

## BRISTOL BAY RED KING CRAB STOCK ASSESSMENT IN SPRING 2021

J. Zheng and M.S.M. Siddeek  
Alaska Department of Fish and Game  
Division of Commercial Fisheries  
P.O. Box 115526  
Juneau, AK 99811-5526, USA  
Phone: (907) 465-6102  
Fax: (907) 465-2604  
Email: jie.zheng@alaska.gov

### Executive Summary

1. Stock: Red king crab (RKC), *Paralithodes camtschaticus*, in Bristol Bay, Alaska.
2. Catches: The domestic RKC fishery began to expand in the late 1960s and peaked in 1980 with a catch of 129.95 million lb (58,943 t). The catch declined dramatically in the early 1980s and remained at low levels during the last three decades. After rationalization, catches were relatively high before the 2010/11 season and have been on a declining trend since 2014. The retained catch in 2019/20 was approximately 3.9 million lb (1,775 t), compared to 4.5 million lb (2,027 t) in 2018/19, following a reduction in total allowable catch (TAC). The magnitude of bycatch from groundfish trawl and fixed gear fisheries has been stable and small relative to stock abundance during the last 10 years.
3. Stock biomass: Estimated mature biomass increased dramatically in the mid-1970s and decreased precipitously in the early 1980s. Estimated mature crab abundance had increased during 1985-2009 with mature females being about three times more abundant in 2009 than in 1985 and mature males being about two times more abundant in 2009 than in 1985. Estimated mature abundance has steadily declined since 2009.
4. Recruitment: Estimated recruitment was high during the 1970s and early 1980s and has generally been low since 1985 (1979-year class). During 1984-2019, estimated recruitment was above the historical average (1976-2019 reference years) only in 1984, 1986, 1995, 1999, 2002 and 2005. Estimated recruitment was extremely low during the last 12 years. Estimated recruitment for 2020 is not reliable due to the lack of trawl survey data.
5. Management performance:

Status and catch specifications (1,000 t) (model 19.3):

Year	MSST	Biomass (MMB)	TAC	Retained Catch	Total Catch	OFL	ABC
2016/17	12.53 <sup>A</sup>	25.81 <sup>A</sup>	3.84	3.92	4.37	6.64	5.97
2017/18	12.74 <sup>B</sup>	24.86 <sup>B</sup>	2.99	3.09	3.60	5.60	5.04
2018/19	10.62 <sup>C</sup>	16.92 <sup>C</sup>	1.95	2.03	2.65	5.34	4.27
2019/20	12.72 <sup>D</sup>	14.24 <sup>D</sup>	1.72	1.78	2.22	3.40	2.72
2020/21		14.93 <sup>D</sup>				2.14	1.61*

The stock was above MSST in 2019/20 and hence was not overfished. Since total catch was below OFL, overfishing did not occur. The relatively low MSST in 2018/19 and  $B_{35\%}$  in 2019/20 below was caused by a problem of the previous GMACS version using the only sex ratio of recruitment in the terminal year for  $B_{35\%}$  computation in 2019. The lower estimated male recruitment ratio in the terminal year in 2019 resulted in a lower mean male recruitment for  $B_{35\%}$  computation. The current version of GMACS uses average of sex ratios of recruitment during the reference period to estimate  $B_{35\%}$ , which results in a stable sex ratio (about 50%) for the reference point calculation.

Status and catch specifications (million lb):

Year	MSST	Biomass (MMB)	TAC	Retained Catch	Total Catch	OFL	ABC
2016/17	27.6 <sup>A</sup>	56.9 <sup>A</sup>	8.47	8.65	9.63	14.63	13.17
2017/18	28.1 <sup>B</sup>	54.8 <sup>B</sup>	6.60	6.82	7.93	12.35	11.11
2018/19	23.4 <sup>C</sup>	37.3 <sup>C</sup>	4.31	4.31	5.85	11.76	9.41
2019/20	28.0 <sup>D</sup>	31.4 <sup>D</sup>	3.80	3.91	4.89	7.50	6.00
2020/21		32.9 <sup>D</sup>				4.72	3.54*

Notes:

A – Calculated from the assessment reviewed by the Crab Plan Team in September 2017

B – Calculated from the assessment reviewed by the Crab Plan Team in September 2018

C – Calculated from the assessment reviewed by the Crab Plan Team in September 2019

D – Calculated from the assessment reviewed by the Crab Plan Team in September 2020

\*ABC is calculated applying a buffer of 25% to OFL

6. Basis for the OFL: Values in 1,000 t (model 19.3):

Year	Tier	B <sub>MSY</sub>	Current MMB	B/B <sub>MSY</sub> (MMB)	F <sub>OFL</sub>	Years to define B <sub>MSY</sub>	Natural Mortality
2016/17	3b	25.8	24.0	0.93	0.27	1984-2016	0.18
2017/18	3b	25.1	21.3	0.85	0.24	1984-2017	0.18
2018/19	3b	25.5	20.8	0.82	0.25	1984-2017	0.18
2019/20	3b	21.2	16.0	0.75	0.22	1984-2018	0.18
2020/21	3b	25.4	14.9	0.59	0.16	1984-2019	0.18

Basis for the OFL: Values in million lb:

Year	Tier	B <sub>MSY</sub>	Current MMB	B/B <sub>MSY</sub> (MMB)	F <sub>OFL</sub>	Years to define B <sub>MSY</sub>	Natural Mortality
2016/17	3b	56.8	52.9	0.93	0.27	1984-2016	0.18
2017/18	3b	55.2	47.0	0.85	0.24	1984-2017	0.18
2018/19	3b	56.2	45.9	0.82	0.25	1984-2017	0.18
2019/20	3b	46.8	35.2	0.75	0.22	1984-2018	0.18
2020/21	3b	56.1	32.9	0.59	0.16	1984-2019	0.18

## ***A. Summary of Major Changes***

**1. Changes to management of the fishery:** None.

**2. Changes to the input data:**

- a. No trawl survey was conducted in 2020.
- b. Updated directed pot fishery catch and bycatch data through 2019 (i.e., completed 2019/20 fishery) (occurred in September 2020), and updated/standardized observer biomass estimates in the directed pot fishery and Tanner crab fishery from 1990 to 2019 (occurred in May 2021) (Appendix D).
- c. Fishing effort data used to estimate red king crab bycatch in years before the observer data in the Tanner crab fishery were available are changed from east of 163° W to east of 166° W, which covers the whole Bristol Bay red king crab stock. Five more years of length composition data with relatively small observed sample sizes from the Tanner crab fishery are also added (occurred in May 2021) (Appendix D).
- d. Updated groundfish fisheries bycatch data during 2014-2019(occurred in September 2020).

**3. Changes to the assessment methodology:**

- a. The analyses of terminal years of recruitment are updated.

b. Eight models are compared in this report (See Section E.3.a for details):

- 19.3:** the base model adopted by the CPT and SSC in September 2020. This model has a constant  $M$  being estimated for males during 1980-1984, while maintaining a constant (base)  $M$  of 0.18 for males during other years, and an estimated constant multiplier being used to multiply male  $M$  for female  $M$ . That is,  $M$  for females is relative to  $M$  for males each year.
- 19.3c:** the same as model 19.3 except for updating/standardizing the observer data in the directed pot and Tanner crab fisheries. Fishing effort data used to estimate red king crab bycatch in years before the observer data in the Tanner crab fishery were available are changed from east of 163° W to east of 166° W, which covers the whole Bristol Bay red king crab stock. Five more years of length composition data with relatively small observed sample sizes from the Tanner crab fishery are also included.
- 19.3d:** the same as model 19.3c except for changing the maximum cap of effective sample size from 100 to 150 for the retained catch and total males in the directed pot fishery. The computed implied effective sample sizes for these two datasets in models 19.3 and 19.3c are close to 200 (Table 4).
- 19.3e:** the same as model 19.3d except for males and females to have different NMFS trawl survey catchabilities. This model is similar to model 19.5 in May 2020 except for changes in model 19.3, 19.3c and 19.3d.
- 19.3f:** the same as model 19.3e except for doubling the CV of the prior for trawl survey catchability.
- 19.3g:** the same as model 19.3d except for VAST-estimated NMFS survey trawl biomass and CV are used. This model is similar to model 19.4a in May 2020 except for the changes in models 19.3, 19.3c and 19.3d.
- 19.3i:** the same as model 19.3g except for an additional CV is estimated for NMFS trawl survey biomass.
- 19.6:** the same as model 19.3d except for base  $M$  is estimated from newly updated equation from Then et al. (2015) as  $0.257 (M = 4.899 t_{max}^{-0.916})$  with max age  $t_{max} = 25$ ). Conjunction with this model, a negative log likelihood profile is computed with a base  $M$  ranging from 0.10 to 0.4 for males.

#### 4. Changes to assessment results:

Eight model scenarios are compared in this draft report. Newly updated and standardized bycatch biomass and length composition estimates from the crab pot fisheries generally have only very minor impacts on the model results, and the minor impacts are likely caused by increasing length composition data from six years to eleven years and use of fishing effort from east of 163° W to east of 166° W in the Tanner crab fishery. Increasing the caps of effective sample sizes from 100 to 150 for retained and total male length compositions also affect very little on the model results. Among the eight models, model 19.3c is model 19.3 with updated pot fishery observer data, and model 19.3d is model 19.3c with a better cap of effective sample sizes for retained catch and total

males. These three models fit the data similarly and have similar results. Thus, model 19.3d is preferred over models 19.3 and 19.3c. Model 19.3e has different NMFS survey catchabilities for males and females, thus with one additional estimated parameter over model 19.3d. Both models 19.3d and 19.3e fit the data similarly and have similar model results. The high CV value for the NMFS survey catchability prior for model 19.3f results in estimated NMFS survey catchability > 1.0 for females. Therefore, model 19.3d is preferred over models 19.3e and 19.3f. VAST-estimated NMFS survey biomasses are generally lower than area-swept estimates until 2014 (Figure 6) and higher during 2014-2019, and thus VAST models 19.3g and 19.3i result in higher mature male biomass estimates than the other models in recent years. With the declining biomass trend during recent years, higher mature male biomass during recent years with VAST models 19.3g and 19.3i would increase the risk of overfishing the stock. Thus, we suggest not adopting VAST models now. Model 19.6 is like a sensitivity study, and the very high base natural mortality would result in an extremely high  $F_{35\%}$  value. Overall, we recommend model 19.3d as the base model for OFL and ABC determination in September 2021.

## ***B. Responses to SSC and CPT Comments***

### **1. Responses to the most recent two sets of SSC and CPT comments on assessments in general:**

#### **Response to SSC Comments (from October 2019):**

*“The SSC reminds authors to use the model numbering protocols that allows the SSC to understand the year in which a particular version of the model was first introduced. Also, when reporting bycatch in tables in each SAFE chapter, the SSC requests authors to be clear whether they report bycatch or bycatch mortality (DMRs have been applied). Further, when reporting bycatch mortality, it would be helpful to report the DMR values used.”*

Response: We have followed these recommendations.

*“The SSC requests that the CPT consider developing a standard approach for projecting the upcoming year’s biomass that does not include removing the entire OFL for stocks where recent mortality has been substantially below the OFL. This may appreciably change the projected biomass levels for stocks such as Tanner crab, where actual catch mortality has been less than 10% of the OFL.”*

Response: Agree to this request and will follow the standard approach developed by the CPT.

### **2. Responses to the most recent two sets of SSC and CPT comments specific to this assessment:**

#### **Response to CPT Comments (from May 2020):**

*“Given the above discussion, the CPT selected model 19.3 as the priority model (in addition to the status quo model, 19.0a) for presentation in September, understanding that time schedules for producing data used in the assessment may be compressed as a result of the global pandemic. Model 19.3 estimated male natural mortality in an early block (1980-1984) and then specified  $M$  as 0.18 thereafter. Female natural mortality was estimated as an offset from males in both periods. Survey selectivity was estimated separately for sexes, but a single catchability was estimated (still with a strong prior). If time allows, a model building from 19.3 in which the prior on catchability is relaxed and estimated separately by sex (and revisited in light of the catchability implied by the BSFRF data) would be useful for comparison.”*

Response: We used model 19.3b to examine the sensitivity of trawl survey catchability estimate when the CV of the prior on catchability was doubled in September 2020. The resulting catchability estimate was greater than 1.0. Different catchabilities for males and females in the NMFS survey were examined in model 19.5 in May 2020 with both sexes having the same survey selectivities. In this report (May 2021), we add model 19.3e with the same prior and model 19.3f with doubled CV value of the prior to examine different survey catchabilities between sexes. BSFRF survey data are used in the model, so we do not use them to estimate a prior for NMFS survey catchabilities.

*“Produce the empirical survey selectivity diagnostics that were produced for Tanner crab at this meeting, but for BBRKC. Specifically, display the ratio of NMFS to BSFRF (rather than  $NMFS/(NMFS+BSFRF)$ ) numbers at size to provide a direct comparison to estimated survey selectivity.”*

Response: Ratios of NMFS to BSFRF numbers at size are plotted in Figure 7 (a, b, and c). Note that the ratios are from combined all haul data due to small amount of crab caught. The abundance-weighted average ratio is 0.891 for crab  $\geq 135$  mm carapace length from all four years (2013-2016) of data, about the same as the double-bag experiment (0.896 at 162.5 mm carapace length), although the ratios changed greatly from year to year.

*“Describe how the sex ratios for OFL calculations were averaged. It is the same as the recruitments, but was difficult to confirm in the document.”*

Response: We added text to explain the sex ratios for OFL calculations in Appendix A (B (b) (2) The proxy for  $B_{MSY}$ ).

*“Check the calculation of total male directed fishery catch as inputted to GMACS to ensure accounting for discard mortality is appropriate. Check the tables for correct numbers and that they match the .DAT files provided. Consider splitting the tables needed by the State of Alaska from those presenting the data used in the assessment. CPT suggests that the methodology for how total catches are calculated should be added to the terms of reference for all assessments.”*

Response: Total male directed fishery catch data in the GMACS input data file are correct. Table 2 is added to include all observer catch and discard data. Methods of bycatch estimation are added to Table 1a caption.

*“Highlight the ‘PriorDensity’ row in the table listing the contribution of likelihoods to the objective function value. Make sure that it is clear that differences in likelihood comparability are well represented in the tables. It appears that modifications will need to be made to the way that GMACS includes or does not include prior densities so that the objective function values from models with different numbers of parameters (but fitting to identical data) are comparable.”*

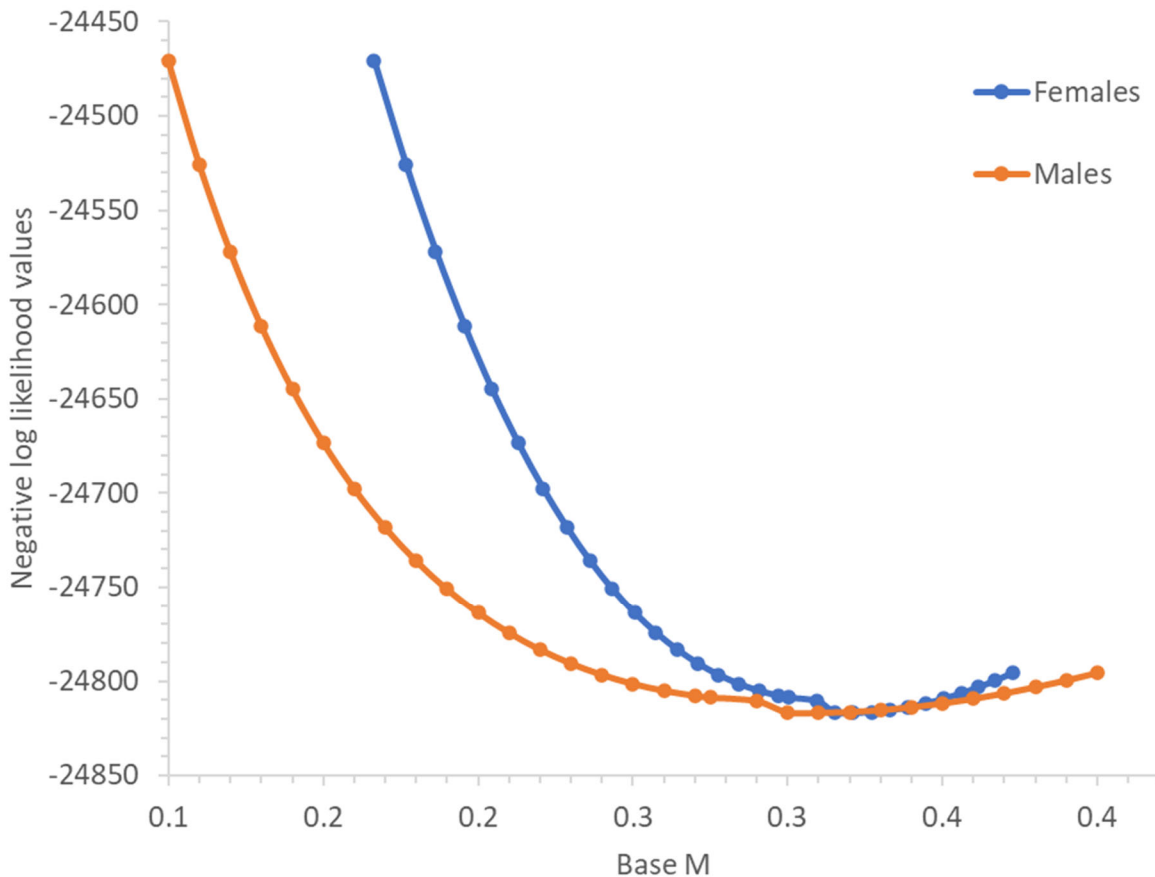
Response: The “PriorDensity” row is highlighted, and a new row is added for total negative log likelihood values without prior densities for easy comparison.

*“Include diagnostics for VAST indices of abundance and provide rationale for accepting or rejecting the index in future iterations (but not for September 2020).”*

Response: Diagnostics for VAST indices are illustrated in Appendix E.

*“Provide justification for the assumed natural mortality for males of 0.18 yr<sup>-1</sup>. How does the 1% rule assumed in the assessment compare to empirical studies on natural mortality and longevity (e.g. Then et al. 2015)?”*

Response: The 1% rule was accepted after very long, several year difficult discussions among the crab overfishing working group, CPT, and SSC. The base  $M$  for females is also higher than 0.18 for model 19.3 and the related models. Model 19.6 uses male base  $M$  of 0.257 estimated by Then et al. (2015), and we also examine a likelihood profile of base  $M$  from 0.1 to 0.4 as follow:



It appears that the maximum likelihood value is achieved with a base  $M$  of 0.31 for males and 0.321 for females.

**Response to CPT Comments (from September 2020):**

*“Include a table in the assessment document providing estimates of  $M$  for all scenarios.”*

Response: Included them in Table 8.

*“Include a table in the assessment document that provide the differences in likelihood values between the base model and each alternative model scenario considered a candidate for status determination.”*

Response: Done (Table 5c).

*“Evaluate different approaches to constraining the terminal year estimate of recruitment for the purpose of developing projections.”*



Response: We plan to compare three approaches to constraining the terminal year estimate of recruitment for the purpose of developing projections: (1) average of recent low recruitment period excluding the terminal year (7 years), (2) average of the period used to estimate  $B_{35\%}$ , (3) increase in the weight on annual recruitment variation. Since the current GMACS version cannot carry work on (1) and (2), we examine only (3) in this report. As shown in the table below with model 19.3d, increasing weight on sex ratio (forcing them close to 50% each year) has some minor impacts on recruitment estimates in the terminal year, while increasing penalty on annual variation seems not to affect the model results. Overall, changes in penalties on recruitment deviation and sex ratio have very minor impacts on terminal recruitment estimates.

Scenario	Penalty values		Total Neg. LL	$B_{35\%}$	$B_{2020}/$ $B_{35\%}$	Male $R_{2020}$	Female $R_{2020}$	R neg. log.like.	
	Deviation	Sex ratio						value	value
Base	2	10	-24735.8	25220.6	0.587	7747253	13422716	67.11	73.73
1	4	10	-24735.8	25220.6	0.587	7747255	13422727	67.11	73.73
2	4	20	-24735.8	25193.9	0.587	7691424	13404168	67.14	73.72
3	4	40	-24735.7	25149.4	0.588	7593412	13360860	67.19	73.70
4	8	10	-24735.8	25220.6	0.587	7747255	13422727	67.11	73.73
5	8	20	-24735.8	25193.9	0.587	7691424	13404168	67.14	73.72
6	8	40	-24735.7	25149.4	0.588	7593412	13360860	67.19	73.70
7	16	10	-24735.8	25220.6	0.587	7747255	13422727	67.11	73.73
8	16	20	-24735.8	25193.9	0.587	7691424	13404168	67.14	73.72
9	16	40	-24735.7	25149.4	0.588	7593412	13360860	67.19	73.70
10	2	100	-24735.5	25058.7	0.590	7376448	13239983	67.31	73.66
11	2	500	-24735.2	24874.1	0.594	6866432	12923606	67.61	73.61

**Response to SSC Comments specific to this assessment (from June 2020):**

*“The SSC agrees with the CPT’s model recommendations for September. Though promising, it is advisable to postpone the use of VAST estimates for this stock assessment until diagnostics for VAST can be more fully analyzed and better-fitting error distributions identified. The SSC also supports the other recommendations on this assessment offered by the CPT.”*

Response: We follow these suggestions.

**Response to SSC Comments specific to this assessment (from October 2020):**

*“Next year’s assessment should estimate the probability that the stock is currently in the overfished condition.”*

Response: Included it in Figure 31 and will include it September 2021 assessment document.

*“Information should be provided on the prevalence of crab bycatch excluders being used in the pot cod fishery, and whether the excluders influence the length composition of bycatch.”*

Response: Will try to collect these data and analyze them in the future. Right now, very limited data are available.

*“The SSC also endorses the Alaska Bering Sea Crabber’s request to include raw numbers used for PSC limits in a table in the BBRKC SAFE consistent with EBS snow crab (see Table 11 in the EBS snow crab SAFE), if it is practical to do so.”*

Response: Included it in this report for years 2010-2020 in Table 8.

*“The SSC looks forward to future analyses of the use of VAST estimates for this stock assessment including diagnostics and better-fitting error distributions.”*

Response: Included VAST results and diagnostics in Appendix E and used VAST-estimated survey biomass in two model scenarios (model 19.3g and 19.3i) in this report (May 2021).

## **C. Introduction**

### **1. Species**

Red king crab (RKC), *Paralithodes camtschaticus*, in Bristol Bay, Alaska.

### **2. General distribution**

Red king crab inhabit intertidal waters to depths >200 m of the North Pacific Ocean from British Columbia, Canada, to the Bering Sea, and south to Hokkaido, Japan, and are found in several areas of the Aleutian Islands, eastern Bering Sea, and the Gulf of Alaska.

### **3. Stock Structure**

The State of Alaska divides the Aleutian Islands and eastern Bering Sea into three management registration areas to manage RKC fisheries: Aleutian Islands, Bristol Bay, and Bering Sea (ADF&G 2012). The Bristol Bay area includes all waters north of the latitude of Cape Sarichef (54°36' N lat.), east of 168°00' W long., and south of the latitude of Cape Newenham (58°39' N lat.) and the fishery for RKC in this area is managed separately from fisheries for RKC outside of this area; i.e., the red king crab in the Bristol Bay area are assumed to be a separate stock from red king crab outside of this area. This report summarizes the stock assessment results for the Bristol Bay RKC stock.

### **4. Life History**

Red king crab have a complex life history. Fecundity is a function of female size, ranging from tens of thousands to hundreds of thousands (Haynes 1968; Swiney et al. 2012). The eggs are extruded by females, fertilized in the spring, and held by females for about 11 months (Powell and

Nickerson 1965). Fertilized eggs are hatched in the spring, most during April-June (Weber 1967). Primiparous females are bred a few weeks earlier in the season than multiparous females.

Larval duration and juvenile crab growth depend on temperature (Stevens 1990; Stevens and Swiney 2007). Male and female RKC mature at 5–12 years old, depending on stock and temperature (Stevens 1990; Loher et al. 2001) and may live >20 years (Matsuura and Takeshita 1990). Males and females attain a maximum size of 227 mm and 195 mm carapace length (CL), respectively (Powell and Nickerson 1965). Female maturity is evaluated by the size at which females are observed to carry egg clutches. Male maturity can be defined by multiple criteria including spermatophore production and size, chelae vs. carapace allometry, and participation in mating *in situ* (reviewed by Webb 2014). For management purposes, females >89 mm CL and males >119 mm CL are assumed to be mature for Bristol Bay RKC. Juvenile RKC molt multiple times per year until age 3 or 4; thereafter, molting continues annually in females for life and in males until maturity. Male molting frequency declines after attaining functional maturity.

## **5. Fishery**

The RKC stock in Bristol Bay, Alaska, supports one of the most valuable fisheries in the United States. A review of the history of the Bristol Bay RKC fishery is provided in Fitch et al. (2012) and Otto (1989). The Japanese fleet started the fishery in the early 1930s, stopped fishing from 1940 to 1952, and resumed the fishery from 1953 until 1974. The Russian fleet fished for RKC from 1959 to 1971. The Japanese fleet employed primarily tanglenets with a very small proportion of catch from trawls and pots. The Russian fleet used only tanglenets. United States trawlers started fishing Bristol Bay RKC in 1947, but the effort and catch declined in the 1950s. The domestic RKC pot fishery began to expand in the late 1960s and peaked in 1980 with a catch of 129.95 million lb (58,943 t), worth an estimated \$115.3 million ex-vessel value. The catch declined dramatically in the early 1980s and has remained at low levels during the last two decades (Tables 1a and 1b). After the early 1980s stock collapse, the Bristol Bay RKC fishery took place during a short period in the fall (usually lasting about a week) with the catch quota based on the stock assessment conducted the previous summer (Zheng and Kruse 2002). Beginning with the 2005/2006 season, new regulations associated with fishery rationalization resulted in an increase in the duration of the fishing season (October 15 to January 15). With the implementation of crab rationalization, historical guideline harvest levels (GHL) were changed to a total allowable catch (TAC). Before rationalization, the implementation errors were quite high for some years and sum of actual catches from 1980 to 2007 was about 6% less than the sum of GHL/TAC over that period.

