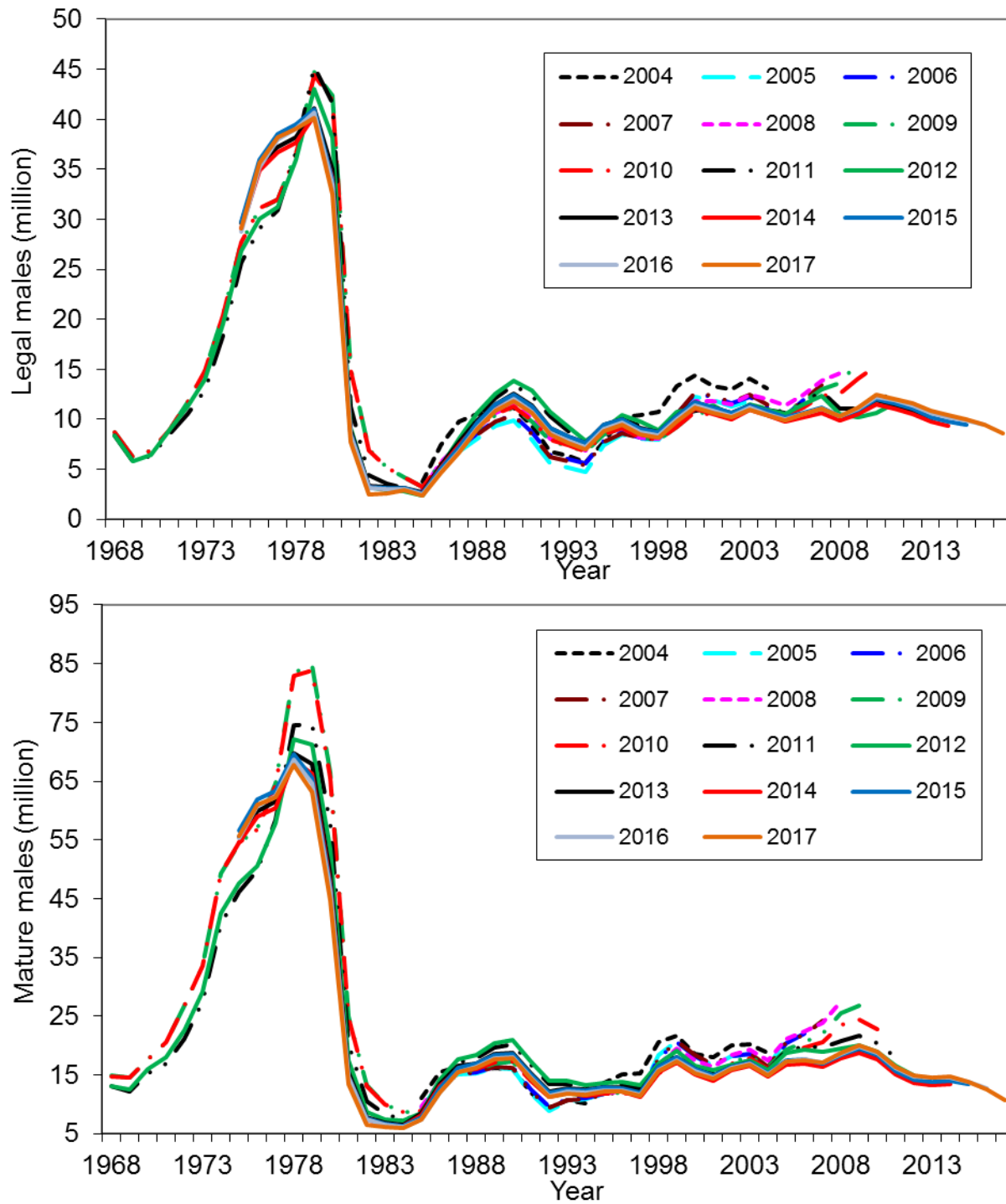


Figure 29. Comparison of estimates of legal male abundance (top) and mature males (bottom) of Bristol Bay red king crab from 1968 to 2017 made with terminal years 2004-2017 with the base scenarios. Scenario 2b is used for 2017. These are results of historical assessments. Legend shows the year in which the assessment was conducted. Pot, fixed gear and trawl handling mortality rates were assumed to be 0.2, 0.5 and 0.8, respectively.

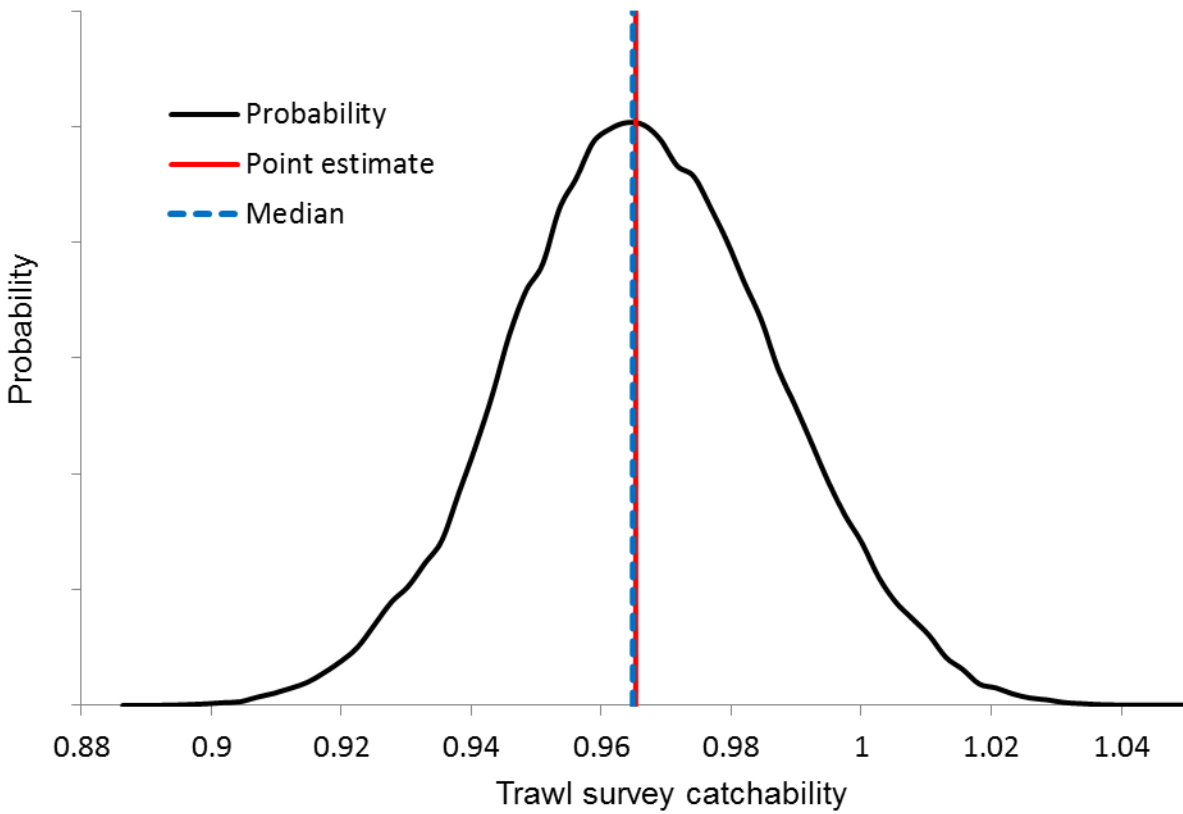


Figure 30. Probability distributions of estimated trawl survey catchability (Q) under scenario 2b with the mcmc approach. Pot, fixed gear and trawl handling mortality rates were assumed to be 0.2, 0.5 and 0.8, respectively.

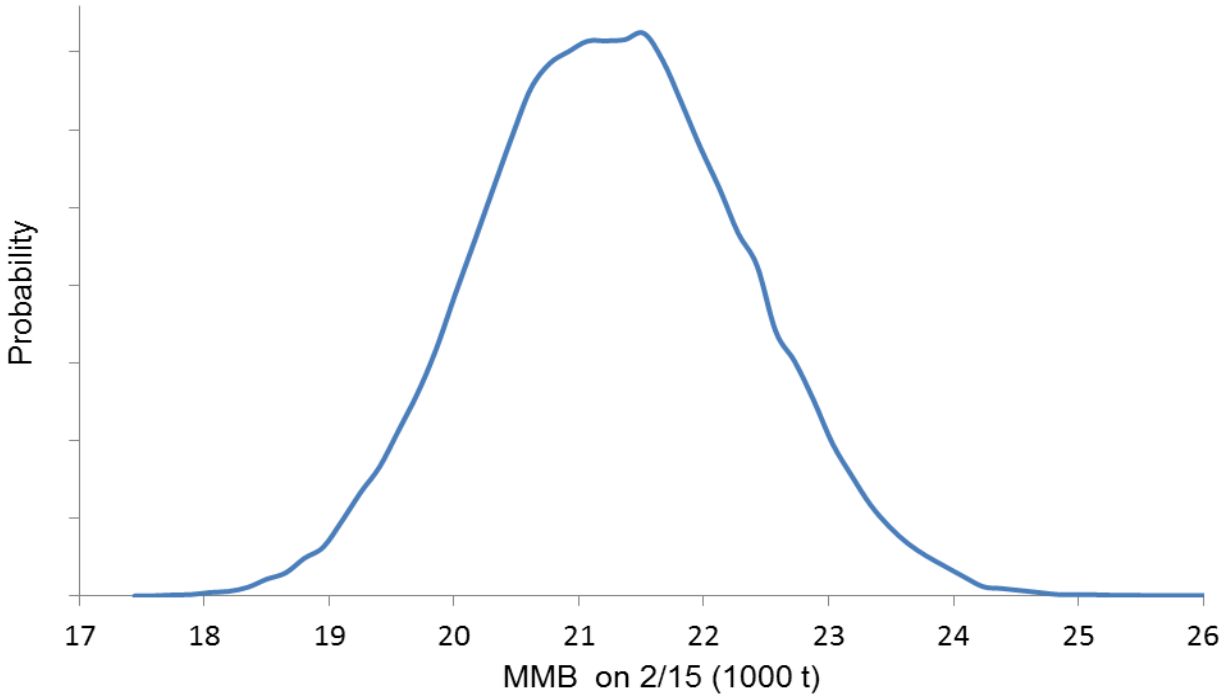


Figure 31a(2b). Probability distributions of estimated mature male biomass on Feb. 15, 2017 with $F_{35\%}$ under scenario 2b with the mcmc approach. Pot, fixed gear and trawl handling mortality rates were assumed to be 0.2, 0.5 and 0.8, respectively.

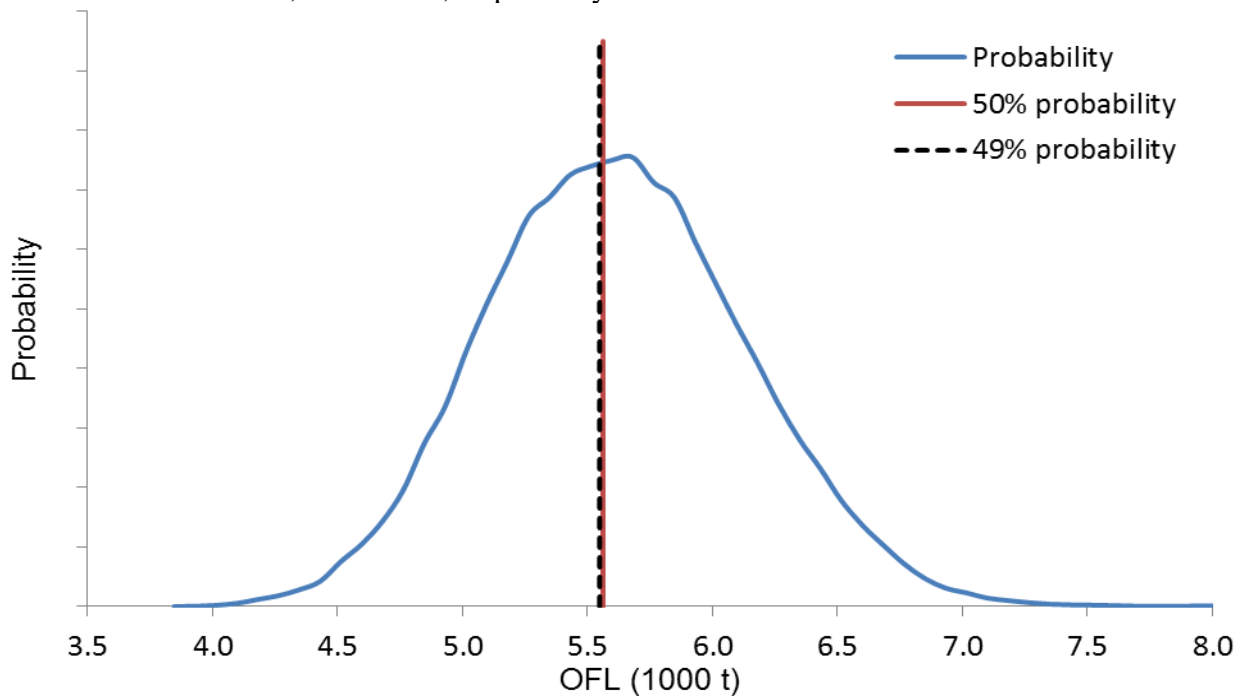


Figure 31b(2b). Probability distributions of the 2017 estimated OFL with scenario 2b with the mcmc approach. Pot, fixed gear and trawl handling mortality rates were assumed to be 0.2, 0.5 and 0.8, respectively.

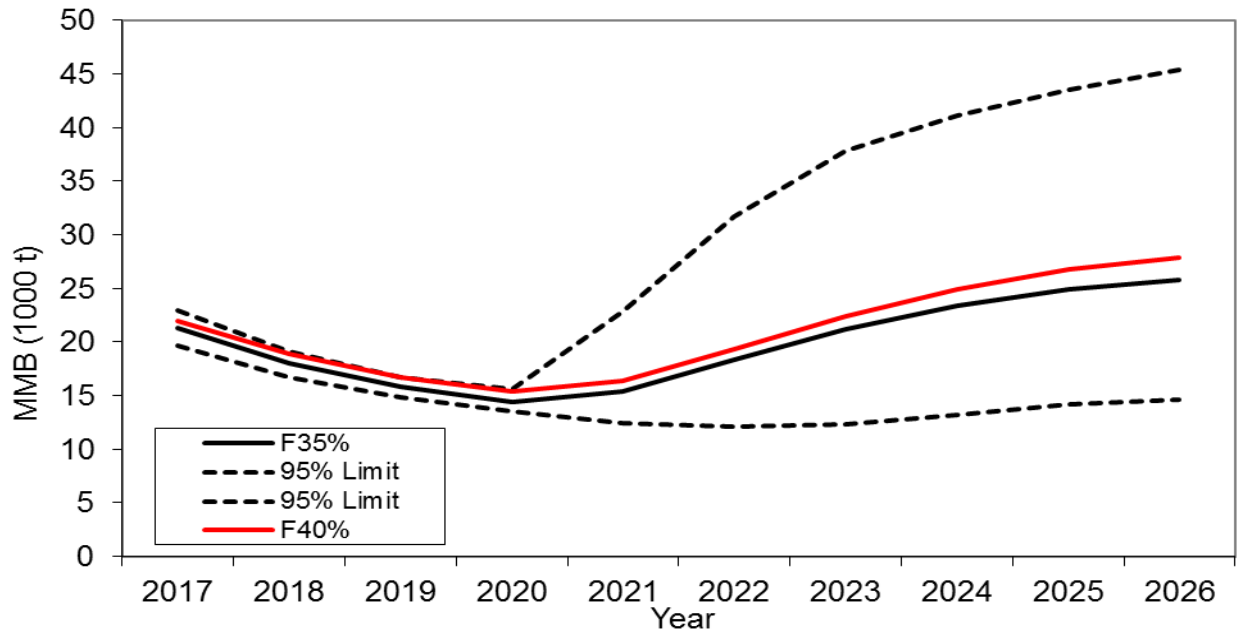


Figure 32(2b). Projected mature male biomass on Feb. 15 with $F_{40\%}$ and $F_{35\%}$ harvest strategy during 2017-2026. Input parameter estimates are based on scenario 2b. Pot, fixed gear and trawl handling mortality rates were assumed to be 0.2, 0.5 and 0.8, respectively, and the confidence limits are for the $F_{35\%}$ harvest strategy.

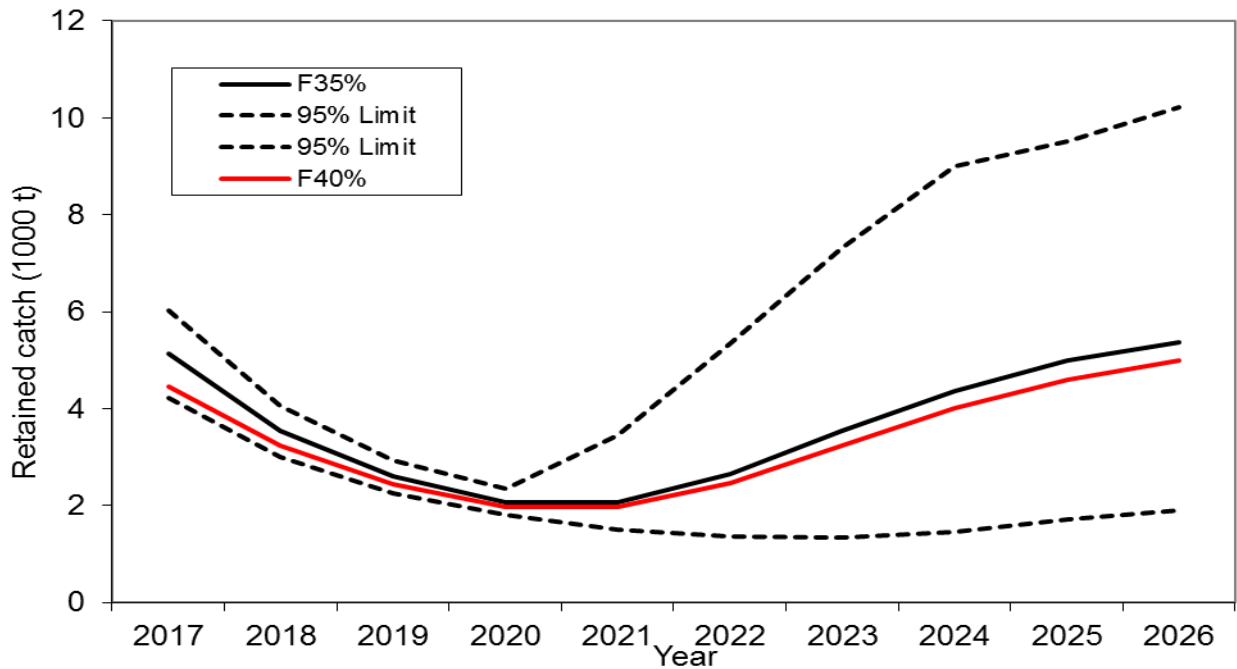


Figure 33(2b). Projected retained catch biomass with $F_{40\%}$ and $F_{35\%}$ harvest strategy during 2017-2026. Input parameter estimates are based on scenario 2b. Pot, fixed gear and trawl handling mortality rates were assumed to be 0.2, 0.5 and 0.8, respectively, and the confidence limits are for the $F_{35\%}$ harvest strategy.

