# Saint Matthew Island blue king crab proposed models for May 2024 

Caitlin Stern ${ }^{1}$ and Katie Palof ${ }^{2}$<br>${ }^{1}$ Alaska Department of Fish and Game, caitlin.stern@alaska.gov<br>${ }^{2}$ Alaska Department of Fish and Game, katie.palof@alaska.gov

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## Executive Summary

1. Stock: Blue king crab, Paralithodes platypus, Saint Matthew Island (SMBKC), Alaska.
2. Catches: Peak historical harvest was $4,288 \mathrm{t}$ ( 9.454 million pounds) in $1983 / 84^{1}$. The fishery was closed for 10 years after the stock was declared overfished in 1999. Fishing resumed in 2009/10 with a fishery-reported retained catch of 209 t ( 0.461 million pounds), less than half the 529.3 t ( 1.167 million pound) TAC. Following three more years of modest harvests supported by a fishery catch per unit effort (CPUE) of around 10 crab per pot lift, the fishery was again closed in 2013/14 due to declining trawl-survey estimates of abundance and concerns about the health of the stock. The directed fishery resumed again in 2014/15 with a TAC of 300 t ( 0.655 million pounds), but the fishery performance was relatively poor with a retained catch of 140 t ( 0.309 million pounds). The retained catch in 2015/16 was even lower at 48 t ( 0.105 million pounds) and the fishery has remained closed since 2016/17.
3. Stock biomass: The 2023 NMFS trawl survey biomass of $\geq 90 \mathrm{~mm}$ carapace length (CL) male crab is $2,114 \mathrm{t}$ ( $44 \% \mathrm{CV} ; 4.66$ million lb), the 12th lowest in the 46 years of the survey and the 9th lowest since 2000. The 2023 biomass is $40 \%$ of the 1978-2023 NMFS trawl survey mean biomass ( $5,272 \mathrm{t}$ ), and an $8 \%$ decrease from the 2022 biomass. The mean biomass NMFS survey biomass over the most recent three years is $41 \%$ of the time series mean value, indicating a decline in biomass compared to historical survey estimates; survey biomass in both 2010 and 2011 was over five times the three-year mean. The ADF\&G pot survey last occurred in 2022, when the relative biomass index was the highest since 2013 , and $70 \%$ of the mean from the 12 surveys conducted since 1995 . The next ADF\&G pot survey is scheduled for 2025 and will be included in the 2026 assessment. The assessment model estimates do not fit either of the survey time series particularly well. For the NMFS trawl survey, model estimates suggest that the stock biomass is $45 \%$ of the mean model-predicted biomass, with a poor fit to the 2019 biomass observation. For the ADF\&G pot survey, model estimates suggest that the stock biomass is $67 \%$ of the mean model-predicted biomass, with a poor fit to the 2016, 2017, and 2018 biomass observations.
4. Recruitment: Recruitment is based on the estimated number of male crab in the $90-104 \mathrm{~mm}$ CL size class in each year. The 2023 trawl-survey area-swept estimate of 511,507 male SMBKC in this size class is ranked 27 th, near the middle of the 46 years of the survey, and down from 25th in 2022. Mean recruitment over the most recent six years (2017-2023) is $37 \%$ of the long-term mean. In the ADF\&G pot survey, the abundance of male crab in the $90-104 \mathrm{~mm}$ CL size class in 2022 ranked 7th in the time series ( $56 \%$ of the mean for the 12 available years of pot survey data) and was the highest abundance observed for this size class since 2013.
5. Management performance: In this assessment, estimated total male catch is the sum of fisheryreported retained catch, estimated male discard mortality in the directed fishery, and estimated male

[^0]bycatch mortality in the groundfish fisheries. Based on the new base model for SMBKC (model 24.0), the estimate for mature male biomass was below the minimum stock-size threshold (MSST) in 2022/23 and is in an "overfished" condition, despite a directed fishery closure since the 2016/17 season (and hence overfishing has not occurred) (Tables 1, 3, and 4). Computations which indicate the relative impact of fishing (i.e., the "dynamic $B_{0}$ ") suggest that the current spawning stock biomass has been reduced to $72 \%$ of what it would have been in the absence of fishing, assuming the same level of recruitment as estimated.

Table 1: Status and catch specifications (1000 t) for the new base model (24.0).

| Year | MSST | Biomass <br> $\left(M M B_{\text {mating }}\right)$ | TAC | Retained <br> catch | Total <br> male catch | OFL | ABC |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $2019 / 20$ | 1.67 | 1.06 | 0.00 | 0.00 | 0.001 | 0.04 | 0.03 |
| $2020 / 21$ | 1.65 | 1.14 | 0.00 | 0.00 | 0.001 | 0.05 | 0.04 |
| $2021 / 22$ | 1.63 | 1.18 | 0.00 | 0.00 | 0.001 | 0.05 | 0.04 |
| $2022 / 23$ | 1.46 | 1.34 | 0.00 | 0.00 | 0.001 | 0.066 | 0.05 |
| $2023 / 24$ |  | 1.39 |  |  |  | 0.085 | 0.064 |

Table 2: Status and catch specifications (million pounds) for the new base model (24.0).

| Year | MSST | Biomass <br> $\left(M M B_{\text {mating }}\right)$ | TAC | Retained <br> catch | Total <br> male catch | OFL | ABC |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $2018 / 19$ | 3.84 | 2.54 | 0.000 | 0.000 | 0.002 | 0.08 | 0.07 |
| $2019 / 20$ | 3.68 | 2.34 | 0.000 | 0.000 | 0.002 | 0.096 | 0.08 |
| $2020 / 21$ | 3.64 | 2.52 | 0.000 | 0.000 | 0.002 | 0.112 | 0.08 |
| $2021 / 22$ | 3.59 | 2.59 | 0.000 | 0.000 | 0.002 | 0.112 | 0.08 |
| $2022 / 23$ | 3.22 | 2.96 | 0.000 | 0.000 | 0.002 | 0.146 | 0.11 |
| $2023 / 24$ |  | 3.07 |  |  |  | 0.188 | 0.141 |

6. Basis for the OFL: Estimated mature-male biomass (MMB) on 15 February is used as the measure of biomass for this Tier 4 stock, with males measuring $\geq 105 \mathrm{~mm}$ CL considered mature. The $B_{M S Y}$ proxy is obtained by averaging estimated MMB over a specific reference period, and current CPT/SSC guidance recommends using the full assessment time frame (1978-2022) as the default reference period.

Table 3: Basis for the OFL (1000 t) from the new base model (24.0).

| Year | Tier | $B_{M S Y}$ | Biomass <br> $\left(M M B_{\text {mating }}\right)$ | $B / B_{M S Y}$ | $F_{O F L}$ | $\gamma$ | Basis for $B_{M S Y}$ | Natural <br> mortality |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $2019 / 20$ | 4 b | 3.48 | 1.06 | 0.31 | 0.042 | 1 | $1978-2018$ | 0.18 |
| $2020 / 21$ | 4 b | 3.34 | 1.14 | 0.32 | 0.047 | 1 | $1978-2019$ | 0.18 |
| $2021 / 22$ | 4 b | 3.30 | 1.34 | 0.34 | 0.048 | 1 | $1978-2020$ | 0.18 |
| $2022 / 23$ | 4 b | 2.92 | 1.39 | 0.48 | 0.076 | 1 | $1978-2022$ | 0.18 |

## A. Summary of Major Changes

## Changes in Management of the Fishery

There are no new changes in management of the fishery.

## Changes to the Input Data

Data used in this assessment have been updated to include the most recently available fishery data and survey data. This assessment includes new biomass data from the 2023 NMFS trawl survey and new relative abundance data from the most recent ADF\&G pot survey, which was conducted in 2022. The assessment was updated with new size composition data from the 2023 NMFS trawl survey and the 2022 ADF\&G pot survey, as well as 2022 groundfish trawl and fixed gear bycatch estimates based on NMFS Alaska Regional Office (AKRO) data. The September 2024 version of this assessment will include the 2024 NMFS trawl survey biomass and size composition data as well as the 2023 groundfish trawl and fixed gear bycatch estimates. The ADF\&G pot survey is on a semi-triennial cycle, with the next survey planned for fall 2025 . The directed fishery has been closed since 2016/17, so no recent fishery data are available.

## Changes in Assessment Methodology

This assessment has used the Generalized size-structured Model for Assessing Crustacean Stocks (GMACS) framework since 2016. The model, which is configured to track three stages of length categories, was first presented in May 2011 by W. Gaeuman, ADF\&G, and accepted by the Crab Plan Team (CPT) in May 2012. A difference from the original approach and that used here is that natural and fishing mortalities are continuous within 5 discrete time blocks within a year, using the appropriate catch equation rather than assuming an applied pulse removal. The time blocks within a year in GMACS are controlled by changing the proportion of natural mortality that is applied to each block. Diagnostic output includes estimates of the "dynamic $B_{0}$ ", which is the ratio of the estimated spawning biomass relative to the spawning biomass that would have occurred had there been no historical fishing mortality. Details of this implementation and other model details are provided in Appendix A.

## Changes in Assessment Results

Both surveys indicate a low population over the past few years. The ADF\&G pot survey showed a declining trend from 2010 to 2018, while the NMFS trawl survey showed a decline from 2014 to 2018. Both the NMFS trawl and ADF\&G pot survey results from 2022 suggested some potential for growth in this stock, but the 2023 NMFS trawl survey results suggest a decrease in biomass.

The reference model (model 16.0 2022) is that which was selected for use in 2022, the year of the last full assessment. The base model presented here is the reference model using the most recent version of GMACS, updated groundfish bycatch data for the $2022 / 23$ crab season, and updated survey data - biomass and size composition - from the 2023 NMFS trawl survey and the 2022 ADF\&G pot survey (model 16.0).

Alternative models presented for consideration in May 2024 are models 16.0a, 16.0b, 24.0, 24.0a, 24.0b, and 24.0 c . Models 16.0 a and 16.0 b represent individual additions of updates that we consider necessary. Model 16.0a is model 16.0 with a fully updated historical time series for the ADF\&G pot survey relative abundance index and size compositions, incorporating error corrections to the historical data sets and a more transparent, replicable data processing procedure. The largest error in the data set that had previously been used in the model was from the 2016 pot survey, when it appears that the total number of males was used rather than the number of males $\geq 90 \mathrm{~mm}$. Compared to model 16.0, model 16.0 a estimates lower values for MMB, $B_{M S Y}$, OFL, and ABC (Table 4).

Model 16.0b is model 16.0 with SSB estimated in season 5 of the model rather than season 4. This change was deemed necessary after the clarification that GMACS estimates SSB at the beginning of a season. As SSB is intended to be estimated on 15 February, and 15 February is the dividing date between seasons 4 and 5 in this model, SSB estimation should take place in season 5. Compared to model 16.0, model 16.0b estimates lower values for MMB and $B_{M S Y}$, but higher values for the OFL and ABC (Table 4).

Model 24.0 includes the changes made in models 16.0 a and 16.0 b : the historical time series for the ADF\&G pot survey relative abundance index and size compositions are fully updated, and SSB is estimated in season 5 of the model rather than season 4 . We propose model 24.0 as the new base model for this stock. Compared to model 16.0, model 24.0 estimates lower values for MMB, $B_{M S Y}$, OFL, and ABC (Table 4).

Model 24.0a is model 24.0 with $M$ estimated using a tight prior (mean $=0.18$, $\mathrm{CV}=0.04$ ), as in the Bristol Bay Red King Crab stock assessment. Model 24.0 b is model 24.0 with $M$ estimated using a less restrictive prior (mean $=0.18, \mathrm{CV}=0.1$ ), allowing greater flexibility in estimation of $M$. Model 24.0 c is model 24.0 with $M$ fixed at the value estimated in model 24.0a. Compared to the estimates from model 24.0, the value estimated for MMB by model 24.0a is lower and model 24.0b's estimated value is lower still, while the values estimated for $B_{M S Y}$, OFL, and ABC are higher in model 24.0a and higher still in 24.0b (Table 4). We also performed a likelihood profile for $M$ using model 24.0 and varying only the fixed value of $M$.

## B. SSC and CPT comments and author responses

## SSC comments on BSAI crab assessments in general (October 2023 and October 2022)

The SSC recommends that risk tables be developed for crab assessments.
The CPT has not chosen SMBKC for risk table development in this assessment cycle, but a risk table for this stock may be developed in a future assessment.

The SSC requests that the CPT develop a process for ensuring that authors have provided a response to all previous (including at least the last assessment) SSC recommendations.

The SSC recommendations from the 2022 SMBKC assessment are included below.
The SSC requests that all crab authors include uncertainty intervals when showing time series of biomass/abundance estimated by the stock assessment models so that alternative models and retrospective patterns can be evaluated in the context of the modeled uncertainty.
We have included uncertainty intervals for these time series (e.g., Figures 1 and 2).
For the inclusion of trawl survey data, the SSC suggests crab assessment authors and the CPT be more explicit about best practices for which standard years are included for bottom trawl survey data.

The CPT decided to use the entire time series of trawl survey data for crab assessment, with time blocks for parameters as necessary, and this assessment follows that practice.

The SSC suggests that the CPT and crab authors continue to evaluate whether VAST or similar approaches, when specified carefully for individual crab stocks, might provide more robust survey time series.

We are working with Jon Richar (NOAA) to develop a single index for SMBKC consisting of output from a Vector Autoregressive Spatio-Temporal (VAST) model (Thorson and Barnett 2017) that uses both the trawl and pot survey data sets, and hope to have model runs using that index ready to present in the near future.

The SSC recommends that all crab authors plot length compositions over years with the most recent year at the bottom of the plot.
We have followed this recommendation (e.g., Figures 3 and 4).

## SSC and CPT comments on SMBKC assessment (October 2022)

Follow up on previous SSC recommendations as time allows, including exploring:

1. Data weighting (Francis and other approaches) and evaluation of models with and without the 1998 natural mortality spike.

We continued with the iterative re-weighting for composition data (Table 5). We did not address models without the natural mortality spike as these have been considered previously.

## 2. Causes of observed retrospective patterns

Discussion of potential causes of the observed retrospective patterns is provided in section E, below, under Results: Retrospective and historical analyses.
3. Potential explanations for the discrepancy in the time trends of the two types of survey data

Exploration of the spatial extent and density differences between the two surveys (NMFS and ADF\&G) was done on all male crab ( $\geq 90 \mathrm{~mm} \mathrm{CL}$ ) and all years of overlap between the two surveys included in the model for May 2022 (May 2022 documentation Appendix C). The authors plan to use this and further analyses to better characterize catchability/availability for the pot survey.
4. Estimates of survey biomass based on VAST compared to design-based estimates, and estimates that combine the two surveys

We are working with Jon Richar to generate VAST model estimates for SMBKC that combine the two surveys, and hope to soon be able to present model runs using a single, VAST model-derived index in place of the two survey indices.

## 5. Random walk on catchability

The initial model of time blocks for Q did not show much potential for this in May 2020, therefore time blocks were not a focus for May 2022 or May 2024.
6. Assumed and estimated life history parameters (e.g., natural mortality, growth, and maturity) to ensure the best available science is being used to assess this stock.

Specific research on St. Matthew Island blue king crab life history parameters is not available and therefore these are borrowed from other stocks/species. Sensitivities of the model to increased natural mortality ( $M$ ) were explored in May 2022. For May 2024, we present model 24.0a, which estimates $M$ using a tight prior (mean $=0.18, \mathrm{CV}=0.04$ ) as in the Bristol Bay Red King Crab stock assessment, model 24.0b, which estimates $M$ using a less restrictive prior (mean $=0.18, \mathrm{CV}=0.1$ ) on $M$, and model 24.0 c , which has $M$ fixed to the value estimated in model 24.0a. We also performed a likelihood profile for $M$ using model 24.0 and varying only the fixed value of $M$. Sensitivities to the model assumptions on growth and maturity will be explored at a later date.

## C. Introduction

## Scientific Name

The blue king crab is a lithodid crab, Paralithodes platypus (Brant 1850).