## **6. Fisheries Management**

King and Tanner crab stocks in the Bering Sea and Aleutian Islands are managed by the State of Alaska through a federal king and Tanner crab fishery management plan (FMP). Under the FMP, management measures are divided into three categories: (1) fixed in the FMP, (2) frameworked in the FMP, and (3) discretion of the State of Alaska. The State of Alaska is responsible for determining and establishing the GHL/TAC under the framework in the FMP.

Harvest strategies for the Bristol Bay RKC fishery have changed over time. Two major management objectives for the fishery are to maintain a healthy stock that ensures reproductive viability and to provide for sustained levels of harvest over the long term (ADF&G 2012). In

attempting to meet these objectives, the GHL/TAC is coupled with size-sex-season restrictions. Only males  $\geq 6.5$ -in carapace width (equivalent to 135-mm carapace length, CL) may be harvested and no fishing is allowed during molting and mating periods (ADF&G 2012). Specification of TAC is based on a harvest rate strategy. Before 1990, harvest rates on legal males were based on population size, abundance of prerecruits to the fishery, and postrecruit abundance, and rates varied from less than 20% to 60% (Schmidt and Pengilly 1990). In 1990, the harvest strategy was modified, and a 20% mature male harvest rate was applied to the abundance of mature-sized ( $\geq 120$ -mm CL) males with a maximum 60% harvest rate cap of legal ( $\geq 135$ -mm CL) males (Pengilly and Schmidt 1995). In addition, a minimum threshold of 8.4 million mature-sized females ( $\geq 90$ -mm CL) was added to existing management measures to avoid recruitment overfishing (Pengilly and Schmidt 1995). Based on a new assessment model and research findings (Zheng et al. 1995a, 1995b, 1997a, 1997b), the Alaska Board of Fisheries adopted a new harvest strategy in 1996. That strategy had two mature male harvest rates: 10% when effective spawning biomass (ESB) is between 14.5 and 55.0 million lb and 15% when ESB is at or above 55.0 million lb (Zheng et al. 1996). The maximum harvest rate cap of legal males was changed from 60% to 50%. A threshold of 14.5 million lb of ESB was also added. In 1997, a minimum threshold of 4.0 million lb was established as the minimum GHL for opening the fishery and maintaining fishery manageability when the stock abundance is low. The Board modified the current harvest strategy in 2003 by adding a mature harvest rate of 12.5% when the ESB is between 34.75 and 55.0 million lb and in 2012 eliminated the minimum GHL threshold. The current harvest strategy is illustrated in Figure 1.

## ***D. Data***

### **1. Summary of New Information**

- a. No trawl survey was conducted in 2020.
- b. Updated directed pot fishery catch and bycatch data through 2019 (i.e., completed 2019/20 fishery) (occurred in September 2020), and updated/standardized observer biomass estimates in the directed pot fishery and Tanner crab fishery from 1990 to 2019 (occurred in May 2021) (Appendix D).
- c. Fishing effort data used to estimate red king crab bycatch in years before the observer data in the Tanner crab fishery were available are changed from east of  $163^{\circ}$  W to east of  $166^{\circ}$  W, which covers the whole Bristol Bay red king crab stock. Five more years of length composition data with relatively small observed sample sizes from the Tanner crab fishery are also added (occurred in May 2021) (Appendix D).
- d. Updated groundfish fisheries bycatch data during 2014-2019.

Data types and ranges are illustrated in Figure 2.

### **2. Catch Data**

Data on landings of Bristol Bay RKC by length and year and catch per unit effort from 1960 to 1973 were obtained from annual reports of the International North Pacific Fisheries Commission (Hoopes et al. 1972; Jackson 1974; Phinney 1975) and from the Alaska Department of Fish and Game from 1974 to 2019 (Tables 1a and 1b). Bycatch data are available starting from 1990 and

were obtained from the ADF&G observer database and reports (Gaeuman 2013) (Tables 2a and 2b). Sample sizes for catch by length and shell condition are summarized in Tables 3a and 3b. Relatively large samples were taken from the retained catch each year. Sample sizes for trawl bycatch were the annual sums of length frequency samples in the National Marine Fisheries Service (NMFS) database.

***(i). Catch Biomass***

Retained catch and estimated bycatch biomasses are summarized in Tables 1a and 1b, and illustrated in Figure 3. Retained catch and estimated bycatch from the directed fishery include the general, open-access fishery (prior to rationalization), or the individual fishery quota (IFQ) fishery (after rationalization), as well as the Community Development Quota (CDQ) fishery and the ADF&G cost-recovery harvest. Starting in 1973, the fishery generally occurred during the late summer and fall. Before 1973, a small portion of retained catch in some years was caught from April to June. The years in Tables 1a and 1b are defined as crab year from July 1 to June 30. Bycatch data for the cost-recovery fishery before 2006 were not available. In this report, pot fisheries include both the directed fishery and RKC bycatch in the Tanner crab pot fishery, and trawl fisheries and fixed gear fisheries are groundfish fisheries. Observers did not separate legal retained and discarded catch after 2017 in the directed pot fishery, so the male discarded biomass from the directed fishery has been estimated by the subtraction method since 2018 (B. Daly, ADF&G, personal communication).

***(ii). Catch Size Composition***

Retained catches by length and shell condition and bycatches by length, shell condition, and sex were obtained for stock assessments. From 1960 to 1966, only retained catch length compositions from the Japanese fishery were available. Retained catches from the Russian and U.S. fisheries were assumed to have the same length compositions as the Japanese fishery during this period. From 1967 to 1969, the length compositions from the Russian fishery were assumed to be the same as those from the Japanese and U.S. fisheries. After 1969, foreign catch declined sharply and only length compositions from the U.S. fishery were used to distribute catch by length.

***(iii). Catch per Unit Effort***

Catch per unit effort (CPUE) is defined as the number of retained crab per tan (a unit fishing effort for tanglenets) for the Japanese and Russian tanglenet fisheries and the number of retained crab per potlift for the U.S. fishery (Table 1b). Soak time, while an important factor influencing CPUE, is difficult to standardize. Furthermore, complete historical soak time data from the U.S. fishery are not available. Based on the approach of Balsiger (1974), all fishing effort from Japan, Russia, and U.S. were standardized to the Japanese tanglenet from 1960 to 1971, and the CPUE was standardized as crab per tan. Except for the peak-to-crash years of the late 1970s and early 1980s, the correspondence between U.S. fishery CPUE and area-swept survey abundance is poor (Figure 4). Due to the difficulty in estimating commercial fishing catchability and crab availability to the NMFS annual trawl survey data, commercial CPUE data were not used in the model.

### 3. NMFS Survey Data

The NMFS has conducted annual trawl surveys of the eastern Bering Sea since 1968. Two vessels, each towing an eastern otter trawl with an 83 ft headrope and a 112 ft footrope, conducted this multispecies, crab-groundfish survey during the summer. Stations were sampled in the center of a systematic 20 X 20 nm grid overlaid in an area of  $\approx 140,000 \text{ nm}^2$ . Since 1972, the trawl survey has covered the full stock distribution except in nearshore waters. The survey in Bristol Bay occurs primarily during late May and June. Tow-by-tow trawl survey data for Bristol Bay RKC during 1975-2019 were provided by NMFS.

Abundance estimates by sex, carapace length, and shell condition were derived from survey data using an area-swept approach (Figures 5a and 5b). Until the late 1980s, NMFS used a post-stratification approach, but subsequently treated Bristol Bay as a single stratum; the estimates shown for Bristol Bay in Figures 4, 5a, and 5b were made without post-stratification. If multiple tows were made for a single station in a given year, the average of the abundances from all tows within that station was used as the estimate of abundance for that station. The new time series since 2015 discards all “hot spot” tows. We used the new area-swept estimates provided by NMFS in 2019. The VAST estimated biomasses are compared to area-swept biomasses in Figure 6.

In addition to the standard surveys conducted in early June (late May to early June in 1999 and 2000), a portion of the distribution of Bristol Bay RKC was re-surveyed in 1999, 2000, 2006-2012, and 2017 to better assess mature female abundance. Resurveys performed in late July, about six weeks after the standard survey, included 31 stations (1999), 23 stations (2000), 31 stations (2006, 1 bad tow and 30 valid tows), 32 stations (2007-2009), 23 stations (2010), and 20 stations (2011 and 2012) with high female densities. The resurveys were necessary because a high proportion of mature females had not yet molted or mated when sampled during the standard survey. Differences in area-swept estimates of abundance between the standard surveys and resurveys of these same stations are attributed to survey measurement errors or to seasonal changes in distribution between survey and resurvey. More large females were observed in the resurveys than during the standard surveys in 1999 and 2000, presumably because most mature females had not molted prior to the standard surveys. As in 2006, area-swept estimates of males  $>89 \text{ mm CL}$ , mature males, and legal males within the 32 resurvey stations in 2007 were not significantly different ( $P=0.74$ ,  $0.74$  and  $0.95$ ; paired  $t$ -test of sample means) between the standard survey and resurvey tows. However, similar to 2006, area-swept estimates of mature females within the 32 resurvey stations in 2007 were significantly different ( $P=0.03$ ; paired  $t$ -test) between the standard survey and resurvey tows. Resurvey stations were close to shore during 2010-2012, and mature and legal male abundance estimates were lower for the re-tow than the standard survey. Following the CPT recommendation, we used the standard survey data for male abundance estimates and only the resurvey data, plus the standard survey data outside the resurveyed stations, to assess female abundances during these resurvey years.

### 4. Bering Sea Fisheries Research Foundation Survey Data

The BSFRF conducted trawl surveys for Bristol Bay RKC in 2007 and 2008 with a small-mesh trawl net and 5-minute tows (S. Goodman, BSFRF, pers. com.). The surveys occurred at similar times as the NMFS standard surveys and covered about 97% of the Bristol Bay survey area. Few Bristol Bay RKC were found outside the BSFRF survey area. Because of the small mesh size, the

BSFRF surveys were expected to catch more RKC within the swept area. Crab abundances of different size groups were estimated by the kriging method. Mature male abundances were estimated to be 22.331 million crab (CV = 0.0634) in 2007 and 19.747 million crab (CV = 0.0765) in 2008. BSFRF also conducted a side-by-side survey concurrent with the NMFS trawl survey during 2013-2016 in Bristol Bay. In May 2017, survey biomass and size composition estimates from 2016 BSFRF side-by-side trawl survey data were updated. Ratios of NMFS survey abundances/total NMFS and BSFRF side-by-side trawl survey abundances are illustrated in Figure 7a, and ratios of NMFS survey abundances/BSFRF side-by-side trawl survey abundances are shown in Figures 7b and 7c.

As a comparison to the estimated NMFS survey catchability (0.896) at 162.5 mm carapace length by the double-bag experiment, we computed an overall ratio ( $q=0.891$ ) of NMFS survey abundances/BSFRF side-by-side trawl survey abundances for legal crab ( $\geq 135$  mm carapace length) as follow:

$$q = \frac{\sum_{y=2013, l=135mm}^{y=2016, l=\infty} r_{y,l} n_{y,l}}{\sum_{y=2013, l=135mm}^{y=2016, l=\infty} n_{y,l}} \quad (1)$$

where  $r_{y,l}$  is the ratio of NMFS survey abundance/BSFRF side-by-side trawl survey abundance in year  $y$  and length group  $l$ , and  $n_{y,l}$  is the combined survey abundance of side-by-side surveys in year  $y$  and length group  $l$ . Due to small catch, all haul data were combined to compute the ratios for each length group and year.

## ***E. Analytic Approach***

### **1. History of Modeling Approaches**

To reduce annual measurement errors associated with abundance estimates derived from the area-swept method, ADF&G developed a length-based analysis (LBA) in 1994 that incorporates multiple years of data and multiple data sources in the estimation procedure (Zheng et al. 1995a). Annual abundance estimates of the Bristol Bay RKC stock from the LBA have been used to manage the directed crab fishery and to set crab bycatch limits in the groundfish fisheries since 1995 (Figure 1). An alternative length-based model (research model) was developed in 2004 to include small size crab to determine federal overfishing limits. Given that the crab abundance declined sharply during the early 1980s, the LBA estimated natural mortality for different periods of years, whereas the research model estimated additional mortality beyond a base constant natural mortality during 1980-1984. In this report, we present only the research model that was fit to the data from 1975 to 2020.

### **2. Model Description**

The original LBA model was described in detail by Zheng et al. (1995a, 1995b) and Zheng and Kruse (2002). The model combines multiple sources of survey, catch, and bycatch data using a maximum likelihood approach to estimate abundance, recruitment, selectivities, catches, and bycatch of the commercial pot fisheries and groundfish trawl fisheries. Since 2019, GMACS (General Model for Alaska Crab Stocks) has been used for this stock assessment. A full model description is provided in Appendix A.

a-f. See Appendix A.

g. Critical assumptions of the model:

- i. The base natural mortality is kept constant at  $0.18\text{yr}^{-1}$  for males, shell condition, and length and was estimated assuming a maximum age of 25 and applying the 1% rule (Zheng 2005).
  - ii. Survey and fisheries selectivities are a function of length and were constant over shell condition. Selectivities may or may not be a function of sex except for groundfish fisheries bycatch selectivities, which are the same for both sexes. Two different NMFS survey selectivities were estimated: (1) 1975-1981 and (2) 1982-2020, based on modifications to the trawl gear used in the assessment survey.
  - iii. Growth is a function of length. For females, growth-per-molt increments as a function of length are estimated for three periods (1975-1982, 1983-1993, and 1994-2020) based on sizes at maturity. Once mature, female red king crab have a much smaller growth increment per molt.
  - iv. Molting probabilities are an inverse logistic function of length for males. Females molt annually.
  - v. Annual fishing seasons for the directed fishery are short.
  - vi. The prior of NMFS survey catchability ( $Q$ ) is estimated to be 0.896 with a standard deviation of 0.025 for some models, based on a trawl experiment by Weinberg et al. (2004);  $Q$  is assumed to be constant over time and is estimated in the model. The BSFRF survey catchability is assumed to be 1.0. The prior of 0.896 for NMFS survey  $Q$  (at 162.5 mm carapace length) is also close to the abundance-weighted average ratio of 0.891 for crab  $\geq 135$  mm carapace length across four years of side-by-side NMFS and BSFRF survey data (Figure 7c).
  - vii. Males mature at sizes  $\geq 120$  mm CL. For convenience, female abundance is summarized at sizes  $\geq 90$  mm CL as an index of mature females.
  - viii. Measurement errors are assumed to be normally distributed for length compositions and are log-normally distributed for biomasses.
- h. Changes to the above since previous assessment: see Section A.3. Changes to the assessment methodology.
- i. Outline of methods used to validate the code used to implement the model and whether the code is available: Assessment results by GMACS have been compared to the previous assessment models, and the code is online and available from the first author.

### 3. Model Selection and Evaluation

a. Alternative model configurations (models):

**19.3:** the base model adopted by the CPT and SSC in September 2020. Basic features of this model include:



- (1) An estimated constant  $M$  for males during 1980-1984, a constant (base)  $M$  of 0.18 for males during the other years, and an estimated constant multiplier being used to multiply male  $M$  for female  $M$ . That is,  $M$  for females is relative to  $M$  for males each year.
- (2) Including BSFRF survey data during 2007-2008 and 2013-2016.
- (3) Estimating a constant NMFS survey catchability over time in the model and assuming BSFRF survey catchability to be 1.0.
- (4) Assuming the BSFRF survey selectivities as the availability to the NMFS trawl survey because the BSFRF survey gear has very small mesh sizes and has tighter contact to the sea floor. This implies that crab occurring in nearshore areas are not available to trawl survey gears.
- (5) Two levels of molting probabilities for males: one before 1980 and one after 1979, based on survey shell condition data. Each level has two parameters.
- (6) Estimating effective sample size from observed sample sizes. Stage-1 effective sample sizes are estimated as  $\min(0.25*n, N)$  for trawl surveys and  $\min(0.05*n, N)$  for catch and bycatch, where  $n$  is the sum of observed sample sizes for two sexes, and  $N$  is the maximum sample size (200 for trawl surveys, 100 for males from the pot fishery and 50 for females from the pot fishery and for both males and females from the groundfish fisheries). There is justification for enforcing a maximum limit to effective sample sizes because the number of length measurements is large (Fournier et al. 1998).
- (7) Standard survey data for males and NMFS survey re-tow data (during cold years) for females.
- (8) Estimating initial year length compositions.
- (9) Using the total observer male biomass and total observer male length composition data in the directed pot fishery to replace discarded male biomass and discarded male length composition data.
- (10) Using total male selectivity and retained proportions in the directed pot fishery to replace retained selectivity and discarded male selectivity; and due to high grading problems in some years since rationalization, estimating two logistic curves for retained proportions: one before rationalization (before 2005) and another after 2004.
- (11) Equal annual effective sample sizes of male and female length compositions.

**19.3c:** the same as model 19.3 except for updating/standardizing the observer data in the directed pot and Tanner crab fishery. Fishing effort data used to estimate red king crab bycatch in years before the observer data in the Tanner crab fishery were available are changed from east of 163° W to east of 166° W, which covers the whole Bristol Bay red king crab stock. Five more years of length composition data with relatively small observed sample sizes from the Tanner crab fishery are also included.

**19.3d:** the same as model 19.3c except for changing the maximum cap (N) of effective sample size from 100 to 150 for the retained catch and total males in the directed pot fishery. The computed implied effective sample sizes for these two datasets in models 19.3 and 19.3c are close to 200 (Table 4).

**19.3e:** the same as model 19.3d except for males and females to have different NMFS trawl survey catchabilities. This model is similar to model 19.5 in May 2020 except for changes in model 19.3, 19.3c and 19.3d.

**19.3f:** the same as model 19.3e except for doubling the CV of the prior for trawl survey catchability.

**19.3g:** the same as model 19.3d except for VAST-estimated NMFS survey trawl biomass and CV are used. This model is similar to model 19.4a in May 2020 except for the changes in models 19.3, 19.3c and 19.3d.

**19.3i:** the same as model 19.3g except for an additional CV is estimated for NMFS trawl survey biomass.

**19.6:** the same as model 19.3d except for base  $M$  is estimated from newly updated equation from Then et al. (2015) as  $0.257 (M = 4.899 t_{max}^{-0.916}$  with max age  $t_{max} = 25$ ). Conjunction with this model, a negative log likelihood profile is computed with a base  $M$  ranging from 0.10 to 0.4 for males.

- b. Progression of results: See the new results at the beginning of the report.
- c. Evidence of search for balance between realistic and simpler models: NA.
- d. Convergence status/criteria: ADMB default convergence criteria.
- e. Sample sizes for length composition data: observed sample sizes are summarized in Tables 3a and 3b.
- f. Credible parameter estimates: All estimated parameters seem to be credible and within bounds.
- g. Model selection criteria: The likelihood values are used to select among alternatives that could be legitimately compared by that criterion.
- h. Residual analysis: Residual plots are illustrated in various figures.
- i. Model evaluation is provided under Results, below.
- j. Jittering: The Stock Synthesis Approach is used to perform jittering to find the optimum:

The *Jitter* factor of 0.1 is multiplied by a random normal deviation  $rdev=N(0,1)$ , to a transformed parameter value based upon the predefined parameter:

$$temp = 0.5 rdev Jitter \ln\left(\frac{P_{max} - P_{min} + 0.0000002}{P_{val} - P_{min} + 0.0000001} - 1\right), \quad (6)$$

with the final jittered starting parameter value back-transformed as:

$$P_{new} = P_{min} + \frac{P_{max} - P_{min}}{1.0 + \exp(-2.0 temp)}, \quad (7)$$

where  $P_{max}$  and  $P_{min}$  are upper and lower bounds of parameters and  $P_{val}$  is the estimated parameter value before the jittering. Jittering results are not updated and presented in this report.

#### 4. Results

- a. Effective sample sizes and weighting factors.
  - i. CVs are assumed to be 0.03 for retained catch biomass, 0.04 for total male biomass, 0.07 for pot bycatch biomasses, 0.10 for groundfish bycatch biomasses, and 0.23 for recruitment sex ratio. Models also estimate sigmaR for recruitment variation and have a penalty  $M$  variation and many prior-densities.
  - ii. Initial trawl survey catchability ( $Q$ ) is estimated to be 0.896 with a standard deviation of 0.025 (CV about 0.03) based on the double-bag experiment results (Weinberg et al. 2004). These values are used to set a prior for estimating  $Q$  in all models.
  - iii. Harmonic means of implied sample sizes and maximum caps of effective sample sizes for models 19.3, 19.3c, and 19.3d are summarized in Table 4.
- b. Tables of estimates.
  - i. Negative log-likelihood values and parameter estimates are summarized in Tables 5a, 5b, 5c, 6a, and 6b for all eight models.
  - ii. Natural mortality estimates are shown in Table 7 for all eight models.
  - iii. Area-swept estimates of mature female abundance and model estimates of effective spawning biomass (Zheng et al. 1995b) during 2010-2020 for groundfish fisheries bycatch calculation are provided in Table 8.
  - iv. Abundance and biomass time series are provided in Tables 9a and 9b for models 19.3 and 19.3d.
  - v. Recruitment time series for models 19.3 and 19.3d are provided in Tables 9a and 9b.
  - vi. Time series of catch biomass is provided in Tables 1a and 1b.

Length-specific fishing mortality is equal to selectivity-at-length times the full fishing mortality. Estimated full pot fishing mortalities for females and full fishing mortalities for groundfish fisheries bycatch are low due to low bycatch and handling mortality rates less than 1.0. Estimated recruits varied greatly among years (Tables 9a and 9b). Estimated selectivities for female pot bycatch are close to 1.0 for all mature females, and the estimated full fishing mortalities for female pot bycatch are lower than for male retained catch and bycatch (Tables 6a and 6b for models 19.3 and 19.3d).

c. Graphs of estimates.

- i. Estimated selectivities and molting probabilities by length are provided in Figures 8a, 8b, 8c, 8d, and 8e and 9a and 9b for models 19.3 and 19.3d.

One of the most important results is estimated trawl survey selectivity (Figures 8a, 8b, and 8c). Survey selectivity affects not only the fitting of the data but also the absolute abundance estimates. Estimated survey selectivities in Figures 8a, 8b, and 8c are generally smaller than the capture probabilities in Figure A1 because survey selectivities include capture probabilities and crab availability. The NMFS survey catchability is estimated to be 0.896 from the trawl experiment. The reliability of estimated survey selectivities will greatly affect the application of the model to fisheries management. Under- or over-estimates of survey selectivities will cause a systematic upward or downward bias of abundance estimates, respectively. Information about crab availability in the survey area at survey times will help estimate the survey selectivities. Among the models, model 19.3f with doubled NMFS survey biomass CV values has the highest selectivities, while model 19.6 with high natural mortality results in lowest selectivities.

For all models, estimated molting probabilities during 1975-2020 (Figures 9a and 9b) are generally lower than those estimated from the 1954-1961 and 1966-1969 tagging data (Balsiger 1974). Lower molting probabilities mean more oldshell crab, possibly due to changes in molting probabilities over time or shell aging errors. Overestimates or underestimates of oldshell crab will result in lower or higher estimates of male molting probabilities.

- ii. Estimated male and female survey biomasses are shown for NMFS surveys (Figures 10a1, 10a2, and 10a3) and BSFRF surveys (Figures 10b and 10c). Absolute mature male biomasses are illustrated in Figures 11a and 11b.

The population biomass estimates in 2020 are slightly higher than those in 2019. Estimated population biomass increased dramatically in the mid-1970s then decreased precipitously in the early 1980s. Estimated biomass had increased during 1985-2009, declined since 2009, and then have steadily declined since the late 2000s (Figures 10a-10c, 11a, and 11b). Absolute mature male biomasses for all models have a similar trend over time (Figures 11a and 11b). Among the eight models, model estimated relative NMFS survey biomasses are similar, especially for models 19.3, 19.3c, 19.3d, 19.3e, 19.3f, and 19.3g. Absolute mature male biomass estimates are higher for model 19.6 with a high base natural mortality than for the other models and higher for models 19.3g and 19.3i (VAST model scenarios) during recent years. As expected, model 19.3f estimates a higher trawl survey catchability ( $>1.0$  for females), thus resulting in overall lower absolute biomass estimates. All eight models fit the catch and bycatch biomasses very well.

The fit to BSFRF survey data and estimated survey selectivities are illustrated in Figures 10b-10f.