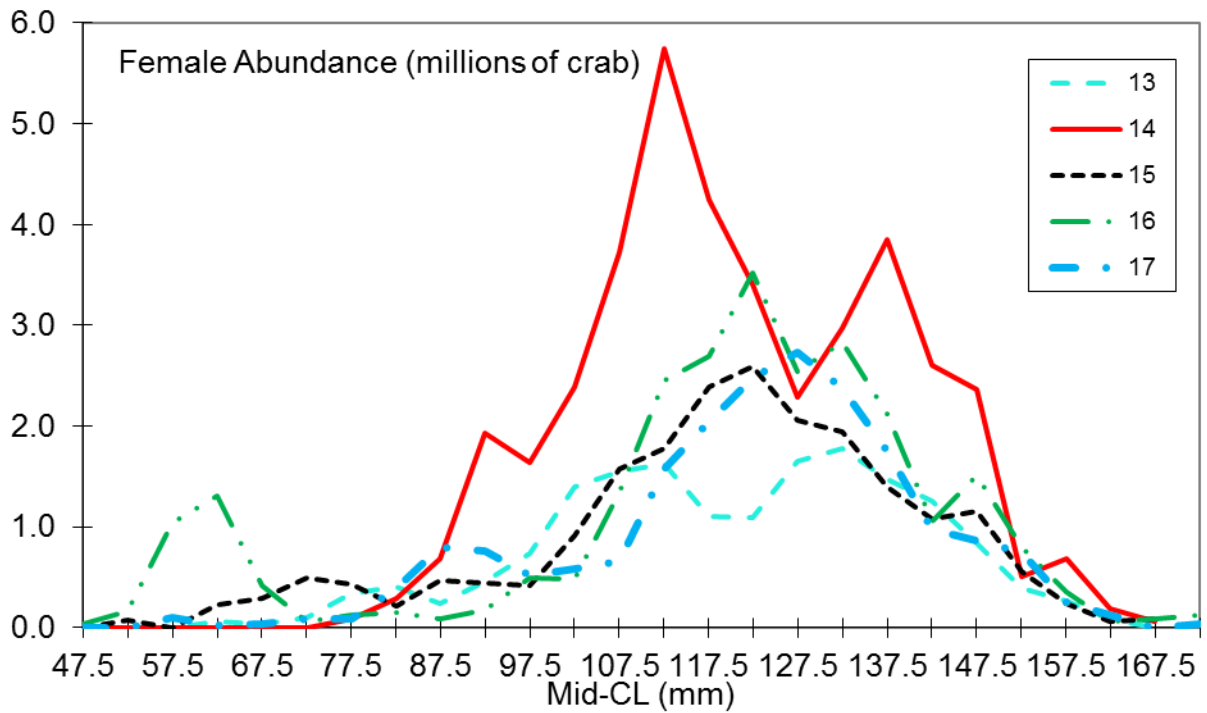
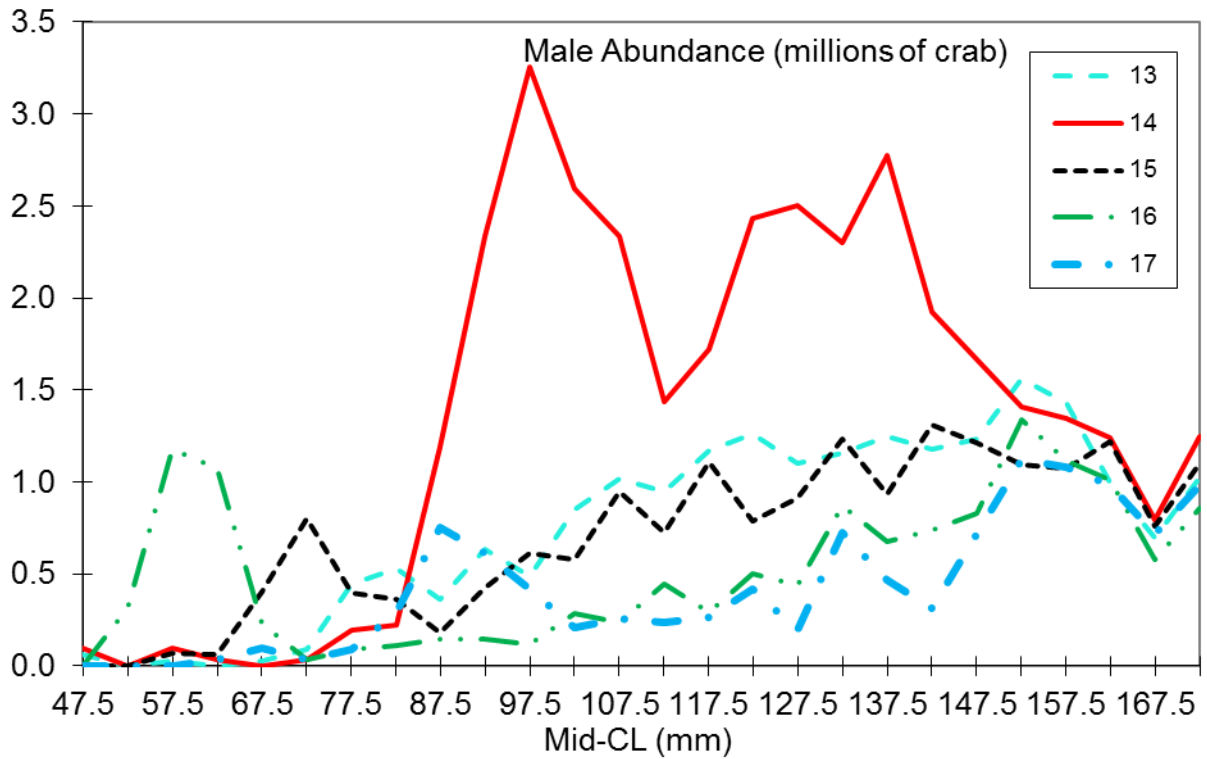


Figure 34. Length frequency distributions of male (top panel) and female (bottom panel) red king crab in Bristol Bay from NMFS trawl surveys during 2013-2017. For purposes of these graphs, abundance estimates are based on area-swept methods.

Appendix A. Description of the Bristol Bay Red King Crab Model

a. Model Description

i. Population model

The original LBA model was described in detail by Zheng et al. (1995a, 1995b) and Zheng and Kruse (2002). Crab abundances by carapace length and shell condition in any one year are modeled to result from abundances in the previous year minus catch and handling and natural mortalities, plus recruitment, and additions to or losses from each length class due to growth:

$$N_{l,t+1}^s = \sum_{l'=1}^l \{P_{l',l,t}^s [(N_{l',t}^s + O_{l',t}^s)e^{-M_i^s} - (C_{l',t}^s + D_{l',t}^s)e^{(y_t-1)M_i^s} - T_{l',t}^s e^{(j_t-1)M_i^s}] m_{l',t}^s\} + R_{t+1}^s U_l^s \quad (\text{A1})$$

$$O_{l,t+1}^s = [(N_{l,t}^s + O_{l,t}^s)e^{-M_i^s} - (C_{l,t}^s + D_{l,t}^s)e^{(y_t-1)M_i^s} - T_{l,t}^s e^{(j_t-1)M_i^s}] (1 - m_{l,t}^s)$$

where $N_{l,t}^s$ is the number of new shell crab of sex s in length-class l at the start of year t , $O_{l,t}^s$ the number of old shell crab of sex s in length-class l at the start of year t , $P_{l',l,t}^s$ the proportion during year t of an animals of sex s in length-class l' which grow into length-class l given that they moulted, M_i^s the rate of natural mortality on animals of sex s during year t , $m_{l,t}^s$ the probability that an animal of sex s in length-class l will moult during year t , R_{t+1}^s the recruitment [to the model] of animals of sex s during year t , U_l^s the proportion of recruits of sex s which recruit to length-class l , $C_{l,t}^s$ the retained catch (in numbers) of animals of sex s in length-class l during year t , $D_{l,t}^s$ the discarded catch of animals of sex s in length-class l during year t in the directed fishery, $T_{l,t}^s$ the discarded catch of animals of sex s in length-class l during year t in the Tanner crab fishery and the groundfish fisheries, y_t the time in years between survey and the directed pot fishery during year t , and j_t the time in years between survey and the Tanner and groundfish fisheries during year t .

The minimum carapace length for both males and females is set at 65 mm, and crab abundance is modeled with a length-class interval of 5 mm. The last length class includes all crab ≥ 160 -mm CL for males and ≥ 140 -mm CL for females. Thus, length classes/groups are 20 for males and 16 for females. Since females moult annually (Powell 1967), females have only the first part of the equation (A1).

The growth increment is assumed to be gamma distributed with mean which depends linearly on pre-moult length, i.e.:

$$P_{l',l,t}^s = \int_{L_l - \Delta L/2}^{L_l + \Delta L/2} \frac{x^{\alpha_{l',t}^s} e^{-x/\beta^s}}{(\beta^s)^{\alpha_{l',t}^s} \Gamma(\alpha_{l',t}^s)} dx \quad \alpha_{l',t}^s \beta^s = a_t^s + b_t^s L_l \quad (\text{A2})$$

where L_l is the mid-point of length-class l , ΔL the width of each size-class (5 mm carapace length), a_t^s, b_t^s the parameters of the length–growth increment relationship for sex s and year t , and β^s the parameter determining the variance of the growth increment. Growth is time-invariant for males, and specified for three time-blocks for females (1968-82; 1983-93; 1994-2017) based on changes to the size at maturity for females. The probability of moulting as a function of length for males is given by an inverse logistic function, i.e.:

$$m_l = \frac{1}{1 + e^{\beta(L_l - L_{50})}} \quad (\text{A3})$$

where β, L_{50} are the parameters which determine the relationship between length and the probability of moulting.

Recruitment is defined as recruitment to the model and survey gear rather than recruitment to the fishery. Recruitment is separated into a time-dependent variable, R_{t+1}^s , and size-dependent variables, U_l^s , representing the proportion of recruits belonging to each length class. R_{t+1}^s is assumed to consist of crab at the recruiting age with different lengths and thus represents year class strength for year t . The proportion of recruits by length-class, U_l^s , is described using a gamma distribution with parameters α_l^s and β_l^s . Because of different growth rates, recruitment is estimated separately for males and females under a constraint of approximately equal sex ratios of recruitment over time.

ii. Catches and Fisheries Selectivities

Before 1990, no observed bycatch data were available in the directed pot fishery; the crab that were discarded and died in those years were estimated as the product of handling mortality rate, legal harvest rates, and mean length-specific selectivities. It is difficult to estimate bycatch from the Tanner crab fishery before 1991. A reasonable index to estimate bycatch fishing mortalities is potlifts of the Tanner crab fishery within the distribution area of Bristol Bay red king crab. Thus, bycatch fishing mortalities from the Tanner crab fishery before 1991 were estimated to be proportional to the smoothing average of potlifts east of 163° W. The smoothing average is equal to $(P_{t-2} + 2P_{t-1} + 3P_t)/6$ for the potlifts in year t . The smoothing process not only smoothes the annual number of potlifts, it also indexes the effects of lost pots during the previous years. All bycatches are death catches because the model fits the estimated observed death bycatches.

The catch (by sex) in numbers by the directed fishery is:

$$G_{l,t}^s = (N_{l,t}^s + O_{l,t}^s) e^{-y_t M_t^s} (1 - e^{-F_{l,t}^s}) \quad (\text{A4})$$

where $F_{l,t}^s$ is the fishing mortality rate during year t on animals of sex s in length-class l due to the directed fishery:

$$F_{l,t}^s = \begin{cases} [(S_l^{dir,land} (1 + h_t \phi) + S_l^{dir,disc,mal}] F_t^{dir} & \text{if } s = \text{mal} \\ S_l^{dir,disc,fem} F_t^{disc,fem} & \text{if } s = \text{fem} \end{cases} \quad (\text{A5})$$

where $S_l^{dir,land}$ is the selectivity pattern for the landings by the directed fishery, $S_l^{dir,disc,s}$ the selectivity pattern for the discards in the directed fishery by sex, F_t^{dir} the fully-selected fishing mortality during year t (on males), $F_t^{disc,fem}$ the fully-selected fishing mortality on female animals during year t related to discards in the directed fishery, ϕ the handling mortality (the proportion of animals which die due to being returned to the water following capture), and h_t the rate of high-grading during year t , i.e. discards of animals which can be legally-retained by the directed pot fishery (non-zero only for 2005-2016).

There are no landings of females in a male-only fishery, while the landings C of males in the directed fishery and discards D of males in the directed fishery are:

$$\begin{aligned} C_{l,t}^{\text{mal}} &= (N_{l,t}^{\text{mal}} + O_{l,t}^{\text{mal}}) e^{-y_t M_t^{\text{mal}}} (1 - e^{-S_l^{dir,land} F_t^{dir}}) \\ D_{l,t}^{\text{mal}} &= G_{l,t}^{\text{mal}} - C_{l,t}^{\text{mal}} \end{aligned} \quad (\text{A6})$$

The catch (by sex) in numbers by the Tanner crab and groundfish fisheries in length-class l during year t is given by:

$$T_{l,t}^s = (N_{l,t}^s + O_{l,t}^s) e^{-j_t M_t^s} e^{-F_{l,t}^s} (1 - e^{-\tilde{F}_{l,t}^s}) \quad (\text{A7})$$

where $\tilde{F}_{l,t}^s$ is the fishing mortality rate during year t on animals of sex s in length-class l due to the Tanner crab and groundfish fisheries:

$$\tilde{F}_{l,t}^s = S_l^{\text{Tanner},s} F_t^{\text{Tanner},s} + S_l^{\text{trawl}} F_t^{\text{trawl}} + S_l^{\text{fix}} F_t^{\text{fix}} \quad (\text{A8})$$

where $S_l^{\text{Tanner},s}$ is the selectivity pattern for the discards in the Tanner crab fishery by sex, $F_t^{\text{Tanner},s}$ the fully-selected fishing mortality during year t on animals of sex s during year t due to this fishery, S_l^{trawl} the selectivity pattern for the bycatch in the groundfish trawl fishery, F_t^{trawl} the fully-selected fishing mortality due to the groundfish trawl fishery, S_l^{fix} the selectivity pattern for the bycatch in the groundfish fixed gear fishery, and F_t^{fix} the fully-selected fishing mortality due to the groundfish fixed gear fishery.

The bycatches by sex are estimated from the Tanner crab fishery, $TC_{l,t}^s$, groundfish trawl fishery, $GT_{l,t}^s$, and groundfish fixed gear fishery, $GF_{l,t}^s$, as follow:

$$\begin{aligned}
TC_{l,t}^s &= (N_{l,t}^s + O_{l,t}^s) e^{-j_t M_t^s} e^{-F_{l,t}^s} (1 - e^{-\tilde{F}_{l,t}^s}) S_l^{\text{Tanner},s} F_t^{\text{Tanner},s} / \tilde{F}_{l,t}^s \\
GT_{l,t}^s &= (N_{l,t}^s + O_{l,t}^s) e^{-j_t M_t^s} e^{-F_{l,t}^s} (1 - e^{-\tilde{F}_{l,t}^s}) S_l^{\text{trawl}} F_t^{\text{trawl}} / \tilde{F}_{l,t}^s \\
GF_{l,t}^s &= (N_{l,t}^s + O_{l,t}^s) e^{-j_t M_t^s} e^{-F_{l,t}^s} (1 - e^{-\tilde{F}_{l,t}^s}) S_l^{\text{fixed}} F_t^{\text{fixed}} / \tilde{F}_{l,t}^s
\end{aligned} \tag{A9}$$

For scenarios separating mature and immature crab, discarded female bycatch in numbers is separated into immature and mature bycatches. The female bycatches in the directed fishery in length-class l and during year t , $D_{l,t}^i$ and $D_{l,t}^m$, and $T_{l,t}^i$ and $T_{l,t}^m$, are:

$$\begin{aligned}
D_{l,t}^i &= N_{l,t}^i e^{-y_t M_t^{\text{fem}}} (1 - e^{-F_{l,t}^{\text{fem}}}) \\
D_{l,t}^m &= N_{l,t}^m e^{-y_t M_t^{\text{fem}}} (1 - e^{-F_{l,t}^{\text{fem}}})
\end{aligned} \tag{A10}$$

The female bycatches (by maturity) in numbers by the Tanner crab and groundfish fisheries in length-class l during year t for scenario 2 are given by:

$$\begin{aligned}
T_{l,t}^i &= N_{l,t}^i e^{-j_t M_t^{\text{fem}}} e^{-F_{l,t}^{\text{fem}}} (1 - e^{-\tilde{F}_{l,t}^{\text{fem}}}) \\
T_{l,t}^m &= N_{l,t}^m e^{-j_t M_t^{\text{fem}}} e^{-F_{l,t}^{\text{fem}}} (1 - e^{-\tilde{F}_{l,t}^{\text{fem}}})
\end{aligned} \tag{A11}$$

Retained selectivity, $S^{\text{dir,land}}$, selectivity for females in the directed fishery, $S^{\text{dir,disc,fem}}$, selectivities for males and females in the groundfish trawl and fixed gear fisheries, S^{trawl} and S^{fix} , and selectivity for males and females in the Tanner crab fishery, $S^{\text{Tanner},s}$, are all assumed to be logistic functions of length:

$$S_l^{\text{type}} = \frac{1}{1 + e^{-\beta^{\text{type}} (l - L_{50}^{\text{type}})}} \tag{A12}$$

Different sets of parameters (β , L_{50}) are estimated for retained males, female pot bycatch, male and female trawl bycatch, and discarded males and females from the Tanner crab fishery.

Male pot bycatch selectivity in the directed fishery is modeled by two linear functions:

$$\begin{aligned}
s_l &= \varphi + \kappa l, \quad \text{if } l < 135 \text{ mm CL}, \\
s_l &= s_{l-1} + 5\gamma, \quad \text{if } l > 134 \text{ mm CL}
\end{aligned} \tag{A13}$$

where φ , κ , γ are parameters.

iii. Trawl Survey Selectivities

Trawl survey selectivities are estimated as

$$S_{l,t}^s = \frac{Q}{1 + e^{-\beta_t^s (l - L_{50,t}^s)}} \tag{A14}$$

with different sets of parameters (β , L_{50}) estimated for males and females as well as two different periods (1975-81 and 1982-17). Survey selectivity for the first length group (67.5 mm) was assumed to be the same for both males and females, so only three parameters (β , L_{50} for females and L_{50} for males) were estimated in the model for each of the four periods. Parameter Q was called the survey catchability that was estimated based on a trawl experiment by Weinberg et al. (2004; Figure A1). Q was assumed to be constant over time.

Assuming that the BSFRF survey caught all crab within the area-swept, the ratio between NMFS abundance and BSFRF abundance is a capture probability for the NMFS survey net. The Delta method was used to estimate the variance for the capture probability. A maximum likelihood method was used to estimate parameters for a logistic function as an estimated capture probability curve (Figure A1). For a given size, the estimated capture probability is smaller based on the BSFRF survey than from the trawl experiment, but the Q value is similar between the trawl experiment and the BSFRF surveys (Figure A1). Because many small-sized crab are likely in the shallow water areas that are not accessible for the trawl survey, NMFS trawl survey selectivity consists of capture probability and crab availability.

iv. Estimating Bycatch Fishing Mortalities for Years without Observer Data

Observer data are not available for the directed pot fishery before 1990 and the Tanner crab fishery before 1991. There are also extremely low observed bycatches in the Tanner crab fishery during 1994 and 2006-2009. Bycatch fishing mortalities for male and females during 1975-1989 in the directed pot fishery were estimated as

$$F_t^{disc,s} = r^s F_t^{dir} \quad (A15)$$

where r^s is the median ratio of estimated bycatch discard fishing mortalities to the estimated directed pot fishing mortalities during 1990-2004 for sex s . Directed pot fishing practice has changed after 2004 due to fishery rationalization.

We used pot fishing effort (potlifts) east of 163° W in the Tanner crab fishery to estimate red king crab bycatch discard fishing mortalities in that fishery when observer data are not available (1975-1990, 1994, 2006-2009):

$$F_t^{Tanner,s} = a^s E_t \quad (A16)$$

where a^s is the mean ratio of estimated Tanner crab fishery bycatch fishing mortalities to fishing efforts during 1991-1993 for sex s , and E_t is Tanner crab fishery fishing efforts east of 163° W in year t . Due to fishery rationalization after 2004, we used the data only during 1991-1993 to estimate the ratio.

b. Software Used: AD Model Builder (Fournier et al. 2012).

c. Likelihood Components

A maximum likelihood approach was used to estimate parameters. For length compositions ($p_{l,t,s,sh}$), the likelihood functions are :

$$Rf = \prod_{l=1}^L \prod_{t=1}^T \prod_{s=1}^2 \prod_{sh=1}^2 \frac{\left\{ \exp \left[-\frac{(p_{l,t,s,sh} - \hat{p}_{l,t,s,sh})^2}{2\sigma^2} \right] + 0.01 \right\}}{\sqrt{2\pi\sigma^2}} \quad (\text{A17})$$

$$\sigma^2 = [\hat{p}_{l,t,s,sh}(1 - \hat{p}_{l,t,s,sh}) + 0.1/L]/n$$

where L is the number of length groups, T the number of years, and n the effective sample size, which was estimated for trawl survey and pot retained catch and bycatch length composition data from the directed pot fishery, and was assumed to be 50 for groundfish trawl and Tanner crab fisheries bycatch length composition data.

The weighted negative log likelihood functions are:

$$\begin{aligned} \text{Length compositions:} & \quad -\sum \ln(Rf_i) \\ \text{Biomasses other than survey:} & \quad \lambda_j \sum [\ln(C_t / \hat{C}_t)^2] \\ \text{NMFS survey biomass:} & \quad \sum [\ln(B_t / \hat{B}_t)^2 / (2\ln(CV_t^2 + 1))] \\ \text{BSFRF mature males:} & \quad \sum [\ln(\ln(CV_t^2 + 1))^{0.5} + \ln(B_t / \hat{B}_t)^2 / (2\ln(CV_t^2 + 1))] \\ \text{R variation:} & \quad \lambda_R \sum [\ln(R_t / \bar{R})^2] \\ \text{R sex ratio:} & \quad \lambda_s [\ln(\bar{R}_M / \bar{R}_F)^2] \\ \text{Trawl bycatch fishing mortalities:} & \quad \lambda_t [\ln(F_{t,t} / \bar{F}_t)^2] \\ \text{Pot female bycatch fishing mortalities:} & \quad \lambda_p [\ln(F_{t,f} / \bar{F}_f)^2] \\ \text{Trawl survey catchability:} & \quad (Q - \hat{Q})^2 / (2\sigma^2) \end{aligned} \quad (\text{A18})$$

where R_t is the recruitment in year t , \bar{R} the mean recruitment, \bar{R}_M the mean male recruitment, \bar{R}_F the mean female recruitment, \bar{F}_t the mean trawl bycatch fishing mortality, \bar{F}_f the mean pot female bycatch fishing mortality, Q summer trawl survey catchability, and σ the estimated standard deviation of Q (all scenarios) or each of six growth increment parameters for scenario 2b.

For BSFRF total survey biomass, CV is the survey CV plus AV , where AV is additional CV and estimated in the model.

Weights λ_j are assumed to be 500 for retained catch biomass, and 100 for all bycatch biomasses, 2 for recruitment variation, 10 for recruitment sex ratio, 0.2 for pot female bycatch fishing mortality, and 0.1 for trawl bycatch fishing mortality. These λ_j values correspond to CV values of 0.03, 0.07, 0.53, 0.23, 3.34, and 12.14, respectively, representing prior assumptions about the accuracy of the observed catch biomass data.

d. Population State in Year 1.

The total abundance and proportions for the first year are estimated in the model.

e. Parameter estimation framework:

i. Parameters estimated independently

Basic natural mortality, length-weight relationships, and mean growth increments per molt were estimated independently outside of the model. Mean length of recruits to the model depends on growth and was assumed to be 72.5 for both males and females. High grading parameters h_t were estimated to be 0.2785 in 2005, 0.0440 in 2006, 0.0197 in 2007, 0.0198 in 2008, 0.0337 in 2009, 0.0153 in 2010, 0.0113 in 2011, 0.0240 in 2012, 0.0632 in 2013, 0.1605 in 2014, 0.07 in 2015, and 0.0826 in 2016, based on the proportions of discarded legal males to total caught legal males. Handling mortality rates were set to 0.2 for the directed pot fishery, 0.25 for the Tanner crab fishery, 0.5 for the groundfish fixed gear fishery, and 0.8 for the groundfish trawl fishery.