## Distribution

Blue king crab are sporadically distributed throughout the North Pacific Ocean from Hokkaido, Japan, to southeastern Alaska (Figure 5). In the eastern Bering Sea, small populations are distributed around St. Matthew Island, the Pribilof Islands, St. Lawrence Island, and Nunivak Island. Isolated populations also exist in some other cold water areas of the Gulf of Alaska (NPFMC 1998). The St. Matthew Island Section for blue king crab is within Area Q (Figure 6), which is the Northern District of the Bering Sea king crab registration area and includes the waters north of Cape Newenham ( $58^{\circ} 39^{\prime} \mathrm{N}$. lat.) and south of Cape Romanzof ( $61^{\circ} 49$ ' N. lat.).

## Stock Structure

The Alaska Department of Fish and Game (ADF\&G) Gene Conservation Laboratory has detected regional population differences between blue king crab collected from St. Matthew Island and the Pribilof Islands ${ }^{2}$. The NMFS tag-return data from studies on blue king crab in the Pribilof Islands and St. Matthew Island support the idea that legal-sized males do not migrate between the two areas (Otto and Cummiskey 1990). St. Matthew Island blue king crab tend to be smaller than their Pribilof conspecifics, and the two stocks are managed separately. An analysis of genetic markers from blue king crab populations in Southeast Alaska, the Pribilof Islands, St. Matthew Island, Little Diomede, Chaunskaya Bay, the western Bering Sea, and two locations from Shelikov Gulf in the Sea of Okhotsk found genetic differences among all locations (Stoutamore 2014).

## Life History

Like the red king crab, Paralithodes camtshaticus, the blue king crab is considered a shallow water species by comparison with other lithodids such as golden king crab, Lithodes aequispinus, and the scarlet king crab, Lithodes couesi (Donaldson and Byersdorfer 2005). Adult male blue king crab are found at an average depth of 70 m (NPFMC 1998). The reproductive cycle appears to be annual for the first two reproductive cycles and biennial thereafter (Jensen and Armstrong 1989), and mature crab seasonally migrate inshore where they molt and mate. Unlike red king crab, juvenile blue king crab do not form pods, but instead rely on cryptic coloration for protection from predators and require suitable habitat such as cobble and shell hash. Somerton and MacIntosh (1983) estimated SMBKC male size at sexual maturity to be 77 mm carapace length (CL). Paul et al. (1991) found that spermatophores were present in the vas deferens of $50 \%$ of the St. Matthew Island blue king crab males examined with sizes of $40-49 \mathrm{~mm}$ CL and in $100 \%$ of the males at least 100 mm CL. Spermataphore diameter also increased with increasing CL with an asymptote at ~ 100 mm CL. It was noted, however, that although spermataphore presence indicates physiological sexual maturity, it may not be an indicator of functional sexual maturity. For purposes of management of the St. Matthew Island blue king crab fishery, the State of Alaska uses 105 mm CL to define the lower size bound of functionally mature males (Pengilly and Schmidt 1995). Otto and Cummiskey (1990) reported an average growth increment of 14.1 mm CL for adult SMBKC males.

## Management History

The SMBKC fishery developed subsequent to baseline ecological studies associated with oil exploration (Otto 1990). Ten U.S. vessels harvested 545 t ( 1.202 million pounds) in 1977, and harvests peaked in 1983 when 164 vessels landed $4,288 \mathrm{t}$ ( 9.454 million pounds) (Fitch et al. 2012; Table 6).

The fishing seasons were generally short, often lasting only a few days. The stock was declared overfished and the fishery was closed in 1999, when the stock biomass estimate was below the minimum stock-size threshold (MSST) of $4,990 \mathrm{t}$ ( 11.0 million pounds) as defined by the Fishery Management Plan (FMP) for

[^1]the Bering Sea/Aleutian Islands King and Tanner crabs ("Crab FMP"; NPFMC 1999). Zheng and Kruse (2002) hypothesized a high level of SMBKC natural mortality from 1998 to 1999 as an explanation for the low catch per unit effort (CPUE) in the 1998/99 commercial fishery and the low numbers across all male crab size groups caught in the annual NMFS eastern Bering Sea trawl survey from 1999 to 2005 (see survey data in next section). In November 2000, Amendment 15 to the Crab FMP was approved to implement a rebuilding plan for the SMBKC stock (NPFMC 2000). The rebuilding plan included a State of Alaska regulatory harvest strategy ( $5 A A C 34.917$ ), area closures, and gear modifications. In addition, commercial crab fisheries near St. Matthew Island were limited to fall and early winter to reduce the potential for bycatch mortality of vulnerable molting and mating crab.

NMFS declared the stock rebuilt on 21 September 2009, and the fishery was reopened after a 10-year closure on 15 October 2009 with a TAC of 529 t ( 1.167 million pounds), closing again by regulation on 1 February 2010. Seven participating vessels landed a catch of 209 t ( 0.461 million pounds) with a reported effort of 10,697 pot lifts and an estimated CPUE of 9.9 retained individual crab per pot lift. The fishery remained open the next three years with modest harvests and similar CPUE, but large declines in the NMFS trawlsurvey estimate of stock abundance raised concerns about the health of the stock. This prompted ADF\&G to close the fishery again for the $2013 / 14$ season. The fishery was reopened for the $2014 / 15$ season with a low TAC of 297 t ( 0.655 million pounds) and in 2015/16 the TAC was further reduced to 186 t ( 0.411 million pounds) then completely closed for the 2016/17 season. The stock was declared overfished in 2018 (NOAA Fisheries 2019).

In October 2020, Amendment 50 to the Crab FMP was approved to implement a second rebuilding plan for the SMBKC stock (NOAA 2020). The primary factors limiting stock rebuilding were identified as warm bottom temperatures, low pre-recruit biomass, and northward shifts in predator populations, rather than fishing mortality. The aim of the rebuilding plan was thus to maintain a low rate of fishing mortality while awaiting ecosystem conditions conducive to stock rebuilding. The lack of stock rebuilding in Pribilof Islands blue king crab despite fisheries closures and abundant juvenile habitat has similarly been attributed to limited larval supply and warm bottom temperatures (Weems et al. 2020).

Although historical observer data are limited due to low sampling effort, bycatch of female and sublegal male crab from the directed blue king crab fishery off St. Matthew Island was relatively high historically, with estimated total bycatch in terms of number of crab captured sometimes more than twice as high as the catch of legal crab (Moore et al. 2000; ADF\&G Crab Observer Database). Pot-lift sampling by ADF\&G crab observers (Gaeuman 2013; ADF\&G Crab Observer Database) indicates similar bycatch rates of discarded male crab since the reopening of the fishery (Table 7), with total male discard mortality in the 2012/13 directed fishery estimated at about $12 \%$ ( 88 t or 0.193 million pounds) of the reported retained catch weight, assuming $20 \%$ handling mortality.

These data suggest a reduction in the bycatch of females, which may be attributable to the later timing of the contemporary fishery and the more offshore distribution of fishery effort since reopening in 2009/10 ${ }^{3}$. Some bycatch of discarded blue king crab has also been observed historically in the eastern Bering Sea snow crab fishery, but in recent years it has generally been negligible. The St. Matthew Island golden king crab fishery, the third commercial crab fishery to have taken place in the area, typically occurred in areas with depths exceeding blue king crab distribution. The NMFS observer data suggest that variable, but mostly limited, SMBKC bycatch has also occurred in the eastern Bering Sea groundfish fisheries (Table 8).

## D. Data

## Summary of New Information

Data used in this assessment were updated to include the most recently available fishery and survey estimates. Since this stock is on a biennial assessment cycle, the new data for these models include 2023 NMFS trawl survey biomass and size composition data as well as 2022 ADF\&G pot survey abundance and size composition

[^2]data. The assessment uses updated 1993-2022 groundfish and fixed gear bycatch estimates based on NMFS Alaska Regional Office data, accessed through the Alaska Fisheries Information Network. The directed fishery has been closed since the $2016 / 17$ season, and therefore no directed fishery catch data are available. The data used in each of the new models are shown in Figure 7.

## Major Data Sources

Major data sources used in this assessment include annual directed-fishery retained-catch statistics from fish tickets (1978/79-1998/99, 2009/10-2012/13, and 2014/15-2015/16; Table 6); results from the annual NMFS eastern Bering Sea trawl survey (1978-2023; Table 9); results from the ADF\&G SMBKC pot survey (every third year during 1995-2013, every year during 2015-2018; every third year during 2022-present; Table 10); mean somatic mass given length category by year ( 0.7 kg for Stage- $1,1.2 \mathrm{~kg}$ for Stage-2, and 1.9 kg for Stage-3; constant across all years and models); size frequency information from ADF\&G crabobserver pot-lift sampling (1990/91-1998/99, 2009/10-2012/13, and 2014/15-2016/17; Table 7); and the NMFS groundfish-observer bycatch biomass estimates (1991/92-2022/23; Table 8).

Figure 8 maps the stations from which SMBKC trawl-survey and pot-survey data were obtained. Further information concerning the NMFS trawl survey as it relates to commercial crab species is available in Daly et al. (2014); see Gish et al. (2012) for a description of ADF\&G SMBKC pot-survey methods. It should be noted that the two surveys cover different geographic regions and that each has in some years encountered proportionally large numbers of male blue king crab in areas not covered by the other survey (Figure 9). Crabobserver sampling protocols are detailed in the crab-observer training manual (ADF\&G 2013). Groundfish SMBKC bycatch data come from the NMFS Regional office and have been compiled to coincide with the SMBKC management area.

## Other Data Sources

The growth transition matrix used is based on Otto and Cummiskey (1990), as in previous assessments for this stock. Other relevant data sources, including assumed population and fishery parameters, are presented in Appendix A, which also provides a detailed description of the model configuration used for this assessment.

## E. Analytic Approach

## History of Modeling Approaches for this Stock

A four-stage catch-survey-analysis (CSA) assessment model was used before 2011 to estimate abundance and biomass and recommend fishery quotas for the SMBKC stock. The four-stage CSA is similar to a full length-based analysis, the major difference being coarser length groups, which are more suited to a small stock with consistently low survey catches. In this approach, the abundance of male crab with a CL $\geq 90$ mm is modeled in terms of four crab stages: stage 1: 90-104 mm CL; stage 2: 105-119 mm CL; stage 3: newshell 120-133 mm CL; and stage 4: oldshell $\geq 120 \mathrm{~mm}$ CL and newshell $\geq 134 \mathrm{~mm}$ CL. Motivation for these stage definitions came from the fact that, for management of the SMBKC stock, male crab measuring $\geq 105 \mathrm{~mm}$ CL are considered mature, whereas 120 mm CL is considered a proxy for the legal size of 5.5 in carapace width, including spines. Additional motivation for these stage definitions came from an estimated average growth increment of about 14 mm per molt for SMBKC (Otto and Cummiskey 1990).
Concerns about the pre-2011 assessment model led to the CPT and SSC recommendations that included development of an alternative model with provisional assessment based on survey biomass or some other index of abundance. An alternative 3-stage model was proposed to the CPT in May 2011, but a survey-based approach was requested for the fall 2011 assessment. In May 2012, the CPT approved a slightly revised and better-documented version of the alternative model for assessment. The model developed and used from 2012 to 2015 was a variant of the previous four-stage SMBKC CSA model and similar in complexity to
that described by Collie et al. (2005). Like the earlier model, it considered only male crab $\geq 90 \mathrm{~mm}$ in CL, but combined stages 3 and 4 of the earlier model, resulting in three stages (male size classes) defined by CL measurements of (1) 90-104 mm, (2) 105-119 mm, and (3) $120 \mathrm{~mm}+$ (i.e., 120 mm and above). This consolidation was driven by concern about the accuracy and consistency of shell-condition information, which had been used in distinguishing stages 3 and 4 of the earlier model.

In 2016, the accepted SMBKC assessment model made use of the modeling framework GMACS, with a three-stage model structure (Webber et al. 2016). In that assessment, an effort was made to match the 2015 SMBKC stock assessment model while bridging to a new framework that provided greater flexibility and opportunity to evaluate model assumptions more fully. The three subsequent assessments, in 2018, 2020, and 2022, also used GMACS.

## Assessment Methodology

This assessment model uses the modeling framework GMACS and is detailed in Appendix A. An updated version of GMACS (version 2.01.M.10, 2024-02-27) was used.

## Model Selection and Evaluation

The base model presented here, model 16.0, is the model accepted in 2022 with updated data including 2023 NMFS trawl survey biomass and size composition data, 2022 ADF\&G pot survey abundance and size composition data, and 1993-2022 groundfish and fixed gear bycatch estimates based on NMFS AKRO data. Six other models provide necessary updates to the base model, and explore natural mortality. The models presented are as follows:

1. 16.0: 2022 accepted model, using GMACS 2.01.M. 10 (2024-02-27), with a fixed $M=0.18$ for all years except the 1998 time block in which $M$ is estimated, updated with 2022/23 groundfish bycatch, 2023 NMFS trawl survey data, and 2022 ADF\&G pot survey data. The base model is model 16.0 since 2016 was the original year of model development and acceptance.
2. 16.0a: model 16.0 with a fully updated historical time series for the ADF\&G pot survey relative abundance index and size compositions, incorporating error corrections to the historical data sets and a more transparent, replicable data processing procedure. An erroneous relative abundance data point from the 2016 ADF\&G pot survey was identified and corrected.
3. 16.0b: model 16.0 with SSB estimated in season 5 of the model rather than season 4 . This change was deemed necessary after the clarification that GMACS estimates SSB at the beginning of a season. As SSB is intended to be estimated on 15 February, and 15 February is the dividing date between seasons 4 and 5 in this model, SSB estimation should take place in season 5.
4. 24.0: proposed new base model, incorporating the changes made in models 16.0 a and 16.0 b : the historical time series for the ADF\&G pot survey relative abundance index and size compositions are fully updated, and SSB is estimated in season 5 of the model rather than season 4.
5. 24.0a: model 24.0 with $M$ estimated using a tight prior (mean $=0.18, \mathrm{CV}=0.04$ ) as in the Bristol Bay Red King Crab stock assessment.
6. 24.0b: model 24.0 with $M$ estimated using a less restrictive prior (mean $=0.18, \mathrm{CV}=0.1$ ).
7. 24.0c: model 24.0 with $M$ fixed at the value estimated in model 24.0a.

## Results

## a. Sensitivity to new data

Model 16.0-2022 is the most recently accepted model for SMBKC, from September 2022. Model 16.0 is the same model with the addition of the following data: groundfish bycatch data for the $2022 / 23$ crab season, NMFS trawl survey data (biomass and length compositions) from the 2023 summer survey, and ADF\&G pot survey data (abundance and length compositions) from the 2022 survey. The 2022 ADF \&G pot survey size composition are relatively similar to previous years (Figure 3), while the 2023 NMFS trawl survey observed a greater relative abundance of larger males than in recent years (Figure 4). Comparisons of the fits of models 16.0-2022 and 16.0 to the NMFS trawl survey index (Figure 10) and the ADF\&G pot survey index (Figure 11) show nearly identical fits to the survey data, which is expected since there is only one new data point for each survey between these models. Estimates from models 16.0-2022 and 16.0 of spawning stock biomass (Figure 12) are also nearly identical, but recruitment estimates differ slightly (Figure 13). As has been noted in the past, this model fits neither survey index well, with particularly poor fits to the ADF\&G pot survey 2017-2018 data points and the 2006-2017 trawl survey data points.

## b. Effective sample sizes and weighting factors

Observed and estimated effective sample sizes are compared in Table 11. Data weighting factors, standard deviation of normalized residuals (SDNRs), and median absolute residual (MAR) are presented in Table 5. Francis (2011) weighting was applied in for this assessment in 2017 but, given the relatively few size bins in this assessment, this approach was suspended. The SDNR values for the trawl survey and pot survey are acceptable and similar across the model scenarios; the values are similar to those for most model scenarios considered in May 2022. The SDNR values for the directed pot fishery and other size compositions are also similar to previous estimates. The MAR values for the trawl survey and pot survey size compositions are adequate, ranging from 0.57 to 0.64 for the proposed new base model (24.0).

## c. Parameter estimates

Model parameter estimates for each of the models are summarized in Tables $12,13,14,15,16,17$, and 18, and compared in Table 19. Negative log-likelihood values and management measures for each of the model configurations are compared in Tables 4 and 20.