Among the eight models, model 19.3c is model 19.3 with updated pot fishery observer data, and model 19.3d is model 19.3c with a better cap of effective sample sizes for retained and total male catches. These three models fit the data similarly and have similar model results. Thus, model 19.3d is preferred over models 19.3 and 19.3c. Model 19.3e has different NMFS survey catchabilities for males and females, thus with one additional estimated parameter over model 19.3d. Both models 19.3d and 19.3e fit the data similarly and have similar model results. The high CV value for the NMFS survey catchability prior for model 19.3f results in estimated NMFS survey catchability  $> 1.0$  for females. Therefore, model 19.3d is preferred over models 19.3e and 19.3f. VAST-estimated NMFS survey biomasses are generally lower than area-swept estimates until 2014 (Figure 6) and higher during 2014-2019, and thus VAST models 19.3g and 19.3i result in higher mature male biomass estimates than the other models in recent years. With the declining biomass trend during recent years, higher mature male biomass during recent years with VAST models 19.3g and 19.3i would increase the risk of overfishing the stock. Thus, we suggest not adopting VAST models now. Model 19.6 is like a sensitivity study, and the very high base natural mortality would result in extremely high  $F_{35\%}$ . Overall, we recommend model 19.3d as the base model for overfishing definition determination in September 2021.

Like the results of model 19.0 in September 2019, the terminal year recruitment analysis with model 19.3 also suggests the estimated recruitment in the last year should not be used for estimating  $B_{35\%}$ .

- iii. Estimated recruitment time series are plotted in Figure 12a and recruitment length distributions in Figure 12b for models 19.3 and 19.3d. Recruitment is estimated at the end of year in GMACS and is moved up one year for the beginning of next year.
- iv. Estimated fishing mortality rates are plotted against mature male biomass in Figures 13a and 13b and estimated  $M$  and directed pot fishing mortality values over time are illustrated in Figure 13c for models 19.3 and 19.3d.

The average of estimated male recruits from 1984 to 2019 (Figure 12a) and mature male biomass per recruit are used to estimate  $B_{35\%}$ . The full fishing mortalities for the directed pot fishery at the time of fishing are plotted against mature male biomass on Feb. 15 (Figures 13a and 13b). Estimated fishing mortalities in most years before the current harvest strategy was adopted in 1996 were above  $F_{35\%}$  (Figures 13a and 13b). Under the current harvest strategy, estimated fishing mortalities were at or above the  $F_{35\%}$  limits in 1998-1999, 2005, 2007-2010, and 2014-2019 for models 19.3 and 19.3d, but below the  $F_{35\%}$  limits in the other post-1995 years.

For model 19.3, estimated full pot fishing mortalities ranged from 0.00 to 2.24 during 1975-2019, with estimated values over 0.40 during 1975-1976, 1978-1982, 1984-1987, 1990-1991, 1993, 1998 and 2007-2008 (Table 6a, Figure 13a). For model 19.3a, estimated full pot fishing mortalities ranged from 0.00 to 2.13 during 1975-2019, with estimated values over 0.40 in the same years as model 19.3 (Table 6b, Figure 13b).

Estimated fishing mortalities for pot female and groundfish fisheries bycatches are generally less than 0.07.

For model 19.3, estimated  $M$  values are 0.8966 during 1980-1984 and 0.18 for the other years for males, and 1.1802 during 1980-1984 and 0.2369 for the other years for females, with estimated female  $M$  values equaling to 1.3163 times male  $M$  values (Figure 13c). For model 19.3d, estimated  $M$  values are 0.9069 during 1980-1984 and 0.18 for the other years for males, and 1.1894 during 1980-1984 and 0.2361 for the other years for females, with estimated female  $M$  values equaling to 1.3117 times male  $M$  values (Figure 13c). Biologically, females mature earlier than males and likely have higher  $M$  values.

- v. Estimated mature male biomass and recruitment are plotted to illustrate their relationships with model 19.3d (Figure 14a). Annual stock productivities are illustrated in Figure 14b.

Stock productivity (recruitment/mature male biomass) is generally lower during the last 20 years (Figure 14b). However, there are high variations for the relation of stock productivity against mature male biomass.

Egg clutch data collected during summer surveys may provide information about mature female reproductive conditions. Although egg clutch data are subject to rating errors as well as sampling errors, data trends over time may be useful. Proportions of empty clutches for newshell mature females >89 mm CL are high in some years before 1990 but have been low since 1990 (Figure 15). The highest proportion of empty clutches (0.2) was in 1986, and primarily involved soft shell females (shell condition 1). Clutch fullness fluctuated annually around average levels during two periods: before 1991 and after 1990 (Figure 15). The average clutch fullness is similar for these two periods (Figure 15). Egg clutch fullness during 2016-2018 was relatively low, then increased in 2019.

d. Graphic evaluation of the fit to the data.

- i. Observed vs. estimated catches are plotted in Figure 16a, with bycatch mortalities from different sources shown in Figure 16b.
- ii. Model fits to NMFS survey biomass are shown in Figure 10 with a standardized residual plot in Figures 17a and 17b for models 19.3 and 19.3d.
- iii. Model fits to catch and survey proportions by length are illustrated in Figures 18-24 and residual bubble plots are shown in Figures 25-26.

All eight models fit the fishery biomass data well and the survey biomass reasonably well (Figures 10 and 16). Because the model estimates annual fishing mortality for directed pot male catch, pot female bycatch, and trawl and fixed gear bycatch, the deviations of observed and predicted (estimated) fishery biomass are mainly due to size composition differences. Models 19.3, 19.3c and 19.3d fit the NMFS area-swept biomass data almost identical (Figure 10a1).

The models also fit the length composition data well (Figures 18-24). Modal progressions are tracked well in the trawl survey data, particularly beginning in mid-1990s (Figures 18 and 19). Cohorts first seen in the trawl survey data in 1975, 1986, 1990, 1995, 1999, 2002 and 2005 can be tracked over time. Some cohorts can be tracked over time in the pot bycatch as well (Figure 21), but the bycatch data did not track the cohorts as well as the survey data. Groundfish bycatch data provide little information to track modal progression (Figures 23 and 24).

Residuals of survey biomasses and proportions of length are plotted to examine their patterns. Residuals were calculated as observed minus predicted and standardized by the estimated standard deviation. Residuals of survey biomasses did not show any consistent patterns for models 19.3 and 19.3d (Figures 17a and 17b). Generally, residuals of proportions of survey males and females appear to be random over length and year for models 19.3 and 19.3d (Figures 25a, 25b, 26a, and 26b).

e. Retrospective and historic analyses.

Two kinds of retrospective analyses were conducted for this report: (1) the 2020 model (model 19.3) hindcast results and (2) historical results. The 2020 model hindcast results are based on sequentially excluding one-year of data to evaluate the current model performance with fewer data. The historical results are the trajectories of biomass and abundance from previous assessments that capture both new data and changes in methodology over time. Treating the 2020 estimates as the baseline values, we can evaluate how well the model had done in the past.

i. Retrospective analysis (retrospective bias in base model or models).

The performance of the 2020 model includes sequentially excluding one-year of data. Models 19.3 and 19.3d produced some upward biases during 2009-2019 with higher terminal year estimates of mature male biomass in 2009-2010 and 2014-2019 (Figures 27-28). Higher than expected BSFRF survey biomass during 2007-2008 and 2013-2016 and NMFS survey biomass in 2014 likely caused these biases. Also, much lower than expected NMFS survey biomass during 2018-2019 results in lower biomass estimates in 2020. The biases for total abundance are much smaller than mature male biomass.

ii. Historic analysis (plot of actual estimates from current and previous assessments).

The model first fit the data from 1985 to 2004 in the terminal year of 2004. Thus, sequentially incrementing the terminal year provided 17 historical assessments for comparison with the 2020 assessment model results (Figure 29). The main differences of the 2004 model were weighting factors and effective sample sizes for the likelihood functions. In 2004, the weighting factors were 1,000 for survey biomass, 2,000 for retained catch biomass and 200 for bycatch biomasses. The effective sample sizes were set to be 200 for all proportion data but weighting factors of 5, 2, and 1 were also respectively applied to retained catch proportions, survey proportions and bycatch

proportions. Estimates of time series of abundance in 2004 were generally higher than those estimated after 2004 (Figure 29).

In 2005, to improve the fit for retained catch data, the weight for retained catch biomass was increased to 3,000 and the weight for retained catch proportions was increased to 6. All other weights were not changed. In 2006, all weights were re-configured. No weights were used for proportion data, and instead, effective sample sizes were set to 500 for retained catch, 200 for survey data, and 100 for bycatch data. Weights for biomasses were changed to 800 for retained catch, 300 for survey, and 50 for bycatch. The weights in 2007 were the same as 2006. Generally, estimates of time series of abundance in 2005 were slightly lower than in 2006 and 2007, and there were few differences between estimates in 2006 and 2007 (Figure 29).

In 2008, estimated coefficients of variation for survey biomass were used to compute likelihood values as suggested by the CPT in 2007. Thus, weights were re-configured to: 500 for retained catch biomass, 50 for survey biomass, and 20 for bycatch biomasses. Effective sample size was lowered to 400 for the retained catch data. These changes were necessary for the estimation to converge and for a relatively good balanced fit to both biomasses and proportion data. Also, sizes at 50% selectivities for all fisheries data were allowed to change annually, subject to a random walk pattern, for all assessments before 2008. The 2008 model did not allow annual changes in any fishery selectivities. Except for higher estimates of abundance during the late 1980s and early 1990s, estimates of time series of abundance in 2008 were generally close to those in 2006 and 2007 (Figure 29).

During 2009-2013, the model was extended to the data through 1968. No weighting factors were used for the NMFS survey biomass during 2009-2013 assessments. Since 2013, the model has fitted the data only back to 1975 for consistency with trawl survey data. Two levels of molting probabilities over time were used, shell conditions for males were combined, and length composition data of the BSFRF survey were used. In 2014 and 2015, the trawl survey time series were re-estimated and a trawl survey catchability was estimated for some models.

Model 19.3 with GMACS was used for 2020. Among many differences from previous models, one main difference is natural mortality structure. Natural mortalities for females are proportional to natural mortalities for males for model 19.3, and one less natural mortality parameter is estimated for females than the previous models. Model 19.3 results in relatively low abundance estimates in recent years.

Overall, both historical results (historic analysis) and the 2020 model results (retrospective analysis) performed reasonably well. No great overestimates or underestimates occurred as was observed in assessments for Pacific halibut (*Hippoglossus stenolepis*) (Parma 1993) and some eastern Bering Sea groundfish stocks (Zheng and Kruse 2002; Ianelli et al. 2003). Since the most recent model was not used to set TAC or overfishing limits until 2009, historical implications for management from the stock assessment errors cannot be



evaluated at the current time. However, management implications of the ADF&G stock assessment model were evaluated by Zheng and Kruse (2002).

Ratios of estimated retrospective recruitments to terminal estimates in 2020 as a function of number of years estimated in the model show converging to 1.0 as the number of years increases (Figure 28). Standard deviations of the ratios drop sharply from one year estimated in the model to two years (Figure 28), showing great uncertainty of recruitment estimates for terminal years. Based on these results, we suggest not using recruitment estimates in a terminal year for overfishing/overfished determination.

f. Uncertainty and sensitivity analyses

- i. Estimated standard deviations of parameters are summarized in Table 6 for models 19.3 and 19.3d. Estimated standard deviations of mature male biomass are listed in Table 9.
- ii. Probabilities for mature male biomass and OFL in 2020 were illustrated in Figures 30a and 30b for model 19.3 using the MCMC approach. The confidence intervals are quite narrow.
- iii. Probabilities for mature male biomass below the minimum threshold ( $0.5 * B_{35\%}$ ) in 2020 were plotted in Figure 31 for model 19.3d using the MCMC approach.
- iv. Sensitivity analysis for handling mortality rate was included in the SAFE report in May 2010. The baseline handling mortality rate for the directed pot fishery was set at 0.2. A 50% reduction and 100% increase respectively resulted in 0.1 and 0.4 as alternatives. Overall, a higher handling mortality rate resulted in slightly higher estimates of mature abundance, and a lower rate resulted in a minor reduction of estimated mature abundance. Differences of estimated legal male abundance and mature male biomass were small for these handling mortality rates.
- v. Sensitivity of weights. Sensitivity of weights was examined in the SAFE report in May 2010. Weights to biomasses (trawl survey biomass, retained catch biomass, and bycatch biomasses) were reduced to 50% or increased to 200% to examine their sensitivity to abundance estimates. Weights to the penalty terms (recruitment variation and sex ratio) were respectively reduced or increased. Overall, estimated biomasses were similar under different weights except during the mid-1970s. The variation of estimated biomasses in the mid-1970s was mainly caused by the changes in estimates of additional mortalities in the early 1980s.

g. Comparison of alternative models

These comparisons, based on the data through 2010, were reported in the SAFE report in May 2011. Estimating length proportions in the initial year (scenario 1a) resulted in a better fit of survey length compositions at an expense of 36 more parameters than model 1. Abundance and biomass estimates with model 1a were similar between models. Using only standard survey data (scenario 1b) resulted in a poorer fit of survey length compositions and biomass than scenarios using both standard and re-tow data (scenarios 1, 1a, and 1c) and had the lowest

likelihood value. Although the likelihood value was higher for using both standard survey and re-tow data for males (scenario 1) than using only standard survey for males (scenario 1c), estimated abundances and biomasses were almost identical. The higher likelihood value for scenario 1 over scenario 1c was due to trawl bycatch length compositions.

In the SAFE report in September 2020, seven models were compared. The population biomass estimates in 2020 were slightly higher than those in 2019. Absolute mature male biomasses for all models had a similar trend over time. Among the seven models, model estimated relative NMFS survey biomasses and mature biomasses were similar, especially for models 19.0a and 19.0b and for models 19.3 and 19.3a. Biomass estimates for model 19.0a and 19.0b were higher during recent years than the other five model scenarios. As expected, model 19.3b estimated a higher trawl survey catchability ( $>1.0$ ), thus resulting in overall lower absolute biomass estimates. Differences of biomass estimates between models 19.0a and 19.0b and models 19.3, 19.3a, 19.3l and 19.3h could largely be explained by different structures of natural mortality. All seven models fitted the catch and bycatch biomasses very well.

In this report (May 2021), eight models are compared. For negative likelihood value comparisons (Tables 5b and 5c), model 19.6 has a higher total likelihood value than the other models due to the high base natural mortality. In fact, the highest total likelihood value is resulted from a base  $M = 0.31$  for males. Between two pairs of models (models 19.3e and 19.3f and models 19.3g and 19.3i), either estimating additional CV for NMFS biomass or increased the CV value for the NMFS survey catchability prior increases the total likelihood value. Model 19.3d has similar total likelihood value to those of models 19.3e and 19.3f.

Among the eight models, model 19.3c is model 19.3 with updated pot fishery observer data, and model 19.3d is model 19.3c with a better cap of effective sample sizes for retained and total male catches. These three models fit the data similarly and have close model results. Thus, model 19.3d is preferred over models 19.3 and 19.3c. Model 19.3e has different NMFS survey catchabilities for males and females, thus with one additional estimated parameter over model 19.3d. Both models 19.3d and 19.3e fit the data similarly and have similar model results. The high CV value for the NMFS survey catchability prior for model 19.3f results in estimated NMFS survey catchability  $>1.0$  for females. Therefore, model 19.3d is preferred over models 19.3e and 19.3f. VAST-estimated NMFS survey biomasses are generally lower than area-swept estimates until 2014 (Figure 6) and higher during 2014-2019, and thus VAST models 19.3g and 19.3i result in higher mature male biomass estimates than the other models in recent years. With the declining biomass trend during recent years, higher mature male biomass during recent years with VAST models 19.3g and 19.3i would increase the risk of overfishing the stock. Thus, we suggest not adopting VAST models now. Model 19.6 is like a sensitivity study, and the very high base natural mortality would result in an extremely high  $F_{35\%}$  value. Overall, we recommend model 19.3d as the base model for overfishing definition determination in September 2021.

## F. Calculation of the OFL and ABC

1. Bristol Bay RKC is currently placed in Tier 3b (NPFMC 2007).
2. For Tier 3 stocks, estimated biological reference points include  $B_{35\%}$  and  $F_{35\%}$ . Estimated model parameters are used to conduct mature male biomass-per-recruit analysis.
3. Specification of the OFL:

The Tier 3 control rule formula is as follows:

$$\begin{aligned}
 \text{a) } \frac{B}{B^*} > 1 & \quad F_{OFL} = F^* \\
 \text{b) } \beta < \frac{B}{B^*} \leq 1 & \quad F_{OFL} = F^* \left( \frac{B/B^* - \alpha}{1 - \alpha} \right) \quad (2) \\
 \text{c) } \frac{B}{B^*} \leq \beta & \quad \text{directed pot fishery } F = 0 \text{ and } F_{OFL} \leq F^*
 \end{aligned}$$

Where

$B$  = a measure of the productive capacity of the stock such as spawning biomass or fertilized egg production. A proxy of  $B$  is mature male biomass (MMB) estimated at the time of primiparous female mating (February 15).

$F^* = F_{35\%}$ , a proxy of  $F_{MSY}$ , which is a full selection instantaneous  $F$  that will produce MSY at the MSY producing biomass.

$B^* = B_{35\%}$ , a proxy of  $B_{MSY}$ , which is the value of biomass at the MSY producing level.

$\beta$  = a parameter with a restriction that  $0 \leq \beta < 1$ . A default value of 0.25 is used.

$\alpha$  = a parameter with a restriction that  $0 \leq \alpha \leq \beta$ . A default value of 0.1 is used.

Because trawl bycatch fishing mortality is not related to pot fishing mortality, average trawl bycatch fishing mortality during 2015 to 2019 is used for the per recruit analysis as well as for projections in the next section. Some discards of legal males occurred after the Individual Fishery Quota (IFQ) fishery started in 2005, but the discard rates were much lower during 2007-2013 than in 2005 after the fishing industry minimized discards of legal males. However, due to high proportions of large oldshell males, the discard rate increased greatly in 2014. The current models estimate two levels of retained proportions before 2005 and after 2004. The retained proportions after 2004 and total male selectivities are used to represent current trends for per recruit analysis and projections. Average molting probabilities during 2014-2019 are used for per recruit analysis and projections. For the models in 2020, the averages are the same since they are constant over time during at least last 15 years.

Average recruitment during 1984-2019 is used to estimate  $B_{35\%}$  (Figure 12a). Estimated  $B_{35\%}$  is compared with historical mature male biomass in Figure 13a. The period of 1984-2019

corresponds to the 1976/77 regime shift, and the recruitment period 1984-present has been used since 2011 to set the overfishing limits. Several factors support our recommendation. First, estimated recruitment was lower after 1983 than before 1984, which corresponded to brood years 1978 and later, after the 1976/77 regime shift. Second, high recruitments during the late 1960s and 1970s generally occurred when the spawning stock was primarily located in the southern Bristol Bay, whereas the current spawning stock is mainly in the middle of Bristol Bay. The current flows favor larvae hatched in the southern Bristol Bay (see the section on Ecosystem Considerations for SAFE reports in 2008 and 2009). Finally, stock productivity (recruitment/mature male biomass) was higher before the 1976/1977 regime shift.

The control rule is used for stock status determination. If total catch exceeds OFL estimated at  $B$ , then “overfishing” occurs. If  $B$  equals or declines below  $0.5 B_{MSY}$  (i.e., MSST), the stock is “overfished.” If  $B/B_{MSY}$  or  $B/B_{MSY}$ -proxy equals or declines below  $\beta$ , then the stock productivity is severely depleted, and the directed fishery is closed.

The estimated probability distribution of MMB in 2020 is illustrated in Figure 30. Based on SSC suggestions in 2011,  $ABC = 0.9 * OFL$  and in October 2018,  $ABC = 0.8 * OFL$ . The CPT then recommended  $ABC = 0.8 * OFL$  in May 2018 (accepted by the SSC), which is used to estimate ABC in this report. Due to the stock close to overfished and lack of survey in 2020, the CPT recommended additional 5% buffer in September 2020, resulting in  $ABC = 0.75 * OFL$  for 2020.

MCMC runs with 500,000 replicates and 500 draws with model 19.3d are used for estimating the probability of estimated mature male biomass below the minimum threshold ( $0.5 * B_{35\%}$ ) (Figure 31). The probability (converted to a percentage) is estimated to be about 10.1%.

Status and catch specifications (1,000 t) (model 19.3):

Year	MSST	Biomass (MMB)	TAC	Retained Catch	Total Catch	OFL	ABC
2016/17	12.53 <sup>A</sup>	25.81 <sup>A</sup>	3.84	3.92	4.37	6.64	5.97
2017/18	12.74 <sup>B</sup>	24.86 <sup>B</sup>	2.99	3.09	3.60	5.60	5.04
2018/19	10.62 <sup>C</sup>	16.92 <sup>C</sup>	1.95	2.03	2.65	5.34	4.27
2019/20	12.72 <sup>D</sup>	14.24 <sup>D</sup>	1.72	1.78	2.22	3.40	2.72
2020/21		14.93 <sup>D</sup>				2.14	1.61

The stock was above MSST in 2019/20 and hence was not overfished. Since total catch was below OFL, overfishing did not occur. The relatively low MSST in 2018/19 and  $B_{MSY}$  in 2019/20 below was caused by a problem of the previous GMACS version using the only sex ratio of recruitment in the terminal year for  $B_{35\%}$  computation in 2019. The lower estimated male recruitment ratio in the terminal year in 2019 resulted in a lower mean male recruitment for  $B_{35\%}$  computation. The current version of GMACS uses average of sex ratios of recruitment during the reference period to estimate  $B_{35\%}$ , which results in a much more stable sex ratio (about 50%) for the reference point calculation.

Status and catch specifications (million lb):

Year	MSST	Biomass (MMB)	TAC	Retained Catch	Total Catch	OFL	ABC
2016/17	27.6 <sup>A</sup>	56.9 <sup>A</sup>	8.47	8.65	9.63	14.63	13.17
2017/18	28.1 <sup>B</sup>	54.8 <sup>B</sup>	6.60	6.82	7.93	12.35	11.11
2018/19	23.4 <sup>C</sup>	37.3 <sup>C</sup>	4.31	4.31	5.85	11.76	9.41
2019/20	28.0 <sup>D</sup>	31.4 <sup>D</sup>	3.80	3.91	4.89	7.50	6.00
2020/21		32.9 <sup>D</sup>				4.72	3.54

Notes:

A – Calculated from the assessment reviewed by the Crab Plan Team in September 2017

B – Calculated from the assessment reviewed by the Crab Plan Team in September 2018

C – Calculated from the assessment reviewed by the Crab Plan Team in September 2019

D – Calculated from the assessment reviewed by the Crab Plan Team in September 2020

Basis for the OFL: Values in 1,000 t (model 19.3):

Year	Tier	B <sub>MSY</sub>	Current MMB	B/B <sub>MSY</sub> (MMB)	F <sub>OFL</sub>	Years to define B <sub>MSY</sub>	Natural Mortality
2016/17	3b	25.8	24.0	0.93	0.27	1984-2016	0.18
2017/18	3b	25.1	21.3	0.85	0.24	1984-2017	0.18
2018/19	3b	25.5	20.8	0.82	0.25	1984-2017	0.18
2019/20	3b	21.2	16.0	0.75	0.22	1984-2018	0.18
2020/21	3b	25.4	14.9	0.59	0.16	1984-2019	0.18

Basis for the OFL: Values in million lb:

Year	Tier	B <sub>MSY</sub>	Current MMB	B/B <sub>MSY</sub> (MMB)	F <sub>OFL</sub>	Years to define B <sub>MSY</sub>	Natural Mortality
2016/17	3b	56.8	52.9	0.93	0.27	1984-2016	0.18
2017/18	3b	55.2	47.0	0.85	0.24	1984-2017	0.18
2018/19	3b	56.2	45.9	0.82	0.25	1984-2017	0.18
2019/20	3b	46.8	35.2	0.75	0.22	1984-2018	0.18
2020/21	3b	56.1	32.9	0.59	0.16	1984-2019	0.18

- Based on the  $B_{35\%}$  estimated from the average male recruitment during 1984-2019, the biological reference points and OFL are illustrated in Table 4.
- Based on the CPT/SSC recommendation of 20% buffer rule in May 2018 and an additional buffer of 5% for 2020 due to lack of survey by the CPT,  $ABC = 0.75 * OFL$  (Table 4).

## ***G. Rebuilding Analyses***

NA

## ***H. Data Gaps and Research Priorities***

1. The following data gaps exist for this stock:
  - a. Information about changes in natural mortality in the early 1980s,
  - b. Un-observed trawl bycatch in the early 1980s,
  - c. Natural mortality,
  - d. Crab availability to the trawl surveys,
  - e. Juvenile crab abundance,
  - f. Female growth per molt as a function of size and maturity,
  - g. Changes in male molting probability over time,
  - h. A better understanding of larval distribution and subsequent recruit distribution.
2. Research priorities:
  - a. Estimating natural mortality,
  - b. Estimating crab availability to the trawl surveys,
  - c. Surveying juvenile crab abundance in nearshore,
  - d. Studying environmental factors that affect the survival rates from larvae to recruitment.

## ***I. Projections and Future Outlook***

### **1. Projections**

Future population projections primarily depend on future recruitment, but crab recruitment is difficult to predict. Therefore, annual recruitment for the projections is a random selection from estimated recruitments during 2012-2019, a low recruitment period. Four levels of fishing mortality for the directed pot fishery are used in the projections: 0, 0.083, 0.167 and 0.25. Fishing mortality of 0.167 corresponds to estimated  $F_{0fl}$  in 2020. MCMC runs with 400,000 replicates and 500 draws are used for projection.

As expected, projected mature male biomasses are much higher without the directed fishing mortality than under other positive mortality values. At the end of 10 years, projected mature male biomass is below  $B_{35\%}$  for all models due to low recruitments (Table 10; Figure 32). Due to the poor recruitment in recent years, the projected biomass and retained catch are expected to decline during the next few years with fishing mortalities of 0.167 and 0.25.