(1). Natural Mortality

Based on an assumed maximum age of 25 years and the 1% rule (Zheng 2005), basic M was estimated to be 0.18 for both males and females. Natural mortality in a given year, M_t , equals to $M + Mm_t$ (for males) or $M + Mf_t$ (females). One value of Mm_t during 1980-1985 was estimated and two values of Mf_t during 1980-1984 and 1976-79, 1985-93 were estimated in the model for scenarios.

(2). Length-weight Relationship

Length-weight relationships for males and females were as follows:

$$\begin{aligned}
 \text{Immature Females: } & W = 0.000408 L^{3.127956} \\
 \text{Ovigerous Females: } & W = 0.003593 L^{2.666076} \\
 \text{Males: } & W = 0.0004031 L^{3.141334}
 \end{aligned}
 \tag{A19}$$

where W is weight in grams, and L CL in mm.

(3). Growth Increment per Molt

A variety of data are available to estimate male mean growth increment per molt for Bristol Bay RKC. Tagging studies were conducted during the 1950s, 1960s and 1990s, and mean growth increment per molt data from these tagging studies in the 1950s and 1960s were analyzed by Weber and Miyahara (1962) and Balsiger (1974). Modal analyses were conducted for the data during 1957-1961 and the 1990s (Weber 1967; Loher et al. 2001). Mean growth increment per molt may be a function of body size and shell condition and vary over time (Balsiger 1974; McCaughran and Powell 1977); however, for simplicity, mean growth increment per molt was assumed to be only a function of body size in the models. Tagging data were used to estimate mean growth increment per molt as a function of pre-molt length for males (Figure A2). The results from modal analyses of 1957-1961 and the 1990s were used to estimate mean growth increment per molt for immature females during 1975-1993 and 1994-2017, respectively, and the data presented in Gray (1963) were used to estimate those for mature females for scenarios 1, 1n and 2 (Figure A2). To make

a smooth transition of growth increment per molt from immature to mature females, weighted growth increment averages of 70% and 30% at 92.5 mm CL pre-molt length and 90% and 10% at 97.5 mm CL were used, respectively, for mature and immature females during 1983-1993. These percentages are roughly close to the composition of maturity. During 1975-1982, females matured at a smaller size, so the growth increment per molt as a function of length was shifted to smaller increments. Likewise, during 1994-2017, females matured at a slightly higher size, so the growth increment per molt was shifted to high increments for immature crab (Figure A2). Once mature, the growth increment per molt for male crab decreases slightly and annual molting probability decreases, whereas the growth increment for female crab decreases dramatically but annual molting probability remains constant at 1.0 (Powell 1967).

(4). Sizes at Maturity for Females

The NMFS collected female reproductive condition data during the summer trawl surveys. Mature females are separated from immature females by a presence of egg clutches or egg cases. Proportions of mature females at 5-mm length intervals were summarized and a logistic curve was fitted to the data each year to estimate sizes at 50% maturity. Sizes at 50% maturity are illustrated in Figure A3 with mean values for three different periods (1975-82, 1983-93, and 1994-2017).

(5). Sizes at Maturity for Males

Although size at sexual maturity for Bristol Bay red king crab males has been estimated (Paul et al. 1991), there are no data for estimating size of functional maturity collected in the natural environment. Sizes at functional maturity for Bristol Bay male RKC have been assumed to be 120 mm CL (Schmidt and Pengilly 1990). This is based on mating pair data collected off Kodiak Island (Figure A4). Sizes at maturity for Bristol Bay female RKC are about 90 mm CL, about 15 mm CL less than Kodiak female RKC (Pengilly et al. 2002). The size ratio of mature males to females is 1.3333 at sizes at maturity for Bristol Bay RKC, and since mature males grow at much larger increments than mature females, the mean size ratio of mature males to females is most likely larger than this ratio. Size ratios of the large majority of Kodiak mating pairs were less than 1.3333, and in some bays, only a small proportion of mating pairs had size ratios above 1.3333 (Figure A4).

In the laboratory, male RKC as small as 80 mm CL from Kodiak and Southeast Alaska can successfully mate with females (Paul and Paul 1990). But few males less than 100 mm CL were observed to mate with females in the wild. Based on the size ratios of males to females in the Kodiak mating pair data, setting 120 mm CL as a minimum size of functional maturity for Bristol Bay male RKC is proper in terms of managing the fishery.

(6). Potential Reasons for High Mortality during the Early 1980s

Bristol Bay red king crab abundance had declined sharply during the early 1980s. Many factors have been speculated for this decline: (i) completely wiped out by fishing: the directed pot fishery, the other directed pot fishery (Tanner crab fishery), and bottom trawling; and (ii) high fishing and natural mortality. With the survey abundance, harvest rates in 1980 and 1981 were among the highest, thus the directed fishing definitely had a

big impact on the stock decline, especially legal and mature males. However, for the sharp decline during 1980-1984 for males, 3 out of 5 years had low mature harvest rates. During the 1981-1984 decline for females, 3 out of 4 years had low mature harvest rates. Also pot catchability for females and immature males are generally much lower than for legal males, so the directed pot fishing alone cannot explain the sharp decline for all segments of the stock during the early 1980s.

Red king crab bycatch in the eastern Bering Sea Tanner crab fishery is another potential factor (Griffin et al. 1983). The main overlap between Tanner crab and Bristol Bay red king crab is east of 163° W. No absolute red king crab bycatch estimates are available until 1991. So there are insufficient data to fully evaluate the impact. Retained catch and potlifts from the eastern Bering Sea Tanner crab fishery are illustrated in Figure A5. The observed red king crab bycatch in the Tanner crab fishery during 1991-1993 and total potlifts east of 163° W during 1968 to 2005 were used to estimate the bycatch mortality in the current model. Because winter sea surface temperatures and air temperatures were warmer (which means a lower handling mortality rate) and there were fewer potlifts during the early 1980s than during the early 1990s, bycatch in the Tanner crab fishery is unlikely to have been a main factor for the sharp decline of Bristol Bay red king crab.

Several factors may have caused increases in natural mortality. Crab diseases in the early 1980s were documented by Sparks and Morado (1985), but inadequate data were collected to examine their effects on the stock. Stevens (1990) speculated that senescence may be a factor because many crab in the early 1980s were very old due to low temperatures in the 1960s and early 1970s. The biomass of the main crab predator, Pacific cod, increased about 10 times during the late 1970s and early 1980s. Yellowfin sole biomass also increased substantially during this period. Predation is primarily on juvenile and molting/softshell crab. But we lack stomach samples in shallow waters (juvenile habitat) and during the period when red king crab molt. Also cannibalism occurs during molting periods for red king crab. High crab abundance in the late 1970s and early 1980s may have increased the occurrence of cannibalism.

Overall, the likely causes for the sharp decline in the early 1980s are combinations of the above factors, such as pot fisheries on legal males, bycatch, and predation on females and juvenile and sublegal males, senescence for older crab, and disease for all crab. In our model, we estimated one mortality parameter for males and another for females during 1980-1984. We also estimated a mortality parameter for females during 1976-1979 and 1985-1993. These three mortality parameters are additional to the basic natural mortality of 0.18yr^{-1} , all directed fishing mortality, and non-directed fishing mortality. These three mortality parameters could be attributed to natural mortality as well as undocumented non-directed fishing mortality. The model fit the data much better with these three parameters than without them.

ii. Parameters estimated conditionally

The following model parameters were estimated for male and female crab: total recruits for each year (year class strength R_t for $t = 1976$ to 2017), total abundance in the first year (1975), growth parameter β , and recruitment parameter β_r for males and females

separately. Molting probability parameters β and L_{50} were also estimated for male crab. Estimated parameters also include β and L_{50} for retained selectivity, β and L_{50} for pot-discarded female selectivity, β and L_{50} for pot-discarded male and female selectivities from the eastern Bering Sea Tanner crab fishery, β and L_{50} for groundfish trawl discarded selectivity, ϕ , κ and γ for pot-discarded male selectivity, and β for trawl survey selectivity and L_{50} for trawl survey male and females separately. The NMFS survey catchabilities Q for some scenarios were also estimated. Three selectivity parameters were estimated for the survey data from the Bering Fisheries Research Foundation. Annual fishing mortalities were also estimated for the directed pot fishery for males (1975-2016), pot-discarded females from the directed fishery (1990-2016), pot-discarded males and females from the eastern Bering Sea Tanner crab fishery (1991-93, 2013-15), and groundfish trawl discarded males and females (1976-2016). Three additional mortality parameters for Mm_t and Mf_t were also estimated. Some estimated parameters were constrained in the model. For example, male and female recruitment estimates were forced to be close to each other for a given year.

f. Definition of model outputs.

- i. Biomass: two population biomass measurements are used in this report: total survey biomass (crab >64 mm CL) and mature male biomass (males >119 mm CL). Mating time is assumed to Feb. 15.
- ii. Recruitment: new entry of number of males in the 1st seven length classes (65- 99 mm CL) and new entry of number of females in the 1st five length classes (65-89 mm CL).
- iii. Fishing mortality: full-selected instantaneous annual fishing mortality rate at the time of fishery.

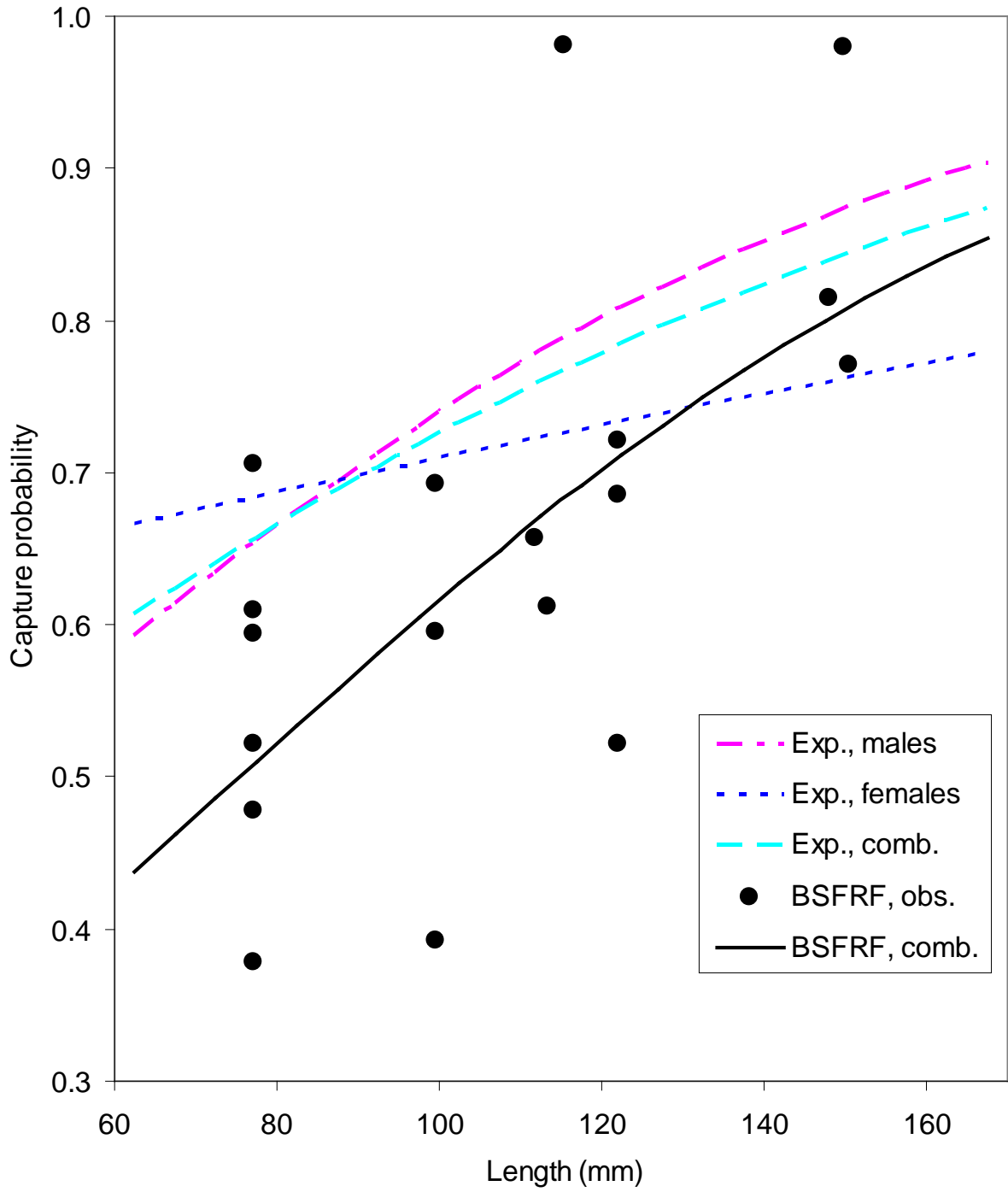


Figure A1. Estimated capture probabilities for NMFS Bristol Bay red king crab trawl surveys by Weinberg et al. (2004) and the Bering Sea Fisheries Research Foundation surveys.

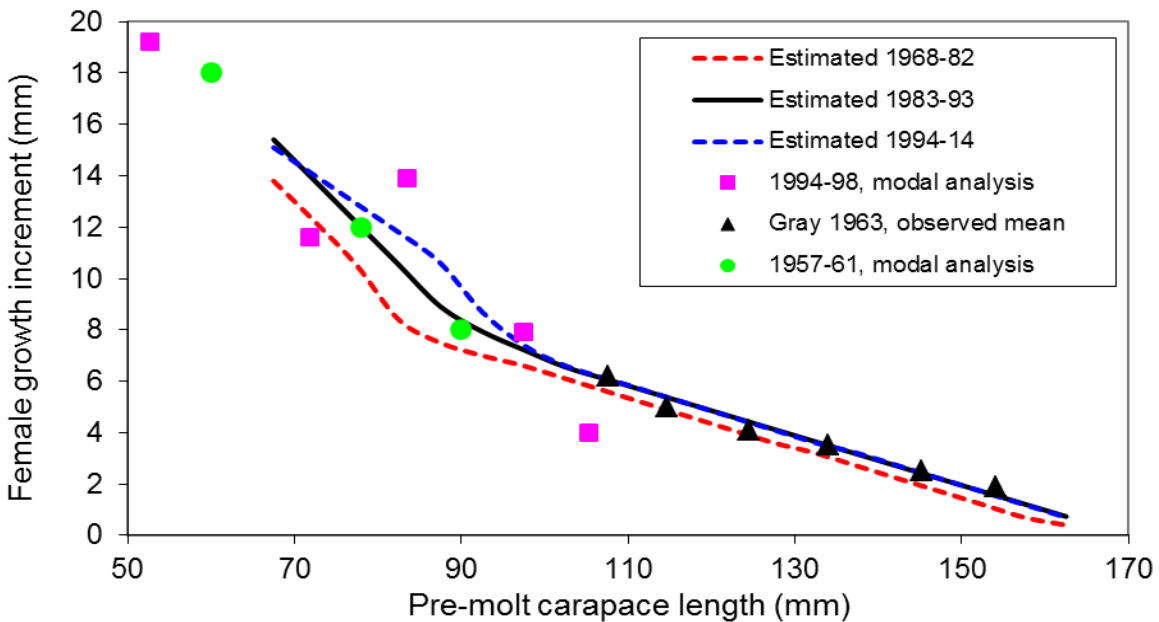
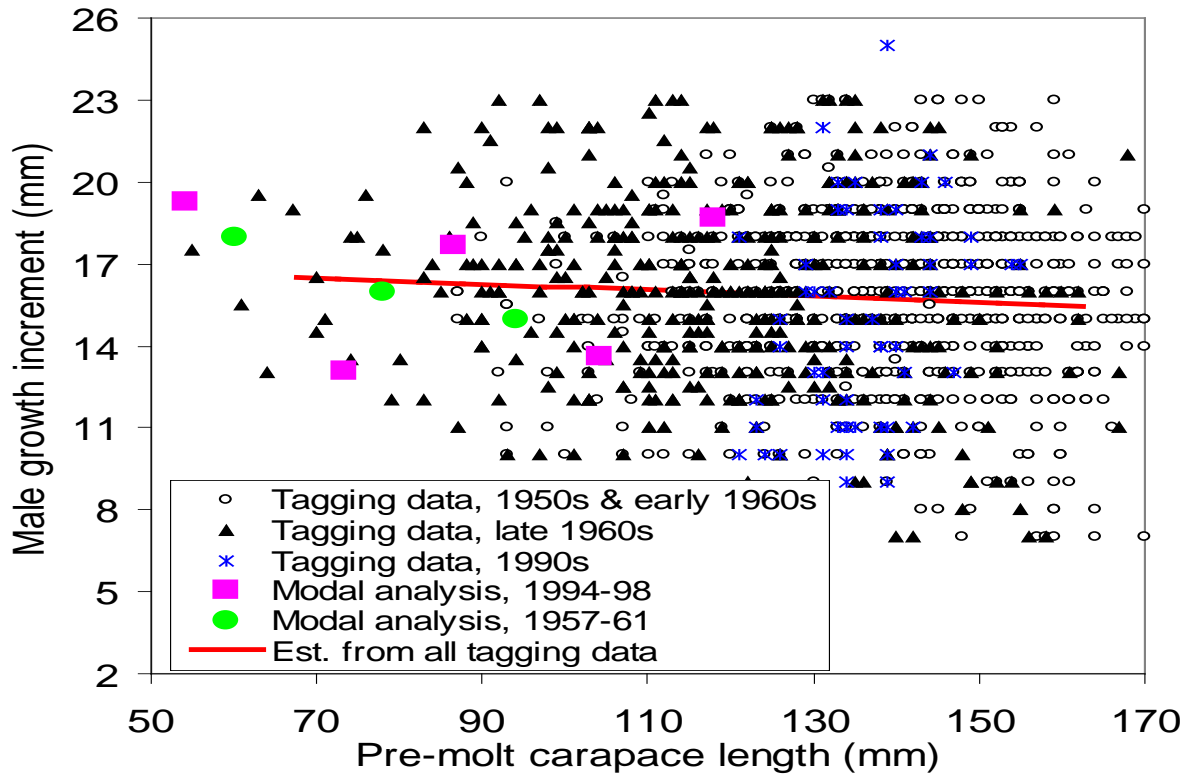


Figure A2. Mean growth increments per molt for Bristol Bay red king crab. Note: “tagging”---based on tagging data; “mode”---based on modal analysis. The female growth increments per molt are for scenarios 1, 1n and 2.

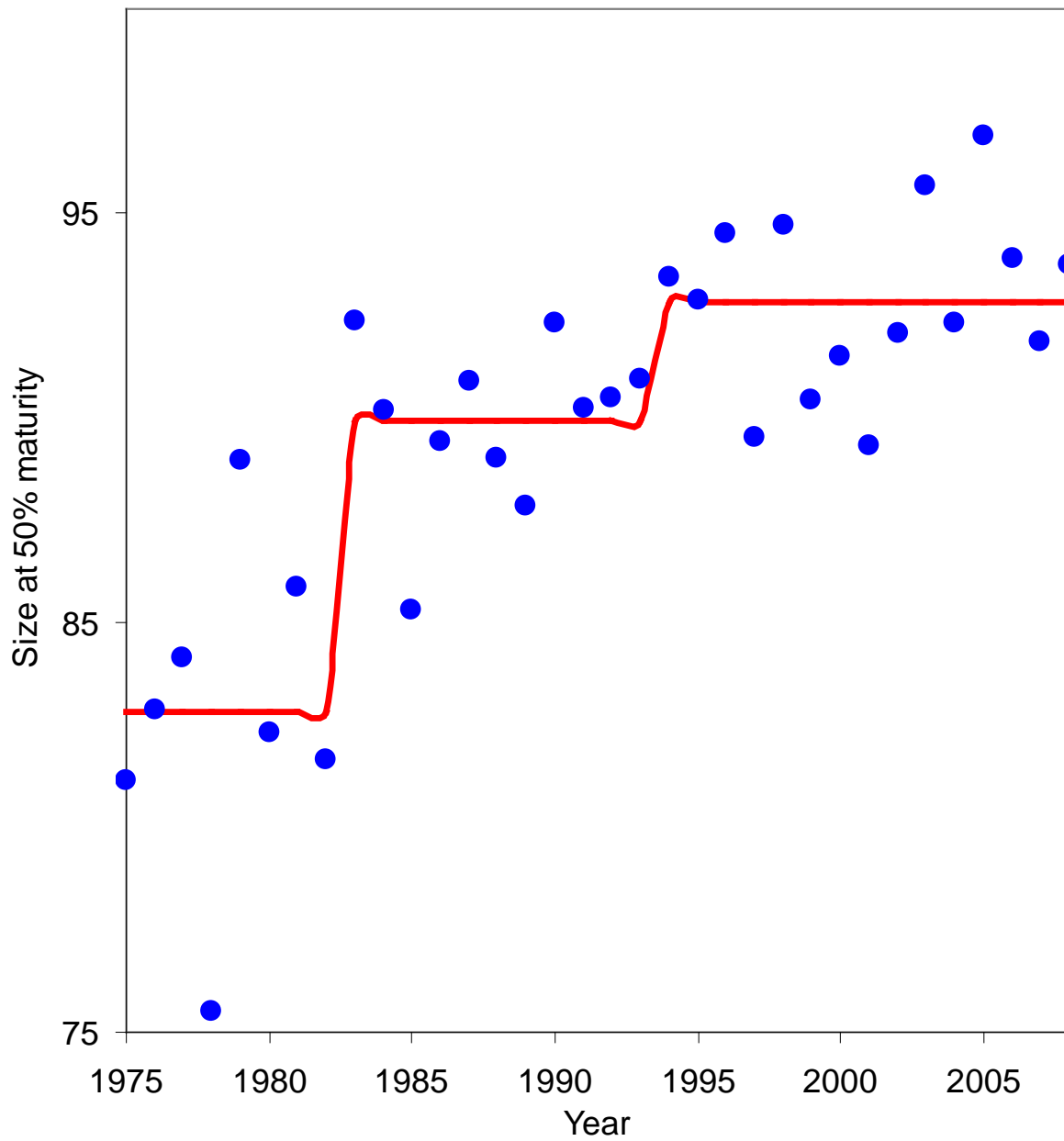


Figure A3. Estimated sizes at 50% maturity for Bristol Bay female red king crab from 1975 to 2008. Averages for three periods (1975-82, 1983-93, and 1994-08) are plotted with a line.