There are differences in parameter estimates among models as reflected in the log-likelihood components and the management quantities. The most drastic of these differences appear in model 24.0 b , which estimates natural mortality as 0.31 , using a less restrictive prior (mean $=0.18, \mathrm{CV}=0.10$ ) than does model 24.0a (which estimates $M=0.20$ ). The negative log-likelihood values for both survey indices are lower for model 24.0 b compared to all the other model scenarios.

Selectivity estimates for the directed fishery (Figure 14) as well as recruitment (Figures 15 and 16) are similar among the models, with the exception of model 24.0b, which estimates lower selectivity and higher recruitment than the other models. Fits to the NMFS trawl survey data (Figures 17 and 18) and estimated mature male biomass (MMB) on 15 February (Figures 1 and 2 ) are nearly identical among the models, with model 24.0 b again diverging from the others.

Estimated natural mortality in each year $\left(M_{t}\right)$ is presented in Figure 19, showing the mortality event in 1998/1999 for all models, as well as the different natural mortality values for the models with $M$ fixed at 0.18 (16.0, 16.0a, 16.0b, 24.0) versus those with $M$ estimated ( 24.0 a and 24.0 b ) or fixed at 0.20 ( 24.0 c ).

Estimates of fishing mortality from the new base model (24.0) are shown to assist with the rebuilding and reference point time frame discussions (Figure 20). Fishing mortality cannot be ruled out as being an influential factor in the current low stock status.

## d. Evaluation of the fit to the data.

The model scenarios do not show large differences in their fits to total male ( $\geq 90 \mathrm{~mm}$ CL) NMFS trawl survey biomass: all tend to miss the recent peak in 2010-2011, and fit recent survey data points on the lower end of their error bars (Figures 17 and 18). These fits are most likely being pulled down by the low abundance in the ADF\&G pot survey data in 2015-2018. The model scenarios also show similar fits to the pot survey relative abundance index, fitting the overall trends in the data but not capturing some of the high and low points (Figures 21 and 22).

Fits to the size compositions for trawl survey, pot survey, and commercial observer data are reasonable but miss the largest size category in some years (Figures 23, 24, and 25). Representative residual plots of the composition data generally have a similar fit to the three composition data sources (Figures 26, 27, and 28). The model fits to different types of retained and discarded catch values performed as expected given the assumed levels of uncertainty in the input data (Figure 29).

## e. Retrospective and historical analyses

The retrospective pattern in MMB for 10 peels for model 24.0 is shown in Figure 30; a positive retrospective bias begins to appear with the 2016 peel. The Mohn's $\rho$ value of 0.579 suggests model misspecification. However, for 5 peels, no retrospective bias is visually apparent and the Mohn's $\rho$ value is 0.037 (Figure 31). This discrepancy may be a result of the stability in NMFS trawl survey biomass estimates in the years following 2016 relative to the larger fluctuations in biomass estimates over 2012-2016.

## f. Uncertainty and sensitivity analyses.

Estimated standard deviations of parameters and selected management measures for the models are summarized for each individual model in Table 12. Model estimates of mature male biomass and OFL in 2023/24 are presented in Section F. A likelihood profile over $M$ suggests that the catch and size composition data are not highly informative for $M$ (Figures 32 and 33 ), while the two survey indices are informative (Figure 34) and likely drive the overall pattern (Figure 35). The biomass trajectories for the $M$ values in the profile show a wide spread of estimates at some points in the time series, e.g., the 2009 spike in biomass, but the biomass estimates for the terminal year are relatively similar (Figure 36).

## g. Comparison of alternative model scenarios.

The authors' recommended new base model is model 24.0 , which provides two needed updates to the old base model (16.0): the ADF\&G pot survey abundance index and size composition time series are fully updated using a consistent data processing procedure, which allowed for the detection and correction of errors in the time series, and SSB estimation is moved to season 5 of the model so that it occurs on the intended date of 15 February. Both models 16.0 and 24.0 have $M$ fixed at 0.18 for most of the time series, only allowing $M$ to be estimated for the 1998/1999 mortality event (Figure 19). Model 24.0 estimates a lower MMB in the most recent year, and in each of the last 10 years, than does model 16.0 (Tables 21 and 22).

Natural mortality is explored in models 24.0a, 24.0b, and 24.0c. When $M$ is estimated with a tight prior (mean $=0.18, \mathrm{CV}=0.04$ ), in 24.0a, the model estimates $M=0.20(\mathrm{SE}=0.01)$. When $M$ is estimated with a less restrictive prior ( mean $=0.18, \mathrm{CV}=0.10$ ), in 24.0 b , the model estimates $M=0.31$ ( $\mathrm{SE}=0.03$ ). In model $24.0 \mathrm{c}, M$ is fixed to the value estimated in model 24.0 a. Compared to the new base model 24.0 , estimated MMB in the final year is lower and $B_{M S Y}$ is higher for models $24.0 \mathrm{a}, 24.0 \mathrm{~b}$, and 24.0 c .

Although the trawl and pot survey data fits remain problematic, work is ongoing to generate a single VAST model index using both survey data sets.
The current Crab FMP (NPFMC 2021) states that "natural mortality of adult red king crab is assumed to be about 18 percent per year $(M=0.2)$ " and, in the absence of species-specific information, $M=0.18$ has
been the value used in SMBKC assessments as well (e.g., Webber et al. 2016). However, the CPT and SSC in 2023 accepted a Bristol Bay red king crab model for harvest specifications that estimates $M=0.23$ (Palof 2023), indicating that views may be shifting on the suitability of the $M=0.18$ value for BSAI king crab stocks. If the CPT decides that using $M$ values departing from that in the Crab FMP is warranted, model 24.0 a and/or 24.0 c may be worth further consideration for SMBKC. The authors recommend that at least one of these models is brought forward for consideration in September 2023, along with model 24.0.

## F. Calculation of the OFL and ABC

The overfishing level (OFL) is the total catch associated with the $F_{O F L}$ fishing mortality. The SMBKC stock is currently managed as Tier 4 , and only a Tier 4 analysis is presented here. Thus, given stock estimates or suitable proxy values of $B_{M S Y}$ and $F_{M S Y}$, along with two additional parameters $\alpha$ and $\beta, F_{O F L}$ is determined by the control rule

$$
\begin{align*}
& F_{O F L}= \begin{cases}F_{M S Y}, & \text { when } B / B_{M S Y}>1 \\
F_{M S Y} \frac{\left(B / B_{M S Y}-\alpha\right)}{(1-\alpha)}, & \text { when } \beta<B / B_{M S Y} \leq 1\end{cases}  \tag{1}\\
& F_{O F L}<F_{M S Y} \text { with directed fishery } F=0 \text { when } B / B_{M S Y} \leq \beta
\end{align*}
$$

where $B$ is quantified as mature-male biomass (MMB) at mating with time of mating assigned a nominal date of 15 February. Note that $B$ is a function of the fishing mortality $F_{O F L}$ (therefore numerical approximation of $F_{O F L}$ is required). As implemented for this assessment, all calculations proceed according to the model equations given in Appendix A. $F_{O F L}$ is taken to be full-selection fishing mortality in the directed pot fishery, and groundfish trawl and fixed-gear fishing mortalities set at their geometric mean values over years for which there are data-based estimates of bycatch-mortality biomass.

The currently recommended Tier 4 convention is to use the full assessment period, currently 1978-2022, to define a $B_{M S Y}$ proxy in terms of average estimated MMB and to set $\gamma=1.0$ with assumed stock natural mortality $M=0.18 \mathrm{yr}^{-1}$ in setting the $F_{M S Y}$ proxy value $\gamma M$. Note that, for models 24.0a, 24.0b, and 24.0 c , the values used for $M$ are $0.20,0.31$, and 0.20 respectively. The parameters $\alpha$ and $\beta$ are assigned their default values $\alpha=0.10$ and $\beta=0.25$. The $F_{\text {OFL }}$, OFL, ABC, and MMB in $2023 / 24$ for all the models are summarized in Table 4. The currently recommended ABC is $75 \%$ of the OFL (ABC buffer $=25 \%$ ).

Table 4: Comparisons of management measures for the models 16.0, 16.0a, 16.0b, 16.0c, and 24.0. Biomass and OFL are in tons.

| Component | Model.16.0 | Model.16.0a | Model.16.0b | Model.24.0 | Model.24.0a | Model.24.0b | Model.24.0c |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathrm{MMB}_{2023}$ | 1498.480 | 1447.045 | 1442.973 | 1393.537 | 1347.944 | 1326.499 | 1347.944 |
| $B_{\mathrm{MSY}}$ | 3212.341 | 3183.997 | 2943.085 | 2916.921 | 2972.172 | 3012.624 | 2972.172 |
| $M_{\mathrm{MM}} / B_{\mathrm{MSY}}$ | 0.466 | 0.454 | 0.490 | 0.478 | 0.454 | 0.440 | 0.454 |
| $F_{\mathrm{OFL}}$ | 0.073 | 0.071 | 0.078 | 0.076 | 0.080 | 0.117 | 0.080 |
| $\mathrm{OFL}_{2023}$ | 85.616 | 80.258 | 90.920 | 85.283 | 86.177 | 115.085 | 86.177 |
| $\mathrm{ABC}_{2023}$ | 64.212 | 60.194 | 68.190 | 63.962 | 64.633 | 86.314 | 64.633 |

## G. Rebuilding Analysis and Update

This stock was declared overfished in fall of 2018, and a rebuilding plan was approved by the NPFMC in June 2020 (NPFMC 2020a). The most updated rebuilding plan can be found on the NPFMC website for the June 2020 meeting (NPFMC 2020b). This assessment was moved to a biannual assessment in early 2021, with full assessments performed in even-numbered years, which falls in line with the two-year rebuilding progress updates required under the rebuilding plan.

The recovery of this stock is highly dependent upon successful recruitment, which is likely linked to climate variability but not well understood. Recruitment was likely negatively impacted by an ecosystem regime shift in the Bering Sea in 1989, and above-average bottom temperatures in recent years (NPFMC 2020b). The 2023 NMFS trawl survey found that water $<-1{ }^{\circ} \mathrm{C}$ extended south of St. Matthew Island, the farthest south water in this temperature range has extended since 2015 (Zacher et al. 2024). However, it is unknown whether and how the colder water temperatures in 2023 will influence the St. Matthew Island blue king crab stock.

NMFS trawl survey biomass of males in the model has been low in 2021-2023 (Figure 10). The most recent ADF\&G pot survey, which occurred in 2022, observed a relative abundance index that was the highest since 2013 (Figure 11). Model-estimated MMB increased in 2022 (Figures 1 and 2), mostly due to an increase in recruitment. Model estimates of recruitment increased both in 2021 and 2022, suggesting some potential for future stock growth (Figures 15 and 16).

## H. Data Gaps and Research Priorities

The following topics have been listed as areas where more research on SMBKC is needed:

1. Growth increments and molting probabilities as a function of size.
2. Trawl survey catchability and selectivities.
3. Pot survey catchability and selectivities.
4. Temporal changes in spatial distributions near the island.
5. Natural mortality.

## I. Projections and outlook

The outlook for recruitment is pessimistic and the abundance relative to the proxy $B_{M S Y}$ is low, although improved compared to recent years. To examine the impact of historical fishing, we conducted a "dynamic$B_{0}$ " analysis, which projects the population based on estimated recruitment but removes the effect of fishing. Using the new base model (24.0), the results of this analysis suggest that the impact of fishing has reduced the stock to about $72 \%$ of what it would have been in the absence of fishing (Figure 37), supporting the hypothesis that fishing pressure is not the sole contributor to the decline of this stock in recent years. The other non-fishing contributors to the observed depleted stock trend (ignoring the stock-recruit relationship) may reflect variable survival rates due to environmental conditions and also range shifts.

## J. Acknowledgements

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## Tables

Table 5: Comparisons of data weights, SDNR and MAR (standard deviation of normalized residuals and median absolute residual) values for the model scenarios.

| Component.wt | Model.16.0 | Model.16.0a | Model.16.0b | Model.24.0 | Model.24.0a | Model.24.0b | Model.24.0c |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| NMFS trawl survey weight | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| ADF\&G pot survey weight | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| Directed pot LF weight | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| NMFS trawl survey LF weight | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| ADF\&G pot survey LF weight | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| SDNR NMFS trawl survey | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| SDNR ADF\&G pot survey | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| SDNR directed pot LF | 0.67 | 0.66 | 0.67 | 0.66 | 0.66 | 0.67 | 0.66 |
| SDNR NMFS trawl survey LF | 1.25 | 1.25 | 1.25 | 1.25 | 1.22 | 1.21 | 1.22 |
| SDNR ADF\&G pot survey LF | 0.93 | 0.90 | 0.93 | 0.90 | 0.90 | 0.92 | 0.90 |
| MAR NMFS trawl survey | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| MAR ADF\&G pot survey | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| MAR directed pot LF | 0.46 | 0.44 | 0.46 | 0.44 | 0.45 | 0.57 | 0.45 |
| MAR NMFS trawl survey LF | 0.59 | 0.64 | 0.59 | 0.64 | 0.60 | 0.60 | 0.60 |
| MAR ADF\&G pot survey LF | 0.68 | 0.57 | 0.68 | 0.57 | 0.54 | 0.72 | 0.54 |

Table 6: Fishery characteristics and update. Columns include the 1978/79 to 2015/16 directed St. Matthew Island blue king crab pot fishery. The Guideline Harvest Level (GHL) and Total Allowable Catch (TAC) are in millions of pounds. Harvest includes deadloss. Catch per unit effort (CPUE) in this table is the harvest number / pot lifts. The average weight is the harvest weight / harvest number in pounds. The average carapace length (CL) is the average of retained crab in mm from dockside sampling of delivered crab. Source: Fitch et al 2012; ADF\&G Dutch Harbor staff, pers. comm. Note that management (GHL) units are in pounds, for conserving space, conversion to tons is ommitted.

| Year | Dates | GHL/TAC | Harvest |  | Pot lifts | CPUE | Avg wt | Avg CL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Crab | Pounds |  |  |  |  |
| 1978/79 | 07/15-09/03 |  | 436,126 | 1,984,251 | 43,754 | 10 | 4.5 | 132.2 |
| 1979/80 | 07/15-08/24 |  | 52,966 | 210,819 | 9,877 | 5 | 4.0 | 128.8 |
| 1980/81 | 07/15-09/03 |  |  | CONFI | ENTIAL |  |  |  |
| 1981/82 | 07/15-08/21 |  | 1,045,619 | 4,627,761 | 58,550 | 18 | 4.4 | NA |
| 1982/83 | 08/01-08/16 |  | 1,935,886 | 8,844,789 | 165,618 | 12 | 4.6 | 135.1 |
| 1983/84 | 08/20-09/06 | 8.0 | 1,931,990 | 9,454,323 | 133,944 | 14 | 4.9 | 137.2 |
| 1984/85 | 09/01-09/08 | 2.0-4.0 | 841,017 | 3,764,592 | 73,320 | 11 | 4.5 | 135.5 |
| 1985/86 | 09/01-09/06 | 0.9-1.9 | 436,021 | 2,175,087 | 46,988 | 9 | 5.0 | 139.0 |
| 1986/87 | 09/01-09/06 | 0.2-0.5 | 219,548 | 1,003,162 | 22,073 | 10 | 4.6 | 134.3 |
| 1987/88 | 09/01-09/05 | 0.6-1.3 | 227,447 | 1,039,779 | 28,230 | 8 | 4.6 | 134.1 |
| 1988/89 | 09/01-09/05 | 0.7-1.5 | 280,401 | 1,236,462 | 21,678 | 13 | 4.4 | 133.3 |
| 1989/90 | 09/01-09/04 | 1.7 | 247,641 | 1,166,258 | 30,803 | 8 | 4.7 | 134.6 |
| 1990/91 | 09/01-09/07 | 1.9 | 391,405 | 1,725,349 | 26,264 | 15 | 4.4 | 134.3 |
| 1991/92 | 09/16-09/20 | 3.2 | 726,519 | 3,372,066 | 37,104 | 20 | 4.6 | 134.1 |
| 1992/93 | 09/04-09/07 | 3.1 | 545,222 | 2,475,916 | 56,630 | 10 | 4.5 | 134.1 |
| 1993/94 | 09/15-09/21 | 4.4 | 630,353 | 3,003,089 | 58,647 | 11 | 4.8 | 135.4 |
| 1994/95 | 09/15-09/22 | 3.0 | 827,015 | 3,764,262 | 60,860 | 14 | 4.9 | 133.3 |
| 1995/96 | 09/15-09/20 | 2.4 | 666,905 | 3,166,093 | 48,560 | 14 | 4.7 | 135.0 |
| 1996/97 | 09/15-09/23 | 4.3 | 660,665 | 3,078,959 | 91,085 | 7 | 4.7 | 134.6 |
| 1997/98 | 09/15-09/22 | 5.0 | 939,822 | 4,649,660 | 81,117 | 12 | 4.9 | 139.5 |
| 1998/99 | 09/15-09/26 | 4.0 | 635,370 | 2,968,573 | 91,826 | 7 | 4.7 | 135.8 |
| 1999/00 | 2008/09 |  |  | FISHER | CLOSED |  |  |  |
| 2009/10 | 10/15-02/01 | 1.17 | 103,376 | 460,859 | 10,697 | 10 | 4.5 | 134.9 |
| 2010/11 | 10/15-02/01 | 1.60 | 298,669 | 1,263,982 | 29,344 | 10 | 4.2 | 129.3 |
| 2011/12 | 10/15-02/01 | 2.54 | 437,862 | 1,881,322 | 48,554 | 9 | 4.3 | 130.0 |
| 2012/13 | 10/15-02/01 | 1.63 | 379,386 | 1,616,054 | 37,065 | 10 | 4.3 | 129.8 |
| 2013/14 |  |  |  | FISHER | CLOSED |  |  |  |
| 2014/15 | 10/15-02/05 | 0.66 | 69,109 | 308,582 | 10,133 | 7 | 4.5 | 132.3 |
| 2015/16 | 10/19-11/28 | 0.41 | 24,076 | 105,010 | 5,475 | 4 | 4.4 | 132.6 |
| 2016/17 |  |  |  | FISHER | CLOSED |  |  |  |
| 2017/18 |  |  |  | FISHER | CLOSED |  |  |  |
| 2018/19 |  |  |  | FISHER | CLOSED |  |  |  |
| 2019/20 |  |  |  | FISHER | CLOSED |  |  |  |
| 2020/21 |  |  |  | FISHER | CLOSED |  |  |  |
| 2021/22 |  |  |  | FISHER | CLOSED |  |  |  |
| 2022/23 |  |  |  | FISHER | CLOSED |  |  |  |
| 2023/24 |  |  |  | FISHER | CLOSED |  |  |  |