### **2. Near Future Outlook**

The near future outlook for the Bristol Bay RKC stock is a declining trend. The three recent above-average year classes (hatching years 1990, 1994, and 1997) had entered the legal population by 2006 (Figure 33). Most individuals from the 1997-year class will continue to gain weight to offset loss of the legal biomass to fishing and natural mortalities. The above-average year class (hatching year 2000) with lengths centered around 87.5 mm CL for both males and females in 2006 and with lengths centered around 112.5-117.5 mm CL for males and around 107.5 mm CL for females in 2008 has largely entered the mature male population in 2009 and the legal population by 2014 (Figure 33). No strong cohorts were observed in the survey data after this cohort through 2010 (Figure 33). A huge tow of juvenile crab of size 45-55 mm in 2011 was not tracked during 2012-2019 surveys and is unlikely to be a strong cohort. The high survey abundance of large males and mature females in 2014 cannot be explained by the survey data during the previous years and were also inconsistent with the 2016-2019 survey results (Figure 33). Due to lack of recruitment, mature and legal crab should continue to decline next year. Current crab abundance is still low relative to the late 1970s, and without favorable environmental conditions, recovery to the high levels of the late 1970s is unlikely.

### ***J. Acknowledgements***

Drs. Andre Punt, James Ianelli, and D'Arcy Webber first applied BBRKC data to GMACS for stock assessments and our GMACS model mainly comes from their work. We thank the Crab Plan Team, William Bechtol, and Katie Palof for reviewing the earlier draft of this manuscript.

### ***K. Literature Cited***

- Alaska Department of Fish and Game (ADF&G). 2012. Commercial king and Tanner crab fishing regulations, 2012-2013. Alaska Department of Fish and Game, Division of Commercial Fisheries, Juneau. 170 pp.
- Balsiger, J.W. 1974. A computer simulation model for the eastern Bering Sea king crab. Ph.D. dissertation, Univ. Washington, Seattle, WA. 198 pp.
- Fitch, H., M. Deiman, J. Shaishnikoff, and K. Herring. 2012. Annual management report for the commercial shellfish fisheries of the Bering Sea, 2010/11. *In* Fitch, H. M. Schwenzfeier, B. Baechler, T. Hartill, M. Salmon, M. Deiman, E. Evans, E. Henry, L. Wald, J. Shaishnikoff, K. Herring, and J. Wilson. 2012. Annual management report for the commercial and subsistence fisheries of the Aleutian Islands, Bering Sea and the Westward Region's shellfish observer program, 2010/11. Alaska Department of Fish and Game, Fishery Management report No. 12-22, Anchorage.
- Fournier, D.A., J. Hampton, and J.R. Sibert. 1998. MULTIFAN-CL: a length-based, age-structured model for fisheries stock assessment, with application to South Pacific albacore, *Thunnus alalunga*. *Can.J.Fish.Aquat. Sci.*, 55: 2105-2116.
- Fournier, D.A., H.J. Skaug, J. Ancheta, J. Ianelli, A. Magnusson, M.N. Maunder, A. Nielsen, and J. Sibert. 2012. AD Model Builder: using automatic differentiation for statistical inference of highly parameterized complex nonlinear models. *Optim. Methods Softw.* 27: 233-249.

- Gaeuman, W.G. 2013. Summary of the 2012/13 mandatory crab observer program database for the Bering Sea/Aleutian Islands commercial crab fisheries. Alaska Department of Fish and Game, Fishery Data Series No. 13-54, Anchorage.
- Gray, G.W. 1963. Growth of mature female king crab *Paralithodes camtschaticus* (Tilesius). Alaska Dept. Fish and Game, Inf. Leaflet. 26.
- Griffin, K. L., M. F. Eaton, and R. S. Otto. 1983. An observer program to gather in-season and post-season on-the-grounds red king crab catch data in the southeastern Bering Sea. Contract 82-2, North Pacific Fishery Management Council, Anchorage.
- Haynes, E.B. 1968. Relation of fecundity and egg length to carapace length in the king crab, *Paralithodes camtschaticus*. Proc. Nat. Shellfish Assoc. 58: 60-62.
- Hoopes, D.T., J.F. Karinen, and M. J. Pelto. 1972. King and Tanner crab research. Int. North Pac. Fish. Comm. Annu. Rep. 1970: 110-120.
- Ianelli, J.N., S. Barbeaux, G. Walters, and N. Williamson. 2003. Eastern Bering Sea walleye pollock stock assessment. Pages 39-126 in Stock assessment and fishery evaluation report for the groundfish resources of the Bering Sea/Aleutian Islands regions. North Pacific Fishery Management Council, Anchorage.
- Jackson, P.B. 1974. King and Tanner crab fishery of the United States in the Eastern Bering Sea, 1972. Int. North Pac. Fish. Comm. Annu. Rep. 1972: 90-102.
- Loher, T., D.A. Armstrong, and B.G. Stevens. 2001. Growth of juvenile red king crab (*Paralithodes camtschaticus*) in Bristol Bay (Alaska) elucidated from field sampling and analysis of trawl-survey data. Fish. Bull. 99: 572-587.
- Matsuura, S., and K. Takeshita. 1990. Longevity of red king crab, *Paralithodes camtschaticus*, revealed by long-term rearing study. Pages 247-266 in Proceedings of the International Symposium on King and Tanner Crabs. University Alaska Fairbanks, Alaska Sea Grant College Program Report 90-04, Fairbanks.
- McCaughran, D.A., and G.C. Powell. 1977. Growth model for Alaskan king crab (*Paralithodes camtschaticus*). J. Fish. Res. Board Can. 34: 989-995.
- North Pacific Fishery Management Council (NPFMC). 2007. Environmental assessment for proposed amendment 24 to the fishery management plan for Bering Sea and Aleutian Islands king and Tanner crabs to revise overfishing definitions.
- Otto, R.S. 1989. An overview of eastern Bering Sea king and Tanner crab fisheries. Pages 9–26 in Proceedings of the International Symposium on King and Tanner Crabs, Alaska Sea Grant College Program Report No. 90-04.
- Parma, A.M. 1993. Retrospective catch-at-age analysis of Pacific halibut: implications on assessment of harvesting policies. Pages 247-266 in G. Kruse, D.M. Eggers, R.J. Marasco, C. Pautzke, and T.J. Quinn II (eds.). Proceedings of the international symposium on management strategies for exploited fish populations. University of Alaska Fairbanks, Alaska Sea Grant Rep. 90-04.



- Paul, J.M., and A.J. Paul. 1990. Breeding success of sublegal size male red king crab *Paralithodes camtschaticus* (Tilesius, 1815) (Decapoda, Lithodidae). *J. Shellfish Res.* 9: 29-32.
- Paul, J.M., A.J. Paul, R.S. Otto, and R.A. MacIntosh. 1991. Spermatophore presence in relation to carapace length for eastern Bering Sea blue king crab (*Paralithodes platypus*, Brandt, 1850) and red king crab (*P. camtschaticus*, Tilesius, 1815). *J. Shellfish Res.* 10: 157-163.
- Pengilly, D., S.F. Blau, and J.E. Blackburn. 2002. Size at maturity of Kodiak area female red king crab. Pages 213-224 in A.J. Paul, E.G. Dawe, R. Elner, G.S. Jamieson, G.H. Kruse, R.S. Otto, B. Sainte-Marie, T.C. Shirley, and D. Woodby (eds.). *Crabs in Cold Water Regions: Biology, Management, and Economics*. University of Alaska Sea Grant, AK-SG-02-01, Fairbanks.
- Pengilly, D., and D. Schmidt. 1995. Harvest strategy for Kodiak and Bristol Bay red king crab and St. Matthew Island and Pribilof Islands blue king crab. Alaska Dep. Fish and Game, Comm. Fish. Manage. and Dev. Div., Special Publication 7. Juneau, AK.
- Phinney, D.E. 1975. United States fishery for king and Tanner crabs in the eastern Bering Sea, 1973. *Int. North Pac. Fish. Comm. Annu. Rep.* 1973: 98-109.
- Powell, G.C. 1967. Growth of king crabs in the vicinity of Kodiak, Alaska. Alaska Dept. Fish and Game, Inf. Leaflet. 92. 106 pp.
- Powell, G. C., and R.B. Nickerson. 1965. Aggregations among juvenile king crab (*Paralithodes camtschaticus*, Tilesius) Kodiak, Alaska. *Animal Behavior* 13: 374-380.
- Schmidt, D., and D. Pengilly. 1990. Alternative red king crab fishery management practices: modeling the effects of varying size-sex restrictions and harvest rates, p.551-566. *In Proc. Int. Symp. King and Tanner Crabs, Alaska Sea Grant Rep.* 90-04.
- Sparks, A.K., and J.F. Morado. 1985. A preliminary report on diseases of Alaska king crabs, p.333-340. *In Proc. Int. Symp. King and Tanner Crabs, Alaska Sea Grant Rep.* 85-12.
- Stevens, B.G. 1990. Temperature-dependent growth of juvenile red king crab (*Paralithodes camtschaticus*), and its effects on size-at-age and subsequent recruitment in the eastern Bering Sea. *Can. J. Fish. Aquat. Sci.* 47: 1307-1317.
- Stevens, B.G., and K. Swiney. 2007. Hatch timing, incubation period, and reproductive cycle for primiparous and multiparous red king crab, *Paralithodes camtschaticus*. *J. Crust. Bio.* 27(1): 37-48.
- Swiney, K. M., W.C. Long, G.L. Eckert, and G.H. Kruse. 2012. Red king crab, *Paralithodes camtschaticus*, size-fecundity relationship, and interannual and seasonal variability in fecundity. *Journal of Shellfish Research*, 31:4, 925-933.
- Then, A. Y., J. M. Hoenig, N. G. Hall, and D. A. Hewitt. 2015. Evaluating the predictive performance of empirical estimators of natural mortality rate using information on over 200 fish species. – *ICES J. Mar. Sci.* 72: 82-92.
- Webb. J. 2014. Reproductive ecology of commercially important Lithodid crabs. Pages 285-314 In B.G. Stevens (ed.): *King Crabs of the World: Biology and Fisheries Management*. CRC Press, Taylor & Francis Group, New York.

- Weber, D.D. 1967. Growth of the immature king crab *Paralithodes camtschaticus* (Tilesius). Int. North Pac. Fish. Comm. Bull. 21:21-53.
- Weber, D.D., and T. Miyahara. 1962. Growth of the adult male king crab, *Paralithodes camtschaticus* (Tilesius). Fish. Bull. U.S. 62:53-75.
- Weinberg, K.L., R.S. Otto, and D.A. Somerton. 2004. Capture probability of a survey trawl for red king crab (*Paralithodes camtschaticus*). Fish. Bull. 102:740-749.
- Zheng, J. 2005. A review of natural mortality estimation for crab stocks: data-limited for every stock? Pages 595-612 in G.H. Kruse, V.F. Gallucci, D.E. Hay, R.I. Perry, R.M. Peterman, T.C. Shirley, P.D. Spencer, B. Wilson, and D. Woodby (eds.). Fisheries Assessment and Management in Data-limited Situation. Alaska Sea Grant College Program, AK-SG-05-02, Fairbanks.
- Zheng, J., and G.H. Kruse. 2002. Retrospective length-based analysis of Bristol Bay red king crabs: model evaluation and management implications. Pages 475-494 in A.J. Paul, E.G. Dawe, R. Elner, G.S. Jamieson, G.H. Kruse, R.S. Otto, B. Sainte-Marie, T.C. Shirley, and D. Woodby (eds.). Crabs in Cold Water Regions: Biology, Management, and Economics. University of Alaska Sea Grant, AK-SG-02-01, Fairbanks.
- Zheng, J., M.C. Murphy, and G.H. Kruse. 1995a. A length-based population model and stock-recruitment relationships for red king crab, *Paralithodes camtschaticus*, in Bristol Bay, Alaska. Can. J. Fish. Aquat. Sci. 52:1229-1246.
- Zheng, J., M.C. Murphy, and G.H. Kruse. 1995b. Updated length-based population model and stock-recruitment relationships for red king crab, *Paralithodes camtschaticus*, in Bristol Bay, Alaska. Alaska Fish. Res. Bull. 2:114-124.
- Zheng, J., M.C. Murphy, and G.H. Kruse. 1996. Overview of population estimation methods and recommended harvest strategy for red king crabs in Bristol Bay. Alaska Department of Fish and Game, Reg. Inf. Rep. 5J96-04, Juneau, Alaska. 37 pp.
- Zheng, J., M.C. Murphy, and G.H. Kruse. 1997a. Analysis of the harvest strategies for red king crab, *Paralithodes camtschaticus*, in Bristol Bay, Alaska. Can. J. Fish. Aquat. Sci. 54:1121-1134.
- Zheng, J., M.C. Murphy, and G.H. Kruse. 1997b. Alternative rebuilding strategies for the red king crab *Paralithodes camtschaticus* fishery in Bristol Bay, Alaska. J. Shellfish Res. 16:205-217.

Table 1a (19.3). Bristol Bay red king crab annual catch and bycatch mortality biomass (t) from July 1 to June 30. A handling mortality rate of 20% for the directed pot, 25% for the Tanner fishery, 80% for trawl, and 50% for fixed gear was assumed to estimate bycatch mortality biomass. Pot bycatch and Tanner crab fishery bycatch are estimated through expanding the mean observer bycatch per pot to total fishery pot. The pot male bycatch after 2017 is estimated through the subtraction method (B. Daly, ADF&G, personal communication). The trawl and fixed gear fishery bycatches are obtained from the NMFS database. The directed pot bycatch before 1990 and Tanner crab fishery bycatch before 1991 are not available from the observer data and thus not included in this table. These results were estimated in 2020.

Year	Retained Catch			Pot Bycatch		Trawl Bycatch	Fixed Bycatch	Tanner Fishery Bycatch	Total Catch
	U.S.	Cost-Recovery	Foreign	Total	Males				
1953	1331.3		4705.6	6036.9					6036.9
1954	1149.9		3720.4	4870.2					4870.2
1955	1029.2		3712.7	4741.9					4741.9
1956	973.4		3572.9	4546.4					4546.4
1957	339.7		3718.1	4057.8					4057.8
1958	3.2		3541.6	3544.8					3544.8
1959	0.0		6062.3	6062.3					6062.3
1960	272.2		12200.7	12472.9					12472.9
1961	193.7		20226.6	20420.3					20420.3
1962	30.8		24618.7	24649.6					24649.6
1963	296.2		24930.8	25227.0					25227.0
1964	373.3		26385.5	26758.8					26758.8
1965	648.2		18730.6	19378.8					19378.8
1966	452.2		19212.4	19664.6					19664.6
1967	1407.0		15257.0	16664.1					16664.1
1968	3939.9		12459.7	16399.6					16399.6
1969	4718.7		6524.0	11242.7					11242.7
1970	3882.3		5889.4	9771.7					9771.7
1971	5872.2		2782.3	8654.5					8654.5
1972	9863.4		2141.0	12004.3					12004.3
1973	12207.8		103.4	12311.2					12311.2
1974	19171.7		215.9	19387.6					19387.6
1975	23281.2		0	23281.2					23281.2
1976	28993.6		0	28993.6			682.8		29676.4
1977	31736.9		0	31736.9			1249.9		32986.8
1978	39743.0		0	39743.0			1320.6		41063.6
1979	48910.0		0	48910.0			1331.9		50241.9
1980	58943.6		0	58943.6			1036.5		59980.1
1981	15236.8		0	15236.8			219.4		15456.2
1982	1361.3		0	1361.3			574.9		1936.2
1983	0.0		0	0.0			420.4		420.4
1984	1897.1		0	1897.1			1094.0		2991.1
1985	1893.8		0	1893.8			390.1		2283.8
1986	5168.2		0	5168.2			200.6		5368.8
1987	5574.2		0	5574.2			186.4		5760.7
1988	3351.1		0	3351.1			598.4		3949.4
1989	4656.0		0	4656.0			175.2		4831.2
1990	9236.2	36.6	0	9272.8	526.9	648.0	259.9		10707.6
1991	7791.8	93.4	0	7885.1	407.8	47.3	349.4	1401.8	10091.5
1992	3648.2	33.6	0	3681.8	552.0	400.2	293.5	244.4	5172.0
1993	6635.4	24.1	0	6659.6	763.2	634.9	401.4	54.6	8513.6
1994	0.0	42.3	0	42.3			87.3	10.8	146.2
1995	0.0	36.4	0	36.4			82.1	0.0	123.3

1996	3812.7	49.0	0	3861.7	164.6	1.0	90.8	41.4	0.0	4159.6
1997	3971.9	70.2	0	4042.1	244.7	37.0	57.5	22.5	0.0	4403.7
1998	6693.8	85.4	0	6779.2	959.7	579.4	186.1	18.5	0.0	8522.8
1999	5293.5	84.3	0	5377.9	314.2	5.6	150.5	50.1	0.0	5898.3
2000	3698.8	39.1	0	3737.9	360.8	166.7	81.7	4.7	0.0	4351.9
2001	3811.5	54.6	0	3866.2	417.9	122.3	192.8	35.3	0.0	4634.4
2002	4340.9	43.6	0	4384.5	442.7	9.2	151.2	29.2	0.0	5016.8
2003	7120.0	15.3	0	7135.3	918.9	360.9	136.9	12.7	0.0	8564.7
2004	6915.2	91.4	0	7006.7	345.5	174.6	173.5	15.2	0.0	7715.5
2005	8305.0	94.7	0	8399.7	1359.5	410.3	124.7	19.9	0.0	10314.1
2006	7005.3	137.9	0	7143.2	563.8	37.5	151.7	19.6	3.8	7919.6
2007	9237.9	66.1	0	9303.9	1001.3	163.3	154.1	32.3	1.8	10656.8
2008	9216.1	0.0	0	9216.1	1165.5	146.9	136.6	15.6	4.0	10684.6
2009	7226.9	45.5	0	7272.5	888.1	93.7	95.1	5.8	1.6	8356.9
2010	6728.5	33.0	0	6761.5	797.5	121.8	83.3	2.4	0.0	7766.5
2011	3553.3	53.8	0	3607.1	395.0	24.7	56.3	10.9	0.0	4093.9
2012	3560.6	61.1	0	3621.7	205.2	12.0	34.2	18.4	0.0	3891.5
2013	3901.1	89.9	0	3991.0	310.6	102.9	67.1	55.5	28.5	4555.5
2014	4530.0	8.6	0	4538.6	584.7	72.4	34.8	118.8	42.0	5391.3
2015	4522.3	91.4	0	4613.7	266.1	216.3	45.3	77.4	84.2	5303.1
2016	3840.4	83.4	0	3923.9	237.4	105.4	67.3	28.9	0.0	4362.9
2017	2994.1	99.6	0	3093.7	225.2	53.3	91.8	127.6	0.0	3591.6
2018	1954.1	72.4	0	2026.5	279.6	114.8	78.3	148.0	0.0	2647.2
2019	1719.8	55.5	0	1775.3	273.8	43.3	80.8	45.1	0.0	2218.3

Table 1a (19.3d). Bristol Bay red king crab annual catch and bycatch mortality biomass (t) from July 1 to June 30. A handling mortality rate of 20% for the directed pot, 25% for the Tanner fishery, 80% for trawl, and 50% for fixed gear was assumed to estimate bycatch mortality biomass. The male bycatch biomass in the directed pot fishery is not estimated outside of a model and not included in this Table. Pot bycatch and Tanner crab fishery bycatch are estimated through expanding the mean observer bycatch per pot to total fishery pot. The pot male bycatch after 2017 is estimated through the subtraction method (B. Daly, ADF&G, personal communication). The trawl and fixed gear fishery bycatches are obtained from the NMFS database. The directed pot bycatch before 1990 and Tanner crab fishery bycatch before 1991 are not available from the observer data and thus not included in this table. These include recently updated estimates from the pot fisheries observer data and are used for models 19.3c-19.6 in 2021.

Year	Retained Catch			Pot Bycatch		Trawl Bycatch	Fixed Bycatch	Tanner Fishery Bycatch
	U.S.	Cost-Recovery	Foreign	Total	Females			
1953	1331.3		4705.6	6036.9				
1954	1149.9		3720.4	4870.2				
1955	1029.2		3712.7	4741.9				
1956	973.4		3572.9	4546.4				
1957	339.7		3718.1	4057.8				
1958	3.2		3541.6	3544.8				
1959	0.0		6062.3	6062.3				
1960	272.2		12200.7	12472.9				
1961	193.7		20226.6	20420.3				
1962	30.8		24618.7	24649.6				
1963	296.2		24930.8	25227.0				
1964	373.3		26385.5	26758.8				
1965	648.2		18730.6	19378.8				
1966	452.2		19212.4	19664.6				
1967	1407.0		15257.0	16664.1				
1968	3939.9		12459.7	16399.6				
1969	4718.7		6524.0	11242.7				
1970	3882.3		5889.4	9771.7				
1971	5872.2		2782.3	8654.5				
1972	9863.4		2141.0	12004.3				
1973	12207.8		103.4	12311.2				
1974	19171.7		215.9	19387.6				
1975	23281.2		0	23281.2				
1976	28993.6		0	28993.6		682.8		
1977	31736.9		0	31736.9		1249.9		
1978	39743.0		0	39743.0		1320.6		
1979	48910.0		0	48910.0		1331.9		
1980	58943.6		0	58943.6		1036.5		
1981	15236.8		0	15236.8		219.4		
1982	1361.3		0	1361.3		574.9		
1983	0.0		0	0.0		420.4		
1984	1897.1		0	1897.1		1094.0		
1985	1893.8		0	1893.8		390.1		
1986	5168.2		0	5168.2		200.6		
1987	5574.2		0	5574.2		186.4		
1988	3351.1		0	3351.1		598.4		
1989	4656.0		0	4656.0		175.2		
1990	9236.2	36.6	0	9272.8	648.0	259.9		
1991	7791.8	93.4	0	7885.1	47.3	349.4		1401.8
1992	3648.2	33.6	0	3681.8	400.2	293.5		244.4

1993	6635.4	24.1	0	6659.6	634.9	401.4		54.6
1994	0.0	42.3	0	42.3		87.3		10.8
1995	0.0	36.4	0	36.4		82.1		0.0
1996	3812.7	49.0	0	3861.7	1.0	90.8	41.4	0.0
1997	3971.9	70.2	0	4042.1	37.0	57.5	22.5	0.0
1998	6693.8	85.4	0	6779.2	579.4	186.1	18.5	0.0
1999	5293.5	84.3	0	5377.9	5.6	150.5	50.1	0.0
2000	3698.8	39.1	0	3737.9	166.7	81.7	4.7	0.0
2001	3811.5	54.6	0	3866.2	122.3	192.8	35.3	0.0
2002	4340.9	43.6	0	4384.5	9.2	151.2	29.2	0.0
2003	7120.0	15.3	0	7135.3	360.9	136.9	12.7	0.0
2004	6915.2	91.4	0	7006.7	174.6	173.5	15.2	0.0
2005	8305.0	94.7	0	8399.7	410.3	124.7	19.9	0.0
2006	7005.3	137.9	0	7143.2	37.5	151.7	19.6	3.8
2007	9237.9	66.1	0	9303.9	163.3	154.1	32.3	1.8
2008	9216.1	0.0	0	9216.1	146.9	136.6	15.6	4.0
2009	7226.9	45.5	0	7272.5	93.7	95.1	5.8	1.6
2010	6728.5	33.0	0	6761.5	121.8	83.3	2.4	0.0
2011	3553.3	53.8	0	3607.1	24.7	56.3	10.9	0.0
2012	3560.6	61.1	0	3621.7	12.0	34.2	18.4	0.0
2013	3901.1	89.9	0	3991.0	102.9	67.1	55.5	28.5
2014	4530.0	8.6	0	4538.6	72.4	34.8	118.8	42.0
2015	4522.3	91.4	0	4613.7	216.3	45.3	77.4	84.2
2016	3840.4	83.4	0	3923.9	105.4	67.3	28.9	0.0
2017	2994.1	99.6	0	3093.7	53.3	91.8	127.6	0.0
2018	1954.1	72.4	0	2026.5	114.8	78.3	148.0	0.0
2019	1719.8	55.5	0	1775.3	43.3	80.8	45.1	0.0

Table 1b. Annual retained catch (millions of crab) and catch per unit effort of the Bristol Bay red king crab fishery.

Year	Japanese Tanglenet		Russian Tanglenet		U.S. Pot		Standardized Crab/tan
	Catch	Crab/tan	Catch	Crab/tan	Catch	Crab/Potlift	
1960	1.949	15.2	1.995	10.4	0.088		15.8
1961	3.031	11.8	3.441	8.9	0.062		12.9
1962	4.951	11.3	3.019	7.2	0.010		11.3
1963	5.476	8.5	3.019	5.6	0.101		8.6
1964	5.895	9.2	2.800	4.6	0.123		8.5
1965	4.216	9.3	2.226	3.6	0.223		7.7
1966	4.206	9.4	2.560	4.1	0.140	52	8.1
1967	3.764	8.3	1.592	2.4	0.397	37	6.3
1968	3.853	7.5	0.549	2.3	1.278	27	7.8
1969	2.073	7.2	0.369	1.5	1.749	18	5.6
1970	2.080	7.3	0.320	1.4	1.683	17	5.6
1971	0.886	6.7	0.265	1.3	2.405	20	5.8
1972	0.874	6.7			3.994	19	
1973	0.228				4.826	25	
1974	0.476				7.710	36	
1975					8.745	43	
1976					10.603	33	
1977					11.733	26	
1978					14.746	36	
1979					16.809	53	
1980					20.845	37	
1981					5.308	10	
1982					0.541	4	
1983					0.000		
1984					0.794	7	
1985					0.796	9	
1986					2.100	12	
1987					2.122	10	
1988					1.236	8	
1989					1.685	8	
1990					3.130	12	
1991					2.661	12	
1992					1.208	6	
1993					2.270	9	
1994					0.015		
1995					0.014		
1996					1.264	16	
1997					1.338	15	
1998					2.238	15	
1999					1.923	12	
2000					1.272	12	
2001					1.287	19	
2002					1.484	20	
2003					2.510	18	
2004					2.272	23	
2005					2.763	30	
2006					2.477	31	
2007					3.154	28	
2008					3.064	22	
2009					2.553	21	
2010					2.410	18	
2011					1.298	28	
2012					1.176	30	
2013					1.272	27	
2014					1.501	26	

2015	1.527	31
2016	1.281	38
2017	0.997	20
2018	0.630	20
2019	0.549	16

Table 2a (19.3). Total observer catch and bycatch (metric ton) of Bristol Bay red king crab. No handling mortality rates are applied. These results were estimated in 2020.