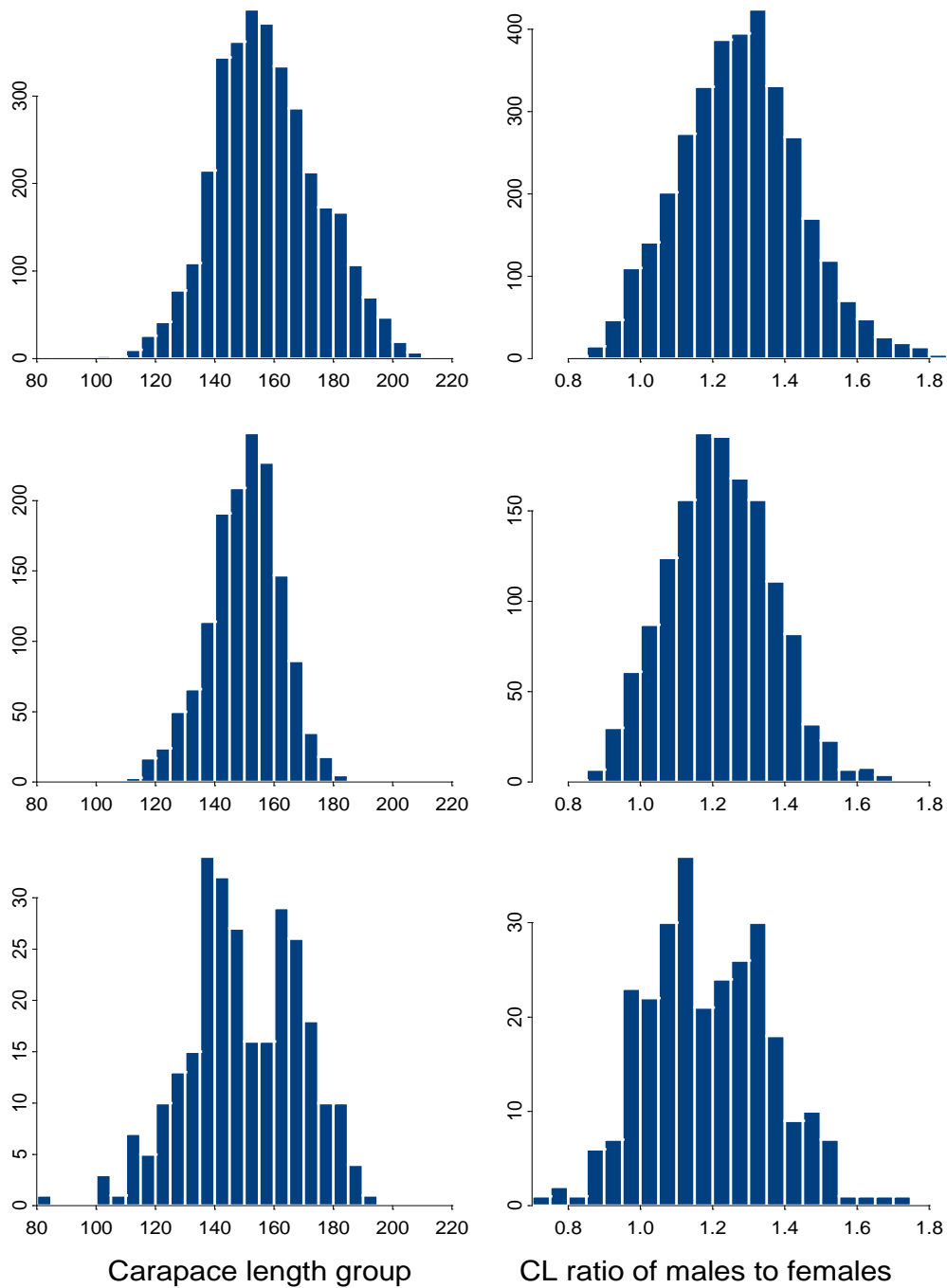


Figure A4. Histograms of carapace lengths (CL) and CL ratios of males to females for male shell ages ≤ 13 months of red king crab males in grasping pairs; Powell's Kodiak data. Upper plot: all locations and years pooled; middle plot: location 11; lower plot: locations 4 and 13. Sizes at maturity for Kodiak red king crab are about 15 mm larger than those for Bristol Bay red king crab. (Doug Pengilly, ADF&G, pers. comm.).

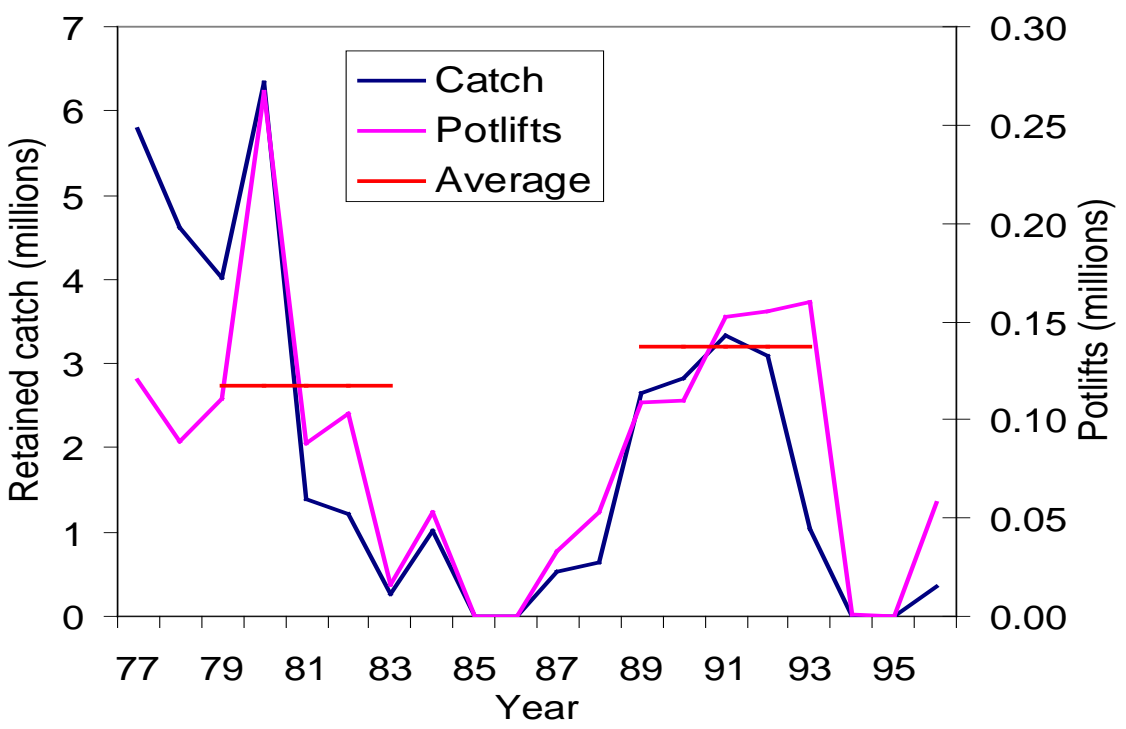
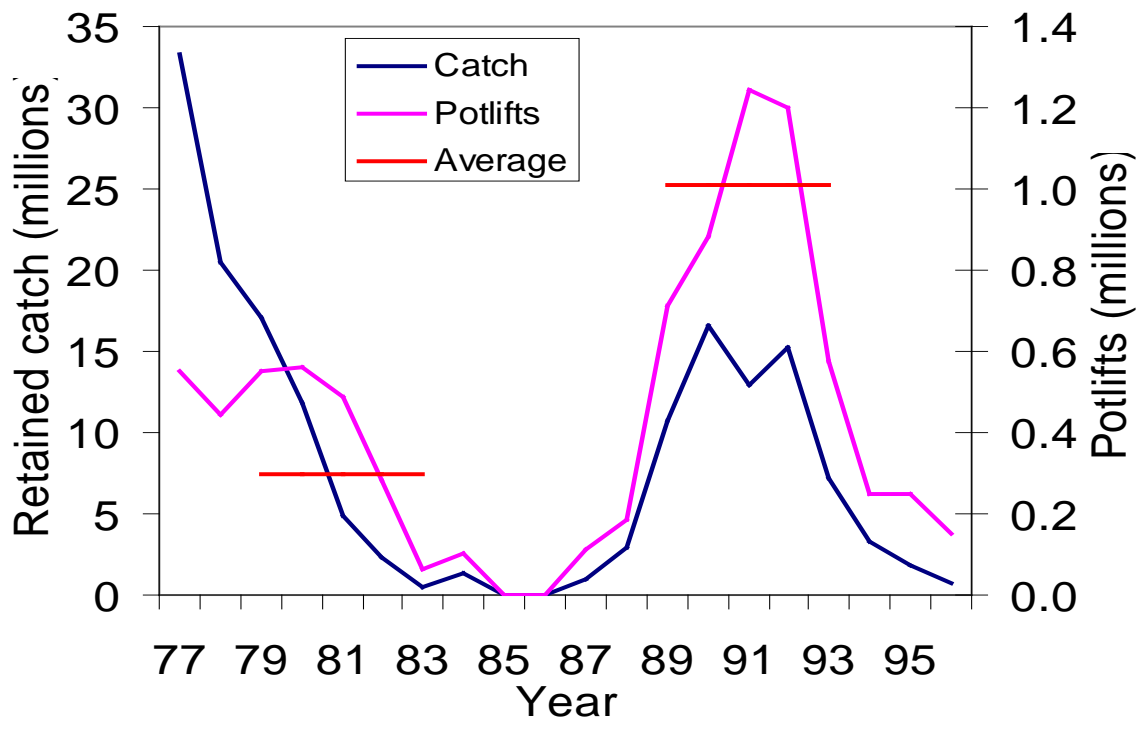


Figure A5. Retained catch and potlifts for total eastern Bering Sea Tanner crab fishery (upper plot) and the Tanner crab fishery east of 163° W (bottom).

Appendix B. Recruitment Breakpoint Analysis

Introduction

SSC asked authors to conduct a recruitment breakpoint analysis similar to that conducted for eastern Bering Sea Tanner crab in 2013 (Stockhausen 2013). We obtained the R codes from Dr. William (Buck) Stockhausen of NMFS and slightly modified them to conduct the analysis for Bristol Bay red king crab for better understanding the temporal change of stock productivity and the recruitment time series used for overfishing/overfished definitions. Results from assessment model scenario 2d were used for this analysis. We are very grateful for the help of Dr. Stockhausen for this analysis.

Methods

The methods are the same as Punt et al. (2014) and Stockhausen (2013). Stock productivity is represented by $\ln(R/MMB)$, where R is recruitment and MMB is mature male biomass, with recruitment lagging to the brood year of mature biomass. Let $y_t = \ln(R/MMB)$ and y_t can be estimated directly from the stock assessment model as observed values or from a stock-recruitment model as \hat{y}_t . For Ricker stock-recruitment models,

$$\begin{aligned}\hat{y}_t &= \alpha_1 + \beta_1 \cdot MMB & t < b, \\ \hat{y}_t &= \alpha_2 + \beta_2 \cdot MMB & t \geq b,\end{aligned}\tag{1}$$

where α_1 and β_1 are the Ricker stock-recruit function parameters for the early time period before the potential breakpoint in year b and α_2 and β_2 are the parameters for the time period after the breakpoint in year b . For Beverton-Holt stock-recruitment models,

$$\begin{aligned}\hat{y}_t &= \alpha_1 - \log(1 + e^{\beta_1} \cdot MMB) & t < b, \\ \hat{y}_t &= \alpha_2 + \log(1 + e^{\beta_2} \cdot MMB) & t \geq b,\end{aligned}\tag{2}$$

where α_1 and β_1 are the Beverton-Holt stock-recruit function log-transformed parameters for the early time period before the potential breakpoint in year b and α_2 and β_2 are the log-transformed parameters for the time period after the breakpoint in year b .

A maximum likelihood approach was used to estimate stock-recruitment model and error parameters. Because y_t is measured with error, the negative log-likelihood function is

$$-\ln(L) = 0.5 \cdot \ln(|\mathbf{\Omega}|) + 0.5 \cdot \sum_t \sum_j (y_t - \hat{y}_t) \cdot [\mathbf{\Omega}^{-1}]_{t,j} \cdot (y_j - \hat{y}_j),\tag{3}$$

where $\mathbf{\Omega}$ contains observation and process error as

$$\mathbf{\Omega} = \mathbf{O} + \mathbf{P},\tag{4}$$

where \mathbf{O} is the observation error covariance matrix estimated from the stock assessment model and \mathbf{P} is the process error matrix and is assumed to reflect a first-order autoregressive process to

have σ^2 on the diagonal and $\sigma^2 \rho^{|t-j|}$ on the off-diagonal elements. σ^2 represents process error variance and ρ represents the degree of autocorrelation.

For each candidate breakpoint year b , the negative log likelihood value of equation (3) was minimized with respect to the six model parameters: $\alpha_1, \beta_1, \alpha_2, \beta_2, \ln(\sigma)$ and $\tan(\rho)$. The minimum time span considered as a potential regime was 5 years. Each brood year from 1980 to 2005 was evaluated as a potential breakpoint b using time series of $\ln(R/MMB)$ and MMB for brood years 1975-2010. A model with no breakpoint was also evaluated. Models with different breakpoints were then ranked using AICc (AIC corrected for small sample size; Burnham and Anderson 2004),

$$AIC_c = -2 \cdot \ln(L) + \frac{2 \cdot k \cdot (k + 1)}{n - k - 1}, \quad (5)$$

where k is the number of parameters and n is the number of observations. Using AICc, the model with the smallest AICc is regarded as the “best” model among the set of models evaluated. Different models can be compared in terms of θ_m , the relative probability (odds) that the model with the minimum AICc score is a better model than model m , where

$$\theta_m = \exp([(AICc_m - AICc_{\min})/2]). \quad (6)$$

Results

Results are summarized in Tables B1-B4 and Figures B1-B6. Discarding the implausible breakpoint year of 1980 for the Ricker model due to implausible stock-recruitment model parameters, both Ricker model and Beverton-Holt model result in the same breakpoint brood year of 1986, which corresponds to recruitment year of 1992. The model with no breakpoint (i.e., a single time period) is about 5 times less probable than the 1986 breakpoint model for Beverton-Holt stock-recruitment models and about eight times less probable for Ricker stock-recruitment relationships, which may suggest a possible change in stock productivity from the early high period to the recent low period. Alternative breakpoint brood years of 1980-1985 for both Ricker and Beverton-Holt models are also reasonably reported. Both Ricker and Beverton-Holt stock-recruitment models fit the data poorly.

Discussion

A recruitment breakpoint analysis was conducted on Bristol Bay red king crab by Punt et al. (2014) with data from 1968 to 2010 to estimate a breakpoint brood year of 1984, corresponding to recruitment year of 1990, which is two years earlier than our estimate, even though our results show that brood year of 1984 is also a likely breakpoint. The different time series of data may explain the different results. Our data start in 1975 and have only two brood-year data points before the regime shift of 1976/77 and thus we cannot detect any stock productivity changes due to the 1976/77 regime shift because of lack of data. Without the early data, the fits of stock-recruitment models to the data are also more poorly.

Time series of estimated recruitment during 1984-present have been used to compute Bmsy proxy. The mean recruitment with scenario 2d during 1984-present is 17.77 million of crab, compared with the mean recruitment of 15.45 million of crab during 1992-present, about 13.0% reduction

(Figure 12(2d)). If the estimated breakpoint year is used to set the new recruitment time series, estimated Bmsy proxy will be correspondingly lower than the current estimated value.

References

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Table B1. Results of the breakpoint analysis, with AICc and the relative probability (odds) against the Ricker stock-recruitment model being correct by breakpoint year. The model with no

breakpoint is listed first in the table. The “best” model is shaded with a plausible stock-recruitment model. Years are brood year.

| Year | AICc | Odds |
|------|---------|---------|
| NA | 46.4933 | 15.0232 |
| 1980 | 41.0741 | 1.0000 |
| 1981 | 43.5372 | 3.4266 |
| 1982 | 43.4335 | 3.2535 |
| 1983 | 43.5460 | 3.4417 |
| 1984 | 43.5839 | 3.5075 |
| 1985 | 43.0025 | 2.6227 |
| 1986 | 42.4169 | 1.9570 |
| 1987 | 45.4294 | 8.8255 |
| 1988 | 46.1588 | 12.7097 |
| 1989 | 49.4106 | 64.6036 |
| 1990 | 46.6891 | 16.5684 |
| 1991 | 47.9850 | 31.6723 |
| 1992 | 48.2826 | 36.7550 |
| 1993 | 48.0169 | 32.1822 |
| 1994 | 48.9392 | 51.0375 |
| 1995 | 48.9373 | 50.9899 |
| 1996 | 49.2335 | 59.1297 |
| 1997 | 48.8284 | 48.2862 |
| 1998 | 48.8394 | 48.5532 |
| 1999 | 48.8440 | 48.6658 |
| 2000 | 46.3349 | 13.8795 |
| 2001 | 45.4607 | 8.9648 |
| 2002 | 45.5360 | 9.3088 |
| 2003 | 45.9752 | 11.5951 |
| 2004 | 46.2300 | 13.1701 |
| 2005 | 45.8085 | 10.6673 |

Table B2. Parameter estimates and standard deviations for the Ricker stock-recruitment model with no breakpoint (first row) and the single breakpoint models (by year of breakpoint). The “best” model is shaded. Years are brood year.

| Year | α_1 | std.dev. | α_2 | std.dev. | β_1 | std.dev. | β_2 | std.dev. | $\ln(\sigma)$ | std.dev. | $\tan(\rho)$ | std.dev. |
|------|------------|----------|------------|----------|-----------|----------|-----------|----------|---------------|----------|--------------|----------|
| | | | -0.523 | 0.319 | | | 0.005 | 0.008 | 0.001 | 0.122 | 0.191 | 0.285 |
| 1980 | -7.356 | 5.342 | 0.708 | 0.505 | -0.077 | 0.061 | 0.061 | 0.021 | -0.117 | 0.122 | -0.052 | 0.286 |
| 1981 | 0.428 | 1.239 | 0.688 | 0.494 | 0.012 | 0.016 | 0.062 | 0.021 | -0.111 | 0.122 | -0.102 | 0.279 |
| 1982 | 0.517 | 0.750 | 0.615 | 0.540 | 0.013 | 0.010 | 0.060 | 0.022 | -0.112 | 0.122 | -0.100 | 0.275 |
| 1983 | 0.337 | 0.582 | 0.675 | 0.602 | 0.011 | 0.008 | 0.062 | 0.024 | -0.111 | 0.122 | -0.107 | 0.273 |
| 1984 | 0.265 | 0.493 | 0.747 | 0.694 | 0.010 | 0.008 | 0.065 | 0.028 | -0.111 | 0.122 | -0.108 | 0.274 |
| 1985 | 0.512 | 0.431 | 0.035 | 0.872 | 0.013 | 0.007 | 0.037 | 0.034 | -0.118 | 0.122 | -0.116 | 0.275 |
| 1986 | 0.500 | 0.397 | -0.677 | 1.148 | 0.013 | 0.007 | 0.011 | 0.044 | -0.132 | 0.122 | -0.083 | 0.281 |
| 1987 | 0.179 | 0.380 | 0.578 | 1.468 | 0.009 | 0.007 | 0.057 | 0.056 | -0.088 | 0.122 | -0.102 | 0.273 |
| 1988 | 0.089 | 0.392 | 0.706 | 1.693 | 0.009 | 0.007 | 0.062 | 0.064 | -0.081 | 0.121 | 0.002 | 0.279 |
| 1989 | -0.174 | 0.384 | 0.819 | 1.738 | 0.007 | 0.007 | 0.063 | 0.066 | -0.038 | 0.121 | -0.029 | 0.281 |
| 1990 | -0.069 | 0.389 | 1.505 | 1.759 | 0.008 | 0.007 | 0.093 | 0.067 | -0.076 | 0.122 | 0.080 | 0.274 |
| 1991 | -0.173 | 0.385 | 1.457 | 1.805 | 0.007 | 0.008 | 0.090 | 0.069 | -0.057 | 0.122 | 0.088 | 0.272 |
| 1992 | -0.342 | 0.374 | 2.270 | 1.875 | 0.005 | 0.008 | 0.118 | 0.071 | -0.051 | 0.122 | 0.090 | 0.271 |
| 1993 | -0.354 | 0.358 | 2.646 | 2.036 | 0.005 | 0.007 | 0.131 | 0.076 | -0.054 | 0.121 | 0.068 | 0.270 |
| 1994 | -0.259 | 0.357 | 1.700 | 2.961 | 0.006 | 0.008 | 0.097 | 0.109 | -0.042 | 0.121 | 0.079 | 0.283 |
| 1995 | -0.290 | 0.344 | 2.037 | 3.181 | 0.006 | 0.007 | 0.109 | 0.116 | -0.041 | 0.121 | 0.064 | 0.276 |
| 1996 | -0.336 | 0.333 | 2.213 | 3.163 | 0.006 | 0.007 | 0.114 | 0.116 | -0.036 | 0.121 | -0.036 | 0.121 |
| 1997 | -0.236 | 0.342 | -0.002 | 3.514 | 0.007 | 0.008 | 0.038 | 0.127 | -0.048 | 0.122 | 0.111 | 0.292 |
| 1998 | -0.293 | 0.322 | 1.265 | 4.351 | 0.006 | 0.007 | 0.082 | 0.156 | -0.044 | 0.121 | 0.060 | 0.272 |
| 1999 | -0.298 | 0.312 | 0.359 | 5.150 | 0.006 | 0.007 | 0.051 | 0.183 | -0.045 | 0.121 | 0.041 | 0.270 |
| 2000 | -0.249 | 0.294 | 2.030 | 5.027 | 0.006 | 0.007 | 0.116 | 0.179 | -0.082 | 0.122 | 0.013 | 0.268 |
| 2001 | -0.260 | 0.275 | 2.972 | 4.984 | 0.006 | 0.006 | 0.153 | 0.178 | -0.096 | 0.122 | -0.060 | 0.268 |
| 2002 | -0.281 | 0.269 | 2.991 | 5.003 | 0.005 | 0.006 | 0.155 | 0.179 | -0.095 | 0.122 | -0.076 | 0.269 |
| 2003 | -0.312 | 0.268 | 3.717 | 5.370 | 0.005 | 0.006 | 0.183 | 0.193 | -0.089 | 0.122 | -0.079 | 0.270 |
| 2004 | -0.336 | 0.266 | 4.122 | 5.359 | 0.005 | 0.006 | 0.200 | 0.193 | -0.086 | 0.122 | -0.078 | 0.267 |
| 2005 | -0.338 | 0.261 | 2.435 | 5.684 | 0.005 | 0.006 | 0.143 | 0.203 | -0.093 | 0.122 | -0.082 | 0.267 |