Table 7: Observed proportion of crab by size class during the ADF\&G crab observer pot-lift sampling. Source: ADF\&G Crab Observer Database.

| Year | Total pot lifts | Pot lifts sampled | Number of crab (90 mm+ CL) | Stage 1 | Stage 2 | Stage 3 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| $1990 / 91$ | 26,264 | 10 | 150 | 0.113 | 0.393 | 0.493 |
| $1991 / 92$ | 37,104 | 125 | 3,393 | 0.133 | 0.177 | 0.690 |
| $1992 / 93$ | 56,630 | 71 | 1,606 | 0.191 | 0.268 | 0.542 |
| $1993 / 94$ | 58,647 | 84 | 2,241 | 0.281 | 0.210 | 0.510 |
| $1994 / 95$ | 60,860 | 48,560 | 203 | 4,735 | 0.294 | 0.271 |
| $1995 / 96$ | 41,085 | 47 | 663 | 0.148 | 0.212 | 0.434 |
| $1996 / 97$ | 81,117 | 96 | 489 | 0.160 | 0.223 | 0.618 |
| $1997 / 98$ | 133 | 3,195 | 0.182 | 0.205 | 0.613 |  |
| $1998 / 99$ | 135 | 1.322 | 0.193 | 0.216 | 0.591 |  |
| $1999 / 00-2008 / 09$ |  | 989 | FISHERY CLOSED |  |  |  |
| $2009 / 10$ | 10,484 | 2,419 | 19,802 | 0.141 | 0.324 | 0.535 |
| $2010 / 11$ | 29,356 | 3,359 | 45,466 | 0.131 | 0.315 | 0.553 |
| $2011 / 12$ | 4,554 | 2,841 | 58,666 | 0.131 | 0.305 | 0.564 |
| $2012 / 13$ | 37,065 |  | 57,298 | 0.141 | 0.318 | 0.541 |
| $2013 / 14$ |  |  | 495 | 9,906 | 0.094 | 0.228 |
| $2014 / 15$ | 10,133 | 5,475 |  | 3,248 | 0.115 | 0.252 |
| $2015 / 16$ |  |  | FISHERY CLOSED | 0.679 |  |  |
| $2016 / 17-2023 / 24$ |  |  |  |  |  |  |

Table 8: Groundfish SMBKC male bycatch biomass ( t ) estimates. Trawl includes pelagic trawl and nonpelagic trawl types. Source: J. Zheng, ADF\&G, and author estimates based on data from R. Foy, NMFS. Estimates used after 2008/09 are from NMFS Alaska Regional Office.

| Year | Trawl bycatch | Fixed gear bycatch |
| :---: | :---: | :---: |
| 1978 | 0.000 | 0.000 |
| 1979 | 0.000 | 0.000 |
| 1980 | 0.000 | 0.000 |
| 1981 | 0.000 | 0.000 |
| 1982 | 0.000 | 0.000 |
| 1983 | 0.000 | 0.000 |
| 1984 | 0.000 | 0.000 |
| 1985 | 0.000 | 0.000 |
| 1986 | 0.000 | 0.000 |
| 1987 | 0.000 | 0.000 |
| 1988 | 0.000 | 0.000 |
| 1989 | 0.000 | 0.000 |
| 1990 | 0.000 | 0.000 |
| 1991 | 3.538 | 0.045 |
| 1992 | 1.996 | 2.268 |
| 1993 | 1.542 | 0.500 |
| 1994 | 0.318 | 0.091 |
| 1995 | 0.635 | 0.136 |
| 1996 | 0.500 | 0.045 |
| 1997 | 0.500 | 0.181 |
| 1998 | 0.500 | 0.907 |
| 1999 | 0.500 | 1.361 |
| 2000 | 0.500 | 0.500 |
| 2001 | 0.500 | 0.862 |
| 2002 | 0.726 | 0.408 |
| 2003 | 0.998 | 1.134 |
| 2004 | 0.091 | 0.635 |
| 2005 | 0.500 | 0.590 |
| 2006 | 2.812 | 1.451 |
| 2007 | 0.045 | 69.717 |
| 2008 | 0.272 | 6.622 |
| 2009 | 0.638 | 7.522 |
| 2010 | 0.360 | 9.564 |
| 2011 | 0.170 | 0.796 |
| 2012 | 0.011 | 0.739 |
| 2013 | 0.163 | 0.341 |
| 2014 | 0.010 | 0.490 |
| 2015 | 0.010 | 0.711 |
| 2016 | 0.229 | 1.630 |
| 2017 | 0.048 | 5.935 |
| 2018 | 0.001 | 1.224 |
| 2019 | 0.030 | 1.124 |
| 2020 | 0.001 | 0.671 |
| 2021 | 0.001 | 0.323 |
| 2022 | 0.001 | 2.118 |

Table 9: NMFS Eastern Bering Sea trawl-survey area-swept estimates of male crab abundance ( $10^{6}$ crab) and male ( $\geq 90 \mathrm{~mm} \mathrm{CL}$ ) biomass ( $10^{6} \mathrm{lbs}$ ). Total number of captured male crab $\geq 90 \mathrm{~mm}$ CL is also given. Source: J.Richar, NMFS. The " + " refer to plus group.

| Year | Abundance |  |  |  |  | Biomass |  | Number of crabs |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\begin{gathered} \text { Stage-1 } \\ (90-104 \mathrm{~mm}) \end{gathered}$ | $\begin{gathered} \text { Stage-2 } \\ (105-119 \mathrm{~mm}) \end{gathered}$ | $\begin{gathered} \text { Stage-3 } \\ (120+\mathrm{mm}) \end{gathered}$ | Total | CV | $\begin{gathered} \text { Total } \\ (90+\mathrm{mm} \text { CL }) \end{gathered}$ | CV |  |
| 1978 | 2.213 | 1.991 | 1.521 | 5.726 | 0.411 | 15.064 | 0.394 | 157 |
| 1979 | 3.061 | 2.281 | 1.808 | 7.150 | 0.472 | 17.615 | 0.463 | 178 |
| 1980 | 2.856 | 2.563 | 2.541 | 7.959 | 0.572 | 22.017 | 0.507 | 185 |
| 1981 | 0.483 | 1.213 | 2.263 | 3.960 | 0.368 | 14.443 | 0.402 | 140 |
| 1982 | 1.669 | 2.431 | 5.884 | 9.984 | 0.401 | 35.763 | 0.344 | 271 |
| 1983 | 1.061 | 1.651 | 3.345 | 6.057 | 0.332 | 21.240 | 0.298 | 231 |
| 1984 | 0.435 | 0.497 | 1.452 | 2.383 | 0.175 | 8.976 | 0.179 | 105 |
| 1985 | 0.379 | 0.376 | 1.117 | 1.872 | 0.216 | 6.858 | 0.210 | 93 |
| 1986 | 0.203 | 0.447 | 0.374 | 1.025 | 0.428 | 3.124 | 0.388 | 46 |
| 1987 | 0.325 | 0.631 | 0.715 | 1.671 | 0.302 | 5.024 | 0.291 | 71 |
| 1988 | 0.410 | 0.816 | 0.957 | 2.183 | 0.285 | 6.963 | 0.252 | 81 |
| 1989 | 2.169 | 1.154 | 1.786 | 5.109 | 0.314 | 13.974 | 0.271 | 208 |
| 1990 | 1.053 | 1.031 | 2.338 | 4.422 | 0.302 | 14.837 | 0.274 | 170 |
| 1991 | 1.147 | 1.665 | 2.233 | 5.046 | 0.259 | 15.318 | 0.248 | 197 |
| 1992 | 1.074 | 1.382 | 2.291 | 4.746 | 0.206 | 15.638 | 0.201 | 220 |
| 1993 | 1.521 | 1.828 | 3.276 | 6.626 | 0.185 | 21.051 | 0.169 | 324 |
| 1994 | 0.883 | 1.298 | 2.257 | 4.438 | 0.187 | 14.416 | 0.176 | 211 |
| 1995 | 1.025 | 1.188 | 1.741 | 3.953 | 0.187 | 12.574 | 0.178 | 178 |
| 1996 | 1.238 | 1.891 | 3.064 | 6.193 | 0.263 | 20.746 | 0.241 | 285 |
| 1997 | 1.165 | 2.228 | 3.789 | 7.182 | 0.367 | 24.084 | 0.337 | 296 |
| 1998 | 0.660 | 1.661 | 2.849 | 5.170 | 0.373 | 17.586 | 0.355 | 243 |
| 1998 | 0.223 | 0.222 | 0.558 | 1.003 | 0.192 | 3.515 | 0.182 | 52 |
| 2000 | 0.282 | 0.285 | 0.740 | 1.307 | 0.303 | 4.623 | 0.310 | 61 |
| 2001 | 0.419 | 0.502 | 0.938 | 1.859 | 0.243 | 6.242 | 0.245 | 91 |
| 2002 | 0.111 | 0.230 | 0.640 | 0.981 | 0.311 | 3.820 | 0.320 | 38 |
| 2003 | 0.449 | 0.280 | 0.465 | 1.194 | 0.399 | 3.454 | 0.336 | 65 |
| 2004 | 0.247 | 0.184 | 0.562 | 0.993 | 0.369 | 3.360 | 0.305 | 48 |
| 2005 | 0.319 | 0.310 | 0.501 | 1.130 | 0.403 | 3.620 | 0.371 | 42 |
| 2006 | 0.917 | 0.642 | 1.240 | 2.798 | 0.339 | 8.585 | 0.334 | 126 |
| 2007 | 2.518 | 2.020 | 1.193 | 5.730 | 0.420 | 14.266 | 0.385 | 250 |
| 2008 | 1.352 | 0.801 | 1.457 | 3.609 | 0.289 | 10.261 | 0.284 | 167 |
| 2009 | 1.573 | 2.161 | 1.410 | 5.144 | 0.263 | 13.892 | 0.256 | 251 |
| 2010 | 3.937 | 3.253 | 2.458 | 9.648 | 0.544 | 24.539 | 0.466 | 388 |
| 2011 | 1.800 | 3.255 | 3.207 | 8.263 | 0.587 | 24.099 | 0.558 | 318 |
| 2012 | 0.705 | 1.970 | 1.808 | 4.483 | 0.361 | 13.669 | 0.339 | 193 |
| 2013 | 0.335 | 0.452 | 0.807 | 1.593 | 0.215 | 5.043 | 0.217 | 74 |
| 2014 | 0.723 | 1.627 | 1.809 | 4.160 | 0.503 | 13.292 | 0.449 | 181 |
| 2015 | 0.992 | 1.269 | 1.979 | 4.240 | 0.774 | 12.958 | 0.770 | 153 |
| 2016 | 0.535 | 0.660 | 1.178 | 2.373 | 0.447 | 7.685 | 0.393 | 108 |
| 2017 | 0.091 | 0.323 | 0.663 | 1.077 | 0.657 | 3.955 | 0.600 | 42 |
| 2018 | 0.154 | 0.232 | 0.660 | 1.047 | 0.298 | 3.816 | 0.281 | 62 |
| 2019 | 0.403 | 0.482 | 1.170 | 2.056 | 0.352 | 6.990 | 0.337 | 105 |
| 2021 | 0.423 | 0.168 | 0.682 | 1.273 | 0.496 | 4.253 | 0.427 | 59 |
| 2022 | 0.620 | 0.372 | 0.763 | 1.754 | 0.452 | 5.216 | 0.497 | 75 |
| 2023 | 0.512 | 0.474 | 0.608 | 1.593 | 0.458 | 4.622 | 0.439 | 76 |

Table 10: Size-class and total CPUE ( $90+\mathrm{mm}$ carapace length) with estimated CV and total number of captured crab ( $90+\mathrm{mm}$ carapace length) from the 96 common stations surveyed during the ADF\&G St. Matthew Island blue king crab pot surveys. Source: ADF\&G.

| Year | Stage-1 <br> $(90-104 \mathrm{~mm})$ | Stage-2 <br> $(105-119 \mathrm{~mm})$ | Stage-3 <br> $(120+\mathrm{mm})$ | Total CPUE | CV | Number of crabs |
| :---: | :---: | :---: | :---: | ---: | ---: | ---: |
| 1995 | 1.919 | 3.198 | 6.924 | 12.042 | 0.13 | 4624 |
| 1998 | 0.964 | 2.763 | 8.805 | 12.531 | 0.06 | 4812 |
| 2001 | 1.266 | 1.737 | 5.474 | 8.477 | 0.08 | 3255 |
| 2004 | 0.112 | 0.414 | 1.141 | 1.667 | 0.15 | 640 |
| 2007 | 1.083 | 2.720 | 4.826 | 8.630 | 0.09 | 3325 |
| 2010 | 1.318 | 3.258 | 5.591 | 10.167 | 0.10 | 3904 |
| 2013 | 0.862 | 1.383 | 3.362 | 5.607 | 0.19 | 2153 |
| 2015 | 0.206 | 0.698 | 1.901 | 2.805 | 0.18 | 1077 |
| 2016 | 0.198 | 0.440 | 1.383 | 2.021 | 0.17 | 776 |
| 2017 | 0.177 | 0.424 | 1.073 | 1.674 | 0.25 | 643 |
| 2018 | 0.076 | 0.161 | 0.508 | 0.745 | 0.14 | 286 |
| 2022 | 0.630 | 1.030 | 2.432 | 4.089 | 0.14 | 1570 |

Table 11: Observed and input sample sizes for observer data from the directed pot fishery, the NMFS trawl survey, and the ADF\&G pot survey.