Year	Total		Pot Bycatch		Trawl Bycatch	Fixed Bycatch	Tanner Bycatch
	Males		Males	Females			
1975					0.000		
1976					853.494		
1977					1,562.313		
1978					1,650.775		
1979					1,664.925		
1980					1,295.625		
1981					274.229		
1982					718.610		
1983					525.554		
1984					1,367.550		
1985					487.576		
1986					250.758		
1987					233.045		
1988					747.996		
1989					219.023		
1990	11,782.900		2,634.570	3,240.200	324.883		
1991	9,974.000		2,039.120	236.600	436.783		5,607.344
1992	6,013.700		2,760.045	2,001.200	366.816		977.750
1993	9,667.700		3,815.785	3,174.400	501.770		218.570
1994	42.300		19.060	9.383	109.129		43.366
1995	36.400		16.369	8.058	102.623		0.000
1996	3,902.300		823.180	5.200	113.495	82.859	0.000
1997	3,847.200		1,223.435	184.800	71.862	44.979	0.000
1998	17,681.400		4,798.560	2,897.100	232.580	36.916	0.000
1999	12,245.200		1,570.855	28.200	188.101	100.242	0.000
2000	6,672.300		1,804.165	833.700	102.161	9.446	0.000
2001	5,797.000		2,089.375	611.400	241.011	70.553	0.000
2002	7,065.300		2,213.290	46.100	189.018	58.382	0.000
2003	12,300.600		4,594.290	1,804.700	171.114	25.351	0.000
2004	10,816.800		1,727.745	873.000	216.889	30.422	0.000
2005	13,753.300		6,797.650	2,051.400	155.924	39.802	0.000
2006	9,170.400		2,818.755	187.700	189.660	39.134	15.232
2007	13,956.600		5,006.550	816.700	192.571	64.655	7.169
2008	15,068.700		5,827.550	734.400	170.754	31.158	15.938
2009	12,300.300		4,440.620	468.500	118.906	11.616	6.499
2010	10,087.400		3,987.380	609.200	104.086	4.736	0.000
2011	5,732.600		1,974.810	123.400	70.419	21.706	0.000
2012	4,568.100		1,025.775	59.800	42.786	36.895	0.000
2013	5,260.700		1,552.895	514.300	83.868	110.970	113.848
2014	8,312.700		2,923.280	362.200	43.460	237.651	168.080
2015	6,706.400		1,330.705	1,081.600	56.686	154.810	336.715
2016	5,557.200		1,187.083	527.000	84.127	57.896	0.000
2017	4,075.760		1,126.025	266.546	114.784	255.155	0.000
2018	3,060.344		1,398.089	574.045	97.891	295.916	0.000
2019	3,143.250		1,369.039	216.739	101.001	90.109	0.000



Table 2b (19.3d). Total observer catch and bycatch (metric ton) of Bristol Bay red king crab. No handling mortality rates are applied. These include recently updated estimates from the pot fishery observer data and are used for models 19.3c-19.6 in 2021.

Year	Directed Pot Total		Trawl	Fixed	Tanner
	Males	Females	Bycatch	Bycatch	Bycatch
1975			0.000		
1976			853.494		
1977			1,562.313		
1978			1,650.775		
1979			1,664.925		
1980			1,295.625		
1981			274.229		
1982			718.610		
1983			525.554		
1984			1,367.550		
1985			487.576		
1986			250.758		
1987			233.045		
1988			747.996		
1989			219.023		
1990	11,621.800	3,196.200	324.883		
1991	9,792.900	233.900	436.783		5,580.843
1992	5,916.200	1,976.300	366.816		962.846
1993	9,516.800	3,141.500	501.770		218.112
1994			109.129		39.395
1995			102.623		0.000
1996	3,845.200	5.100	113.495	82.859	0.000
1997	3,758.800	182.700	71.862	44.979	0.000
1998	15,644.800	2,769.300	232.580	36.916	0.000
1999	12,112.300	28.000	188.101	100.242	0.000
2000	6,579.700	821.900	102.161	9.446	0.000
2001	5,711.500	604.000	241.011	70.553	0.000
2002	6,961.400	45.600	189.018	58.382	0.000
2003	12,166.500	1,784.400	171.114	25.351	0.000
2004	10,692.000	859.200	216.889	30.422	0.000
2005	13,615.900	2,027.100	155.924	39.802	0.000
2006	9,254.000	187.400	189.660	39.134	15.217
2007	13,871.900	799.400	192.571	64.655	7.142
2008	14,895.900	723.000	170.754	31.158	16.070
2009	12,215.600	441.100	118.906	11.616	6.499
2010	10,092.100	592.200	104.086	4.736	0.000
2011	5,668.700	124.700	70.419	21.706	0.000
2012	4,506.400	55.800	42.786	36.895	0.000
2013	5,317.900	490.200	83.868	110.970	113.063
2014	8,113.800	424.300	43.460	237.651	137.786
2015	6,734.900	1,193.000	56.686	154.810	639.573
2016	5,651.800	617.200	84.127	57.896	0.000
2017	4,077.200	266.900	114.784	255.155	0.000

2018	3,423.200	750.400	97.891	295.916	0.000
2019	3,144.600	218.000	101.001	90.109	0.000

Table 3a (19.3). Annual sample sizes (>64 mm CL) in numbers of crab for trawl surveys, retained catch, directed pot, Tanner crab, trawl, and fixed gear fishery bycatches of Bristol Bay red king crab. The data were compiled in 2020.

Year	Trawl Survey		Retained Catch	Pot Total	Pot Bycatch	Trawl & Fixed Gear Bycatch		Tanner Fishery Bycatch	
	Males	Females		Males	Females	Males	Females	Males	Females
1975	2,815	2,042	29,570						
1976	2,699	1,466	26,450			676	2,327		
1977	2,734	2,424	32,596			689	14,014		
1978	2,735	2,793	27,529			1,456	8,983		
1979	1,158	1,456	27,900			2,821	7,228		
1980	1,917	1,301	34,747			39,689	47,463		
1981	591	664	18,029			49,634	42,172		
1982	1,911	1,948	11,466			47,229	84,240		
1983	1,343	733	0			104,910	204,464		
1984	1,209	778	4,404			147,134	357,981		
1985	790	414	4,582			30,693	169,767		
1986	959	341	5,773			1,199	927		
1987	1,123	1,011	4,230			723	275		
1988	708	478	9,833			437	194		
1989	764	403	32,858			3,140	1,566		
1990	729	535	7,218	2,571	1,416	756	375		
1991	1,180	490	36,820	5,024	366	236	90	885	2,198
1992	509	357	23,552	4,769	3,238	212	228	280	685
1993	725	576	32,777	10,334	6,187	24	3	232	265
1994	416	239	0	0	0	327	245		
1995	685	407	0	0	0	120	40		
1996	755	753	8,896	1,778	11	1,035	971		
1997	1,280	702	15,747	11,089	939	1,200	445		
1998	1,067	1,123	16,131	31,432	10,236	1,623	913		
1999	765	618	17,666	13,519	57	2,025	843		
2000	734	730	14,091	32,711	8,470	957	661		
2001	599	736	12,854	26,460	5,474	3,444	2,406		
2002	972	826	15,932	32,612	714	3,262	1,435		
2003	1,360	1,250	16,212	45,583	12,971	1,518	1,008		
2004	1,852	1,271	20,038	38,782	6,667	1,656	1,508		
2005	1,198	1,563	21,938	94,794	26,824	1,814	1,871		
2006	1,178	1,432	18,027	66,529	3,646	1,461	1,979		
2007	1,228	1,305	22,387	111,575	12,457	1,018	1,099		
2008	1,228	1,183	14,567	90,331	8,737	1,794	979		
2009	837	941	16,708	92,616	6,050	1,424	853		
2010	708	1,004	20,137	66,659	6,862	612	843		
2011	531	912	10,706	40,226	1,752	563	1,071		
2012	585	707	8,956	20,161	562	1,507	1,752		
2013	647	569	10,197	30,261	6,070	4,806	4,198	218	596
2014	1,107	1,257	9,618	28,540	1,953	1,966	2,580	256	381
2015	615	681	11,746	22,022	5,927	1,150	3,731	726	2,163

2016	378	812	10,811	26,510	4,315	1,935	3,011
2017	385	508	9,867	27,219	3,834	996	1,137
2018	285	359	7,626	22,480	7,386	2,806	3,389
2019	273	299	8,034	21,712	2,819	713	909

Table 3b (19.3d). Annual sample sizes (>64 mm CL) in numbers of crab for trawl surveys, retained catch, directed pot, Tanner crab, trawl, and fixed gear fishery bycatches of Bristol Bay red king crab. These include recently updated estimates from the pot fisheries observer data and are used for models 19.3c-19.6 in 2021.

Year	Trawl Survey		Retained Catch	Pot Total	Pot Bycatch	Trawl Bycatch	Fixed G. Bycatch	Tanner Fishery Bycatch
	Males	Females		Males	Females	Combined	Combined	Combined
1975	2,815	2,042	29,570					
1976	2,699	1,466	26,450			3,003		
1977	2,734	2,424	32,596			14,703		
1978	2,735	2,793	27,529			10,439		
1979	1,158	1,456	27,900			10,049		
1980	1,917	1,301	34,747			87,152		
1981	591	664	18,029			91,806		
1982	1,911	1,948	11,466			131,469		
1983	1,343	733	0			309,374		
1984	1,209	778	4,404			505,115		
1985	790	414	4,582			200,460		
1986	959	341	5,773			2,126		
1987	1,123	1,011	4,230			998		
1988	708	478	9,833			631		
1989	764	403	32,858			4,706		
1990	729	535	7,218	2,483	696	1,131		
1991	1,180	490	36,928	4,693	375	326		3,086
1992	509	357	25,550	4,747	2,379	440		965
1993	725	576	32,942	10,199	5,944	27		497
1994	416	239	0	0	0	572		13
1995	685	407	0	0	0	160		
1996	755	753	8,896	642	11	1,226	780	
1997	1,280	702	16,143	9,971	903	349	1,296	
1998	1,067	1,123	17,116	24,479	9,600	1,445	1,091	
1999	765	618	18,685	6,887	40	643	2,225	
2000	734	730	14,143	32,709	8,470	734	884	
2001	599	736	13,735	25,135	5,435	802	5,048	
2002	972	826	16,837	32,315	705	1,142	3,555	
2003	1,360	1,250	18,178	44,600	12,473	525	2,001	
2004	1,852	1,271	22,465	38,738	6,666	666	2,498	
2005	1,198	1,563	27,971	94,591	26,775	1,043	2,642	
2006	1,178	1,432	18,451	73,112	3,980	1,180	2,260	140
2007	1,228	1,305	22,809	115,364	12,661	1,265	852	53
2008	1,228	1,183	24,997	89,765	8,488	1,609	1,164	144
2009	837	941	19,336	97,574	6,039	1,188	1,089	193
2010	708	1,004	20,347	69,264	6,868	907	548	
2011	531	912	10,904	43,012	1,920	443	1,191	
2012	585	707	9,084	21,364	563	282	2,977	
2013	647	569	10,396	32,474	6,030	481	8,523	814
2014	1,107	1,257	9,718	31,192	2,663	261	4,285	637
2015	615	681	11,971	24,589	7,422	409	4,472	2,889
2016	378	812	11,003	30,005	5,734	617	4,329	

2017	385	508	10,067	29,821	3,939	718	1,415
2018	285	359	7,825	25,573	9,718	893	5,302
2019	273	299	8,134	25,991	2,879	713	909

Table 4. Comparison of harmonic means of implied sample sizes and maximum caps (N) of effective sample sizes for models 19.3, 19.3c and 19.3d.

	Model 19.3		Model 19.3c		Model 19.3d	
	Harmonic mean	N	Harmonic mean	N	Harmonic mean	N
Retained catch	177.9	100	182.5	100	190.7	150
Pot total males	199.2	100	225.7	100	238.4	150
Pot total females	30.4	50	30.0	50	30.0	50
Trawl bycatch	62.0	50	61.7	50	61.7	50
Tanner fishery bycatch	65.6	50	25.5	50	25.4	50
Fixed gear bycatch	46.0	50	46.8	50	46.8	50
NMFS survey	181.1	200	180.6	200	179.4	200
BSFRF survey	133.5	200	133.4	200	132.1	200

Table 5a. Number of parameters for the model (Models 19.3, 19.3c, 19.3d, 19.3e, 19.3f, 19.3g, 19.3i, and 19.6). Red values indicate different values among models.

<b>Parameter counts</b>	<b>19.3</b>	<b>19.3c</b>	<b>19.3d</b>	<b>19.3e</b>	<b>19.3f</b>	<b>19.3g</b>	<b>19.3i</b>	<b>19.6</b>
Fixed growth parameters	9	9	9	9	9	9	9	9
Fixed recruitment parameters	2	2	2	2	2	2	2	2
Fixed length-weight relationship parameters	6	6	6	6	6	6	6	6
Fixed mortality parameters	4	4	4	4	4	4	4	4
Fixed survey catchability parameter	1	1	1	1	1	1	1	1
Fixed high grading parameters	0	0	0	0	0	0	0	0
Total number of fixed parameters	22	22	22	22	22	22	22	22
Free survey catchability parameter	1	1	1	2	2	1	1	1
Free growth parameters	6	6	6	6	6	6	6	6
Initial abundance (1975)	1	1	1	1	1	1	1	1
Recruitment-distribution parameters	2	2	2	2	2	2	2	2
Mean recruitment parameters	1	1	1	1	1	1	1	1
Male recruitment deviations	45	45	45	45	45	45	45	45
Female recruitment deviations	45	45	45	45	45	45	45	45
Natural mortality parameters	2	2	2	2	2	2	2	2
Mean & offset fishing mortality parameters	6	6	6	6	6	6	6	6
Pot male fishing mortality deviations	45	45	45	45	45	45	45	45
Bycatch mortality from the Tanner crab fishery	50	50	50	50	50	50	50	50
Pot female bycatch fishing mortality deviations	30	30	30	30	30	30	30	30
Trawl bycatch fishing mortality deviations	44	44	44	44	44	44	44	44
Fixed gear bycatch fishing mortality deviations	24	24	24	24	24	24	24	24
Initial (1975) length compositions	35	35	35	35	35	35	35	35
Survey extra CV	1	1	1	1	1	1	2	1
Free selectivity parameters	28	28	28	28	28	28	28	28
Total number of free parameters	366	366	366	367	367	366	367	366
Total number of fixed and free parameters	388	388	388	389	389	388	389	388

Table 5b. Negative log likelihood components for Models 119.3, 19.3c, 19.3d, 19.3e, 19.3f, 19.3g, 19.3i, and 19.6 and some management quantities. Highlighted cells in yellow color show prior density values and total negative likelihood values without prior density. Highlighted cells in red color show values not comparable to the other models.

	Models							
	19.3	19.3c	19.3d	19.3e	19.3f	19.3g	19.3i	19.6
Pot-ret-catch	-59.87	-58.38	-57.32	-57.03	-57.67	-54.05	-56.74	-59.19
Pot-totM-catch	25.90	28.58	30.61	31.17	29.99	32.61	31.40	28.20
Pot-F-discC	-52.21	-52.21	-52.21	-52.21	-52.21	-52.19	-52.21	-52.23
Trawl-discC	-60.98	-60.98	-60.98	-60.98	-60.98	-60.97	-60.98	-60.98
Tanner-M-discC	-43.54	-43.54	-43.54	-43.54	-43.54	-43.54	-43.54	-43.54
Tanner-F-discC	-43.49	-43.47	-43.47	-43.48	-43.47	-43.44	-43.48	-43.51
Fixed-discC	-33.27	-33.27	-33.27	-33.27	-33.27	-33.26	-33.27	-33.27
Trawl-suv-bio	-33.82	-33.68	-33.38	-30.19	-30.59	104.60	-61.31	-35.71
BSFRF-sur-bio	-4.80	-4.71	-4.59	-3.02	-3.84	-6.19	-5.81	-6.56
Pot-ret-comp	-3643.89	-3645.56	-3787.11	-3787.12	-3787.02	-3789.11	-3788.56	-3793.65
Pot-totM-comp	-2150.62	-2150.73	-2223.64	-2223.09	-2224.19	-2219.77	-2224.48	-2224.66
Pot-discF-comp	-1353.14	-1346.66	-1346.70	-1346.68	-1346.74	-1345.73	-1346.88	-1346.10
Trawl-disc-comp	-5583.78	-5583.87	-5582.35	-5582.80	-5583.19	-5586.35	-5589.49	-5588.38
TC-disc-comp	-790.17	-1272.75	-1273.08	-1272.84	-1273.43	-1271.77	-1274.29	-1276.23
Fixed-disc-comp	-3168.76	-3171.53	-3171.32	-3171.47	-3170.67	-3169.04	-3171.91	-3171.00
Trawl-sur-comp	-6717.35	-6714.06	-6710.20	-6708.68	-6711.97	-6696.49	-6701.23	-6715.51
BSFRF-sur-comp	-851.44	-851.60	-851.13	-850.76	-851.97	-852.22	-852.22	-858.23
Recruit-dev	67.03	67.22	67.11	67.05	67.09	67.93	65.71	68.79
Recruit-sex-R	73.72	73.71	73.73	73.74	73.74	73.87	73.75	73.74
Log_fdev=0	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00
M-deviation	44.12	44.27	44.41	44.38	44.41	44.92	44.76	40.00
Sex-specific-R	0.06	0.05	0.05	0.05	0.05	0.08	0.01	0.00
Ini-size-struct.	31.46	31.11	31.65	31.42	31.80	30.90	32.68	38.29
PriorDensity	297.16	291.06	290.91	284.87	289.53	286.97	285.37	255.70
Tot-likelihood	-24051.7	-24531.0	-24735.8	-24737.4	-24739.6	-24582.2	-24770.4	-24804.0
Tot-likeli-no-PD	-24348.9	-24822.1	-25026.7	-25022.3	-25029.1	-24869.2	-25055.8	-25059.7
Tot-parameter	366	366	366	367	367	366	367	366
MMB35%	25444.7	25283.9	25220.6	25550.9	24790.2	25533.8	25571.1	21577.5
MMB-terminal	14928.4	14743.0	14802.0	15366.3	14152.8	16538.3	17876.8	12866.6
F35%	0.291	0.292	0.293	0.293	0.292	0.293	0.293	0.443
Fofl	0.157	0.157	0.158	0.163	0.153	0.178	0.195	0.244
OFL	2140.72	2099.17	2114.53	2266.73	1950.18	1950.18	3120.54	2714.94
ABC	1605.54	1574.38	1585.90	1700.05	1462.63	1462.63	2340.41	2036.21
Q-male	0.959	0.958	0.962	0.921	0.987	0.949	0.937	0.930
Q-female	0.959	0.958	0.962	0.945	1.027	0.949	0.937	0.930

Table 5c. Differences of negative log likelihood components and some management quantities between model 19.3 and models 19.3c and 19.3d, and between model 19.3d and models 19.3c, 19.3e, 19.3f, 19.3g, 19.3i, and 19.6. Highlighted cells in red color show values not comparable to the other models.

	19.3 - 19.3c	19.3 - 19.3d	19.3d - 19.3c	19.3d- 19.3e	19.3d- 19.3f	19.3d- 19.3g	19.3d- 19.3i	19.3d- 19.6
Pot-ret-catch	-1.49	-2.54	1.05	-0.29	0.35	-3.28	-0.59	1.86
Pot-totM-catch	-2.68	-4.71	2.03	-0.56	0.62	-2.00	-0.79	2.41
Pot-F-discC	0.00	0.00	0.00	0.00	0.00	-0.02	0.00	0.02
Trawl-discC	0.00	0.00	0.00	0.00	0.00	-0.01	0.00	0.00
Tanner-M-discC	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00
Tanner-F-discC	-0.01	-0.01	0.00	0.00	-0.01	-0.03	0.00	0.04
Fixed-discC	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00
Traw-suv-bio	-0.14	-0.44	0.30	-3.19	-2.79	-137.98	27.93	2.33
BSFRF-sur-bio	-0.09	-0.20	0.12	-1.57	-0.75	1.60	1.22	1.97
Pot-ret-comp	1.67	143.22	-141.55	0.01	-0.09	2.00	1.45	6.54
Pot-totM-comp	0.11	73.02	-72.91	-0.55	0.55	-3.87	0.84	1.02
Pot-discF-comp	-6.48	-6.44	-0.04	-0.02	0.04	-0.97	0.18	-0.60
Trawl-disc-comp	0.09	-1.43	1.52	0.45	0.84	4.00	7.14	6.03
Tanner-disc-comp	482.58	482.91	-0.33	-0.24	0.35	-1.31	1.21	3.15
Fixed-disc-comp	2.77	2.56	0.21	0.15	-0.65	-2.28	0.59	-0.32
Trawl-sur-comp	-3.29	-7.15	3.86	-1.52	1.77	-13.71	-8.97	5.31
BSFRF-sur-comp	0.16	-0.32	0.47	-0.36	0.85	1.09	1.10	7.11
Recruit-dev	-0.20	-0.09	-0.11	0.06	0.02	-0.82	1.40	-1.68
Recruit-sex-R	0.01	-0.01	0.02	-0.01	-0.01	-0.14	-0.03	-0.01
Log_fdev=0	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00
M-deviation	-0.15	-0.29	0.14	0.04	0.00	-0.51	-0.35	4.41
Sex-specific-R	0.02	0.02	0.00	0.00	0.00	-0.03	0.04	0.05
Ini-size-structure	0.36	-0.19	0.54	0.23	-0.15	0.76	-1.02	-6.63
PriorDensity	6.09	6.25	-0.16	6.04	1.38	3.93	5.53	35.21
Tot-likelihood	479.3	684.1	-47.6	1.60	3.80	-153.60	34.60	68.20
Tot-like-no-PD	473.2	677.8	-52.3	-4.44	2.42	-157.53	29.07	32.99
Tot-parameter	0	0	0	-1	-1	0	-1	0
MMB35%	160.79	224.03	-63.25	-330.27	430.40	-313.18	-350.45	3643.18
MMB-terminal	185.36	126.40	58.96	-564.33	649.25	-1736.3	-3074.8	1935.36
F35%	0.00	0.00	0.00	0.00	0.00	0.00	0.00	-0.15
Fofl	0.00	0.00	0.00	-0.01	0.01	-0.02	-0.04	-0.09
OFL	41.55	26.19	15.36	-152.20	164.35	164.35	-1006.0	-600.41
ABC	31.16	19.64	11.52	-114.15	123.27	123.27	-754.51	-450.31
Q-male	0.00	0.00	0.00	0.04	-0.02	0.01	0.03	0.03
Q-female	0.00	0.00	0.00	0.02	-0.06	0.01	0.03	0.03

Table 6a. Summary of estimated model parameter values and standard deviations for model 19.3 for Bristol Bay red king crab.