Table B3. Results of the breakpoint analysis, with AICc and the relative probability (odds) against the Beverton-Holt stock-recruitment model being correct by breakpoint year. The model with no breakpoint is listed first in the table. The “best” model is shaded. Years are brood year.

| Year | AICc | Odds |
|------|---------|---------|
| NA | 45.3981 | 5.0697 |
| 1980 | 43.8995 | 2.3964 |
| 1981 | 42.3954 | 1.1297 |
| 1982 | 42.3742 | 1.1177 |
| 1983 | 42.5415 | 1.2153 |
| 1984 | 42.6196 | 1.2637 |
| 1985 | 42.6775 | 1.3008 |
| 1986 | 42.1516 | 1.0000 |
| 1987 | 45.3144 | 4.8618 |
| 1988 | 45.9970 | 6.8395 |
| 1989 | 49.1365 | 32.8664 |
| 1990 | 47.0869 | 11.7947 |
| 1991 | 48.2198 | 20.7824 |
| 1992 | 49.4103 | 37.6892 |
| 1993 | 49.4378 | 38.2106 |
| 1994 | 49.0962 | 32.2110 |
| 1995 | 49.2897 | 35.4830 |
| 1996 | 49.7282 | 44.1816 |
| 1997 | 48.3534 | 22.2179 |
| 1998 | 48.8959 | 29.1420 |
| 1999 | 48.7480 | 27.0641 |
| 2000 | 46.5764 | 9.1378 |
| 2001 | 45.9210 | 6.5844 |
| 2002 | 45.8966 | 6.5046 |
| 2003 | 46.4147 | 8.4280 |
| 2004 | 46.6195 | 9.3366 |
| 2005 | 45.6408 | 5.7238 |

Table B4. Parameter estimates and standard deviations for the Beverton-Holt stock-recruitment model with no breakpoint (first row) and the single breakpoint models (by year of breakpoint). The “best” model is shaded. Years are brood year.

| Year | α_1 | std.dev. | α_2 | std.dev. | β_1 | std.dev. | β_2 | std.dev. | $\ln(\sigma)$ | std.dev. | $\tan(\rho)$ | std.dev. |
|------|------------|----------|------------|----------|-----------|----------|-----------|----------|---------------|----------|--------------|----------|
| | | | -0.159 | 0.894 | | | -3.713 | 2.225 | -0.005 | 0.123 | 0.215 | 0.295 |
| 1980 | -0.625 | 0.391 | 7.820 | 66.239 | -11.19 | 60.247 | 5.471 | 66.254 | -0.101 | 0.123 | -0.164 | 0.282 |
| 1981 | 1.500 | 4.577 | 7.493 | 50.669 | -2.440 | 5.381 | 5.185 | 50.685 | -0.129 | 0.122 | -0.078 | 0.287 |
| 1982 | 0.796 | 1.109 | 6.982 | 47.358 | -3.321 | 1.661 | 4.681 | 47.381 | -0.129 | 0.122 | -0.097 | 0.276 |
| 1983 | 0.460 | 0.724 | 7.357 | 43.960 | -3.817 | 1.354 | 5.044 | 43.974 | -0.126 | 0.122 | -0.108 | 0.275 |
| 1984 | 0.349 | 0.586 | 8.411 | 65.301 | -3.999 | 1.241 | 6.091 | 65.308 | -0.126 | 0.122 | -0.111 | 0.274 |
| 1985 | 0.666 | 0.573 | 0.959 | 3.804 | -3.492 | 1.065 | -1.508 | 4.519 | -0.123 | 0.122 | -0.108 | 0.276 |
| 1986 | 0.647 | 0.530 | -0.690 | 1.307 | -3.514 | 1.031 | -4.454 | 5.662 | -0.135 | 0.122 | -0.080 | 0.280 |
| 1987 | 0.292 | 0.483 | 5.501 | 41.505 | -3.983 | 1.175 | 3.163 | 41.573 | -0.092 | 0.122 | -0.096 | 0.274 |
| 1988 | 0.227 | 0.528 | 6.910 | 83.603 | -3.992 | 1.316 | 4.571 | 83.636 | -0.084 | 0.121 | 0.031 | 0.276 |
| 1989 | -0.005 | 0.560 | 5.507 | 42.863 | -4.127 | 1.569 | 3.080 | 42.939 | -0.042 | 0.121 | 0.007 | 0.280 |
| 1990 | 0.103 | 0.571 | 5.404 | 31.615 | -4.034 | 1.491 | 3.066 | 31.672 | -0.071 | 0.122 | 0.107 | 0.279 |
| 1991 | 0.016 | 0.593 | 5.997 | 43.869 | -4.059 | 1.603 | 3.631 | 43.913 | -0.054 | 0.122 | 0.107 | 0.276 |
| 1992 | -0.179 | 0.584 | 6.277 | 42.024 | -4.316 | 1.863 | 3.830 | 42.059 | -0.037 | 0.122 | 0.115 | 0.277 |
| 1993 | -0.194 | 0.571 | 6.265 | 41.986 | -4.334 | 1.867 | 3.820 | 42.021 | -0.037 | 0.122 | 0.121 | 0.277 |
| 1994 | -0.049 | 0.608 | 4.133 | 30.922 | -4.054 | 1.719 | 1.753 | 31.120 | -0.040 | 0.122 | 0.135 | 0.282 |
| 1995 | -0.090 | 0.592 | 4.862 | 43.254 | -4.112 | 1.752 | 2.481 | 43.386 | -0.038 | 0.122 | 0.118 | 0.279 |
| 1996 | -0.143 | 0.583 | 4.980 | 43.179 | -4.170 | 1.810 | 2.577 | 43.299 | -0.033 | 0.121 | -0.033 | 0.121 |
| 1997 | -0.027 | 0.598 | 0.689 | 17.930 | -4.018 | 1.685 | -1.771 | 21.766 | -0.052 | 0.122 | 0.129 | 0.297 |
| 1998 | -0.112 | 0.548 | 3.575 | 39.931 | -4.175 | 1.718 | 1.269 | 40.335 | -0.047 | 0.122 | 0.078 | 0.275 |
| 1999 | -0.124 | 0.528 | 1.114 | 24.395 | -4.213 | 1.703 | -1.266 | 27.474 | -0.050 | 0.121 | 0.051 | 0.273 |
| 2000 | -0.096 | 0.481 | 3.838 | 44.284 | -4.274 | 1.592 | 1.729 | 44.563 | -0.084 | 0.122 | 0.030 | 0.272 |
| 2001 | -0.117 | 0.449 | 5.966 | 109.07 | -4.344 | 1.556 | 3.936 | 109.14 | -0.094 | 0.122 | -0.033 | 0.270 |
| 2002 | -0.133 | 0.450 | 4.710 | 58.628 | -4.345 | 1.571 | 2.726 | 58.765 | -0.094 | 0.122 | -0.038 | 0.269 |
| 2003 | -0.150 | 0.470 | 4.518 | 51.104 | -4.308 | 1.611 | 2.561 | 51.245 | -0.086 | 0.122 | -0.031 | 0.269 |
| 2004 | -0.169 | 0.476 | 4.207 | 43.439 | -4.307 | 1.638 | 2.300 | 43.595 | -0.082 | 0.121 | -0.036 | 0.269 |
| 2005 | -0.176 | 0.459 | 2.668 | 27.512 | -4.331 | 1.609 | 0.892 | 27.915 | -0.096 | 0.122 | -0.058 | 0.268 |

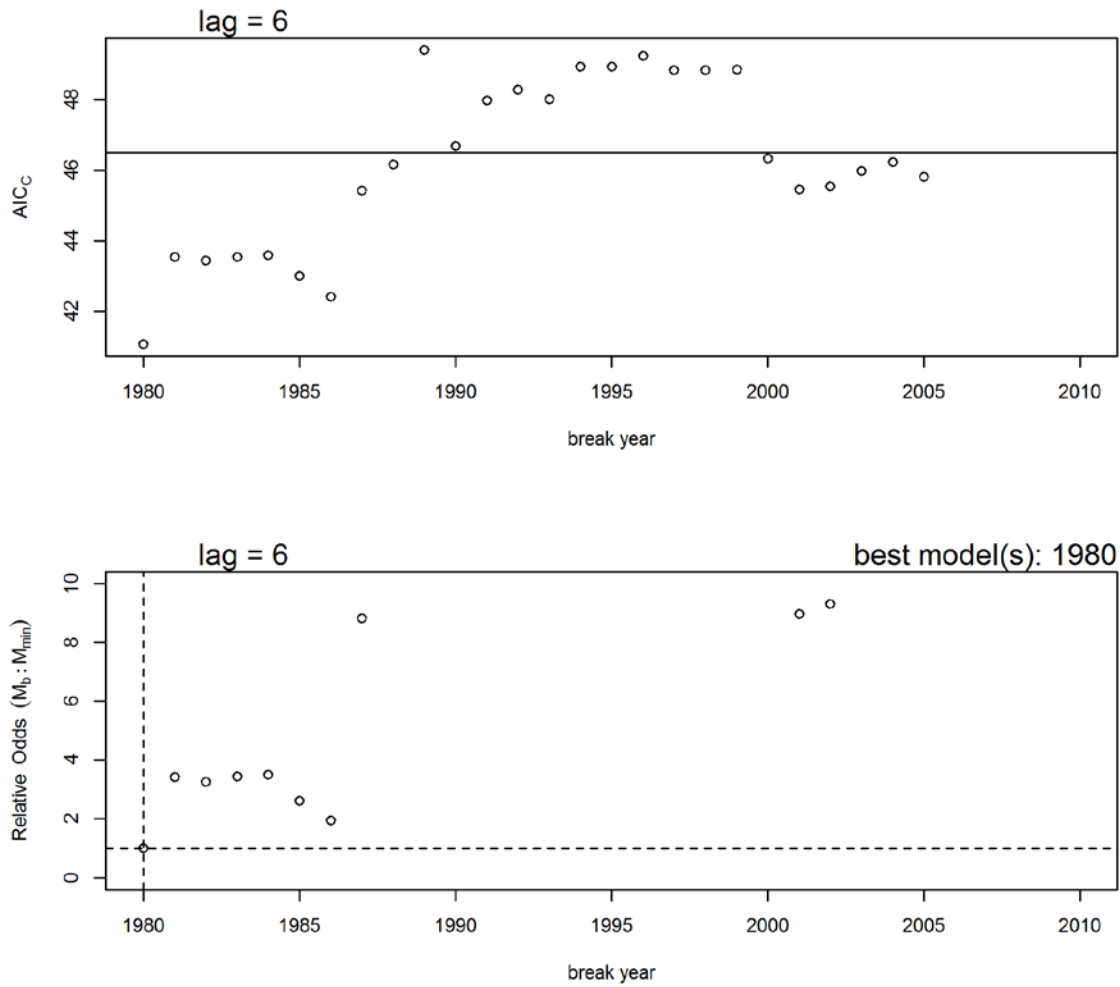


Figure B1. Results from the Ricker stock-recruit breakpoint analysis. Upper graph: AIC_c vs. year of breakpoint for the 1-breakpoint models (circles) and AIC_c for the model with no breakpoint (horizontal line). Lower graph: probabilistic odds for all 1-breakpoint models (circles) and the no breakpoint model (horizontal solid line) relative to the model with the smallest AIC_c score. The dashed lines indicate the value for the model with the lowest AIC_c score. Not shown are 1-breakpoint models with high odds (>10) of being incorrect.

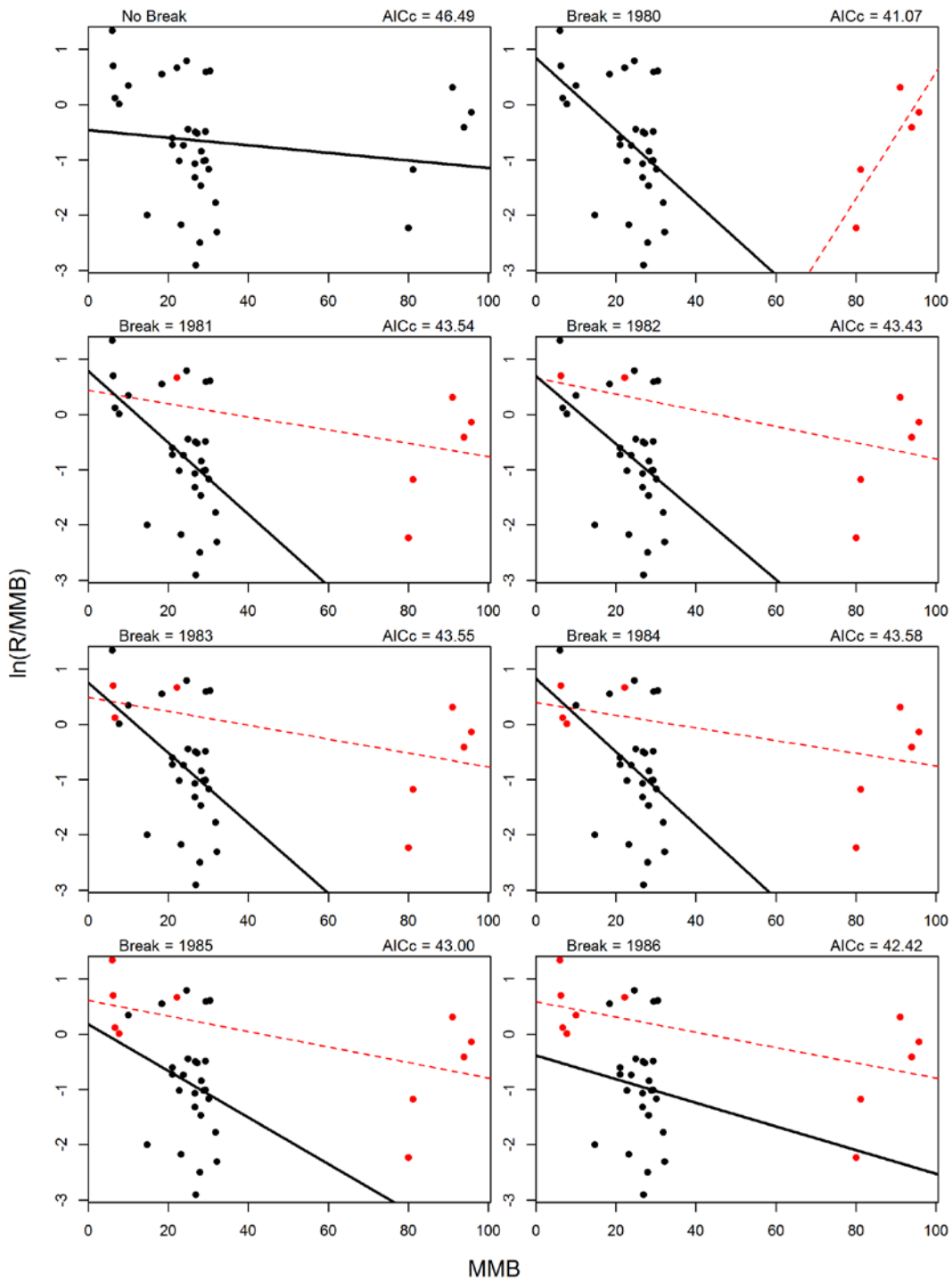


Figure B2. Fits for Ricker models with no breakpoint (upper left graph) and with 1-breakpoint for break years 1975-2005. For 1-breakpoint models, the pre-break data (circles) and model fit (line) are shown in red, whereas the post-break data and fit are shown in black.

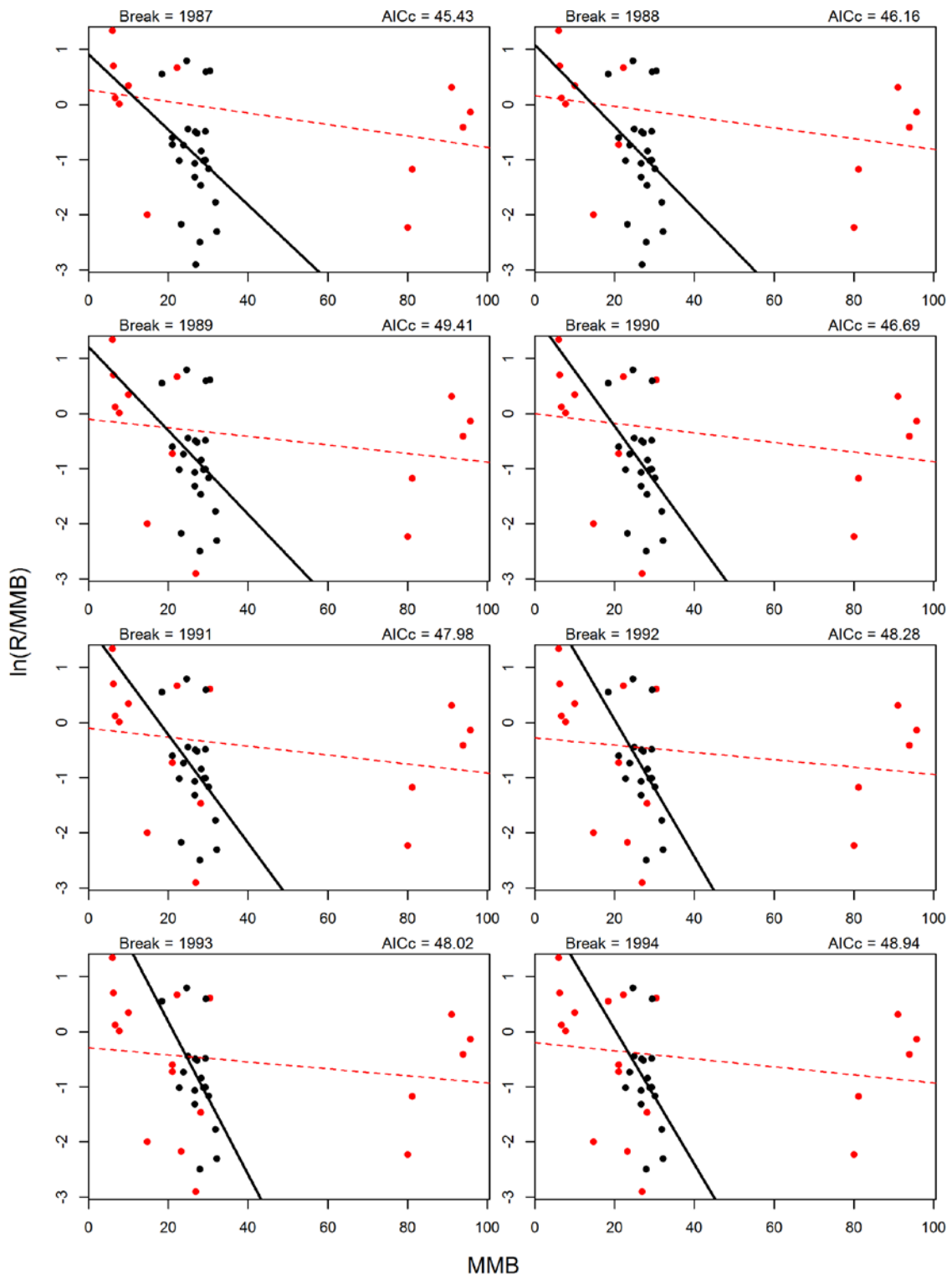


Figure B2. Continue.