| Year | Number measured |  |  | Input sample sizes |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Observer pot | NMFS trawl | ADF\&G pot | Observer pot | NMFS trawl | ADF\&G pot |
| 1978 |  | 157 |  |  | 50 |  |
| 1979 |  | 178 |  |  | 50 |  |
| 1980 |  | 185 |  |  | 50 |  |
| 1981 |  | 140 |  |  | 50 |  |
| 1982 |  | 271 |  |  | 50 |  |
| 1983 |  | 231 |  |  | 50 |  |
| 1984 |  | 105 |  |  | 50 |  |
| 1985 |  | 93 |  |  | 46.5 |  |
| 1986 |  | 46 |  |  | 23 |  |
| 1987 |  | 71 |  |  | 35.5 |  |
| 1988 |  | 81 |  |  | 40.5 |  |
| 1989 |  | 208 |  |  | 50 |  |
| 1990 | 150 | 170 |  | 15 | 50 |  |
| 1991 | 3393 | 197 |  | 25 | 50 |  |
| 1992 | 1606 | 220 |  | 25 | 50 |  |
| 1993 | 2241 | 324 |  | 25 | 50 |  |
| 1994 | 4735 | 211 |  | 25 | 50 |  |
| 1995 | 663 | 178 | 4624 | 25 | 50 | 100 |
| 1996 | 489 | 285 |  | 25 | 50 |  |
| 1997 | 3195 | 296 |  | 25 | 50 |  |
| 1998 | 1323 | 243 | 4812 | 25 | 50 | 100 |
| 1999 |  | 52 |  |  | 26 |  |
| 2000 |  | 61 |  |  | 30.5 |  |
| 2001 |  | 91 | 3255 |  | 45.5 | 100 |
| 2002 |  | 38 |  |  | 19 |  |
| 2003 |  | 65 |  |  | 32.5 |  |
| 2004 |  | 48 | 640 |  | 24 | 100 |
| 2005 |  | 42 |  |  | 21 |  |
| 2006 |  | 126 |  |  | 50 |  |
| 2007 |  | 250 | 3319 |  | 50 | 100 |
| 2008 |  | 167 |  |  | 50 |  |
| 2009 | 19802 | 251 |  | 50 | 50 |  |
| 2010 | 45466 | 388 | 3920 | 50 | 50 | 100 |
| 2011 | 58667 | 318 |  | 50 | 50 |  |
| 2012 | 57282 | 193 |  | 50 | 50 |  |
| 2013 |  | 74 | 2167 |  | 37 | 100 |
| 2014 | 9906 | 181 |  | 50 | 50 |  |
| 2015 | 3248 | 153 | 1077 | 50 | 50 | 100 |
| 2016 |  | 108 | 777 |  | 50 | 100 |
| 2017 |  | 42 | 643 |  | 21 | 100 |
| 2018 |  | 62 | 286 |  | 31 | 100 |
| 2019 |  | 105 |  |  | 50 |  |
| 2020 |  |  |  |  |  |  |
| 2021 |  | 59 |  |  | 50 |  |
| 2022 |  | 75 | 1570 |  | 50 | 100 |
| 2023 |  | 76 |  |  | 50 |  |

Table 12: Model parameter estimates, selected derived quantities, and their standard errors (SE) for model 16.0, the 2022 accepted model with recent data added.

| Parameter | Estimate | SE |
| :--- | ---: | ---: |
| Natural mortality deviation in 1998/99 $\left(\delta_{1998}^{M}\right)$ | 1.581 | 0.136 |
| $\log (\bar{R})$ | 13.864 | 0.192 |
| $\log \left(n_{1}^{0}\right)$ | 14.954 | 0.175 |
| $\log \left(n_{2}^{0}\right)$ | 14.528 | 0.210 |
| $\log \left(n_{3}^{0}\right)$ | 14.339 | 0.207 |
| $q_{p o t}$ | 3.827 | 0.245 |
| $\log \left(\bar{F}^{\mathrm{df}}\right)$ | -2.134 | 0.051 |
| $\log \left(\bar{F}^{\mathrm{tb}}\right)$ | -9.927 | 0.071 |
| $\log \left(\bar{F}^{\mathrm{fb}}\right)$ | -8.049 | 0.071 |
| $\log$ Stage-1 directed pot selectivity 1978-2008 | -0.925 | 0.180 |
| $\log$ Stage-2 directed pot selectivity 1978-2008 | -0.562 | 0.131 |
| $\log$ Stage-1 directed pot selectivity $2009-2017$ | -0.545 | 0.163 |
| $\log$ Stage-2 directed pot selectivity $2009-2017$ | -0.000 | 0.000 |
| $\log$ Stage-1 NMFS trawl selectivity | -0.294 | 0.064 |
| $\log$ Stage-2 NMFS trawl selectivity | -0.000 | 0.000 |
| $\log$ Stage-1 ADF\&G pot selectivity | -0.809 | 0.121 |
| $\log$ Stage-2 ADF\&G pot selectivity | -0.000 | 0.000 |
| $F_{\text {OFL }}$ | 0.073 | 0.009 |
| OFL | 85.616 | 17.664 |

Table 13: Model parameter estimates, selected derived quantities, and their standard errors (SE) for model 16.0a, with updated historical pot survey data.

| Parameter | Estimate | SE |
| :--- | ---: | ---: |
| Natural mortality deviation in 1998/99 $\left(\delta_{1998}^{M}\right)$ | 1.566 | 0.138 |
| $\log (\bar{R})$ | 13.852 | 0.191 |
| $\log \left(n_{1}^{0}\right)$ | 14.955 | 0.175 |
| $\log \left(n_{2}^{0}\right)$ | 14.527 | 0.209 |
| $\log \left(n_{3}^{0}\right)$ | 14.338 | 0.207 |
| $q_{p o t}$ | 3.905 | 0.239 |
| $\log \left(\bar{F}^{\text {df }}\right)$ | -2.121 | 0.051 |
| $\log \left(\bar{F}^{\mathrm{tb}}\right)$ | -9.903 | 0.070 |
| $\log \left(\bar{F}^{\mathrm{fb}}\right)$ | -8.026 | 0.070 |
| $\log$ Stage-1 directed pot selectivity 1978-2008 | -0.914 | 0.180 |
| $\log$ Stage-2 directed pot selectivity 1978-2008 | -0.559 | 0.132 |
| $\log$ Stage-1 directed pot selectivity $2009-2017$ | -0.585 | 0.162 |
| $\log$ Stage-2 directed pot selectivity $2009-2017$ | -0.000 | 0.000 |
| $\log$ Stage-1 NMFS trawl selectivity | -0.297 | 0.061 |
| $\log$ Stage-2 NMFS trawl selectivity | -0.000 | 0.000 |
| $\log$ Stage-1 ADF\&G pot selectivity | -0.762 | 0.119 |
| $\log$ Stage-2 ADF\&G pot selectivity | -0.000 | 0.000 |
| $F_{\text {OFL }}$ | 0.071 | 0.009 |
| OFL | 80.258 | 16.570 |

Table 14: Model parameter estimates, selected derived quantities, and their standard errors (SE) for model 16.0b, with spawning stock biomass (SSB) estimated in season 5 rather than season 4.

| Parameter | Estimate | SE |
| :--- | ---: | ---: |
| Natural mortality deviation in 1998/99 $\left(\delta_{1998}^{M}\right)$ | 1.581 | 0.136 |
| $\log (\bar{R})$ | 13.864 | 0.192 |
| $\log \left(n_{1}^{0}\right)$ | 14.954 | 0.175 |
| $\log \left(n_{2}^{0}\right)$ | 14.528 | 0.210 |
| $\log \left(n_{3}^{0}\right)$ | 14.339 | 0.207 |
| $q_{p o t}$ | 3.827 | 0.245 |
| $\log \left(\bar{F}^{\text {df }}\right)$ | -2.134 | 0.051 |
| $\log \left(\bar{F}^{\mathrm{tb}}\right)$ | -9.927 | 0.071 |
| $\log \left(\bar{F}^{\mathrm{fb}}\right)$ | -8.049 | 0.071 |
| $\log$ Stage-1 directed pot selectivity $1978-2008$ | -0.925 | 0.180 |
| $\log$ Stage-2 directed pot selectivity 1978-2008 | -0.562 | 0.131 |
| $\log$ Stage-1 directed pot selectivity 2009-2017 | -0.545 | 0.163 |
| $\log$ Stage-2 directed pot selectivity $2009-2017$ | -0.000 | 0.000 |
| $\log$ Stage-1 NMFS trawl selectivity | -0.294 | 0.064 |
| $\log$ Stage-2 NMFS trawl selectivity | -0.000 | 0.000 |
| $\log$ Stage-1 ADF\&G pot selectivity | -0.809 | 0.121 |
| $\log$ Stage-2 ADF\&G pot selectivity | -0.000 | 0.000 |
| $F_{\text {OFL }}$ | 0.078 | 0.009 |
| OFL | 90.920 | 18.567 |

Table 15: Model parameter estimates, selected derived quantities, and their standard errors (SE) for model 24.0, the proposed new base model, with updated historical pot survey data and spawning stock biomass (SSB) estimated in season 5.

| Parameter | Estimate | SE |
| :--- | ---: | ---: |
| Natural mortality deviation in 1998/99 $\left(\delta_{1998}^{M}\right)$ | 1.566 | 0.138 |
| $\log (\bar{R})$ | 13.852 | 0.191 |
| $\log \left(n_{1}^{0}\right)$ | 14.955 | 0.175 |
| $\log \left(n_{2}^{0}\right)$ | 14.527 | 0.209 |
| $\log \left(n_{3}^{0}\right)$ | 14.338 | 0.207 |
| $q_{p o t}$ | 3.905 | 0.239 |
| $\log \left(\bar{F}^{\mathrm{df}}\right)$ | -2.121 | 0.051 |
| $\log \left(\bar{F}^{\mathrm{tb}}\right)$ | -9.903 | 0.070 |
| $\log \left(\bar{F}^{\mathrm{fb}}\right)$ | -8.026 | 0.070 |
| $\log$ Stage-1 directed pot selectivity 1978-2008 | -0.914 | 0.180 |
| $\log$ Stage-2 directed pot selectivity 1978-2008 | -0.559 | 0.132 |
| $\log$ Stage-1 directed pot selectivity $2009-2017$ | -0.585 | 0.162 |
| $\log$ Stage-2 directed pot selectivity $2009-2017$ | -0.000 | 0.000 |
| $\log$ Stage-1 NMFS trawl selectivity | -0.297 | 0.061 |
| $\log$ Stage-2 NMFS trawl selectivity | -0.000 | 0.000 |
| $\log$ Stage-1 ADF\&G pot selectivity | -0.762 | 0.119 |
| $\log$ Stage-2 ADF\&G pot selectivity | -0.000 | 0.000 |
| $F_{\text {OFL }}$ | 0.076 | 0.009 |
| OFL | 85.283 | 17.423 |

Table 16: Model parameter estimates, selected derived quantities, and their standard errors (SE) for model 24.0a, with natural mortality estimated using a tight prior.

| Parameter | Estimate | SE |
| :--- | ---: | ---: |
| Natural mortality deviation in 1998/99 $\left(\delta_{1998}^{M}\right)$ | 1.413 | 0.150 |
| $\log (\bar{R})$ | 13.962 | 0.194 |
| $\log \left(n_{1}^{0}\right)$ | 15.052 | 0.177 |
| $\log \left(n_{2}^{0}\right)$ | 14.561 | 0.213 |
| $\log \left(n_{3}^{0}\right)$ | 14.383 | 0.210 |
| $q_{p o t}$ | 3.736 | 0.237 |
| $\log \left(\bar{F}^{\text {df }}\right)$ | -2.137 | 0.052 |
| $\log \left(\bar{F}^{\mathrm{tb}}\right)$ | -9.949 | 0.072 |
| $\log \left(\bar{F}^{\mathrm{fb}}\right)$ | -8.073 | 0.072 |
| $\log$ Stage-1 directed pot selectivity 1978-2008 | -1.005 | 0.183 |
| $\log$ Stage-2 directed pot selectivity 1978-2008 | -0.613 | 0.132 |
| $\log$ Stage-1 directed pot selectivity $2009-2017$ | -0.671 | 0.167 |
| $\log$ Stage-2 directed pot selectivity $2009-2017$ | -0.000 | 0.000 |
| $\log$ Stage-1 NMFS trawl selectivity | -0.368 | 0.066 |
| $\log$ Stage-2 NMFS trawl selectivity | -0.000 | 0.000 |
| $\log$ Stage-1 ADF\&G pot selectivity | -0.854 | 0.123 |
| $\log$ Stage-2 ADF\&G pot selectivity | -0.000 | 0.000 |
| $F_{\text {OFL }}$ | 0.080 | 0.010 |
| OFL | 86.177 | 17.958 |

Table 17: Model parameter estimates, selected derived quantities, and their standard errors (SE) for model 24.0 b , with natural mortality estimated using a less restrictive prior.

| Parameter | Estimate | SE |
| :--- | ---: | ---: |
| Natural mortality deviation in 1998/99 $\left(\delta_{1998}^{M}\right)$ | 1.088 | 0.161 |
| $\log (\bar{R})$ | 14.443 | 0.233 |
| $\log \left(n_{1}^{0}\right)$ | 15.364 | 0.198 |
| $\log \left(n_{2}^{0}\right)$ | 14.684 | 0.223 |
| $\log \left(n_{3}^{0}\right)$ | 14.271 | 0.248 |
| $q_{p o t}$ | 3.457 | 0.251 |
| $\log \left(\bar{F}^{\text {df }}\right)$ | -2.128 | 0.058 |
| $\log \left(\bar{F}^{\mathrm{tb}}\right)$ | -10.117 | 0.086 |
| $\log \left(\bar{F}^{\mathrm{fb}}\right)$ | -8.243 | 0.086 |
| $\log$ Stage-1 directed pot selectivity $1978-2008$ | -1.421 | 0.211 |
| $\log$ Stage-2 directed pot selectivity $1978-2008$ | -0.856 | 0.144 |
| $\log$ Stage-1 directed pot selectivity $2009-2017$ | -1.069 | 0.209 |
| $\log$ Stage-2 directed pot selectivity $2009-2017$ | -0.082 | 0.119 |
| $\log$ Stage-1 NMFS trawl selectivity | -0.785 | 0.141 |
| $\log$ Stage-2 NMFS trawl selectivity | -0.233 | 0.095 |
| $\log$ Stage-1 ADF\&G pot selectivity | -1.310 | 0.181 |
| $\log$ Stage-2 ADF\&G pot selectivity | -0.278 | 0.108 |
| $F_{\text {OFL }}$ | 0.117 | 0.021 |
| OFL | 115.085 | 28.957 |

Table 18: Model parameter estimates, selected derived quantities, and their standard errors (SE) for model 24.0 c , with natural mortality fixed at the value estimated by model 24.0 a .

| Parameter | Estimate | SE |
| :--- | ---: | ---: |
| Natural mortality deviation in 1998/99 $\left(\delta_{1998}^{M}\right)$ | 1.413 | 0.143 |
| $\log (\bar{R})$ | 13.962 | 0.191 |
| $\log \left(n_{1}^{0}\right)$ | 15.052 | 0.174 |
| $\log \left(n_{2}^{0}\right)$ | 14.561 | 0.213 |
| $\log \left(n_{3}^{0}\right)$ | 14.383 | 0.210 |
| $q_{p o t}$ | 3.736 | 0.232 |
| $\log \left(\bar{F}^{\text {df }}\right)$ | -2.137 | 0.052 |
| $\log \left(\bar{F}^{\mathrm{tb}}\right)$ | -9.949 | 0.071 |
| $\log \left(\bar{F}^{\mathrm{fb}}\right)$ | -8.073 | 0.071 |
| $\log$ Stage-1 directed pot selectivity $1978-2008$ | -1.005 | 0.181 |
| $\log$ Stage-2 directed pot selectivity 1978-2008 | -0.613 | 0.131 |
| $\log$ Stage-1 directed pot selectivity $2009-2017$ | -0.671 | 0.165 |
| $\log$ Stage-2 directed pot selectivity $2009-2017$ | -0.000 | 0.000 |
| $\log$ Stage-1 NMFS trawl selectivity | -0.368 | 0.063 |
| $\log$ Stage-2 NMFS trawl selectivity | -0.000 | 0.000 |
| $\log$ Stage-1 ADF\&G pot selectivity | -0.854 | 0.120 |
| $\log$ Stage-2 ADF\&G pot selectivity | -0.000 | 0.000 |
| $F_{\text {OFL }}$ | 0.080 | 0.010 |
| OFL | 86.177 | 17.958 |