index	name	value	std.dev	index	name	value	std.dev
1	theta[2]	0.2749	0.0173	47	log_slx_pars[1]	4.7444	0.0083
2	theta[4]	19.8860	0.0569	48	log_slx_pars[2]	2.1890	0.0583
3	theta[5]	16.3000	0.1429	49	log_slx_pars[3]	4.5081	0.0295
4	theta[7]	0.6590	0.1257	50	log_slx_pars[4]	2.0856	0.1812
5	theta[9]	-0.4401	0.2572	51	log_slx_pars[5]	5.1519	0.0566
6	theta[13]	0.9628	0.3826	52	log_slx_pars[6]	2.8465	0.0460
7	theta[14]	0.6174	0.4329	53	log_slx_pars[7]	4.6374	0.0651
8	theta[15]	0.8052	0.3219	54	log_slx_pars[8]	2.1786	0.6064
9	theta[16]	0.6510	0.3010	55	log_slx_pars[9]	4.5128	0.0168
10	theta[17]	0.4889	0.2941	56	log_slx_pars[10]	0.9159	0.4156
11	theta[18]	0.4465	0.2788	57	log_slx_pars[11]	4.7991	0.0261
12	theta[19]	0.3027	0.2819	58	log_slx_pars[12]	2.3519	0.0920
13	theta[20]	0.3306	0.2712	59	log_slx_pars[13]	4.0859	0.5844
14	theta[21]	0.3533	0.2661	60	log_slx_pars[14]	3.1951	1.5504
15	theta[22]	0.1478	0.2865	61	log_slx_pars[15]	4.1851	0.2052
16	theta[23]	0.1432	0.2807	62	log_slx_pars[16]	3.1842	0.3813
17	theta[24]	0.0240	0.2912	63	log_slx_pars[17]	4.0735	0.2493
18	theta[25]	0.0904	0.2740	64	log_slx_pars[18]	2.1854	0.4853
19	theta[26]	-0.0117	0.2182	65	log_slx_pars[19]	3.7549	236.6700
20	theta[27]	-0.2226	0.2111	66	log_slx_pars[20]	0.3179	410.7200
21	theta[28]	-0.3853	0.2138	67	log_slx_pars[21]	4.3551	0.0450
22	theta[29]	-0.7165	0.2288	68	log_slx_pars[22]	2.3047	0.1459
23	theta[30]	-1.1582	0.2498	69	log_slx_pars[23]	4.4858	0.0145
24	theta[31]	-1.1849	0.2518	70	log_slx_pars[24]	2.4915	0.0696
25	theta[52]	1.2533	0.9311	71	log_slx_pars[25]	4.9217	0.0016
26	theta[53]	1.5687	0.5268	72	log_slx_pars[26]	0.6855	0.0650
27	theta[54]	1.5399	0.4050	73	log_slx_pars[27]	4.9283	0.0022
28	theta[55]	1.2891	0.3561	74	log_slx_pars[28]	0.6763	0.1275
29	theta[56]	1.1377	0.3118	75	log_fbar[1]	-1.5043	0.0428
30	theta[57]	0.6097	0.3388	76	log_fbar[2]	-4.2897	0.0775
31	theta[58]	0.2224	0.3645	77	log_fbar[3]	-5.4585	0.0989
32	theta[59]	-0.0187	0.3664	78	log_fbar[4]	-6.6075	0.0837
33	theta[60]	-0.2084	0.3541	79	log_fdev[1]	0.6427	0.1226
34	theta[61]	-0.5465	0.3714	80	log_fdev[1]	0.6494	0.0929
35	theta[62]	-0.9352	0.3819	81	log_fdev[1]	0.5870	0.0750
36	theta[63]	-1.1947	0.3863	82	log_fdev[1]	0.7065	0.0617
37	theta[64]	-1.4263	0.3848	83	log_fdev[1]	0.9335	0.0553
38	theta[65]	-1.8059	0.3740	84	log_fdev[1]	1.8165	0.0614
39	theta[66]	-1.9123	0.3701	85	log_fdev[1]	2.3108	0.1365
40	theta[67]	-1.8529	0.3494	86	log_fdev[1]	0.6701	0.1759
41	Grwth[21]	0.8870	0.1854	87	log_fdev[1]	-9.0309	0.1185
42	Grwth[42]	1.4192	0.1224	88	log_fdev[1]	1.0063	0.1052
43	Grwth[85]	140.970	1.7806	89	log_fdev[1]	1.1137	0.0932
44	Grwth[86]	0.0596	0.0103	90	log_fdev[1]	1.2936	0.0756
45	Grwth[87]	140.110	0.6511	91	log_fdev[1]	0.8411	0.0661



46	Grwth[88]	0.0729	0.0037	92	log_fdev[1]	-0.0909	0.0545
93	log_fdev[1]	0.0275	0.0490	143	log_fdev[2]	-0.8520	0.1036
94	log_fdev[1]	0.6682	0.0405	144	log_fdev[2]	-0.7779	0.1038
95	log_fdev[1]	0.6733	0.0433	145	log_fdev[2]	-1.2343	0.1037
96	log_fdev[1]	0.1482	0.0476	146	log_fdev[2]	0.0863	0.1042
97	log_fdev[1]	0.8191	0.0517	147	log_fdev[2]	-0.1993	0.1040
98	log_fdev[1]	-4.3245	0.0493	148	log_fdev[2]	-0.9709	0.1032
99	log_fdev[1]	-4.7230	0.0425	149	log_fdev[2]	-0.2103	0.1031
100	log_fdev[1]	-0.2379	0.0413	150	log_fdev[2]	-0.5125	0.1028
101	log_fdev[1]	-0.1767	0.0419	151	log_fdev[2]	-0.6062	0.1026
102	log_fdev[1]	0.7894	0.0451	152	log_fdev[2]	-0.3762	0.1025
103	log_fdev[1]	0.3819	0.0438	153	log_fdev[2]	-0.6571	0.1024
104	log_fdev[1]	-0.2162	0.0423	154	log_fdev[2]	-0.4930	0.1021
105	log_fdev[1]	-0.3014	0.0417	155	log_fdev[2]	-0.4231	0.1022
106	log_fdev[1]	-0.1917	0.0406	156	log_fdev[2]	-0.4598	0.1025
107	log_fdev[1]	0.2737	0.0393	157	log_fdev[2]	-0.8254	0.1027
108	log_fdev[1]	0.2300	0.0393	158	log_fdev[2]	-0.9867	0.1029
109	log_fdev[1]	0.5087	0.0397	159	log_fdev[2]	-1.4550	0.1028
110	log_fdev[1]	0.2488	0.0388	160	log_fdev[2]	-1.9816	0.1032
111	log_fdev[1]	0.6134	0.0388	161	log_fdev[2]	-1.2798	0.1037
112	log_fdev[1]	0.7772	0.0409	162	log_fdev[2]	-1.8574	0.1045
113	log_fdev[1]	0.5760	0.0419	163	log_fdev[2]	-1.5055	0.1061
114	log_fdev[1]	0.4312	0.0421	164	log_fdev[2]	-1.0216	0.1086
115	log_fdev[1]	-0.2039	0.0416	165	log_fdev[2]	-0.6217	0.1119
116	log_fdev[1]	-0.2809	0.0412	166	log_fdev[2]	-0.7132	0.1150
117	log_fdev[1]	-0.1157	0.0419	167	log_fdev[2]	-0.6279	0.1185
118	log_fdev[1]	0.2040	0.0440	168	log_fdev[3]	-0.0389	0.0685
119	log_fdev[1]	0.2318	0.0486	169	log_fdev[3]	-0.0388	0.0685
120	log_fdev[1]	0.1762	0.0559	170	log_fdev[3]	1.7536	0.0685
121	log_fdev[1]	0.0390	0.0652	171	log_fdev[3]	1.4488	0.0685
122	log_fdev[1]	-0.2324	0.0743	172	log_fdev[3]	1.6753	0.0685
123	log_fdev[1]	-0.2629	0.0820	173	log_fdev[3]	2.5538	0.0685
124	log_fdev[2]	0.1418	0.1261	174	log_fdev[3]	1.4425	0.0685
125	log_fdev[2]	0.6032	0.1168	175	log_fdev[3]	1.6003	0.0685
126	log_fdev[2]	0.6008	0.1111	176	log_fdev[3]	-0.2471	0.0685
127	log_fdev[2]	0.6844	0.1094	177	log_fdev[3]	0.9278	0.0685
128	log_fdev[2]	1.3961	0.1135	178	log_fdev[3]	0.4542	0.0685
129	log_fdev[2]	1.1126	0.1313	179	log_fdev[3]	0.9392	0.0685
130	log_fdev[2]	2.3962	0.1289	180	log_fdev[3]	1.6522	0.0685
131	log_fdev[2]	2.1357	0.1170	181	log_fdev[3]	1.6600	0.0685
132	log_fdev[2]	3.3701	0.1155	182	log_fdev[3]	2.9993	0.0720
133	log_fdev[2]	2.1852	0.1123	183	log_fdev[3]	1.0492	0.0729
134	log_fdev[2]	1.1270	0.1121	184	log_fdev[3]	0.3264	0.0792
135	log_fdev[2]	0.6761	0.1096	185	log_fdev[3]	-2.9934	0.0685
136	log_fdev[2]	1.4522	0.1052	186	log_fdev[3]	-3.9508	0.0685
137	log_fdev[2]	0.0183	0.1042	187	log_fdev[3]	-3.7276	0.0685
138	log_fdev[2]	0.4656	0.1043	188	log_fdev[3]	-3.7276	0.0685
139	log_fdev[2]	0.8772	0.1056	189	log_fdev[3]	-4.6439	0.0685
140	log_fdev[2]	0.7061	0.1056	190	log_fdev[3]	-1.1276	0.0702

141	log_fdev[2]	1.1851	0.1081	191	log_fdev[3]	-0.2264	0.0723
142	log_fdev[2]	-0.5717	0.1051	192	log_fdev[3]	0.2395	0.0772
193	log_fdev[4]	0.6887	0.1037	243	log_fdov[1]	-0.3031	0.0796
194	log_fdev[4]	0.0364	0.1022	244	log_fdov[1]	0.8545	0.0812
195	log_fdev[4]	-0.1681	0.1028	245	log_fdov[1]	0.2983	0.0841
196	log_fdev[4]	0.7408	0.1019	246	log_fdov[1]	-0.1485	0.0875
197	log_fdev[4]	-1.6971	0.1013	247	log_fdov[1]	0.9944	0.0918
198	log_fdev[4]	0.2552	0.1009	248	log_fdov[1]	0.1632	0.0959
199	log_fdev[4]	-0.0024	0.1005	249	log_fdov[3]	-0.0002	0.0967
200	log_fdev[4]	-0.8381	0.1004	250	log_fdov[3]	-0.0004	0.0967
201	log_fdev[4]	-0.6665	0.1001	251	log_fdov[3]	0.0002	0.0967
202	log_fdev[4]	-0.3943	0.0999	252	log_fdov[3]	0.0006	0.0967
203	log_fdev[4]	-0.4464	0.0996	253	log_fdov[3]	0.0006	0.0967
204	log_fdev[4]	0.0951	0.0996	254	log_fdov[3]	-0.0016	0.0966
205	log_fdev[4]	-0.6118	0.1001	255	log_fdov[3]	-0.0007	0.0967
206	log_fdev[4]	-1.6194	0.0999	256	log_fdov[3]	-0.0003	0.0967
207	log_fdev[4]	-2.5090	0.0995	257	log_fdov[3]	-0.0005	0.0967
208	log_fdev[4]	-0.9955	0.0992	258	log_fdov[3]	0.0002	0.0967
209	log_fdev[4]	-0.4479	0.0993	259	log_fdov[3]	0.0003	0.0967
210	log_fdev[4]	0.6876	0.0995	260	log_fdov[3]	0.0015	0.0967
211	log_fdev[4]	1.5158	0.1000	261	log_fdov[3]	0.0026	0.0967
212	log_fdev[4]	1.1726	0.1010	262	log_fdov[3]	0.0038	0.0967
213	log_fdev[4]	0.2879	0.1025	263	log_fdov[3]	0.5057	0.0988
214	log_fdev[4]	1.8747	0.1047	264	log_fdov[3]	0.7525	0.0978
215	log_fdev[4]	2.0949	0.1067	265	log_fdov[3]	-0.4482	0.1022
216	log_fdev[4]	0.9467	0.1090	266	log_fdov[3]	-0.0006	0.0967
217	log_foff[1]	-2.8529	0.0537	267	log_fdov[3]	-0.0006	0.0967
218	log_foff[3]	0.5009	0.0929	268	log_fdov[3]	-0.0006	0.0967
219	log_fdov[1]	2.0679	0.0841	269	log_fdov[3]	-0.0006	0.0967
220	log_fdov[1]	-0.5974	0.0832	270	log_fdov[3]	-0.0006	0.0967
221	log_fdov[1]	2.0825	0.0847	271	log_fdov[3]	0.0182	0.0966
222	log_fdov[1]	1.9121	0.0858	272	log_fdov[3]	-0.7141	0.0973
223	log_fdov[1]	-0.3400	0.0844	273	log_fdov[3]	-0.1175	0.0997
224	log_fdov[1]	-0.1270	0.0827	274	rec_dev_est	1.0794	0.2976
225	log_fdov[1]	-3.6240	0.0827	275	rec_dev_est	0.7311	0.2950
226	log_fdov[1]	-0.2733	0.0845	276	rec_dev_est	1.1263	0.2445
227	log_fdov[1]	1.4941	0.0829	277	rec_dev_est	1.7291	0.2113
228	log_fdov[1]	-2.7279	0.0813	278	rec_dev_est	1.9904	0.2231
229	log_fdov[1]	1.2165	0.0805	279	rec_dev_est	1.1519	0.2681
230	log_fdov[1]	0.9443	0.0805	280	rec_dev_est	2.3399	0.1690
231	log_fdov[1]	-1.8064	0.0798	281	rec_dev_est	1.3687	0.1839
232	log_fdov[1]	1.2767	0.0805	282	rec_dev_est	0.9960	0.1708
233	log_fdov[1]	0.4918	0.0809	283	rec_dev_est	-0.8590	0.2556
234	log_fdov[1]	1.0262	0.0796	284	rec_dev_est	0.2556	0.1674
235	log_fdov[1]	-1.1644	0.0791	285	rec_dev_est	-0.8849	0.2447
236	log_fdov[1]	-0.1117	0.0793	286	rec_dev_est	-1.3230	0.2789
237	log_fdov[1]	-0.3832	0.0795	287	rec_dev_est	-1.1210	0.2339
238	log_fdov[1]	-0.5928	0.0798	288	rec_dev_est	-0.1322	0.1713
239	log_fdov[1]	-0.1359	0.0803	289	rec_dev_est	-0.5997	0.1933

240	log_fdov[1]	-1.0767	0.0793	290	rec_dev_est	-2.0873	0.3716
241	log_fdov[1]	-1.7165	0.0787	291	rec_dev_est	-1.0340	0.2076
242	log_fdov[1]	0.3028	0.0788	292	rec_dev_est	-2.3004	0.5003
293	rec_dev_est	0.9320	0.1518	339	logit_rec_prop_es	1.4330	0.7775
294	rec_dev_est	-1.0433	0.2655	340	logit_rec_prop_es	0.6054	0.6934
295	rec dev est	-1.6231	0.3342	341	logit_rec_prop_es	0.4621	0.3267
296	rec_dev_est	-0.6536	0.2037	342	logit_rec_prop_es	-0.1146	0.1462
297	rec_dev_est	0.3285	0.1611	343	logit_rec_prop_es	0.2329	0.3548
298	rec_dev_est	-0.5955	0.2220	344	logit_rec_prop_es	-0.4851	0.3715
299	rec_dev_est	-0.5981	0.2419	345	logit_rec_prop_es	-0.5161	0.1317
300	rec_dev_est	0.7746	0.1599	346	logit_rec_prop_es	-0.3856	0.4374
301	rec dev est	-0.7101	0.2737	347	logit_rec_prop_es	-0.0832	0.4245
302	rec_dev_est	-0.6874	0.2618	348	logit_rec_prop_es	-0.4556	0.1413
303	rec_dev_est	0.5600	0.1615	349	logit_rec_prop_es	-0.0760	0.2474
304	rec_dev_est	-0.1755	0.1895	350	logit_rec_prop_es	0.1947	0.2815
305	rec_dev_est	-0.5592	0.1953	351	logit_rec_prop_es	-0.2368	0.3697
306	rec_dev_est	-1.1078	0.2414	352	logit_rec_prop_es	-0.3192	0.3748
307	rec dev est	-1.0323	0.2465	353	logit_rec_prop_es	-0.8485	0.1925
308	rec_dev_est	-0.0045	0.1799	354	logit_rec_prop_es	-0.3224	0.3105
309	rec_dev_est	-0.5554	0.2233	355	logit_rec_prop_es	-0.5481	0.3173
310	rec_dev_est	-0.9540	0.2248	356	logit_rec_prop_es	-0.0122	0.3469
311	rec_dev_est	-1.3618	0.2286	357	logit_rec_prop_es	-0.2385	0.4730
312	rec_dev_est	-1.9292	0.2923	358	logit_rec_prop_es	-0.1864	0.3287
313	rec dev est	-1.4162	0.2269	359	logit_rec_prop_es	0.2586	0.2467
314	rec_dev_est	-0.8414	0.1882	360	logit_rec_prop_es	0.6521	0.5618
315	rec_dev_est	-1.6911	0.2850	361	logit_rec_prop_es	0.4341	0.4426
316	rec_dev_est	-1.2456	0.2701	362	logit_rec_prop_es	0.7423	0.9166
317	rec_dev_est	-1.8541	0.4577	363	logit_rec_prop_es	-0.3395	1.6742
318	rec_dev_est	-0.2405	1.3063	364	m_dev_est[1]	1.6056	0.0288
319	logit_rec_prop_es	-0.1738	0.4779	365	survey_q[1]	0.9592	0.0280
320	logit_rec_prop_es	-0.7552	0.4696	366	log_add_cv[2]	-0.9615	0.2885
321	logit_rec_prop_es	-0.2946	0.3618	367	sd_rbar	16133000	521640
322	logit_rec_prop_es	-0.5530	0.2706	368	sd_ssbF0	72699.0	2135.60
323	logit_rec_prop_es	-0.0626	0.2743	369	sd_Bmsy	25445.0	747.440
324	logit_rec_prop_es	0.0951	0.3784	370	sd_depl	0.5867	0.0405
325	logit_rec_prop_es	0.3407	0.1569	371	sd_fmsy	0.2907	0.0043
326	logit_rec_prop_es	0.3958	0.2409	372	sd_fmsy	0.0059	0.0006
327	logit_rec_prop_es	-0.0992	0.1810	373	sd_fmsy	0.0011	0.0001
328	logit_rec_prop_es	0.5050	0.4900	374	sd_fmsy	0.0059	0.0006
329	logit_rec_prop_es	-0.4662	0.1645	375	sd_fmsy	0.0000	0.0000
330	logit_rec_prop_es	0.2581	0.4222	376	sd_fmsy	0.0000	0.0000
331	logit_rec_prop_es	-0.0528	0.4617	377	sd_fofl	0.1572	0.0137
332	logit_rec_prop_es	0.4767	0.4221	378	sd_fofl	0.0059	0.0006
333	logit_rec_prop_es	-0.1924	0.1754	379	sd_fofl	0.0011	0.0001
334	logit_rec_prop_es	0.1362	0.2614	380	sd_fofl	0.0059	0.0006
335	logit_rec_prop_es	0.9226	0.8947	381	sd_fofl	0.0000	0.0000
336	logit_rec_prop_es	0.0337	0.2920	382	sd_fofl	0.0000	0.0000
337	logit_rec_prop_es	-0.0668	0.8645	383	sd_ofl	2140.7000	334.4400
338	logit_rec_prop_es	-0.2947	0.0904				

Table 6b. Summary of estimated model parameter values and standard deviations for model 19.3d for Bristol Bay red king crab.

index	name	std.de		index	name	value	std.dev
		value	v				
1	theta[2]	0.2713	0.0165	47	log_slx_pars[1]	4.7479	0.0081
2	theta[4]	19.8670	0.0559	48	log_slx_pars[2]	2.2543	0.0497
3	theta[5]	16.2910	0.1432	49	log_slx_pars[3]	4.5086	0.0178
4	theta[7]	0.7335	0.1388	50	log_slx_pars[4]	2.0335	0.1184
5	theta[9]	-0.5733	0.2506	51	log_slx_pars[5]	5.1395	0.0538
6	theta[13]	0.9402	0.3912	52	log_slx_pars[6]	2.8479	0.0466
7	theta[14]	0.6093	0.4306	53	log_slx_pars[7]	4.7220	0.2127
8	theta[15]	0.8111	0.3206	54	log_slx_pars[8]	2.1621	0.3066
9	theta[16]	0.6681	0.3000	55	log_slx_pars[9]	4.7306	0.0827
10	theta[17]	0.5109	0.2958	56	log_slx_pars[10]	0.9006	0.3030
11	theta[18]	0.4734	0.2811	57	log_slx_pars[11]	4.7990	0.0259
12	theta[19]	0.3232	0.2825	58	log_slx_pars[12]	2.3573	0.0916
13	theta[20]	0.3634	0.2686	59	log_slx_pars[13]	3.9746	1.0975
14	theta[21]	0.3963	0.2617	60	log_slx_pars[14]	3.3182	2.1299
15	theta[22]	0.1712	0.2840	61	log_slx_pars[15]	4.1815	0.2231
16	theta[23]	0.1494	0.2788	62	log_slx_pars[16]	3.2184	0.4120
17	theta[24]	0.0433	0.2876	63	log_slx_pars[17]	4.0719	0.2606
18	theta[25]	0.1593	0.2629	64	log_slx_pars[18]	2.1911	0.4900
19	theta[26]	-0.0195	0.2031	65	log_slx_pars[19]	3.7377	504.3600
20	theta[27]	-0.2475	0.1962	66	log_slx_pars[20]	0.0455	125.5900
21	theta[28]	-0.4033	0.1987	67	log_slx_pars[21]	4.3563	0.0479
22	theta[29]	-0.7501	0.2122	68	log_slx_pars[22]	2.3193	0.1469
23	theta[30]	-1.2101	0.2328	69	log_slx_pars[23]	4.4799	0.0138
24	theta[31]	-1.2592	0.2357	70	log_slx_pars[24]	2.4654	0.0688
25	theta[52]	1.1907	1.0092	71	log_slx_pars[25]	4.9207	0.0015
26	theta[53]	1.5700	0.5376	72	log_slx_pars[26]	0.6739	0.0546
27	theta[54]	1.5415	0.4094	73	log_slx_pars[27]	4.9284	0.0019
28	theta[55]	1.2870	0.3571	74	log_slx_pars[28]	0.6713	0.1075
29	theta[56]	1.1298	0.3124	75	log_fbar[1]	-1.5066	0.0423
30	theta[57]	0.5999	0.3396	76	log_fbar[2]	-4.3132	0.0774
31	theta[58]	0.2168	0.3652	77	log_fbar[3]	-5.6295	0.2865
32	theta[59]	-0.0212	0.3670	78	log_fbar[4]	-6.6052	0.0828
33	theta[60]	-0.2077	0.3547	79	log_fdev[1]	0.7046	0.1204
34	theta[61]	-0.5434	0.3722	80	log_fdev[1]	0.6733	0.0932
35	theta[62]	-0.9300	0.3828	81	log_fdev[1]	0.5896	0.0759
36	theta[63]	-1.1896	0.3872	82	log_fdev[1]	0.6931	0.0616
37	theta[64]	-1.4206	0.3857	83	log_fdev[1]	0.9156	0.0548
38	theta[65]	-1.7988	0.3748	84	log_fdev[1]	1.7859	0.0579
39	theta[66]	-1.9047	0.3708	85	log_fdev[1]	2.2615	0.1249
40	theta[67]	-1.8438	0.3499	86	log_fdev[1]	0.6253	0.1754
41	Grwth[21]	0.9570	0.1887	87	log_fdev[1]	-9.0439	0.1260
42	Grwth[42]	1.4224	0.1232	88	log_fdev[1]	1.0441	0.1054
43	Grwth[85]	141.940	1.9324	89	log_fdev[1]	1.1367	0.0906
44	Grwth[86]	0.0601	0.0115	90	log_fdev[1]	1.2979	0.0748
45	Grwth[87]	139.920	0.6309	91	log_fdev[1]	0.8366	0.0646

46	Grwth[88]	0.0722	0.0035	92	log_fdev[1]	-0.0937	0.0532
93	log_fdev[1]	0.0288	0.0477	143	log_fdev[2]	-0.8531	0.1034
94	log_fdev[1]	0.6777	0.0393	144	log_fdev[2]	-0.7854	0.1035
95	log_fdev[1]	0.6884	0.0423	145	log_fdev[2]	-1.2510	0.1034
96	log_fdev[1]	0.1656	0.0471	146	log_fdev[2]	0.0443	0.1037
97	log_fdev[1]	0.8265	0.0518	147	log_fdev[2]	-0.2317	0.1035
98	log_fdev[1]	-4.3296	0.0490	148	log_fdev[2]	-0.9867	0.1029
99	log_fdev[1]	-4.7351	0.0421	149	log_fdev[2]	-0.2155	0.1027
100	log_fdev[1]	-0.2623	0.0407	150	log_fdev[2]	-0.5144	0.1025
101	log_fdev[1]	-0.2142	0.0412	151	log_fdev[2]	-0.6076	0.1022
102	log_fdev[1]	0.6965	0.0437	152	log_fdev[2]	-0.3746	0.1022
103	log_fdev[1]	0.3470	0.0428	153	log_fdev[2]	-0.6544	0.1021
104	log_fdev[1]	-0.2342	0.0412	154	log_fdev[2]	-0.4890	0.1018
105	log_fdev[1]	-0.3126	0.0407	155	log_fdev[2]	-0.4192	0.1020
106	log_fdev[1]	-0.2014	0.0395	156	log_fdev[2]	-0.4581	0.1024
107	log_fdev[1]	0.2645	0.0382	157	log_fdev[2]	-0.8253	0.1026
108	log_fdev[1]	0.2250	0.0382	158	log_fdev[2]	-0.9851	0.1027
109	log_fdev[1]	0.5092	0.0386	159	log_fdev[2]	-1.4506	0.1025
110	log_fdev[1]	0.2568	0.0380	160	log_fdev[2]	-1.9746	0.1028
111	log_fdev[1]	0.6169	0.0381	161	log_fdev[2]	-1.2702	0.1032
112	log_fdev[1]	0.7783	0.0406	162	log_fdev[2]	-1.8469	0.1040
113	log_fdev[1]	0.5769	0.0417	163	log_fdev[2]	-1.4889	0.1057
114	log_fdev[1]	0.4348	0.0414	164	log_fdev[2]	-0.9955	0.1082
115	log_fdev[1]	-0.2041	0.0402	165	log_fdev[2]	-0.5893	0.1115
116	log_fdev[1]	-0.2814	0.0396	166	log_fdev[2]	-0.6744	0.1148
117	log_fdev[1]	-0.1055	0.0403	167	log_fdev[2]	-0.5933	0.1184
118	log_fdev[1]	0.2040	0.0424	168	log_fdev[3]	-0.1164	0.0682
119	log_fdev[1]	0.2487	0.0471	169	log_fdev[3]	0.6699	0.0682
120	log_fdev[1]	0.2098	0.0548	170	log_fdev[3]	1.2285	0.0682
121	log_fdev[1]	0.0761	0.0647	171	log_fdev[3]	1.0928	0.0682
122	log_fdev[1]	-0.1508	0.0745	172	log_fdev[3]	1.3825	0.0682
123	log_fdev[1]	-0.2265	0.0828	173	log_fdev[3]	1.4242	0.0682
124	log_fdev[2]	0.1597	0.1266	174	log_fdev[3]	0.9927	0.0682
125	log_fdev[2]	0.6029	0.1178	175	log_fdev[3]	0.4764	0.0682
126	log_fdev[2]	0.5869	0.1114	176	log_fdev[3]	-0.9874	0.0682
127	log_fdev[2]	0.6622	0.1093	177	log_fdev[3]	-0.5788	0.0682
128	log_fdev[2]	1.3565	0.1121	178	log_fdev[3]	-1.0994	0.0682
129	log_fdev[2]	1.0729	0.1297	179	log_fdev[3]	-0.2563	0.0682
130	log_fdev[2]	2.3685	0.1291	180	log_fdev[3]	0.9401	0.0682
131	log_fdev[2]	2.1245	0.1173	181	log_fdev[3]	1.4181	0.0682
132	log_fdev[2]	3.3721	0.1147	182	log_fdev[3]	3.2775	0.0763
133	log_fdev[2]	2.1828	0.1117	183	log_fdev[3]	1.3179	0.0948
134	log_fdev[2]	1.1189	0.1114	184	log_fdev[3]	0.6099	0.1163
135	log_fdev[2]	0.6688	0.1088	185	log_fdev[3]	-0.7287	0.0836
136	log_fdev[2]	1.4501	0.1045	186	log_fdev[3]	-2.1053	0.0733
137	log_fdev[2]	0.0231	0.1036	187	log_fdev[3]	-2.9662	0.0939
138	log_fdev[2]	0.4773	0.1038	188	log_fdev[3]	-2.4001	0.1114
139	log_fdev[2]	0.8991	0.1051	189	log_fdev[3]	-3.4930	0.0747
140	log_fdev[2]	0.7312	0.1053	190	log_fdev[3]	-0.8825	0.0927