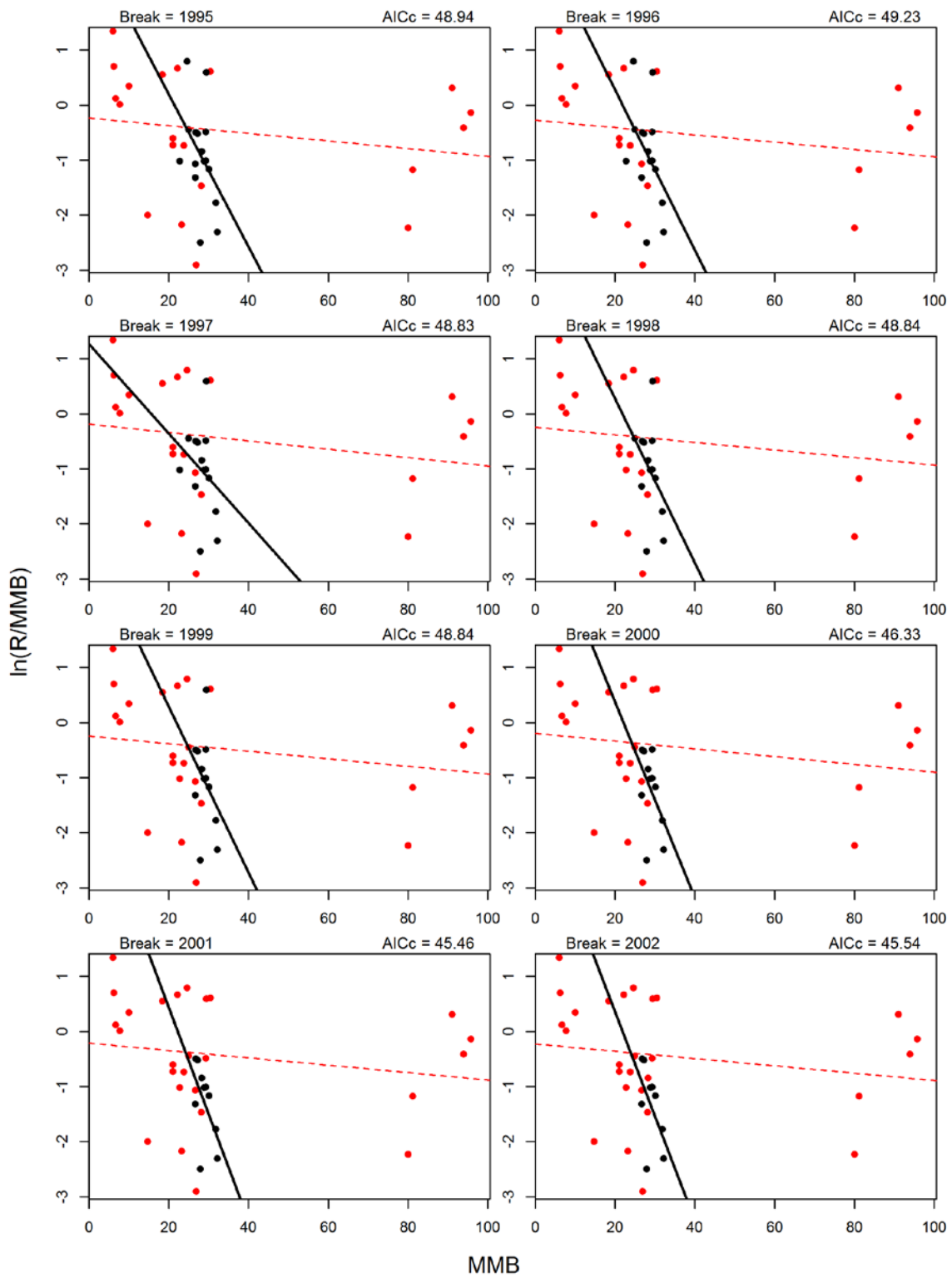
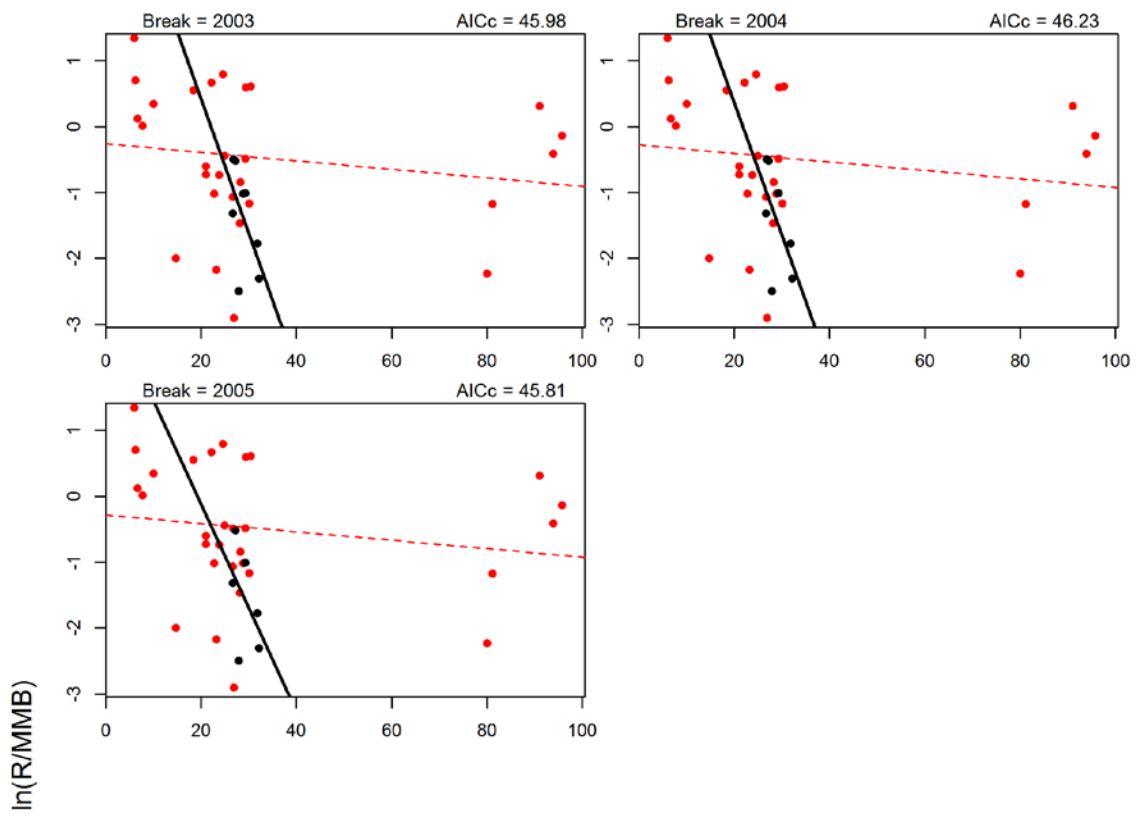


Figure B2. Continue.



MMB

Figure B2. Continue.

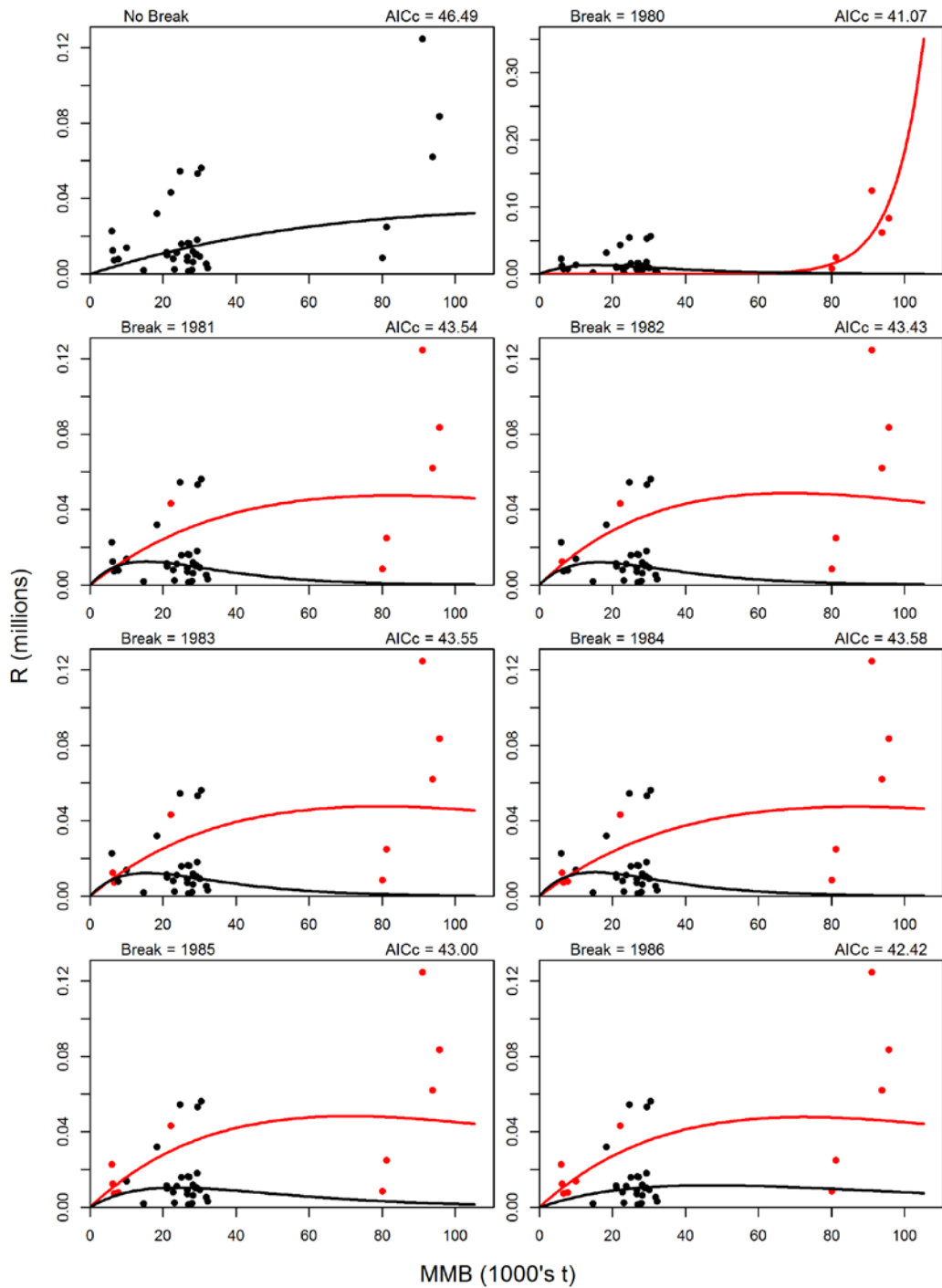


Figure B3. Fits on the arithmetic scale for Ricker models with no breakpoint (upper left graph) and with 1-breakpoint for break years 1975-2005. For 1-breakpoint models, the pre-break data (circles) and model fit (line) are shown in red, whereas the post-break data and fit are shown in black.

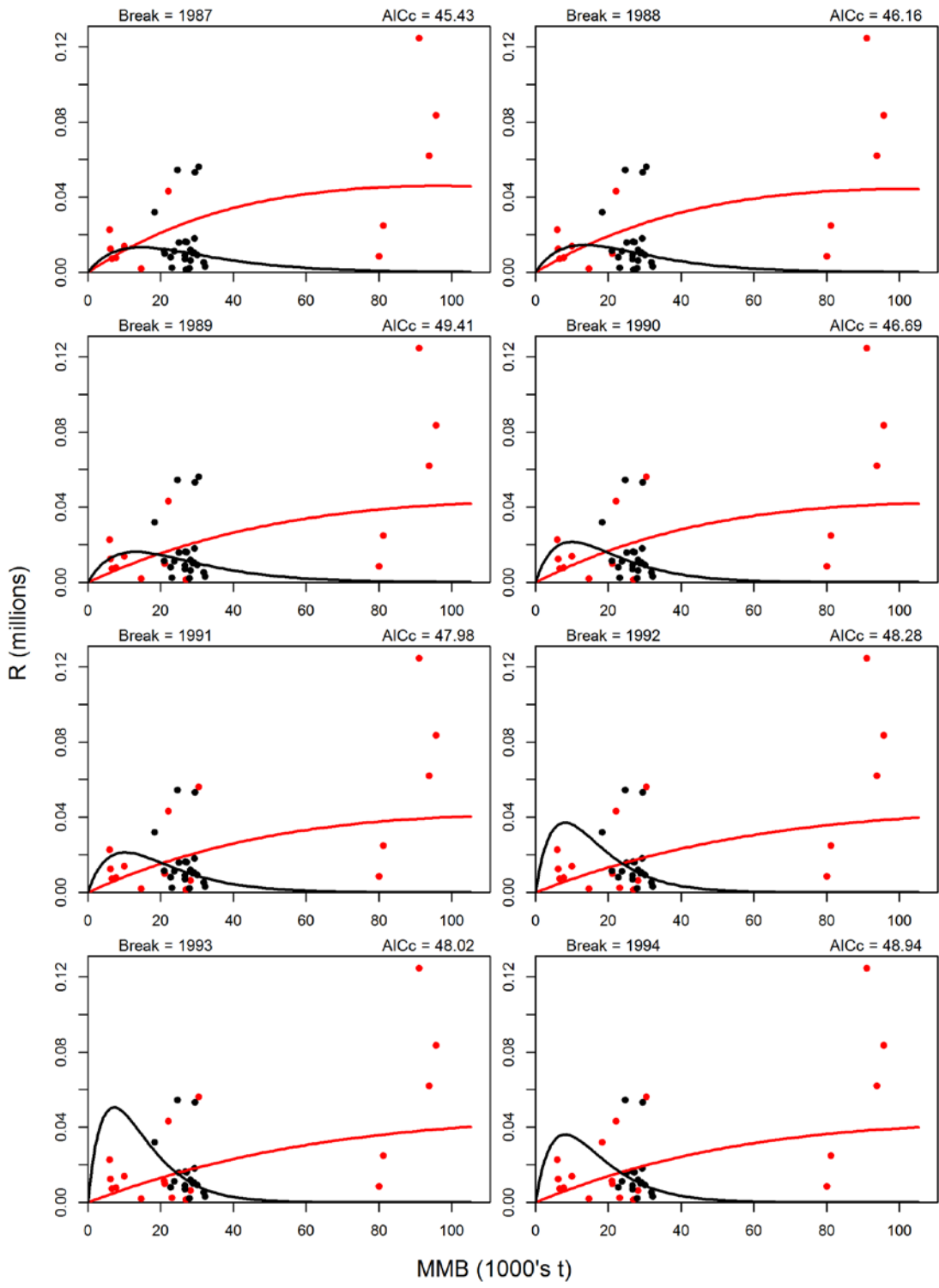


Figure B3. Continue.

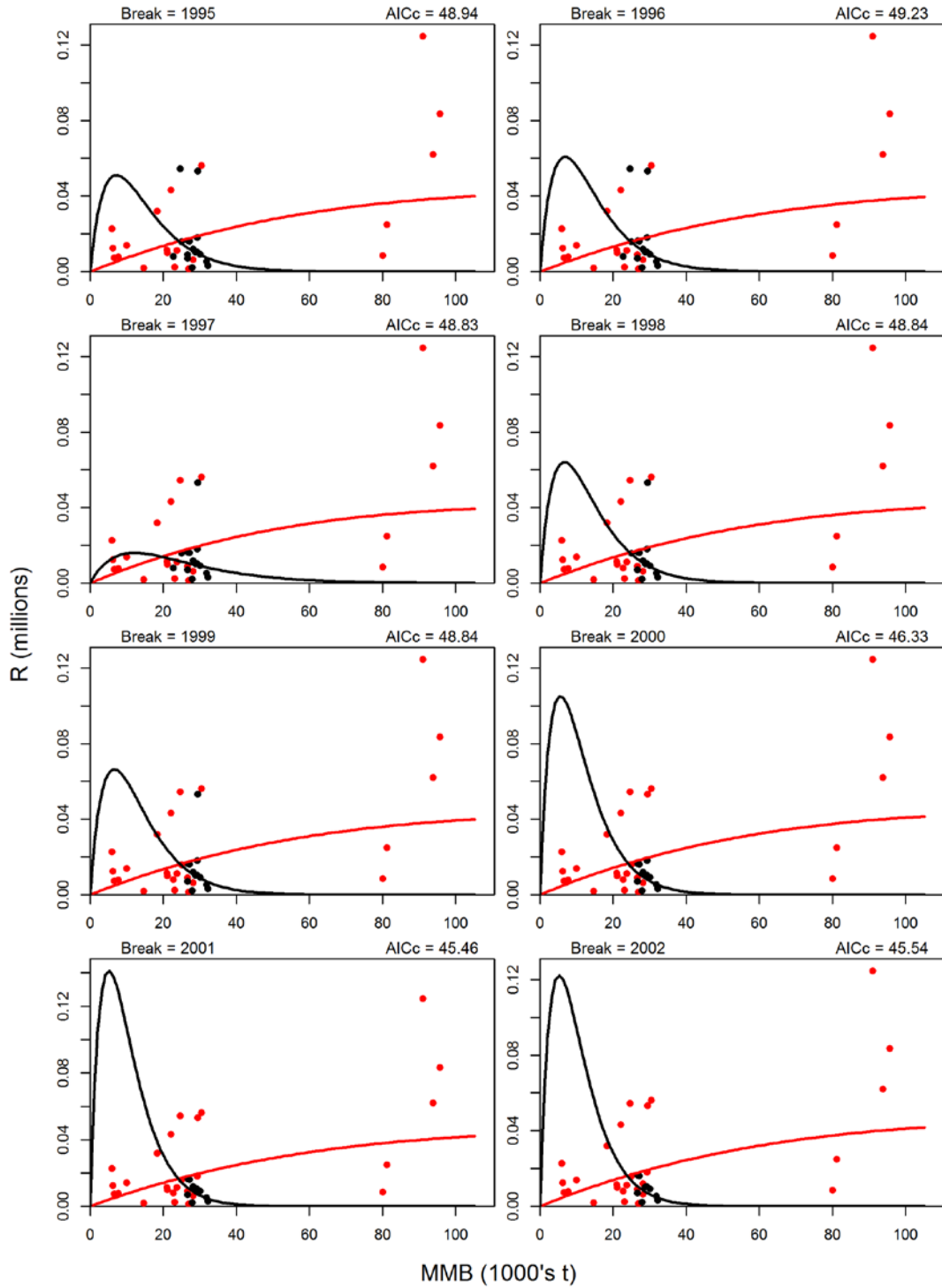
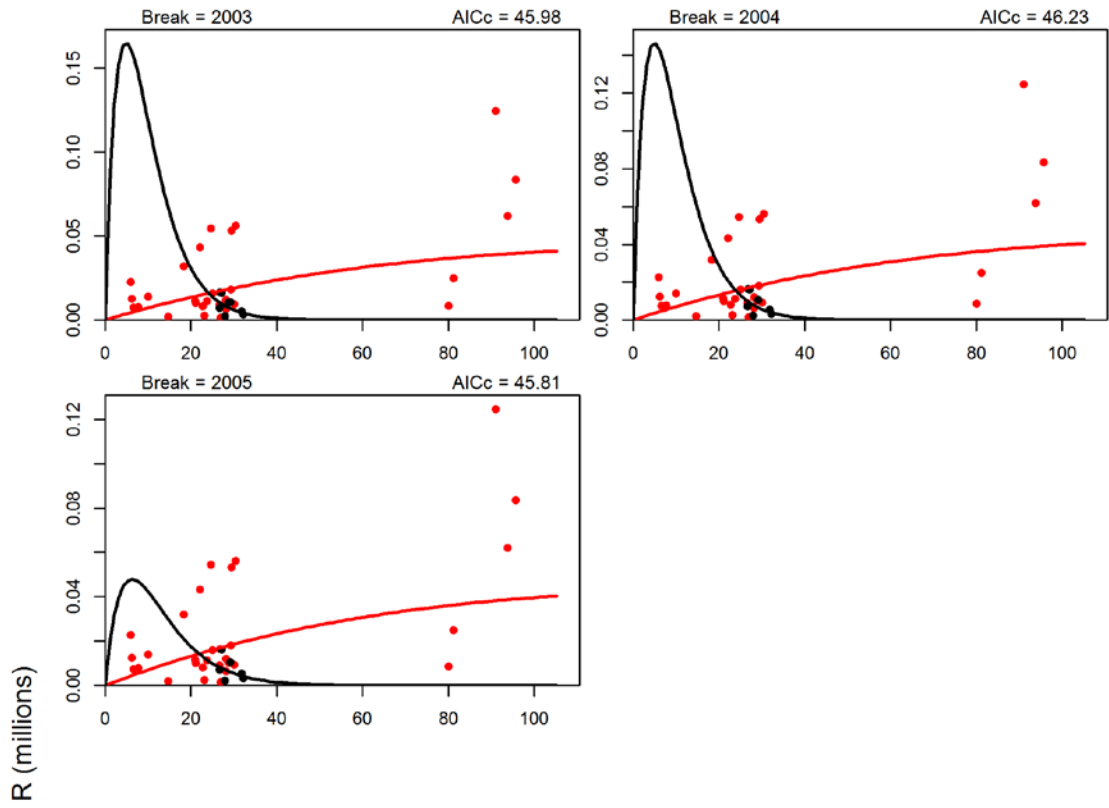


Figure B3. Continue.



MMB (1000's t)

Figure B3. Continue.

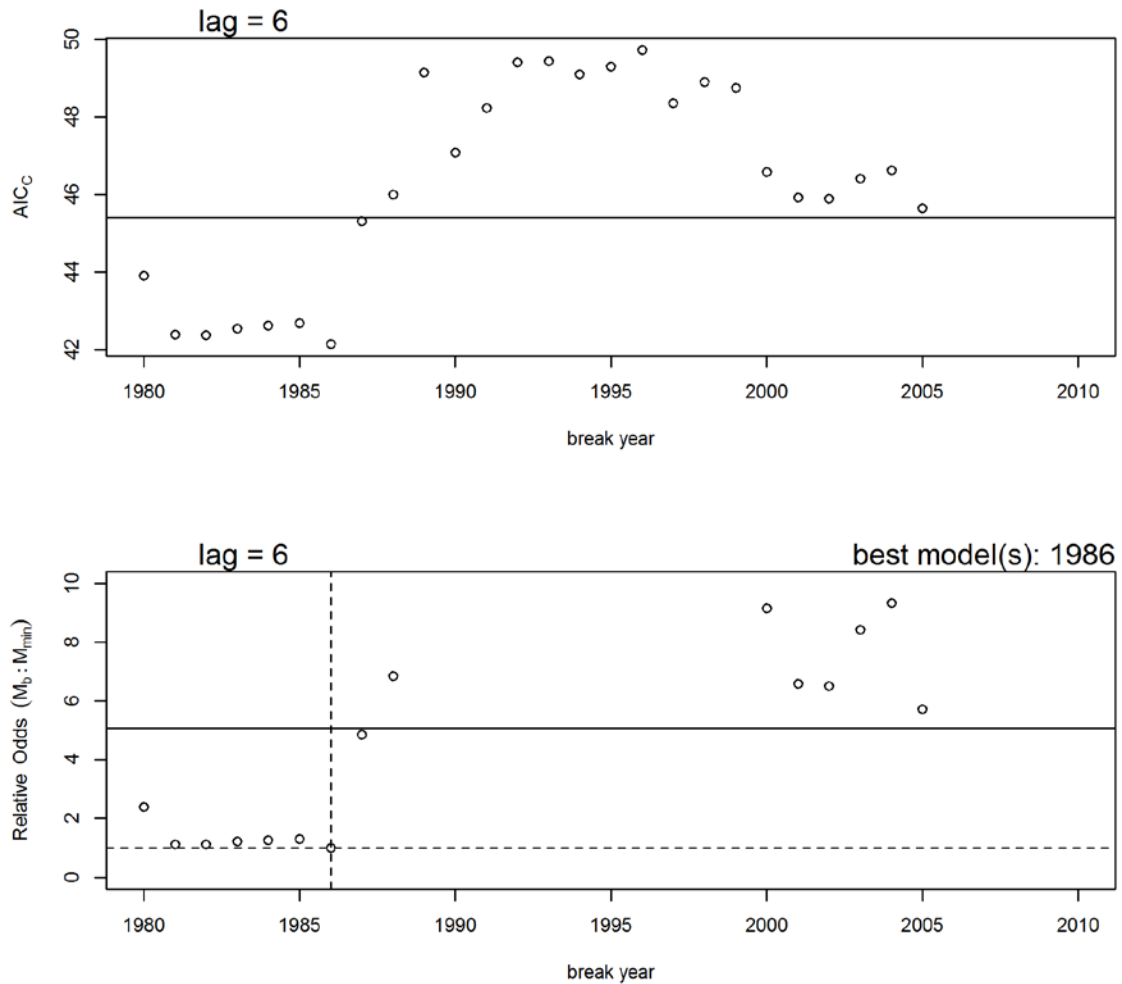


Figure B4. Results from the B-H stock-recruit breakpoint analysis. Upper graph: AICc vs. year of breakpoint for the 1-breakpoint models (circles) and AICc for the model with no breakpoint (horizontal line). Lower graph: probabilistic odds for all 1-breakpoint models (circles) and the no breakpoint model (horizontal solid line) relative to the model with the smallest AICc score. The dashed lines indicate the value for the model with the lowest AICc score (breakpoint in 1986). Not shown are 1-breakpoint models with high odds (>10) of being incorrect.