Table 19: Comparisons of parameter estimates for the model scenarios.

| Parameter | Model 16.0 | Model 16.0a | Model 16.0b | Model 24.0 | Model 24.0a | Model 24.0b | Model 24.0c |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Natural mortality deviation in 1998/99 ( $\delta_{1998}^{M}$ ) | 1.581 | 1.566 | 1.581 | 1.566 | 1.413 | 1.088 | 1.413 |
| $\log (\bar{R}) \quad$ | 13.864 | 13.852 | 13.864 | 13.852 | 13.962 | 14.443 | 13.962 |
| $\log \left(n_{1}^{0}\right)$ | 14.954 | 14.955 | 14.954 | 14.955 | 15.052 | 15.364 | 15.052 |
| $\log \left(n_{2}^{0}\right)$ | 14.528 | 14.527 | 14.528 | 14.527 | 14.561 | 14.684 | 14.561 |
| $\log \left(n_{3}^{0}\right)$ | 14.339 | 14.338 | 14.339 | 14.338 | 14.383 | 14.271 | 14.383 |
| $q_{p o t}$ | 3.827 | 3.905 | 3.827 | 3.905 | 3.736 | 3.457 | 3.736 |
| $\log \left(\bar{F}^{\mathrm{df}}\right)$ | -2.134 | -2.121 | -2.134 | -2.121 | -2.137 | -2.128 | -2.137 |
| $\log \left(\bar{F}^{\mathrm{tb}}\right)$ | -9.927 | -9.903 | -9.927 | -9.903 | -9.949 | -10.117 | -9.949 |
| $\log \left(\bar{F}^{\mathrm{fb}}\right)$ | -8.049 | -8.026 | -8.049 | -8.026 | -8.073 | -8.243 | -8.073 |
| log Stage-1 directed pot selectivity 1978-2008 | -0.925 | -0.914 | -0.925 | -0.914 | -1.005 | -1.421 | -1.005 |
| log Stage-2 directed pot selectivity 1978-2008 | -0.562 | -0.559 | -0.562 | -0.559 | -0.613 | -0.856 | -0.613 |
| log Stage-1 directed pot selectivity 2009-2017 | -0.545 | -0.585 | -0.545 | -0.585 | -0.671 | -1.069 | -0.671 |
| log Stage-2 directed pot selectivity 2009-2017 | -0.000 | -0.000 | -0.000 | -0.000 | -0.000 | -0.082 | -0.000 |
| log Stage-1 NMFS trawl selectivity | -0.294 | -0.297 | -0.294 | -0.297 | -0.368 | -0.785 | -0.368 |
| log Stage-2 NMFS trawl selectivity | -0.000 | -0.000 | -0.000 | -0.000 | -0.000 | -0.233 | -0.000 |
| log Stage-1 ADF\&G pot selectivity | -0.809 | -0.762 | -0.809 | -0.762 | -0.854 | -1.310 | -0.854 |
| log Stage-2 ADF\&G pot selectivity | -0.000 | -0.000 | -0.000 | -0.000 | -0.000 | -0.278 | -0.000 |
| $F_{\text {OFL }}$ | 0.073 | 0.071 | 0.078 | 0.076 | 0.080 | 0.117 | 0.080 |
| OFL | 85.616 | 80.258 | 90.920 | 85.283 | 86.177 | 115.085 | 86.177 |

Table 20: Comparisons of negative log-likelihood values for the selected model scenarios. It is important to note that comparisons among models may be limited since the number of parameters between models changes (e.g., models in which natural mortality is estimated have an additional estimated parameter).

| Component | Model.16.0 | Model.16.0a | Model.16.0b | Model.24.0 | Model.24.0a | Model.24.0b | Model.24.0c |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Pot retained catch | -68.44 | -68.28 | -68.44 | -68.28 | -68.67 | -69.36 | -68.67 |
| Pot discarded catch | 6.69 | 5.96 | 6.69 | 5.96 | 5.31 | 3.26 | 5.31 |
| Trawl bycatch discarded catch | -8.81 | -8.81 | -8.81 | -8.81 | -8.81 | -8.81 | -8.81 |
| Fixed bycatch discarded catch | -8.77 | -8.77 | -8.77 | -8.77 | -8.79 | -8.81 | -8.79 |
| NMFS trawl survey | 5.15 | 8.57 | 5.15 | 8.57 | 5.56 | -3.99 | 5.56 |
| ADF\&G pot survey CPUE | 84.94 | 86.62 | 84.94 | 86.62 | 81.52 | 67.42 | 81.52 |
| Directed pot LF | -104.57 | -104.90 | -104.57 | -104.90 | -104.91 | -104.55 | -104.91 |
| NMFS trawl LF | -276.19 | -277.79 | -276.19 | -277.79 | -279.25 | -277.97 | -279.25 |
| ADF\&G pot LF | -99.87 | -99.98 | -99.87 | -99.98 | -100.10 | -99.42 | -100.10 |
| Recruitment deviations | 63.95 | 64.00 | 63.95 | 64.00 | 63.75 | 62.74 | 63.75 |
| F penalty | 9.66 | 9.66 | 9.66 | 9.66 | 9.66 | 9.66 | 9.66 |
| M penalty | 6.46 | 6.46 | 6.46 | 6.46 | 6.45 | 6.45 | 6.45 |
| Prior | 13.71 | 13.71 | 13.71 | 13.71 | 14.58 | 25.72 | 13.71 |
| Total | -385.77 | -383.21 | -385.77 | -383.21 | -393.36 | -407.32 | -394.22 |
| Total estimated parameters | 156.00 | 156.00 | 156.00 | 156.00 | 157.00 | 157.00 | 156.00 |

Table 21: Population abundances $(\boldsymbol{n})$ by crab stage in numbers of crab at the time of the survey and mature male biomass (MMB) in tons on 15 February for the model configuration used in 2022 with updated data and GMACS version (model 16.0).

| Year | $n_{1}$ | $n_{2}$ | $n_{3}$ | MMB | CV MMB |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 1978 | 3122519 | 2039873 | 1687952 | 4645 | 0.162 |
| 1979 | 4224930 | 2383406 | 2326233 | 6538 | 0.115 |
| 1980 | 3787728 | 3181785 | 3500179 | 10239 | 0.080 |
| 1981 | 1428548 | 3201459 | 4860114 | 10682 | 0.061 |
| 1982 | 1601647 | 1826210 | 4883584 | 7588 | 0.070 |
| 1983 | 806247 | 1443581 | 3461295 | 4535 | 0.095 |
| 1984 | 655815 | 862524 | 1982633 | 3032 | 0.117 |
| 1985 | 912252 | 623891 | 1410998 | 2670 | 0.134 |
| 1986 | 1363829 | 700416 | 1191588 | 2605 | 0.131 |
| 1987 | 1315767 | 988173 | 1280667 | 3080 | 0.121 |
| 1988 | 1225021 | 1055127 | 1486163 | 3359 | 0.118 |
| 1989 | 2904786 | 1024788 | 1636341 | 3838 | 0.114 |
| 1990 | 1862124 | 1960586 | 1936270 | 4974 | 0.088 |
| 1991 | 1909207 | 1670367 | 2421417 | 4999 | 0.089 |
| 1992 | 2084804 | 1582555 | 2383036 | 5170 | 0.082 |
| 1993 | 2353898 | 1666641 | 2491861 | 5420 | 0.074 |
| 1994 | 1591284 | 1837720 | 2569203 | 5193 | 0.068 |
| 1995 | 1811559 | 1456227 | 2467494 | 5065 | 0.069 |
| 1996 | 1730683 | 1467145 | 2367512 | 4829 | 0.070 |
| 1997 | 902822 | 1424153 | 2284582 | 4195 | 0.088 |
| 1998 | 626100 | 933682 | 1860950 | 2765 | 0.102 |
| 1999 | 368573 | 315951 | 715337 | 1693 | 0.096 |
| 2000 | 411463 | 313726 | 791632 | 1833 | 0.080 |
| 2001 | 373439 | 337416 | 858509 | 1985 | 0.073 |
| 2002 | 129410 | 323647 | 922604 | 2089 | 0.068 |
| 2003 | 293687 | 180554 | 945869 | 1971 | 0.069 |
| 2004 | 199731 | 226412 | 908621 | 1954 | 0.069 |
| 2005 | 465178 | 188336 | 892049 | 1880 | 0.070 |
| 2006 | 703175 | 326421 | 884272 | 2019 | 0.071 |
| 2007 | 440990 | 506802 | 969309 | 2351 | 0.070 |
| 2008 | 841244 | 412441 | 1092130 | 2504 | 0.057 |
| 2009 | 691996 | 613585 | 1199282 | 2542 | 0.051 |
| 2010 | 626304 | 586833 | 1274130 | 2157 | 0.052 |
| 2011 | 463313 | 525590 | 1122994 | 1575 | 0.064 |
| 2012 | 229951 | 401427 | 808563 | 1020 | 0.097 |
| 2013 | 258553 | 234504 | 518313 | 1177 | 0.087 |
| 2014 | 207808 | 224408 | 575781 | 1111 | 0.093 |
| 2015 | 170699 | 188298 | 547491 | 1091 | 0.094 |
| 2016 | 179288 | 158039 | 545032 | 1142 | 0.091 |
| 2017 | 135886 | 154011 | 551623 | 1147 | 0.089 |
| 2018 | 155046 | 127846 | 550069 | 1117 | 0.088 |
| 2019 | 249543 | 130286 | 537980 | 1100 | 0.088 |
| 2020 | 178914 | 184762 | 538676 | 1160 | 0.093 |
| 2021 | 389956 | 162722 | 560074 | 1174 | 0.097 |
| 2022 | 465788 | 275293 | 587085 | 1343 | 0.098 |
|  |  |  |  |  |  |

Table 22: Population abundances ( $\boldsymbol{n}$ ) by crab stage in numbers of crab at the time of the survey (1 July, season 1) and mature male biomass (MMB) in tons on 15 February for the recommended new base model (model 24.0).

| Year | $n_{1}$ | $n_{2}$ | $n_{3}$ | MMB | CV MMB |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 1978 | 3124126 | 2036532 | 1686167 | 4192 | 0.161 |
| 1979 | 4227968 | 2382634 | 2323034 | 5895 | 0.115 |
| 1980 | 3790548 | 3183189 | 3497385 | 9254 | 0.080 |
| 1981 | 1428477 | 3203500 | 4858747 | 9622 | 0.061 |
| 1982 | 1603649 | 1826561 | 4883261 | 6860 | 0.069 |
| 1983 | 806064 | 1444294 | 3461154 | 4137 | 0.095 |
| 1984 | 656259 | 862205 | 1982578 | 2756 | 0.116 |
| 1985 | 914060 | 623809 | 1410735 | 2444 | 0.134 |
| 1986 | 1366195 | 701192 | 1191427 | 2386 | 0.131 |
| 1987 | 1319765 | 989590 | 1281089 | 2822 | 0.121 |
| 1988 | 1229940 | 1057680 | 1487540 | 3080 | 0.118 |
| 1989 | 2918312 | 1028229 | 1639166 | 3522 | 0.114 |
| 1990 | 1866004 | 1969084 | 1941566 | 4571 | 0.088 |
| 1991 | 1913030 | 1675127 | 2430365 | 4631 | 0.089 |
| 1992 | 2090633 | 1586017 | 2393206 | 4755 | 0.081 |
| 1993 | 2351701 | 1670881 | 2502634 | 5022 | 0.074 |
| 1994 | 1547965 | 1837723 | 2580128 | 4808 | 0.067 |
| 1995 | 1719558 | 1432318 | 2472775 | 4656 | 0.068 |
| 1996 | 1709520 | 1408385 | 2351649 | 4366 | 0.069 |
| 1997 | 894757 | 1393074 | 2240522 | 3754 | 0.086 |
| 1998 | 626571 | 918058 | 1807507 | 1810 | 0.092 |
| 1999 | 360066 | 317058 | 698546 | 1533 | 0.096 |
| 2000 | 400957 | 309260 | 777352 | 1661 | 0.079 |
| 2001 | 379107 | 329973 | 843325 | 1797 | 0.072 |
| 2002 | 129397 | 324400 | 906710 | 1900 | 0.066 |
| 2003 | 287625 | 180794 | 932972 | 1795 | 0.067 |
| 2004 | 227838 | 223050 | 897389 | 1779 | 0.067 |
| 2005 | 465791 | 203183 | 883660 | 1734 | 0.068 |
| 2006 | 678896 | 331681 | 884811 | 1868 | 0.067 |
| 2007 | 383831 | 494766 | 970086 | 2157 | 0.063 |
| 2008 | 845711 | 376394 | 1081102 | 2253 | 0.055 |
| 2009 | 687266 | 604188 | 1172326 | 2401 | 0.050 |
| 2010 | 607788 | 581101 | 1246391 | 2033 | 0.051 |
| 2011 | 442774 | 513496 | 1094538 | 1461 | 0.063 |
| 2012 | 230499 | 386310 | 775556 | 914 | 0.097 |
| 2013 | 227484 | 229126 | 481871 | 1069 | 0.087 |
| 2014 | 207536 | 204984 | 539659 | 992 | 0.093 |
| 2015 | 177517 | 181700 | 507483 | 981 | 0.094 |
| 2016 | 177914 | 159654 | 508913 | 1043 | 0.091 |
| 2017 | 131955 | 153760 | 522141 | 1057 | 0.088 |
| 2018 | 150318 | 125521 | 524943 | 1034 | 0.087 |
| 2019 | 243914 | 126831 | 515375 | 1020 | 0.087 |
| 2020 | 176399 | 180421 | 517515 | 1080 | 0.092 |
| 2021 | 379523 | 159856 | 539970 | 1097 | 0.096 |
| 2022 | 440858 | 268420 | 567850 | 1257 | 0.098 |
|  |  |  |  |  |  |

## Figures



Figure 1: Comparisons of estimated mature male biomass (MMB) time series on 15 February during 19782022 for models 16.0, 16.0a, 16.0b, and 24.0.


Figure 2: Comparisons of estimated mature male biomass (MMB) time series on 15 February during 19782022 for models 24.0, 24.0a, 24.0b, and 24.0c.


Figure 3: ADFG pot survey abundances by carapace length for male St. Matthew Island blue king crab from 1995 to 2022.


Figure 4: NMFS trawl survey abundances by carapace length for male St. Matthew Island blue king crab from 1978 to 2023.


Figure 5: Distribution of blue king crab (Paralithodes platypus) in the Gulf of Alaska, Bering Sea, and Aleutian Islands waters (shown in blue).


Figure 6: Blue king crab Registration Area Q (Bering Sea)


Figure 7: Data extent for the SMBKC assessment.


Figure 8: Trawl and pot-survey stations used in the SMBKC stock assessment.


Figure 9: Catches (in numbers) of male blue king crab > 90mm CL from the 2013-2022 NMFS trawl-survey at the 56 stations used to assess the SMBKC stock.


Figure 10: Fits to NMFS area-swept trawl estimates of total ( $>90 \mathrm{~mm}$ ) male survey biomass for models 16.0 - 2022 (without new data added) and 16.0 (with new data added). Error bars are plus and minus 2 standard deviations.


Figure 11: Comparisons of fits to CPUE from the ADFG pot surveys for models 16.0-2022 (without new data added) and 16.0 (with new data added). Error bars are plus and minus 2 standard deviations.


Figure 12: Estimated mature male biomass (MMB) over 1978-2022 from models 16.0 - 2022 and 16.0; model 16.0-2022 is the model accepted in 2022 using the data available in September 2022, while model 16.0 is the same model updated with the data available in April 2024.


Figure 13: Estimated recruitment over 1979-2022 from models $16.0-2022$ and 16.0; model 16.0-2022 is the model accepted in 2022 using the data available in September 2022, while model 16.0 is the same model updated with the data available in April 2024. The dashed horizontal lines represent the estimate of the average recruitment parameter $(\bar{R})$ in each model scenario.


Figure 14: Comparisons of the estimated stage-1 and stage-2 selectivities for the different model scenarios (the stage-3 selectivities are all fixed at 1). Estimated selectivities are shown for the directed pot fishery, the trawl bycatch fishery, the fixed bycatch fishery, the NMFS trawl survey, and the ADFG pot survey. Two selectivity periods are estimated in the directed pot fishery, from 1978-2008 and 2009-2022. Solid lines are capture selectivities while dashed lines are retained selectivities.