141	log_fdev[2]	1.2002	0.1080	191	log_fdev[3]	-0.1832	0.1092
142	log_fdev[2]	-0.5674	0.1049	192	log_fdev[3]	0.9667	0.1313
193	log_fdev[4]	0.6684	0.1033	243	log_fdov[1]	-0.1534	0.0787
194	log_fdev[4]	0.0117	0.1020	244	log_fdov[1]	0.9281	0.0800
195	log_fdev[4]	-0.1961	0.1024	245	log_fdov[1]	0.4190	0.0826
196	log_fdev[4]	0.7253	0.1016	246	log_fdov[1]	-0.1885	0.0863
197	log_fdev[4]	-1.7057	0.1011	247	log_fdov[1]	1.1760	0.0910
198	log_fdev[4]	0.2495	0.1007	248	log_fdov[1]	0.1268	0.0953
199	log_fdev[4]	-0.0069	0.1003	249	log_fdov[3]	0.0000	0.0962
200	log_fdev[4]	-0.8407	0.1002	250	log_fdov[3]	0.0001	0.0962
201	log_fdev[4]	-0.6675	0.1000	251	log_fdov[3]	0.0001	0.0963
202	log_fdev[4]	-0.3952	0.0998	252	log_fdov[3]	0.0001	0.0963
203	log_fdev[4]	-0.4466	0.0995	253	log_fdov[3]	0.0003	0.0963
204	log_fdev[4]	0.0948	0.0995	254	log_fdov[3]	0.0000	0.0963
205	log_fdev[4]	-0.6123	0.1000	255	log_fdov[3]	-0.0002	0.0963
206	log_fdev[4]	-1.6191	0.0998	256	log_fdov[3]	-0.0002	0.0962
207	log_fdev[4]	-2.5074	0.0993	257	log_fdov[3]	-0.0001	0.0962
208	log_fdev[4]	-0.9934	0.0991	258	log_fdov[3]	-0.0001	0.0962
209	log_fdev[4]	-0.4455	0.0992	259	log_fdov[3]	-0.0001	0.0962
210	log_fdev[4]	0.6916	0.0993	260	log_fdov[3]	0.0001	0.0962
211	log_fdev[4]	1.5223	0.0999	261	log_fdov[3]	0.0003	0.0962
212	log_fdev[4]	1.1855	0.1009	262	log_fdov[3]	0.0009	0.0963
213	log_fdev[4]	0.3079	0.1024	263	log_fdov[3]	1.5183	0.1436
214	log_fdev[4]	1.8971	0.1046	264	log_fdov[3]	1.8010	0.1185
215	log_fdev[4]	2.1179	0.1067	265	log_fdov[3]	0.5805	0.1378
216	log_fdev[4]	0.9641	0.1091	266	log_fdov[3]	-3.4377	0.1091
217	log_foff[1]	-2.8291	0.0410	267	log_fdov[3]	-2.1784	0.1506
218	log_foff[3]	-0.1444	0.4096	268	log_fdov[3]	-0.8048	0.1228
219	log_fdov[1]	2.0327	0.0834	269	log_fdov[3]	0.0265	0.1317
220	log_fdov[1]	-0.6368	0.0826	270	log_fdov[3]	0.3755	0.1025
221	log_fdov[1]	2.0434	0.0838	271	log_fdov[3]	0.9644	0.1531
222	log_fdov[1]	1.8857	0.0855	272	log_fdov[3]	0.2010	0.1418
223	log_fdov[1]	-0.3437	0.0840	273	log_fdov[3]	0.9525	0.1650
224	log_fdov[1]	-0.1185	0.0821	274	rec_dev_est	1.0748	0.2984
225	log_fdov[1]	-3.6226	0.0810	275	rec_dev_est	0.6922	0.3000
226	log_fdov[1]	-0.2549	0.0816	276	rec_dev_est	1.1259	0.2437
227	log_fdov[1]	1.5335	0.0816	277	rec_dev_est	1.7307	0.2120
228	log_fdov[1]	-2.7064	0.0808	278	rec_dev_est	1.9814	0.2295
229	log_fdov[1]	1.2151	0.0801	279	rec_dev_est	1.1071	0.2866
230	log_fdov[1]	0.9380	0.0800	280	rec_dev_est	2.3693	0.1684
231	log_fdov[1]	-1.8108	0.0795	281	rec_dev_est	1.4016	0.1834
232	log_fdov[1]	1.2680	0.0796	282	rec_dev_est	0.9997	0.1712
233	log_fdov[1]	0.4680	0.0797	283	rec_dev_est	-0.8854	0.2601
234	log_fdov[1]	1.0025	0.0792	284	rec_dev_est	0.2307	0.1682
235	log_fdov[1]	-1.1823	0.0788	285	rec_dev_est	-0.9147	0.2478
236	log_fdov[1]	-0.1450	0.0788	286	rec_dev_est	-1.3290	0.2777
237	log_fdov[1]	-0.4078	0.0793	287	rec_dev_est	-1.1097	0.2310
238	log_fdov[1]	-0.6609	0.0795	288	rec_dev_est	-0.1255	0.1684
239	log_fdov[1]	-0.1740	0.0793	289	rec_dev_est	-0.5981	0.1890

240	log_fdov[1]	-1.0739	0.0784	290	rec_dev_est	-2.1159	0.3754
241	log_fdov[1]	-1.7957	0.0781	291	rec_dev_est	-0.9618	0.2016
242	log_fdov[1]	0.2344	0.0782	292	rec_dev_est	-2.1549	0.4556
293	rec_dev_est	0.9043	0.1523	339	logit_rec_prop_e	1.4632	0.7644
294	rec_dev_est	-1.0234	0.2681	340	logit_rec_prop_e	0.4894	0.6837
295	rec_dev_est	-1.6638	0.3460	341	logit_rec_prop_e	0.4657	0.3258
296	rec_dev_est	-0.6549	0.2034	342	logit_rec_prop_e	-0.0656	0.1443
297	rec_dev_est	0.3338	0.1606	343	logit_rec_prop_e	0.2439	0.3668
298	rec_dev_est	-0.6347	0.2267	344	logit_rec_prop_e	-0.5694	0.3816
299	rec_dev_est	-0.6209	0.2440	345	logit_rec_prop_e	-0.4877	0.1275
300	rec_dev_est	0.7884	0.1591	346	logit_rec_prop_e	-0.3932	0.4268
301	rec_dev_est	-0.6835	0.2685	347	logit_rec_prop_e	-0.0418	0.4555
302	rec_dev_est	-0.7679	0.2740	348	logit_rec_prop_e	-0.4050	0.1395
303	rec_dev_est	0.5593	0.1610	349	logit_rec_prop_e	-0.0713	0.2434
304	rec_dev_est	-0.1849	0.1880	350	logit_rec_prop_e	0.2583	0.2816
305	rec_dev_est	-0.5628	0.1939	351	logit_rec_prop_e	-0.2074	0.3698
306	rec_dev_est	-1.1264	0.2405	352	logit_rec_prop_e	-0.4067	0.3604
307	rec_dev_est	-1.0074	0.2403	353	logit_rec_prop_e	-0.8321	0.1907
308	rec_dev_est	0.0006	0.1796	354	logit_rec_prop_e	-0.3538	0.3122
309	rec_dev_est	-0.5645	0.2254	355	logit_rec_prop_e	-0.4678	0.3372
310	rec_dev_est	-1.0624	0.2319	356	logit_rec_prop_e	-0.0856	0.3281
311	rec_dev_est	-1.3591	0.2231	357	logit_rec_prop_e	-0.2073	0.4597
312	rec_dev_est	-1.9157	0.2840	358	logit_rec_prop_e	-0.1264	0.3249
313	rec_dev_est	-1.4057	0.2237	359	logit_rec_prop_e	0.3055	0.2409
314	rec_dev_est	-0.8047	0.1861	360	logit_rec_prop_e	0.6063	0.5625
315	rec_dev_est	-1.7289	0.2892	361	logit_rec_prop_e	0.4574	0.4302
316	rec_dev_est	-1.2117	0.2639	362	logit_rec_prop_e	0.6559	0.8912
317	rec_dev_est	-1.8569	0.4579	363	logit_rec_prop_e	-0.5496	1.6529
318	rec_dev_est	-0.1160	1.3557	364	m_dev_est[1]	1.6170	0.0286
319	logit_rec_prop_e	-0.1714	0.4778	365	survey_q[1]	0.9624	0.0288
320	logit_rec_prop_e	-0.8295	0.4885	366	log_add_cv[2]	-0.9397	0.2873
321	logit_rec_prop_e	-0.2797	0.3623	367	sd_rbar	15908000	510700
322	logit_rec_prop_e	-0.5324	0.2808	368	sd_ssbF0	72059.000	2094.400
323	logit_rec_prop_e	-0.0809	0.2779	369	sd_Bmsy	25221.000	733.0500
324	logit_rec_prop_e	0.0175	0.3949	370	sd_depl	0.5869	0.0408
325	logit_rec_prop_e	0.3783	0.1533	371	sd_fmsy	0.2926	0.0040
326	logit_rec_prop_e	0.4141	0.2366	372	sd_fmsy	0.0059	0.0006
327	logit_rec_prop_e	-0.0631	0.1821	373	sd_fmsy	0.0019	0.0004
328	logit_rec_prop_e	0.4733	0.4895	374	sd_fmsy	0.0060	0.0006
329	logit_rec_prop_e	-0.4640	0.1683	375	sd_fmsy	0.0000	0.0000
330	logit_rec_prop_e	0.2745	0.4305	376	sd_fmsy	0.0000	0.0000
331	logit_rec_prop_e	-0.0788	0.4568	377	sd_fofl	0.1583	0.0139
332	logit_rec_prop_e	0.4625	0.4087	378	sd_fofl	0.0059	0.0006
333	logit_rec_prop_e	-0.0551	0.1691	379	sd_fofl	0.0019	0.0004
334	logit_rec_prop_e	0.1730	0.2465	380	sd_fofl	0.0060	0.0006
335	logit_rec_prop_e	0.9104	0.8664	381	sd_fofl	0.0000	0.0000
336	logit_rec_prop_e	0.1949	0.2840	382	sd_fofl	0.0000	0.0000
337	logit_rec_prop_e	-0.1169	0.7766	383	sd_ofl	2114.5000	333.3800
338	logit_rec_prop_e	-0.3015	0.0939	384	sd log recruits	17.2760	0.2115

Table 7. Natural mortality estimates for eight model scenarios.

Model	Sex	Year	
		1975-1979, 1985-2020	1980-1984
19.3	Males	0.180	0.897
	Females	0.237	1.180
19.3c	Males	0.180	0.902
	Females	0.237	1.185
19.3d	Males	0.180	0.907
	Females	0.236	1.189
19.3e	Males	0.180	0.906
	Females	0.236	1.186
19.3f	Males	0.180	0.907
	Females	0.237	1.193
19.3g	Males	0.180	0.925
	Females	0.234	1.200
19.3i	Males	0.180	0.919
	Females	0.233	1.189
19.6	Males	0.257	1.081
	Females	0.289	1.215

Table 8. Area-swept estimates of mature female abundance (million crab) and model estimates of effective spawning biomass (ESB, Zheng et al. 1995b) (1000 t) during 2010-2020 for groundfish fisheries bycatch (PSC) calculation. (\*mature female abundance in 2020 is the model projected value). Note that PSC limits apply to previous-year ESB.

Year	Mature female abundance	Effective spawning biomass
2010	31.603	30.573
2011	28.520	19.541
2012	21.121	20.029
2013	15.694	22.382
2014	38.580	23.272
2015	18.666	21.098
2016	22.633	19.147
2017	18.497	18.042
2018	9.106	15.093
2019	8.587	12.705
2020	9.668*	11.394



Table 9a. Annual abundance estimates (million crab), mature male biomass (MMB, 1000 t), and total survey biomass (1000 t) for red king crab in Bristol Bay estimated by length-based model 19.3 during 1975-2020. Mature male biomass for year  $t$  is on Feb. 15, year  $t+1$ . Size measurements are mm carapace length. The highlighted cell shows a very unreliable recruitment estimate.

Year (t)	Males				Females		Total Recruits	Total Survey Biomass	
	Mature (>119 mm)	Legal (>134mm)	MMB (>119 mm)	SD MMB	Mature (>89 mm)			Model Est. (>64 mm)	Area-Swept (>64 mm)
1975	57.510	30.033	88.074	9.093	57.640		233.362	202.731	
1976	66.807	36.605	102.546	8.584	91.349	70.625	272.161	331.868	
1977	73.512	41.868	114.496	7.479	124.005	49.849	294.567	375.661	
1978	78.735	46.378	120.111	5.979	128.207	74.012	299.709	349.545	
1979	69.672	47.182	100.043	4.316	123.110	135.246	291.714	167.627	
1980	52.117	37.842	30.293	1.835	126.594	175.629	280.477	249.322	
1981	15.211	8.130	6.866	1.141	55.764	75.931	112.334	132.669	
1982	7.114	2.252	6.873	0.927	24.830	249.089	69.540	143.740	
1983	6.447	2.252	7.689	0.680	15.709	94.311	59.842	49.320	
1984	6.169	2.354	5.258	0.465	14.618	64.973	51.154	155.312	
1985	7.520	1.854	9.605	0.671	9.902	10.165	34.527	34.535	
1986	12.079	4.594	14.870	1.005	13.818	30.986	45.010	48.158	
1987	14.241	6.584	20.087	1.220	17.184	9.906	50.786	70.263	
1988	14.314	8.328	24.736	1.292	21.684	6.391	54.268	55.372	
1989	15.555	9.606	27.738	1.255	20.408	7.822	57.078	55.941	
1990	15.152	10.379	24.181	1.188	18.069	21.026	57.209	60.321	
1991	11.710	8.694	18.709	1.122	17.428	13.175	52.137	85.055	
1992	9.364	6.555	17.471	1.079	18.700	2.976	47.443	37.687	
1993	10.405	6.199	15.788	1.128	17.408	8.533	46.767	53.703	
1994	10.172	5.936	21.438	1.224	14.799	2.405	41.945	32.335	
1995	10.677	7.764	24.504	1.215	13.665	60.942	47.884	38.396	
1996	10.786	8.388	22.669	1.155	19.834	8.454	57.153	44.649	
1997	10.056	7.500	21.044	1.132	29.204	4.734	63.220	85.277	
1998	15.657	7.336	23.885	1.358	25.554	12.482	67.192	85.176	
1999	16.755	9.402	27.888	1.542	21.571	33.329	65.731	65.604	
2000	14.529	10.426	28.358	1.544	23.110	13.230	67.546	68.102	
2001	14.323	10.074	28.833	1.513	26.337	13.196	71.214	53.188	
2002	17.013	10.241	32.689	1.538	25.600	52.068	76.387	69.786	
2003	17.939	11.804	32.330	1.510	31.356	11.798	82.650	116.794	
2004	16.252	11.448	30.031	1.436	38.727	12.068	84.330	131.910	
2005	18.170	10.707	30.673	1.410	35.976	42.013	85.769	107.341	
2006	17.287	11.331	31.150	1.379	36.928	20.136	86.267	95.676	
2007	15.646	11.114	26.295	1.299	41.524	13.719	88.489	104.841	
2008	16.198	9.486	25.265	1.346	39.154	7.926	85.461	114.430	
2009	16.245	9.567	26.531	1.435	34.624	8.548	79.898	91.673	
2010	15.168	9.939	26.053	1.425	30.370	23.889	75.264	81.642	
2011	12.925	9.459	25.889	1.359	29.952	13.771	71.252	67.053	
2012	11.643	8.947	24.502	1.283	32.030	9.244	70.123	61.248	
2013	11.670	8.256	23.728	1.241	30.405	6.148	67.944	62.410	
2014	11.658	8.069	22.187	1.225	27.191	3.486	63.732	114.103	
2015	10.360	7.575	19.786	1.220	23.252	5.823	57.246	64.240	
2016	8.772	6.674	17.238	1.224	19.789	10.346	50.807	61.231	
2017	7.197	5.709	14.783	1.214	17.900	4.423	45.776	52.922	
2018	6.362	4.800	13.580	1.226	16.240	6.906	42.167	28.932	
2019	6.983	4.493	14.237	1.348	14.118	3.758	39.853	28.744	
2020	7.305	4.896	14.928	1.185	12.471	18.867			

Table 9b. Annual abundance estimates (million crab), mature male biomass (MMB, 1000 t), and total survey biomass (1000 t) for red king crab in Bristol Bay estimated by length-based model 19.3d during 1975-2020. Mature male biomass for year  $t$  is on Feb. 15, year  $t+1$ . Size measurements are mm carapace length. The highlighted cell shows a very unreliable recruitment estimate.

Year (t)	Males				Females	Total Recruits	Total Survey Biomass	
	Mature (>119 mm)	Legal (>134mm)	MMB (>119 mm)	SD MMB	Mature (>89 mm)		Model Est. (>64 mm)	Area-Swept (>64 mm)
1975	56.540	28.706	85.209	8.897	56.435		231.627	202.731
1976	66.313	35.957	101.126	8.640	89.457	69.644	270.954	331.868
1977	73.140	41.714	114.504	7.601	121.229	47.506	293.353	375.661
1978	78.167	46.593	120.570	6.130	126.064	73.296	299.720	349.545
1979	69.198	47.378	100.952	4.330	121.678	134.204	292.939	167.627
1980	51.700	38.048	31.062	1.800	125.782	172.433	283.215	249.322
1981	15.239	8.202	7.167	1.165	55.535	71.933	113.178	132.669
1982	7.205	2.325	7.034	0.984	24.495	254.141	69.794	143.740
1983	6.296	2.276	7.529	0.659	15.406	96.563	60.212	49.320
1984	6.104	2.281	5.144	0.452	14.348	64.609	51.301	155.312
1985	7.492	1.834	9.544	0.674	9.664	9.808	34.397	34.535
1986	12.107	4.578	14.926	1.009	13.476	29.943	44.799	48.158
1987	14.257	6.613	20.203	1.221	16.717	9.525	50.438	70.263
1988	14.290	8.353	24.800	1.286	21.095	6.294	53.798	55.372
1989	15.390	9.609	27.622	1.243	19.919	7.838	56.540	55.941
1990	14.924	10.297	23.869	1.173	17.822	20.971	56.818	60.321
1991	11.516	8.557	18.311	1.109	17.197	13.073	51.949	85.055
1992	9.274	6.417	17.192	1.078	18.384	2.865	47.367	37.687
1993	10.555	6.146	15.932	1.151	17.113	9.087	46.945	53.703
1994	10.402	6.039	21.799	1.265	14.544	2.756	42.434	32.335
1995	10.901	7.917	24.937	1.266	13.495	58.729	48.557	38.396
1996	11.154	8.569	23.391	1.213	19.460	8.544	57.831	44.649
1997	10.627	7.790	22.173	1.187	28.542	4.503	63.737	85.277
1998	15.835	7.765	24.912	1.399	25.011	12.351	67.488	85.176
1999	16.835	9.661	28.460	1.556	21.175	33.196	66.182	65.604
2000	14.577	10.534	28.717	1.548	22.662	12.603	67.813	68.102
2001	14.406	10.138	29.131	1.510	25.768	12.777	71.317	53.188
2002	17.145	10.331	33.035	1.532	25.033	52.299	76.434	69.786
2003	17.975	11.902	32.546	1.496	30.829	12.003	82.756	116.794
2004	16.250	11.484	30.130	1.420	38.340	11.031	84.464	131.910
2005	18.202	10.741	30.798	1.403	35.599	41.592	85.872	107.341
2006	17.370	11.365	31.286	1.376	36.213	19.761	86.329	95.676
2007	15.714	11.159	26.428	1.306	40.650	13.543	88.462	104.841
2008	16.288	9.550	25.452	1.365	38.326	7.708	85.370	114.430
2009	16.312	9.638	26.699	1.444	33.843	8.682	79.778	91.673
2010	15.221	9.991	26.181	1.412	29.765	23.790	75.160	81.642
2011	12.967	9.497	25.999	1.335	29.538	13.519	71.181	67.053
2012	11.674	8.981	24.581	1.257	31.686	8.217	70.003	61.248
2013	11.693	8.289	23.753	1.214	29.949	6.107	67.677	62.410
2014	11.591	8.078	22.120	1.195	26.582	3.500	63.268	114.103
2015	10.211	7.534	19.491	1.186	22.785	5.830	56.700	64.240
2016	8.556	6.554	16.759	1.188	19.373	10.633	50.152	61.231
2017	7.005	5.539	14.291	1.179	17.541	4.220	45.170	52.922
2018	6.272	4.651	13.142	1.201	15.966	7.078	41.712	28.932
2019	6.965	4.393	14.003	1.340	13.889	3.713	39.442	28.744
2020	7.293	4.861	14.802	1.186	12.325	21.170		

Table 10. Comparison of projected mature male biomass (1000 t) on Feb. 15 and their 95% limits with four levels of fishing mortality during 2020-2030. Parameter estimates with model 19.3a are used for the projection with recruitments randomly drawn from estimated recruitments from 2012 to 2019. Fishing mortality of 0.167 is about estimated  $F_{off}$  for Model 19.3a for 2020.

	F=0			F=0.083		
	Mean	2.5% limit	97.5% limit	Mean	2.5% limit	97.5% limit
2020	16.559	15.055	17.985	15.562	14.142	16.896
2021	18.365	16.408	20.181	16.365	14.543	18.058
2022	19.274	17.074	21.720	16.340	14.399	18.530
2023	19.876	17.551	22.607	16.136	14.145	18.508
2024	20.567	18.082	23.657	16.154	13.986	18.811
2025	21.251	18.268	24.662	16.273	13.670	19.145
2026	21.883	18.439	25.880	16.425	13.441	19.680
2027	22.451	18.484	26.760	16.579	13.304	20.149
2028	22.906	18.886	27.598	16.678	13.385	20.426
2029	23.305	19.103	28.054	16.772	13.439	20.390
2030	23.677	19.278	28.473	16.881	13.420	20.644

	F=0.167			F=0.250		
	Mean	2.5% limit	97.5% limit	Mean	2.5% limit	97.5% limit
2020	14.638	13.299	15.885	13.780	12.514	14.939
2021	14.629	12.942	16.223	13.122	11.551	14.613
2022	13.950	12.205	15.930	11.996	10.410	13.832
2023	13.267	11.564	15.364	11.051	9.580	12.925
2024	12.951	10.999	15.183	10.597	8.846	12.625
2025	12.833	10.581	15.242	10.409	8.396	12.557
2026	12.809	10.170	15.613	10.346	8.016	12.819
2027	12.829	10.086	15.747	10.340	7.946	12.939
2028	12.821	10.045	15.907	10.314	7.852	12.899
2029	12.833	10.068	15.891	10.312	7.945	12.854
2030	12.877	10.035	16.016	10.346	7.908	12.898

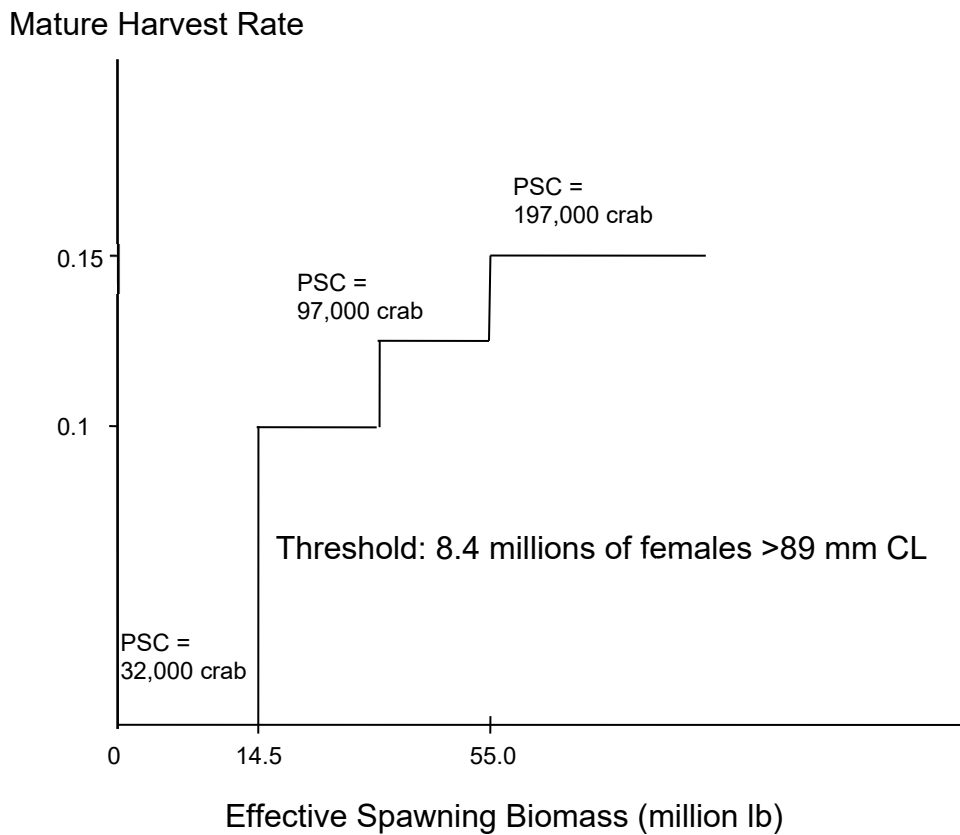


Figure 1. Current harvest rate strategy (line) for the Bristol Bay red king crab fishery and annual prohibited species catch (PSC) limits (numbers of crab) of Bristol Bay red king crab in the groundfish fisheries in zone 1 in the eastern Bering Sea. Harvest rates are based on current-year estimates of effective spawning biomass (ESB, Zheng et al. 1995b), whereas PSC limits apply to previous-year ESB.