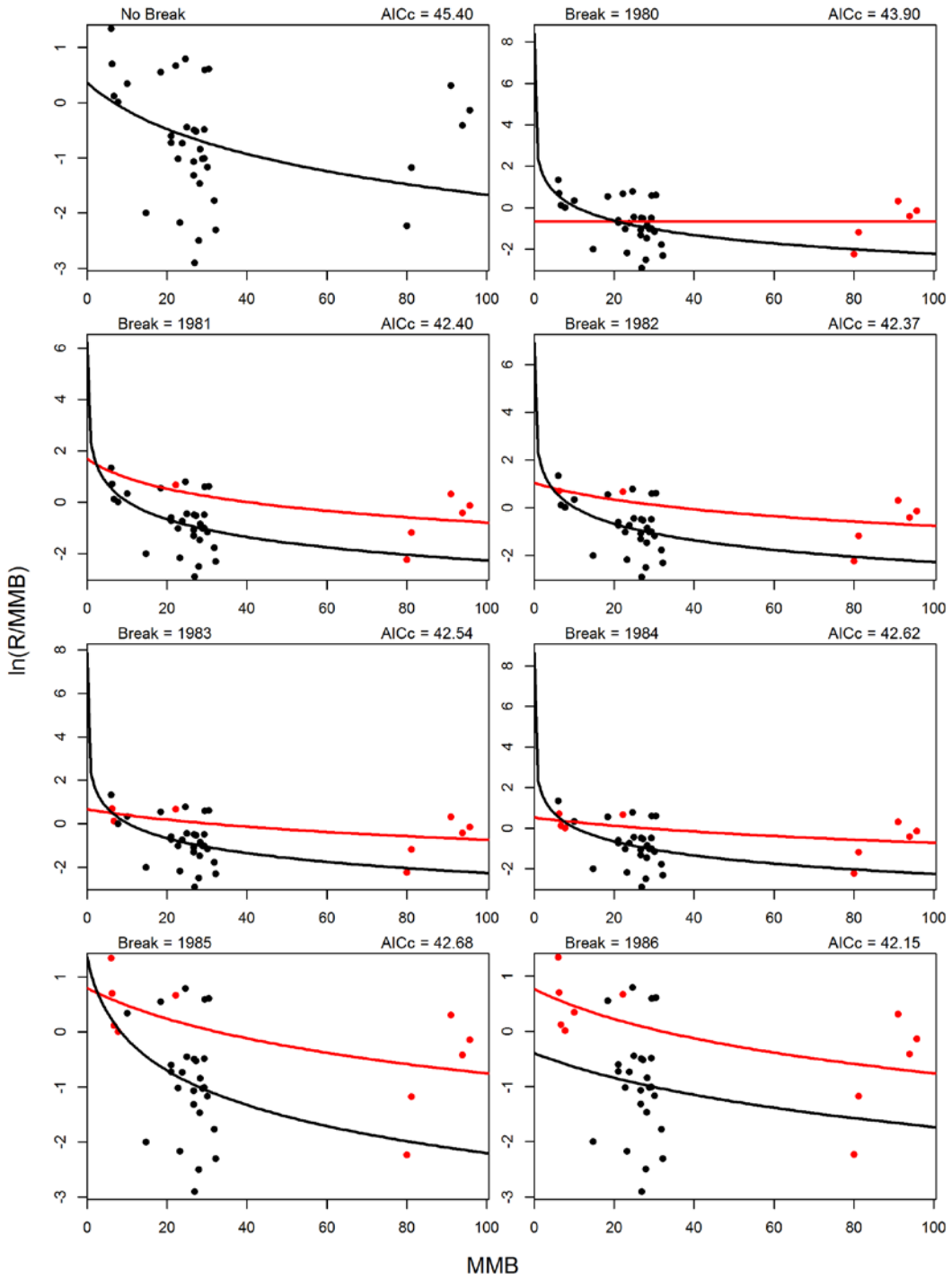


Figure B5. Fits for B-H models with no breakpoint (upper left graph) and with 1-breakpoint for break years 1975-2005. For 1-breakpoint models, the pre-break data (circles) and model fit (line) are shown in red, whereas the post-break data and fit are shown in black.

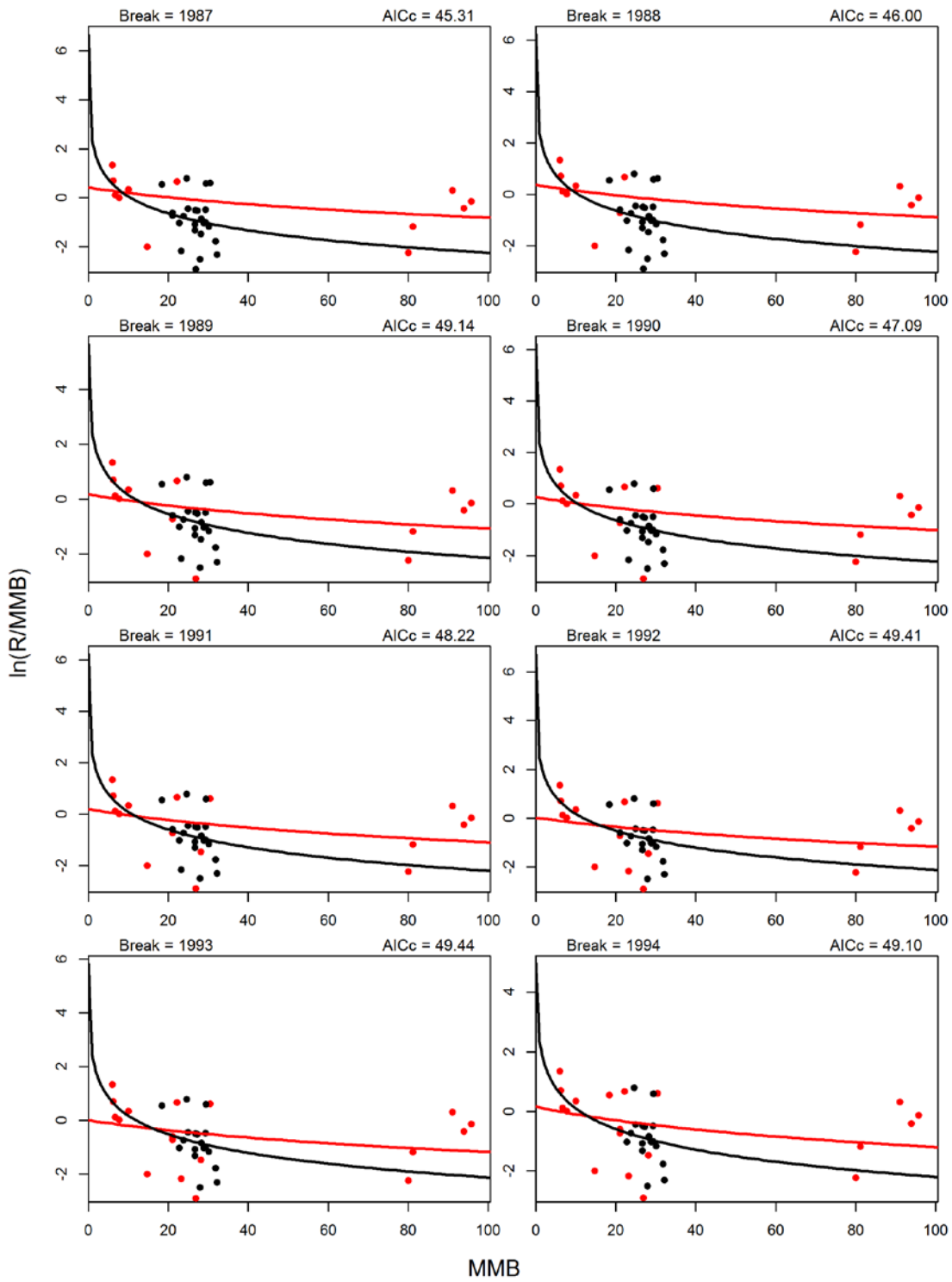


Figure B5. Continue.

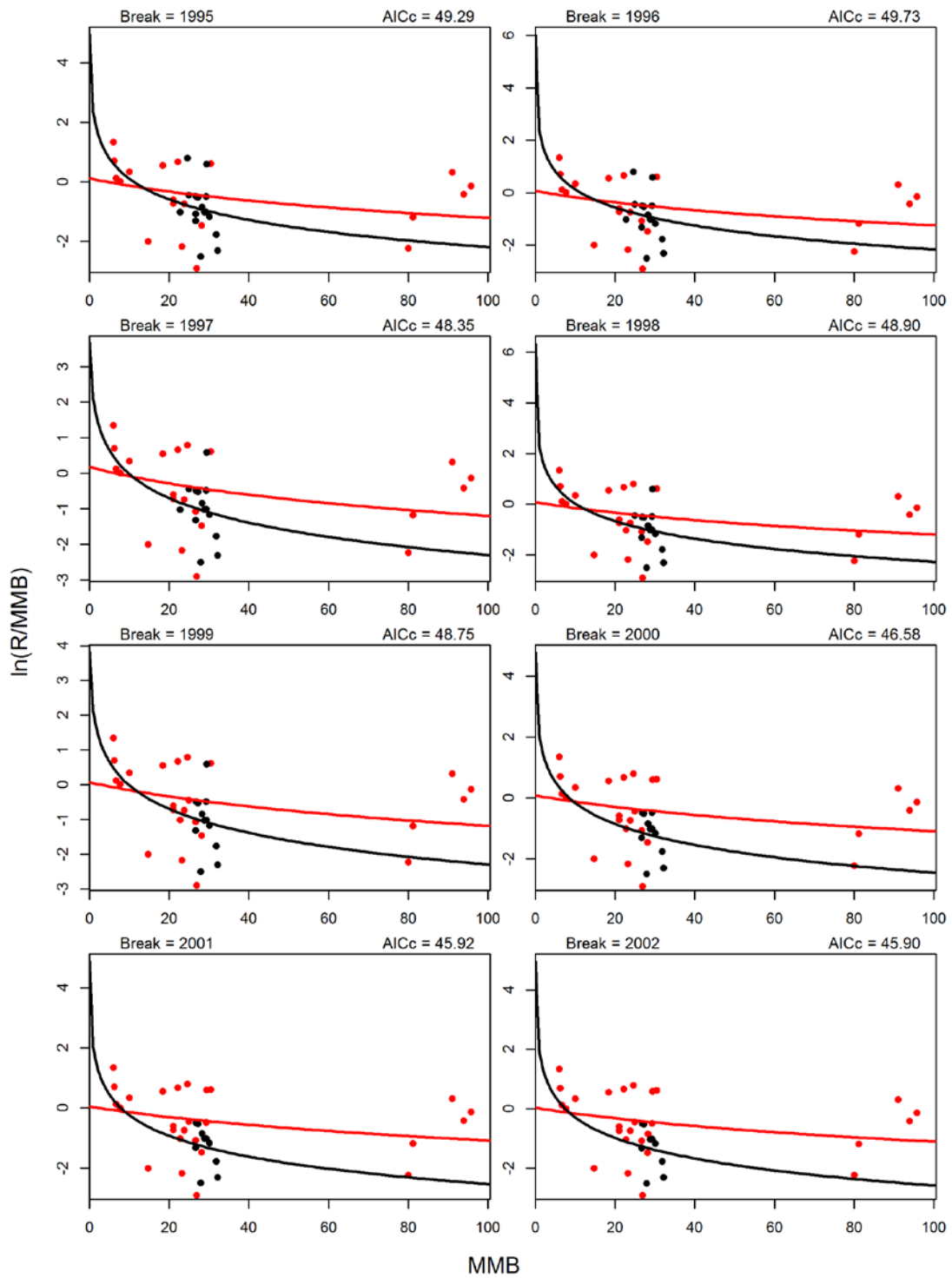
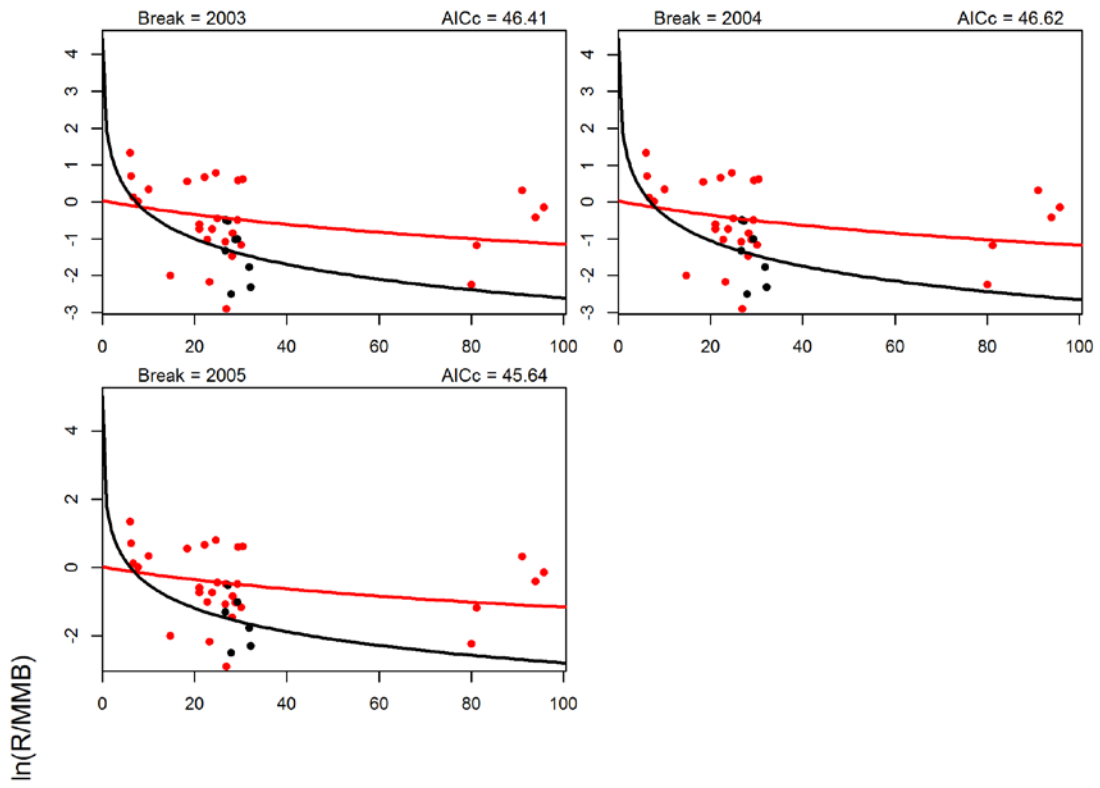


Figure B5. Continue.



MMB

Figure B5. Continue.

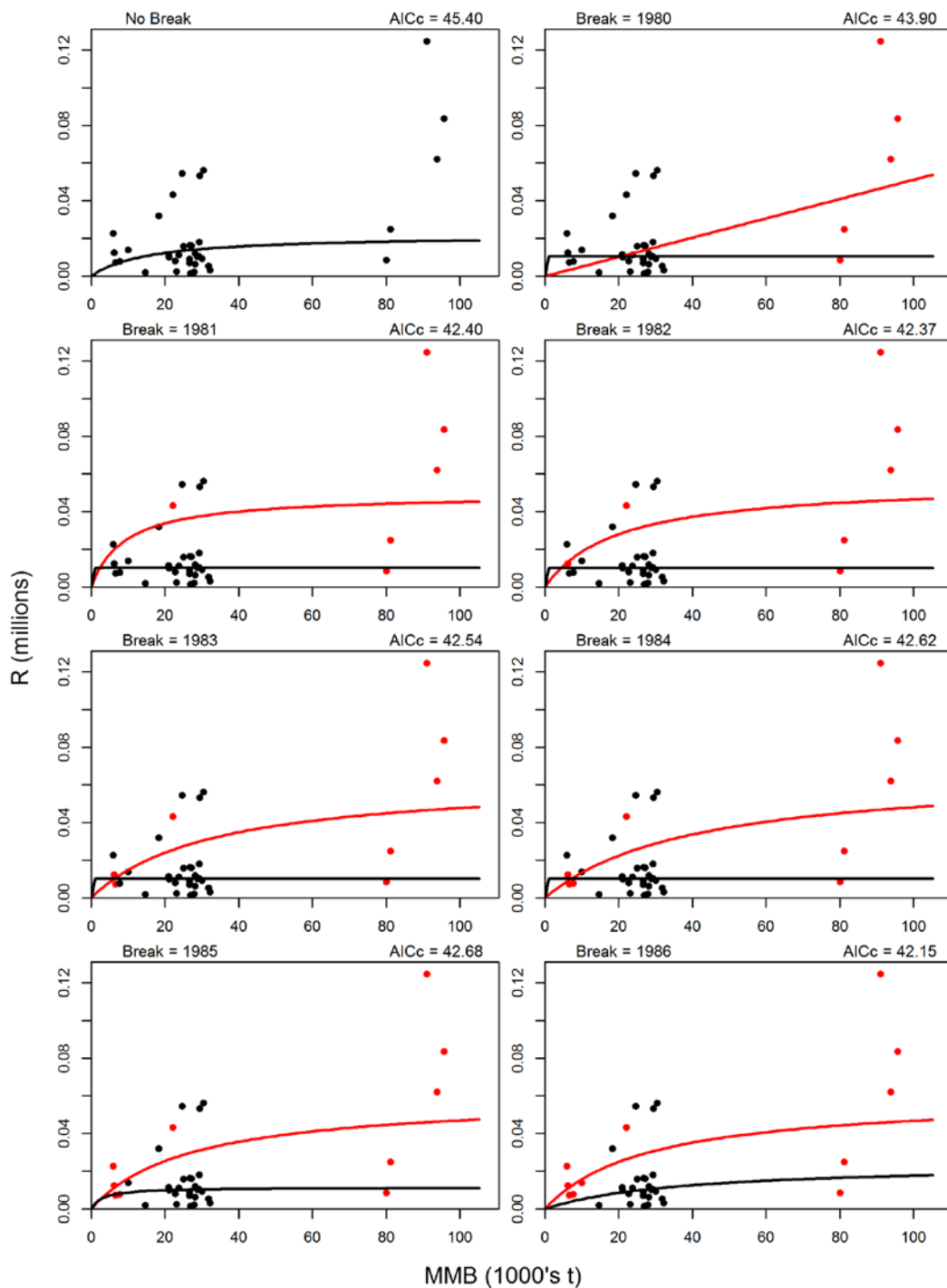


Figure B6. Fits on the arithmetic scale for B-H models with no breakpoint (upper left graph) and with 1-breakpoint for break years 1975-2005. For 1-breakpoint models, the pre-break data (circles) and model fit (line) are shown in red, whereas the post-break data and fit are shown in black.