Figure 15: Estimated recruitment 1979-2022 comparing models 16.0, 16.0a, 16.0b, and 16.0c. The dashed horizontal lines represent the estimate of the average recruitment parameter $(\bar{R})$ in each model scenario.


Figure 16: Estimated recruitment 1979-2022 comparing models 24.0, 24.0a, 24.0b, and 24.0c. The dashed horizontal lines represent the estimate of the average recruitment parameter $(\bar{R})$ in each model scenario.


Figure 17: Comparisons of area-swept estimates of total ( $90+\mathrm{mm}$ CL) male survey biomass (tons) and model predictions for models $16.0,16.0 \mathrm{a}, 16.0 \mathrm{~b}$, and 24.0 . The error bars are plus and minus 2 standard deviations.


Figure 18: Comparisons of area-swept estimates of total ( $90+\mathrm{mm} \mathrm{CL}$ ) male survey biomass (tons) and model predictions for models 24.0, 24.0a, 24.0b, and 24.0c. The error bars are plus and minus 2 standard deviations.


Figure 19: Time-varying natural mortality $\left(M_{t}\right)$. Estimated pulse period occurs in 1998/99 (i.e. $M_{1998}$ ).




Figure 20: Fishing mortality estimates from the new base model (24.0) for directed and bycatch fleets


Figure 21: Comparisons of total ( $90+\mathrm{mm}$ CL) male pot survey CPUEs and model predictions for models 16.0, 16.0a, 16.0b, and 24.0. The error bars are plus and minus 2 standard deviations.


Figure 22: Comparisons of total ( $90+\mathrm{mm}$ CL) male pot survey CPUEs and model predictions for models 24.0 , 24.0a, 24.0b, and 24.0 c . The error bars are plus and minus 2 standard deviations.


Figure 23: Observed and model estimated size frequencies of SMBKC by year retained in the directed pot fishery for the model scenarios.


Figure 24: Observed and model estimated size frequencies of discarded male SMBKC by year in the NMFS trawl survey for the model scenarios.


Figure 25: Observed and model estimated size frequencies of discarded SMBKC by year in the ADFG pot survey for the model scenarios.


Figure 26: Line plots of residuals by size and year for the directed pot fishery size composition data set for all model scenarios.


Figure 27: Line plots of residuals by size and year for the NMFS trawl survey size composition data set for all model scenarios.


Figure 28: Line plots of residuals by size and year for the ADFG pot survey size composition data set for all models scenarios.


Figure 29: Comparison of observed and model predicted retained catch and bycatches in each of the models. Note the difference in units among the panels: some panels are expressed in numbers of crab, some in biomass (tons).


Figure 30: Retrospective pattern in mature male biomass (MMB (t)) for model 24.0 using 10 peels.


Figure 31: Retrospective pattern in mature male biomass (MMB ( t ) ) for model 24.0 using 5 peels.


Figure 32: Catch likelihood components for the values of natural mortality considered in the likelihood profile.


Figure 33: Size composition likelihood components for the values of natural mortality considered in the likelihood profile.


Figure 34: Survey index likelihood components for the values of natural mortality considered in the likelihood profile.


Figure 35: Total likelihood for the values of natural mortality considered in the likelihood profile.


Figure 36: Mature male biomass (MMB) trajectories for the natural mortality likelihood profile models.


Figure 37: Comparison of spawning stock biomass (SSB) to the dynamic $B_{0}$ value ( 15 February, 1978-2022) for model 16.0 (2022).

## Appendix A: SMBKC Model Description

## 1. Introduction

The GMACS model has been specified to account only for male crab $\geq 90 \mathrm{~mm}$ in carapace length (CL). These are partitioned into three stages (size-classes) determined by CL measurements of (1) 90-104 mm, (2) 105-119 mm, and (3) $120+\mathrm{mm}$. For management of the St. Matthew Island blue king crab (SMBKC) fishery, 120 mm CL is used as the proxy value for the legal measurement of 5.5 inch carapace width (CW), whereas 105 mm CL is the management proxy for mature-male size (state regulation 5 AAC 34.917 (d)). Accordingly, within the model only stage-3 crab are retained in the directed fishery, and stage- 2 and stage- 3 crab together comprise the collection of mature males. Some justification for the 105 mm value is presented in Pengilly and Schmidt (1995), who used it in developing the current regulatory SMBKC harvest strategy. The term "recruit" here designates recruits to the model, i.e., annual new stage- 1 crab, rather than recruits to the fishery. The following description of model structure reflects the GMACS base model configuration.

## 2. Model Population Dynamics

Within the model, the beginning of the crab year is assumed contemporaneous with the NMFS trawl survey, nominally assigned a date of 1 July. Although the timing of the fishery is different each year, MMB is estimated at 15 February, which is the reference date for calculation of federal management biomass quantities. To accommodate this, each model year is split into 5 seasons $(t)$ and a proportion of the natural mortality $\left(\tau_{t}\right)$, scaled relative to the portions of the year, is applied in each of these seasons where $\sum_{t=1}^{t=5} \tau_{t}=1$. Each model year consists of the following processes with time-breaks denoted here by "Seasons." However, it is important to note that actual seasons are survey-to-fishery, fishery-to Feb 15, and Feb 15 to July 1. The following breakdown accounts for events and fishing mortality treatments:

1. Season 1 (survey period)

- Beginning of the SMBKC fishing year (1 July)
- $\tau_{1}=0$
- Surveys

2. Season 2 (natural mortality until pulse fishery)

- $\tau_{2}$ ranges from 0.05 to 0.44 depending on the time of year the fishery begins each year (i.e., a higher value indicates the fishery begins later in the year).

3. Season 3 (pulse fishery)

- $\tau_{3}=0$
- fishing mortality applied

4. Season 4 (natural mortality until spawning)

- $\tau_{4}=0.63-\sum_{i=1}^{i=4} \tau_{i}$
- Calculate MMB (15 February). Note: in model 16.0 b and all 24.0 series models, MMB is calculated in Season 5.

5. Season 5 (natural mortality and somatic growth through to June 30th)

- $\tau_{5}=0.37$
- Growth and molting
- Recruitment (all to stage-1)

The proportion of natural mortality $\left(\tau_{t}\right)$ applied during each season in the model is provided in Table 23, see Table 6 for season 2 interaction with directed fishery timing. The beginning of the year ( 1 July) to the date that MMB is measured ( 15 February) is $63 \%$ of the year. Therefore $63 \%$ of the natural mortality must be applied before the MMB is calculated. Because the timing of the fishery is different each year, $\tau_{2}$ varies and thus $\tau_{4}$ varies also.

With boldface lower-case letters indicating vector quantities we designate the vector of stage abundances during season $t$ and year $y$ as

$$
\begin{equation*}
\boldsymbol{n}_{t, y}=n_{l, t, y}=\left[n_{1, t, y}, n_{2, t, y}, n_{3, t, y}\right]^{\top} \tag{2}
\end{equation*}
$$

The number of new crab, or recruits, of each stage entering the model each season $t$ and year $y$ is represented as the vector $\boldsymbol{r}_{t, y}$. The SMBKC formulation of GMACS specifies recruitment to stage-1 only during season $t=5$, thus the recruitment size distribution is

$$
\begin{equation*}
\phi_{l}=[1,0,0]^{\top}, \tag{3}
\end{equation*}
$$

and the recruitment is

$$
\boldsymbol{r}_{t, y}= \begin{cases}0 & \text { for } \quad t<5  \tag{4}\\ \bar{R} \phi_{l} \delta_{y}^{R} & \text { for } \quad t=5\end{cases}
$$

where $\bar{R}$ is the average annual recruitment and $\delta_{y}^{R}$ are the recruitment deviations each year $y$

$$
\begin{equation*}
\delta_{y}^{R} \sim \mathcal{N}\left(0, \sigma_{R}^{2}\right) \tag{5}
\end{equation*}
$$

Using boldface upper-case letters to indicate a matrix, we describe the size transition matrix $\boldsymbol{G}$ as

$$
\boldsymbol{G}=\left[\begin{array}{ccc}
1-\pi_{12}-\pi_{13} & \pi_{12} & \pi_{13}  \tag{6}\\
0 & 1-\pi_{23} & \pi_{23} \\
0 & 0 & 1
\end{array}\right]
$$

with $\pi_{j k}$ equal to the proportion of stage- $j$ crab that molt and grow into stage- $k$ within a season or year.
The natural mortality each season $t$ and year $y$ is

$$
\begin{equation*}
M_{t, y}=\bar{M} \tau_{t}+\delta_{y}^{M} \text { where } \delta_{y}^{M} \sim \mathcal{N}\left(0, \sigma_{M}^{2}\right) \tag{7}
\end{equation*}
$$

Fishing mortality by year $y$ and season $t$ is denoted $F_{t, y}$ and calculated as

$$
\begin{equation*}
F_{t, y}=F_{t, y}^{\mathrm{df}}+F_{t, y}^{\mathrm{tb}}+F_{t, y}^{\mathrm{fb}} \tag{8}
\end{equation*}
$$

where $F_{t, y}^{\mathrm{df}}$ is the fishing mortality associated with the directed fishery, $F_{t, y}^{\mathrm{tb}}$ is the fishing mortality associated with the trawl bycatch fishery, $F_{t, y}^{\mathrm{fb}}$ is the fishing mortality associated with the fixed bycatch fishery. Each of these are derived as

$$
\begin{array}{lll}
F_{t, y}^{\mathrm{df}}=\bar{F}^{\mathrm{df}}+\delta_{t, y}^{\mathrm{df}} \quad \text { where } & \delta_{t, y}^{\mathrm{df}} \sim \mathcal{N}\left(0, \sigma_{\mathrm{df}}^{2}\right), \\
F_{t, y}^{\mathrm{tb}}=\bar{F}^{\mathrm{tb}}+\delta_{t, y}^{\mathrm{tb}} & \text { where } & \delta_{t, y}^{\mathrm{df}} \sim \mathcal{N}\left(0, \sigma_{\mathrm{tb}}^{2}\right), \\
F_{t, y}^{\mathrm{fb}}=\bar{F}^{\mathrm{fb}}+\delta_{t, y}^{\mathrm{fb}} & \text { where } & \delta_{t, y}^{\mathrm{df}} \sim \mathcal{N}\left(0, \sigma_{\mathrm{fb}}^{2}\right), \tag{9}
\end{array}
$$

where $\delta_{t, y}^{\mathrm{df}}, \delta_{t, y}^{\mathrm{tb}}$, and $\delta_{t, y}^{\mathrm{fb}}$ are the fishing mortality deviations for each of the fisheries, each season $t$ during each year $y, \bar{F}^{\text {df }}, \bar{F}^{\mathrm{tb}}$, and $\bar{F}^{\mathrm{fb}}$ are the average fishing mortalities for each fishery. The total mortality $Z_{l, t, y}$ represents the combination of natural mortality $M_{t, y}$ and fishing mortality $F_{t, y}$ during season $t$ and year $y$

$$
\begin{equation*}
Z_{t, y}=Z_{l, t, y}=M_{t, y}+F_{t, y} \tag{10}
\end{equation*}
$$

The survival matrix $\boldsymbol{S}_{t, y}$ during season $t$ and year $y$ is

$$
\boldsymbol{S}_{t, y}=\left[\begin{array}{ccc}
1-e^{-Z_{1, t, y}} & 0 & 0  \tag{11}\\
0 & 1-e^{-Z_{2, t, y}} & 0 \\
0 & 0 & 1-e^{-Z_{3, t, y}}
\end{array}\right]
$$

The basic population dynamics underlying GMACS can thus be described as

$$
\begin{array}{lc}
\boldsymbol{n}_{t+1, y}=\boldsymbol{S}_{t, y} \boldsymbol{n}_{t, y}, & \text { if } t<5 \\
\boldsymbol{n}_{t, y+1}=\boldsymbol{G} \boldsymbol{S}_{t, y} \boldsymbol{n}_{t, y}+\boldsymbol{r}_{t, y} & \text { if } t=5 .
\end{array}
$$

## 3. Model Data

Data inputs used in model estimation are listed in Table 24 . The mean weight ( kg ) by stage, provided as a vector of weights at length each year to GMACS, is the same for all years and all models: 0.7 kg for Stage-1, 1.2 kg for Stage-2, and 1.9 kg for Stage-3.

## 4. Model Parameters

Table 25 lists fixed (externally determined) parameters used in model computations. In all scenarios, the stage-transition matrix is

$$
\boldsymbol{G}=\left[\begin{array}{ccc}
0.2 & 0.7 & 0.1  \tag{13}\\
0 & 0.2 & 0.8 \\
0 & 0 & 1
\end{array}\right]
$$

which is the combination of the growth matrix and molting probabilities.
Estimated parameters are listed in Table 26 and include an estimated natural mortality deviation parameter in 1998/99 ( $\delta_{1998}^{M}$ ) assuming an anomalous mortality event in that year, as hypothesized by Zheng and Kruse (2002), with natural mortality otherwise fixed at $0.18 \mathrm{yr}^{-1}$.

## 5. Model Objective Function and Weighting Scheme

The objective function consists of the sum of several negative log-likelihood terms characterizing the hypothesized error structure of the principal data inputs (Table 20). A log-normal distribution is assumed to characterize the catch data and is modeled as

$$
\begin{align*}
\sigma_{t, y}^{\text {catch }} & =\sqrt{\log \left(1+\left(C V_{t, y}^{\text {catch }}\right)^{2}\right)}  \tag{14}\\
\delta_{t, y}^{\text {catch }} & =\mathcal{N}\left(0,\left(\sigma_{t, y}^{\text {catch }}\right)^{2}\right) \tag{15}
\end{align*}
$$

where $\delta_{t, y}^{\text {catch }}$ is the residual catch. The relative abundance data is also assumed to be log-normally distributed

$$
\begin{align*}
\sigma_{t, y}^{\mathrm{I}} & =\frac{1}{\lambda} \sqrt{\log \left(1+\left(C V_{t, y}^{\mathrm{I}}\right)^{2}\right)}  \tag{16}\\
\delta_{t, y}^{\mathrm{I}} & =\log \left(I^{\mathrm{obs}} / I^{\mathrm{pred}}\right) / \sigma_{t, y}^{\mathrm{I}}+0.5 \sigma_{t, y}^{\mathrm{I}} \tag{17}
\end{align*}
$$

and the likelihood is

$$
\begin{equation*}
\sum \log \left(\delta_{t, y}^{\mathrm{I}}\right)+\sum 0.5\left(\sigma_{t, y}^{\mathrm{I}}\right)^{2} \tag{18}
\end{equation*}
$$

GMACS calculates standard deviation of the normalised residual (SDNR) values and median of the absolute residual (MAR) values for all abundance indices and size compositions to help the user come up with reasonable likelihood weights. For an abundance data set to be well fitted, the SDNR should not be much greater than 1 (a value much less than 1 , which means that the data set is fitted better than was expected, is not a cause for concern). What is meant by "much greater than 1 " depends on $m$ (the number of years in the data set). Francis (2011) suggests upper limits of $1.54,1.37$, and 1.26 for $m=5,10$, and 20, respectively. Although an SDNR not much greater than 1 is a necessary condition for a good fit, it is not sufficient. It is important to plot the observed and expected abundances to ensure that the fit is good.
GMACS also calculates Francis weights for each of the size composition data sets supplied (Francis 2011). If the user wishes to use the Francis iterative re-weighting method, first the weights applied to the abundance indices should be adjusted by trial and error until the SDNR (and/or MAR) are adequate. Then the Francis weights supplied by GMACS should be used as the new likelihood weights for each of the size composition data sets the next time the model is run. The user can then iteratively adjust the abundance index and size composition weights until adequate SDNR (and/or MAR) values are achieved, given the Francis weights.

## 6. Estimation

The model was implemented using the software AD Model Builder (Fournier et al. 2012), with parameter estimation by minimization of the model objective function using automatic differentiation. Parameter estimates and standard deviations provided in this document are AD Model Builder reported values assuming maximum likelihood theory asymptotics.