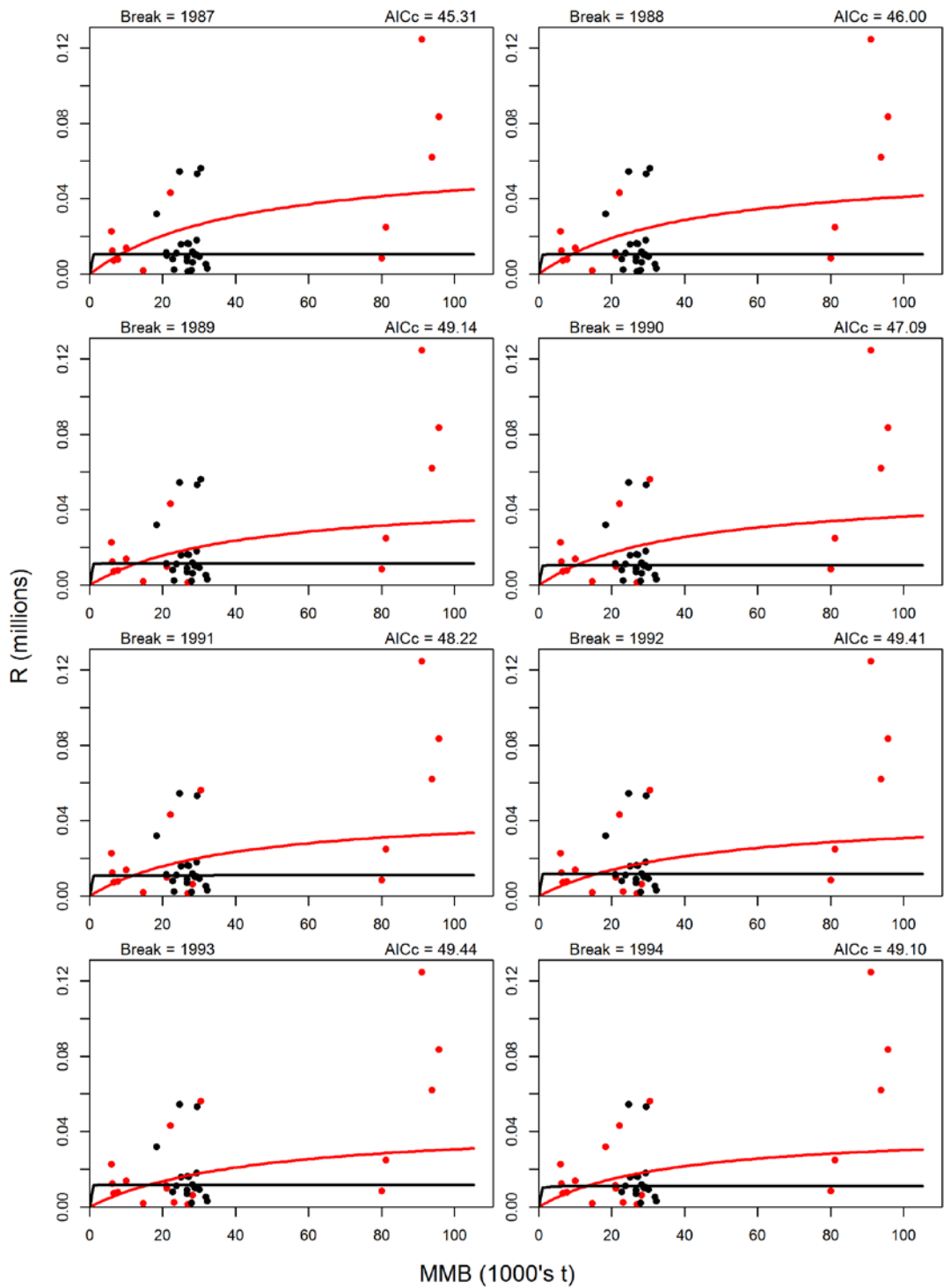


Figure B6. Continue.

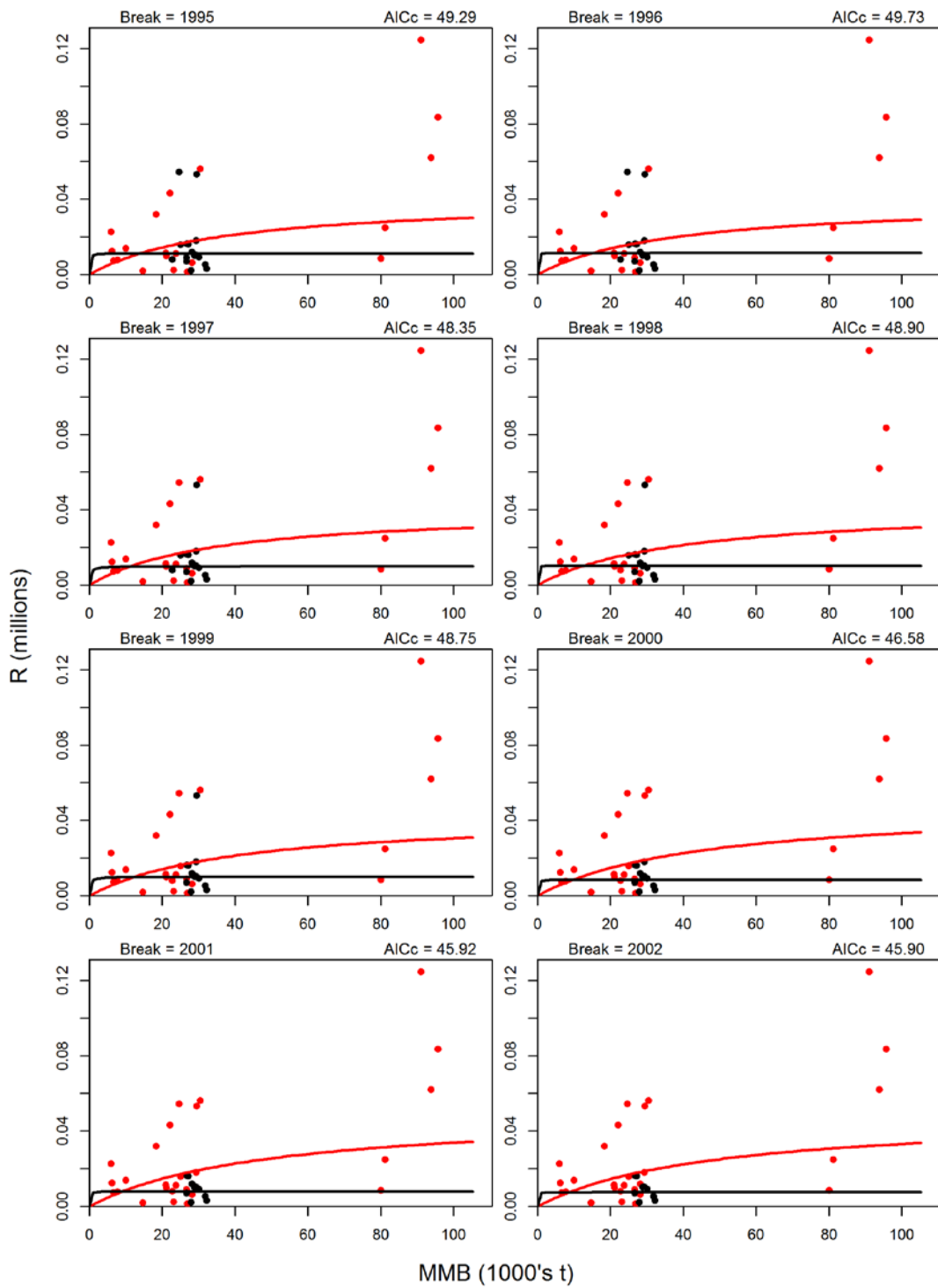
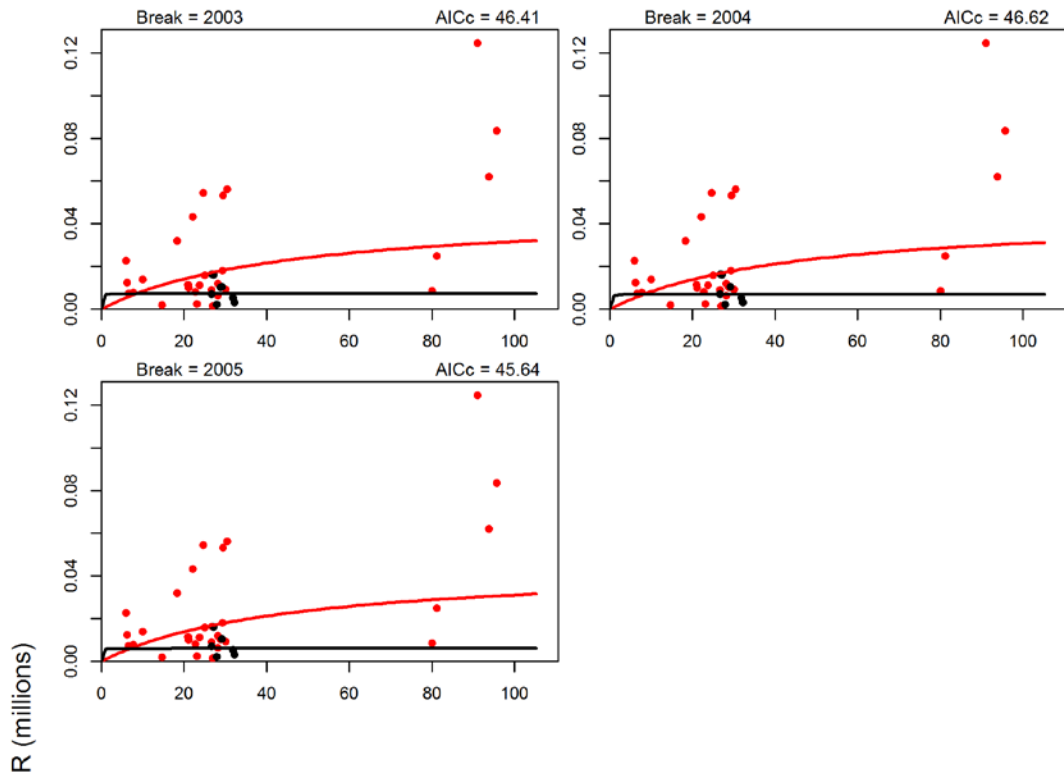


Figure B6. Continue.



MMB (1000's t)

Figure B6. Continue.

Appendix C. Discard Male Biomass Estimation When Legal Discard Information is not available

Before 2018, observed discard male abundance and biomass of Bristol Bay red king crab were estimated directly from the observer data, which contain length frequencies of legal males that were retained or not retained. Starting from 2018, observers will not guess the retained status of legal males, and we need to find an approach to estimate observed discard male abundance and biomass.

We propose to use the approach of AI golden king crab stock assessments, naming it as “Subtraction” approach here. Based on this approach, observed discard male abundance and biomass are equal to total male abundance and biomass estimated from observer data minus the total commercial catch abundance and biomass for a given year. This approach is affected by sampling errors of observer data. The negative log likelihood value for discard male length compositions is replaced by the value for total male length positions. Table C1 summarizes the observer data and estimated observer legal discard rates during 2005-2017, fishing years after rationalization. Observer legal discard rates are the ratios of non-retained legal males (Legal_NR) to total legal males with retained status (sum of Legal_NR and Legal_Ret) from the observer data. Table C2 summarizes estimated abundances of sublegal, legal and total males, commercial catch abundance and biomass, and estimated discard male biomass and discard rates of legal males.

Under Subtraction approach, total selectivity for males is assumed to be a logistic curve:

$$S_l^{dir, mal} = \frac{1}{1+e^{-\beta^{dir, mal} (l-L_{50}^{dir, mal})}} \quad (C1)$$

and the retained proportions of legal males also follow a logistic function:

$$S_l^{dir, ret} = \frac{1}{1+e^{-\beta^{dir, ret} (l-L_{50}^{dir, ret})}} \quad (C2)$$

Two approaches (scenarios 2bn1 and 2bn2) are used to modify the annual retained proportions of legal males. Approach one (scenario 2bn1) has two logistic curves for retained proportions: the first one for a period of 1975-2004, before rationalization, and the second one for a period of 2005-present. There have been some high gradings during the second period, retained proportions should be lower than those during the first period. Approach two (scenario 2bn2) has only one logistic curve for retained proportions for all years; however, to deal with annual high gradings, annual adjusted factor parameter, x_t , is estimated for each year after 2004 and a logit transformation is used to make sure the adjusted factor, u_t , be <1.0:

$$u_t = \frac{e^{x_t}}{1+e^{x_t}} \quad (C3)$$

Annual retained proportions after 2004 are estimated as:

$$S_{l,t}^{dir, ret} = u_t S_l^{dir, ret} \quad (C4)$$

To avoid overfitting the data, a negative likelihood value is computed as:

$$\sum_t (u_t - 1.0)^2 / (2\sigma^2) \quad (C5)$$

where σ is the standard deviation of u_t and is assumed to be 0.1. The model results hardly change with either 0.1 or 0.2.

Table C1. Numbers of sampled crab and potlifts, total number of potlifts and observer legal discard rates by the Bristol Bay red king crab directed pot fishery and cost-recovery fishery based on observer data for crab ≥ 65 mm carapace length since rationalization for the Bristol Bay. Notations: NR = none retained, UR = unknown retained status.

| Year | Fishery | Female | Sublegal | Legal_NR | Legal_UR | Legal_Ret | Samp.pots | Tot.pots | Leg-disc-rate |
|------|----------|--------|----------|----------|----------|-----------|-----------|----------|---------------|
| 2005 | Directed | 26824 | 44103 | 11052 | 0 | 39639 | 1855 | 114944 | 21.80% |
| 2006 | Directed | 3646 | 28504 | 1602 | 0 | 36423 | 1214 | 71735 | 4.21% |
| 2007 | Directed | 12457 | 57295 | 1049 | 0 | 53231 | 1918 | 113214 | 1.93% |
| 2008 | Directed | 8737 | 48903 | 808 | 0 | 40620 | 1849 | 139937 | 1.95% |
| 2009 | Directed | 6050 | 49023 | 1420 | 0 | 42173 | 1950 | 118521 | 3.26% |
| 2010 | Directed | 6862 | 34935 | 471 | 0 | 31253 | 1891 | 131627 | 1.48% |
| 2011 | Directed | 1752 | 19277 | 240 | 0 | 20709 | 696 | 45166 | 1.15% |
| 2012 | Directed | 562 | 6894 | 320 | 0 | 12947 | 437 | 38159 | 2.41% |
| 2013 | Directed | 6070 | 12723 | 1111 | 0 | 16427 | 657 | 45927 | 6.33% |
| 2014 | Directed | 1953 | 10082 | 2964 | 0 | 15494 | 520 | 58702 | 16.06% |
| 2015 | Directed | 5927 | 6975 | 1061 | 0 | 13986 | 413 | 48008 | 7.05% |
| 2016 | Directed | 4315 | 8087 | 1413 | 1298 | 15712 | 413 | 33126 | 8.25% |
| 2017 | Directed | 3834 | 11416 | 1178 | 0 | 14625 | 803 | 48222 | 7.45% |
| 2005 | Cost-rec | | | | | | | | |
| 2006 | Cost-rec | 386 | 2787 | 82 | 0 | 4738 | 110 | 1067 | 1.70% |
| 2007 | Cost-rec | 250 | 1638 | 32 | 0 | 2436 | 61 | 734 | 1.30% |
| 2008 | Cost-rec | 0 | 0 | 0 | 0 | 0 | 0 | 0 | |
| 2009 | Cost-rec | 17 | 2337 | 6 | 0 | 3233 | 114 | 740 | 0.19% |
| 2010 | Cost-rec | 29 | 1283 | 4 | 0 | 1364 | 46 | 556 | 0.29% |
| 2011 | Cost-rec | 179 | 1141 | 5 | 0 | 1680 | 51 | 618 | 0.30% |
| 2012 | Cost-rec | 2 | 133 | 18 | 0 | 1203 | 35 | 683 | 1.47% |
| 2013 | Cost-rec | 7 | 466 | 68 | 0 | 1720 | 35 | 662 | 3.80% |
| 2014 | Cost-rec | 714 | 1006 | 11 | 0 | 1666 | 38 | 665 | 0.66% |
| 2015 | Cost-rec | 1533 | 660 | 0 | 0 | 1932 | 42 | 755 | 0.00% |
| 2016 | Cost-rec | 1519 | 1038 | 11 | 0 | 2473 | 40 | 482 | 0.44% |
| 2017 | Cost-rec | 214 | 738 | 38 | 0 | 2018 | 51 | 927 | 1.85% |

Table C2. Estimated abundances of sublegal, legal and total males, commercial catch abundance and biomass, and estimated discard male biomass and discard rates of legal males for the Bristol Bay red king crab fishery, for crab ≥ 65 mm carapace length. Abundance is in millions of crab and biomass is 1000 tons.

| Year | Abundance | | | Biomass | Abundance | Biomass | Biomass | Disc-rate |
|------|-----------|--------|------------|------------|-----------|-----------|-----------|-----------|
| | Sublegal | Legal | Total male | Total male | Com.catch | Com.catch | Disc.male | Legal |
| 2005 | 2.7652 | 3.1716 | 5.9368 | 13.7533 | 2.7627 | 8.3997 | 5.3536 | 12.89% |
| 2006 | 1.7111 | 2.2936 | 4.0047 | 9.1704 | 2.4769 | 7.1432 | 2.0272 | -7.99% |
| 2007 | 3.4005 | 3.2337 | 6.6342 | 13.9566 | 3.1543 | 9.3039 | 4.6526 | 2.46% |
| 2008 | 3.7009 | 3.1354 | 6.8363 | 15.0687 | 3.0640 | 9.2161 | 5.8526 | 2.28% |
| 2009 | 2.9944 | 2.6706 | 5.6650 | 12.3003 | 2.5529 | 7.2725 | 5.0278 | 4.41% |
| 2010 | 2.4464 | 2.2248 | 4.6711 | 10.0874 | 2.4099 | 6.7615 | 3.3258 | -8.32% |
| 2011 | 1.2632 | 1.3799 | 2.6431 | 5.7326 | 1.2978 | 3.6071 | 2.1255 | 5.95% |
| 2012 | 0.6041 | 1.1823 | 1.7864 | 4.5681 | 1.1756 | 3.6217 | 0.9464 | 0.57% |
| 2013 | 0.8980 | 1.2598 | 2.1578 | 5.2607 | 1.2720 | 3.9910 | 1.2697 | -0.97% |
| 2014 | 1.1553 | 2.1130 | 3.2683 | 8.3127 | 1.5012 | 4.5386 | 3.7741 | 28.96% |
| 2015 | 0.8225 | 1.7838 | 2.6064 | 6.7064 | 1.5270 | 4.6137 | 2.0928 | 14.40% |
| 2016 | 0.6612 | 1.5076 | 2.1688 | 5.5572 | 1.2812 | 3.9239 | 1.6333 | 15.02% |
| 2017 | 0.6986 | 0.9864 | 1.6850 | 3.9865 | 0.9972 | 3.0937 | 0.8928 | -1.10% |

Appendix D. Simple B0 Analysis

Ideally, a stock-recruitment relationship and impacts of environmental factors on recruitment are developed before doing B0 analysis. For Bristol Bay red king crab, there is hardly any relationship between estimated recruits and MMB (Figure 14a). The impacts of environmental factors on recruitment have not been quantified. We simply computed B0 values over time using the same recruitment time series estimated from the assessment model through setting all fishing mortality to be zero. Figure D1 shows the time series of estimated B0, MMB with fishing, and ratios of MMB to B0 for scenario 2bn2. As expected, estimated B0 values change greatly over time.

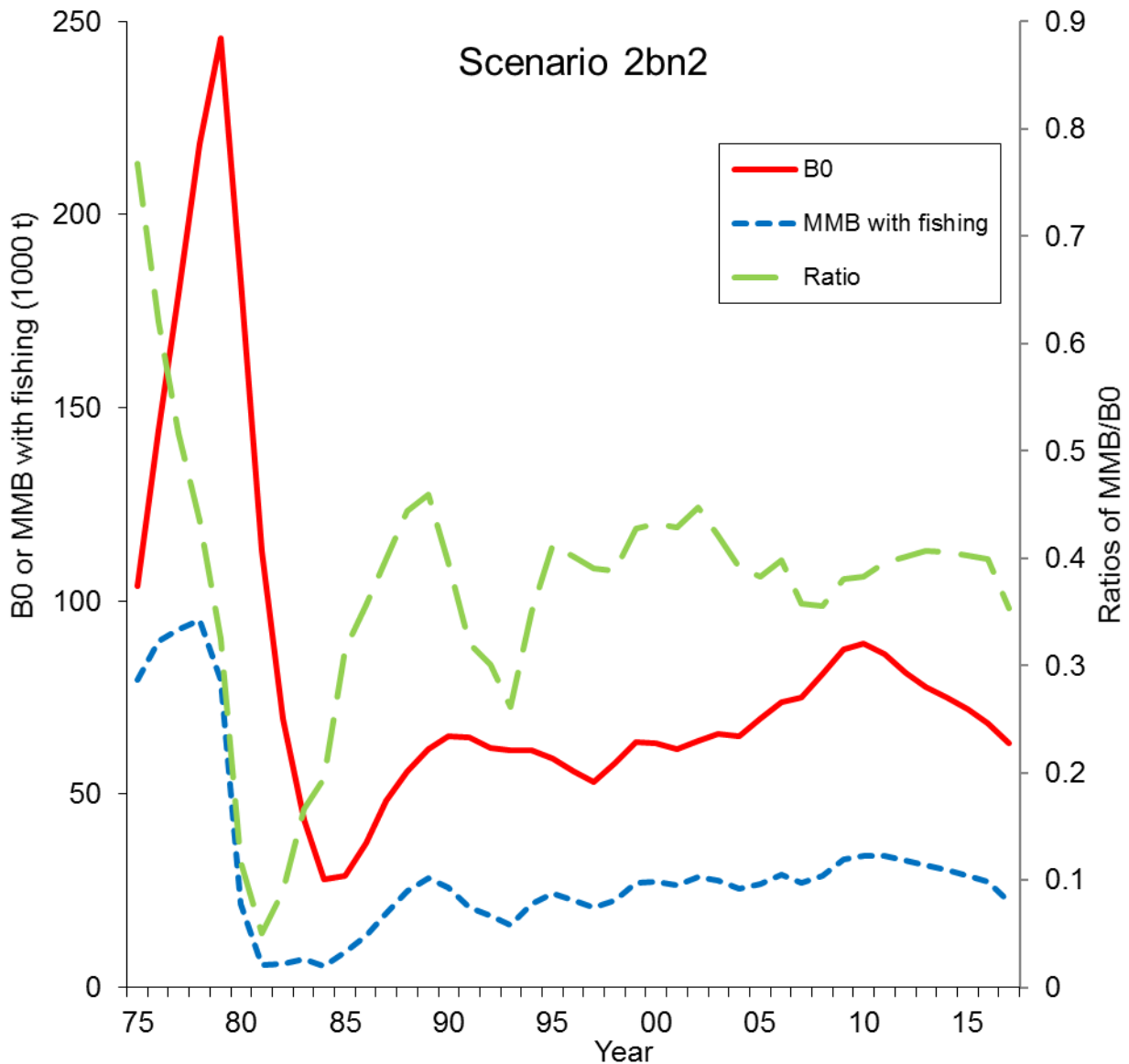


Figure D1. Estimated B0, MMB with fishing, and ratios of MMB/B0 from 1975 to 2017 for scenario 2bn2 for Bristol Bay red king crab.