Table 23: Proportion of the natural mortality $\left(\tau_{t}\right)$ that is applied during each season $(t)$ in the model.

| Year | Season 1 | Season 2 | Season 3 | Season 4 | Season 5 |
| :--- | ---: | ---: | ---: | ---: | ---: |
| 1978 | 0.00 | 0.07 | 0.00 | 0.56 | 0.37 |
| 1979 | 0.00 | 0.06 | 0.00 | 0.57 | 0.37 |
| 1980 | 0.00 | 0.07 | 0.00 | 0.56 | 0.37 |
| 1981 | 0.00 | 0.05 | 0.00 | 0.58 | 0.37 |
| 1982 | 0.00 | 0.07 | 0.00 | 0.56 | 0.37 |
| 1983 | 0.00 | 0.12 | 0.00 | 0.51 | 0.37 |
| 1984 | 0.00 | 0.10 | 0.00 | 0.53 | 0.37 |
| 1985 | 0.00 | 0.14 | 0.00 | 0.49 | 0.37 |
| 1986 | 0.00 | 0.14 | 0.00 | 0.49 | 0.37 |
| 1987 | 0.00 | 0.14 | 0.00 | 0.49 | 0.37 |
| 1988 | 0.00 | 0.14 | 0.00 | 0.49 | 0.37 |
| 1989 | 0.00 | 0.14 | 0.00 | 0.49 | 0.37 |
| 1990 | 0.00 | 0.14 | 0.00 | 0.49 | 0.37 |
| 1991 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 1992 | 0.00 | 0.14 | 0.00 | 0.49 | 0.37 |
| 1993 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 1994 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 1995 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 1996 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 1997 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 1998 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 1999 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 2000 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 2001 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 2002 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 2003 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 2004 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 2005 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 2006 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 2007 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 2008 | 0.00 | 0.18 | 0.00 | 0.45 | 0.37 |
| 2009 | 0.00 | 0.44 | 0.00 | 0.19 | 0.37 |
| 2010 | 0.00 | 0.44 | 0.00 | 0.19 | 0.37 |
| 2011 | 0.00 | 0.44 | 0.00 | 0.19 | 0.37 |
| 2012 | 0.00 | 0.44 | 0.00 | 0.19 | 0.37 |
| 2013 | 0.00 | 0.44 | 0.00 | 0.19 | 0.37 |
| 2014 | 0.00 | 0.44 | 0.00 | 0.19 | 0.37 |
| 2015 | 0.00 | 0.44 | 0.00 | 0.19 | 0.37 |
| 2016 | 0.00 | 0.44 | 0.00 | 0.19 | 0.37 |
| 2017 | 0.00 | 0.44 | 0.00 | 0.19 | 0.37 |
| 2018 | 0.00 | 0.44 | 0.00 | 0.19 | 0.37 |
| 2019 | 0.00 | 0.44 | 0.00 | 0.19 | 0.37 |
| 2020 | 0.00 | 0.44 | 0.00 | 0.19 | 0.37 |
| 2022 | 0.00 | 0.44 | 0.00 | 0.19 | 0.37 |
|  | 0.00 | 0.44 | 0.00 | 0.19 | 0.37 |
|  |  |  |  |  |  |
| 18 | 0 |  |  |  | 0 |

Table 24: Data inputs used in model estimation.

| Data | Years | Source |
| :--- | :--- | :--- |
| Directed pot-fishery retained-catch number <br> (not biomass) | $1978 / 79-1998 / 99$ <br> $2009 / 10-2015 / 16$ | Fish tickets <br> (fishery closed 1999/00-2008/09 <br> and $2016 / 17-2018 / 19)$ |
| Groundfish trawl bycatch biomass | $1992 / 93-2022 / 23$ | NMFS groundfish observer program |
| Groundfish fixed-gear bycatch biomass | $1992 / 93-2022 / 23$ | NMFS groundfish observer program |
| NMFS trawl-survey biomass index <br> (area-swept estimate) and CV | $1978-2023$ | NMFS EBS trawl survey |
| ADF\&G pot-survey abundance index <br> (CPUE) and CV | $1995-2022$ | ADF\&G SMBKC pot survey |

Table 25: Fixed model parameters for models 16.0, 16.0a, 16.0b, and 24.0. Models 24.0a and 24.0b estimate $M$, while model 24.0 c uses a fixed $M=0.20$.

| Parameter | Symbol | Value | Source/rationale |
| :---: | :---: | :---: | :---: |
| Trawl-survey catchability | $q$ | 1.0 | Default |
| Natural mortality | M | $0.18 \mathrm{yr}^{-1}$ | NPFMC (2007) |
| Size transition matrix | G | Equation 13 | Otto and Cummiskey (1990) |
| Stage-1 and stage-2 mean weights | $w_{1}, w_{2}$ | $0.7,1.2 \mathrm{~kg}$ | Length-weight equation (B. Foy, NMFS) applied to stage midpoints |
| Stage-3 mean weight | $w_{3, y}$ | Depends on year | Fishery reported average retained weight from fish tickets, or its average, and mean weights of legal males |
| Recruitment SD | $\sigma_{R}$ | 1.2 | High value |
| Natural mortality SD | $\sigma_{M}$ | 10.0 | High value (basically free parameter) |
| Directed fishery |  | 0.2 | 2010 Crab SAFE |
| handling mortality Groundfish trawl handling mortality |  | 0.8 | 2010 Crab SAFE |
| Groundfish fixed-gear handling mortality |  | 0.5 | 2010 Crab SAFE |

Table 26: The lower bound (LB), upper bound (UB), initial value, prior, and estimation phase for each estimated model parameter.

| Parameter | LB | Initial value | UB | Prior | Phase |
| :--- | ---: | ---: | ---: | :--- | ---: |
| Average recruitment $\log (\bar{R})$ | -7 | 10.0 | 20 | Uniform $(-7,20)$ | 1 |
| Stage-1 initial numbers $\log \left(n_{1}^{0}\right)$ | 5 | 14.5 | 20 | Uniform $(5,20)$ | 1 |
| Stage-2 initial numbers $\log \left(n_{2}^{0}\right)$ | 5 | 14.0 | 20 | Uniform $(5,20)$ | 1 |
| Stage-3 initial numbers $\log \left(n_{3}^{0}\right)$ | 5 | 13.5 | 20 | Uniform $(5,20)$ | 1 |
| ADF\&G pot survey catchability $q$ | 0 | 3.0 | 5 | Uniform $(0,5)$ | 1 |
| Stage-1 directed fishery selectivity 1978-2008 | 0 | 0.4 | 1 | Uniform $(0,1)$ | 3 |
| Stage-2 directed fishery selectivity 1978-2008 | 0 | 0.7 | 1 | Uniform $(0,1)$ | 3 |
| Stage-1 directed fishery selectivity 2009-2017 | 0 | 0.4 | 1 | Uniform $(0,1)$ | 3 |
| Stage-2 directed fishery selectivity 2009-2017 | 0 | 0.7 | 1 | Uniform $(0,1)$ | 3 |
| Stage-1 NMFS trawl survey selectivity | 0 | 0.4 | 1 | Uniform $(0,1)$ | 4 |
| Stage-2 NMFS trawl survey selectivity | 0 | 0.7 | 1 | Uniform $(0,1)$ | 4 |
| Stage-1 ADF\&G pot survey selectivity | 0 | 0.4 | 1 | Uniform $(0,1)$ | 4 |
| Stage-2 ADF\&G pot survey selectivity | 0 | 0.7 | 1 | Uniform $(0,1)$ | 4 |
| Natural mortality deviation during 1998 $\delta_{1998}^{M}$ | -3 | 0.0 | 3 | Normal $\left(0, \sigma_{M}^{2}\right)$ | 4 |
| Recruitment deviations $\delta_{y}^{R}$ | -7 | 0.0 | 7 | Normal $\left(0, \sigma_{R}^{2}\right)$ | 3 |
| Average directed fishery fishing mortality $\bar{F}^{\text {df }}$ | - | 0.2 | - | - | 1 |
| Average trawl bycatch fishing mortality $\bar{F}^{\mathrm{tb}}$ | - | 0.001 | - | - | 1 |
| Average fixed gear bycatch fishing mortality $\bar{F}^{\mathrm{fb}}$ | - | 0.001 | - | - | 1 |

## Appendix B. Data files for the new base model (24.0)

## The new base model (24.0) data file for 2024







```
2009 14 1 0 0 0 50 0.3057 0.4202 0.2741
2010}1441000050 0.4081 0.3371 0.2548
2011 14 1 0 0 0 50 0.2179 0.3940}00.388
2012 14 1 0 0 0 50 0.1573 0.4393 0.4034
2013 14 1 0 0 0 37 0.2100 0.2834 0.5065
2014 141 0 0 0 50 0.1738 0.3912 0.4350
2015 14 1 0 0 0 50 0.2340 0.2994 0.4666
2016 14 1 0 0 0 50 0.2255 0.2780}00.4965
2017 14 1 0 0 0 21 0.0849 0.2994 0.6157
2018 1441000 31 0.1475 0.2219 0.6306 #55
2019 141000 50 0.1961 0.2346 0.5692 #105 no survey so not updated
2021 14 1 0 0 0 50 0.3323 0.1320 0.5357 #59 updated 4-14-22
2022 144100050 0.3531}00.2121 0.4348 #75 updated 8-25-22
2023 14 1 0 0 0 50 0.3211 0.2974 0.3815 # updated 2-14-24
##length proportions of pot survey
##Year, Seas, Fleet, Sex, Type, Shell, Maturity, Nsamp, DataVec
1995 1 5 1 0 0 0 100 0.151581 0.257254 0.591165
1998 1 5 1 0 0 0 100 0.086263 0.223461 0.690276
2001 1 5 1 0 0 0 100 0.158784 0.202492 0.638725
2004 15 1 1 0 0 0 100 0.092476 0.240596 0.666928
2007 15 1 0 0 0 100 0.116245 0.301801 0.581954
2010}115100000100 0.12951 0.316528 0.553962
2013 15 1 0 0 0 100 0.137872 0.275316 0.586812
2015 1 5 1 0 0 0 100 0.090959 0.271233 0.637808
2016 1 5 1 0 0 0 100 0.097938
2017 15 1 0 0 0 100 0.110656 0.262295 0.627049
2018 15 1 0 0 0 100 0.109272 0.215232 0.675497 # no survey so not updated
2022 1 5 1 0 0 0 100 0.15414 0.250955 0.594904 # updated 2-14-2024
## Growth data (increment)
# Type of growth increment (0=ignore;1=growth increment with a CV;2=size-at-release; size-at)
O
# nobs_growth
O
#3
# MidPoint Sex Increment CV
# 97.5 1 1 14.1 0.2197
#112.5 1 14.1 0.2197
#127.5 1 14.1 0.2197
# 97.5 1 13.8 0.2197
# 112.5 1 14.1 0.2197
# 127.5 1 14.4 0.2197
# MidPoint Sex MidPoint Time-at-liberty Size-trans matrix Number of points
# Release Recapture
# Environmental data
# Number of series
O
# Year ranges
# Indices
# Index Year Value
## eof
9999
```


## The new base model (24.0) control file for 2024

```
\#\# May 2024 smbkc base model 24.0 version: combining updated data (16.0a) and SSB phase 5 (16.0b)
\#\# ===========================================1 updated for May 2024 base model \#\#
\#\# LEADING PARAMETER CONTROLS
\# Controls for leading parameter vector theta
\# LEGEND FOR PRIOR:
\# \(0=\) uniform, 1 = normal, 2 = lognormal, 3 = beta, 4 = gamma
```



```
\# ntheta
    12
\#\# =================================================================================142
\# ival lb ub phz prior p1 p2 \# parameter \#
```



```
##
# Use custom transition matrix ( 0=no, 1=growth matrix, 2=transition matrix, i.e. growth and molting)
```



```
# option 8 is normal distributed growth incerment, size after incrment is normal
1
# growth increment model (0=prespecified;1=alpha/beta; 2=estimated by size-class;3=pre-specified/emprical)
O
# molt probability function (0=pre-specified; 1=flat;2=declining logistic)
2
# Maximum size-class for recruitment(males then females)
1
## number of size-increment periods
1
## Year(s) molt period changes (blank if no change)
## Two lines for each parameter if split sex, one line if not ##
## number of molt periods
1
## Year(s) molt period changes (blank if no changes)
## Beta parameters are relative (1=Yes;0=no)
1
# AEP Growth parameters
\begin{tabular}{|c|c|c|c|c|c|c|c|}
\hline \# ival & lb & ub & phz prior & & p1 & p2 & \# parameter \# \\
\hline \# 14.1 & 10.0 & 30.0 & -3 & 0 & 0.0 & 999.0 & \# alpha males or combined \\
\hline \# 0.0001 & 0.0 & 0.01 & -3 & 0 & 0.0 & 999.0 & \# beta males or combined \\
\hline \# 0.45 & 0.01 & 1.0 & -3 & 0 & 0.0 & 999.0 & \# gscale males or combined \\
\hline 121.5 & 65.0 & 145.0 & -4 & 0 & 0.0 & 999.0 & \# molt_mu males or combined \\
\hline 0.060 & 0.0 & 1.0 & -3 & 0 & 0.0 & 999.0 & \# molt_cv males or combined \\
\hline
\end{tabular}
# The custom growth matrix (if not using just fill with zeros)
# Alternative TM (loosely) based on Otto and Cummiskey (1990)
# Size1....Size2....Size3
    0.1761 0.0000 0.0000
    0.7052 0.2206 0.0000
    0.1187 0.7794 1.0000
# 0.1761 0.7052 0.1187
# 0.0000 0.2206 0.7794
# 0.0000 0.0000 1.0000
\# custom molt probability matrix
```





```
## Year position of the knots (vector must be equal to the number of nodes)
19981999
## Number of Breakpoints in M by size
0
## Size-class of breakpoint
#3
## Specific initial values for the natural mortality devs (0-no, 1=yes)
1
```



```
# maturity specific natural mortality? (yes = 1; no = 0; only for use if nmature > 1) # NEW april 22?
0
```



```
## ============================================================================================= ##
## OTHER CONTROLS
## =============================================================================================== ##
1978 # First rec_dev 
    0 # Terminal molting ( }0=0\mathrm{ off, 1 = on). If on, the calc_stock_recruitment_relationship() isn't called in the procedure
    # Terminal molting (0 = off, 1 = on). If on, the calc_stock_recruitment_relationship() isn't called in the procedure
    # Estimated rec_dev phase
    # Estimated sex_ratio
    # initial sex-ratio
    # Estimated rec_ini phase
    # Initial conditions (0 = Unfished, 1 = Steady-state fished, 2 = Free parameters)
    # Lambda (proportion of mature male biomass for SPR reference points)
    # Stock-Recruit-Relationship (0 = None, 1 = Beverton-Holt)
    # Use years specified to computed average sex ratio in the calculation of average recruitment for reference points (0 = off -i.e. Rec
    200 # Years to compute equilibria
## =============================================================================================== ##
## EMPHASIS FACTORS (CATCH)
## =========================================================================================== ##
#Ret_POT Disc_POT Disc_trawl Disc_fixed
    1 1 1 1
## EMPHASIS FACTORS (Priors) by fleet: Fdev_total, Fdov_total, Fdev_year, Fdov_year
100.000 0 # Pot fishery
1 0 0.000 0 # Trawl bycatch
100.000 0 # fixed gear bycatch
1 0 0.000 0 # NMFS survey
10 0.000 0 # ADF&G survey
## ========================================================================================== ##
## EMPHASIS FACTORS (Priors)
```



```
10000 \# Log_fdevs
0 # meanF
1.0 # Mdevs
    # Rec_devs
    # Initial_devs
    # Fst_dif_dev
    # Mean_sex-Ratio
    # Molt_prob
    # Free selectivity
```

| 0 | \# Init_n_at_len |
| :--- | :--- |
| 0 | \# Fdevs |
| 0 | \# Fdovs |
| 0 | \# Sel_devs |
|  |  |
| \#\# EOF |  |
| 9999 |  |


[^0]:    ${ }^{1} 1983 / 84$ refers to a fishing year that extends from 1 July 1983 to 30 June 1984.

[^1]:    ${ }^{2}$ NOAA grant Bering Sea Crab Research II, NA16FN2621, 1997.

[^2]:    ${ }^{3}$ D. Pengilly, ADF\&G, pers. comm.