## Data by type and year

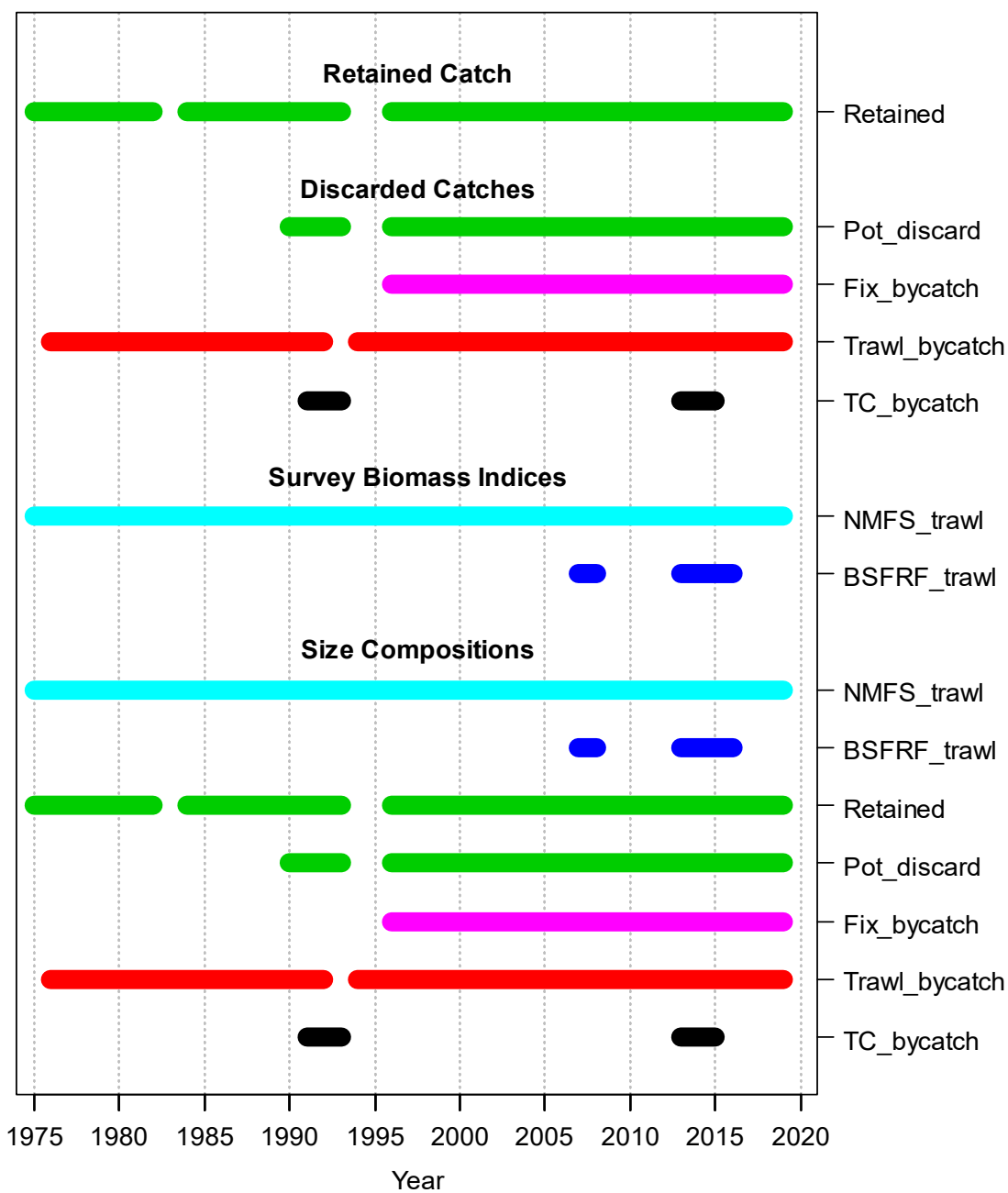


Figure 2a. Data types and ranges used for the stock assessment with model 19.3.

## Data by type and year

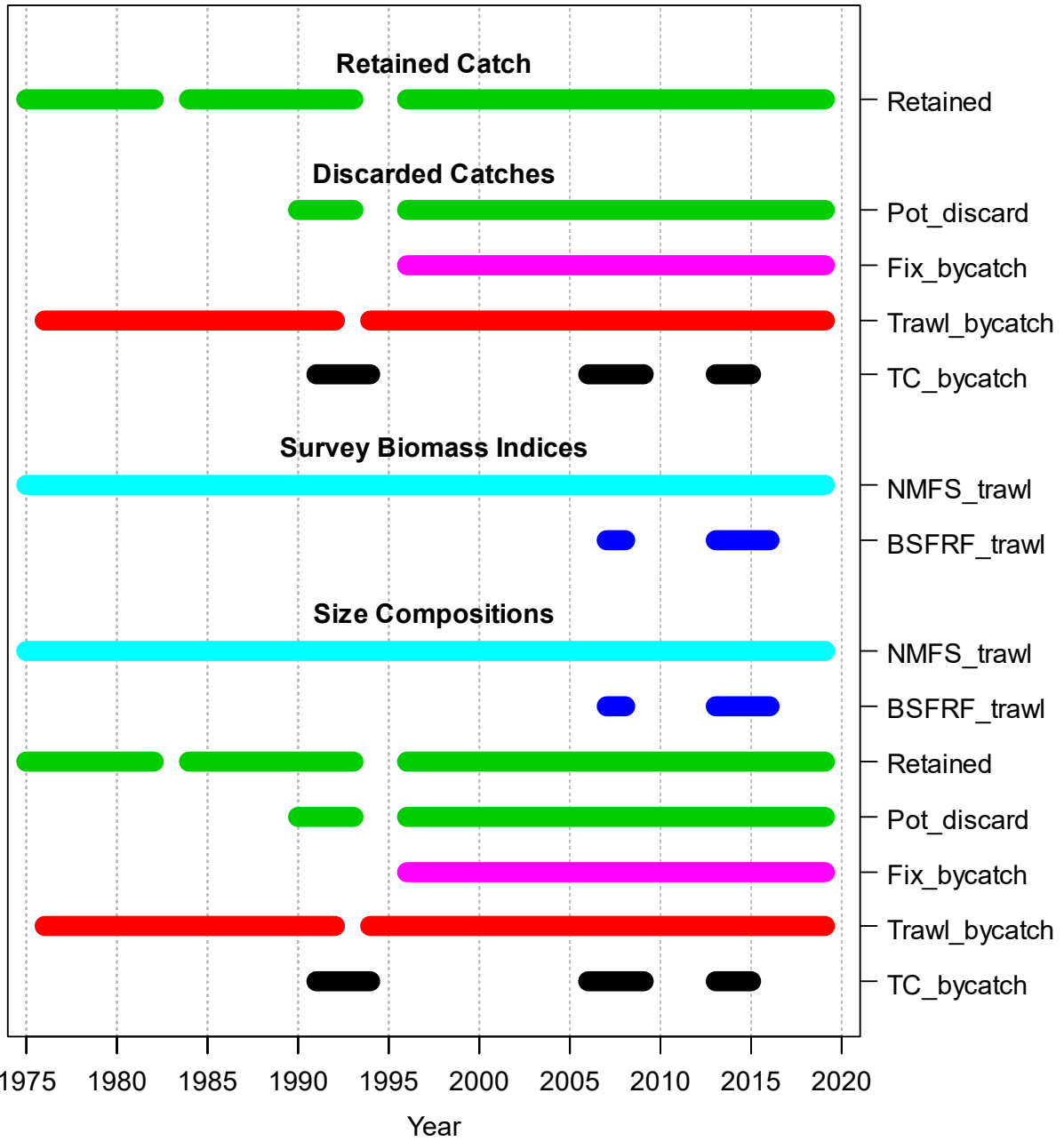


Figure 2b. Data types and ranges used for the stock assessment with the other seven models (19.3c-19.6).

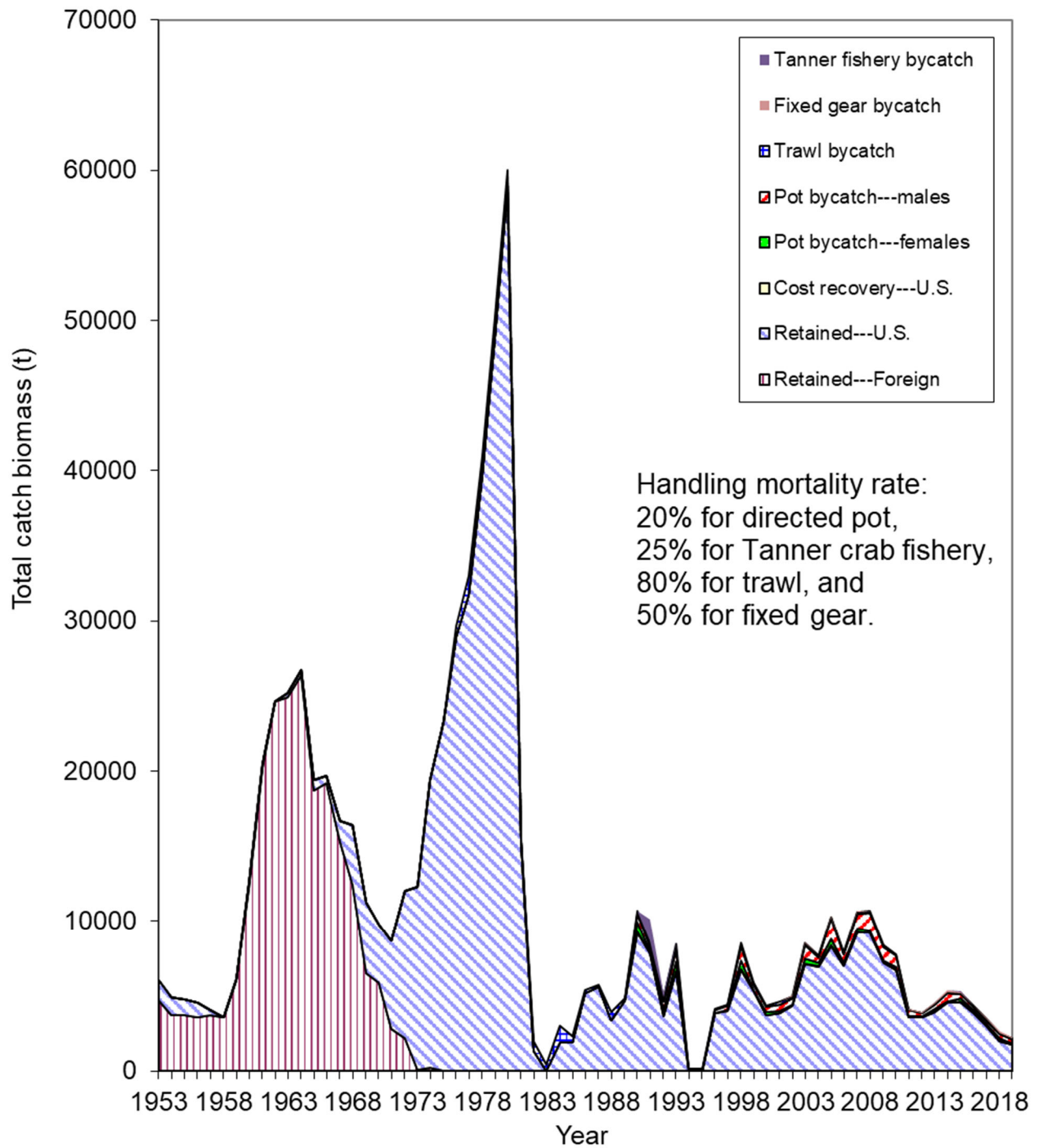


Figure 3. Retained catch biomass and bycatch mortality biomass (t) for Bristol Bay red king crab from 1953 to 2019. Directed pot bycatch data were not available from the observer program before 1990 and are not included in this figure.

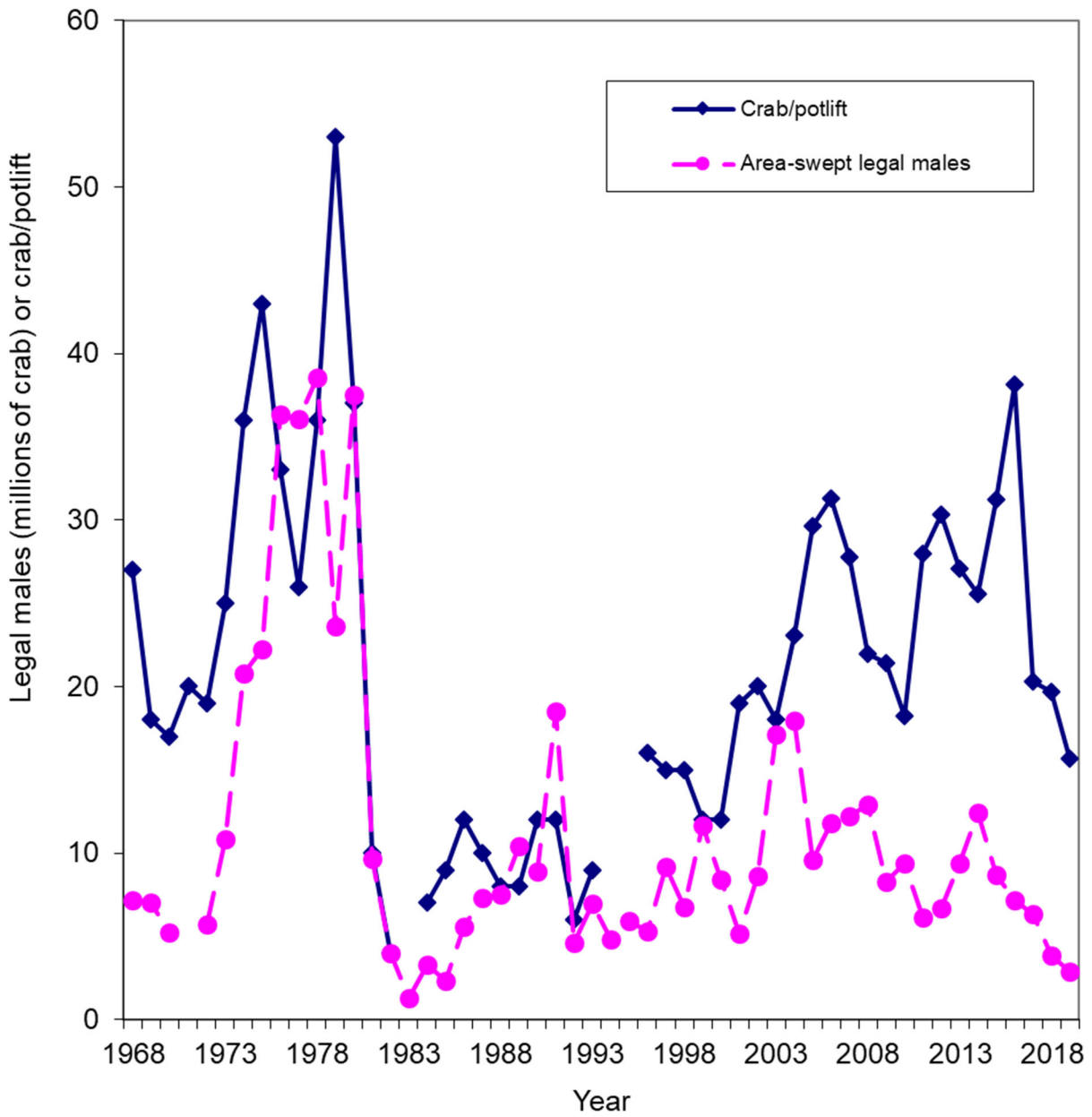


Figure 4. Comparison of survey legal male abundances and catches per unit effort for Bristol Bay red king crab from 1968 to 2019.



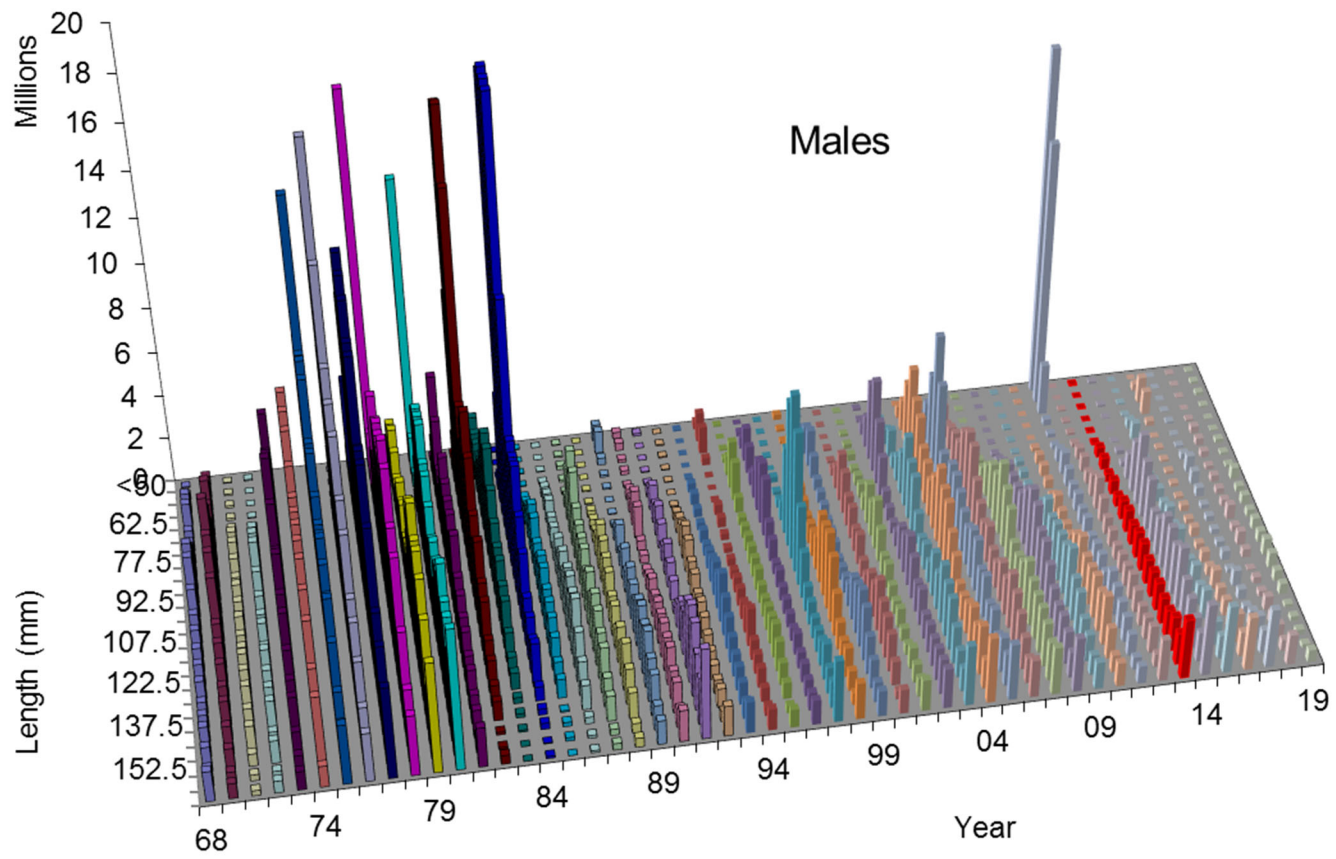


Figure 5a. Survey abundances by 5-mm carapace length bin for male Bristol Bay red king crab from 1968 to 2019.

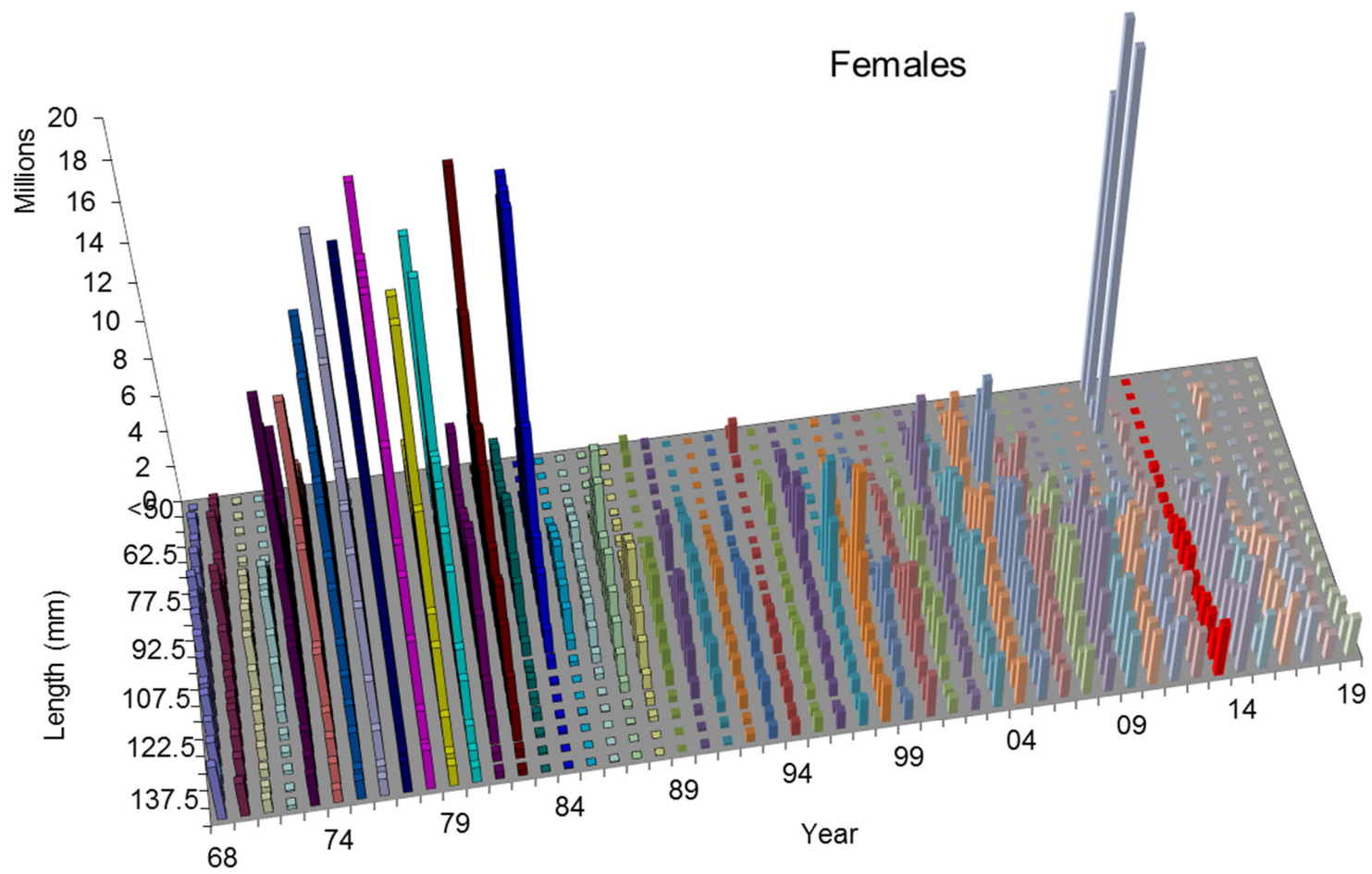


Figure 5b. Survey abundances by 5 mm carapace length bin for female Bristol Bay red king crab from 1968 to 2019.

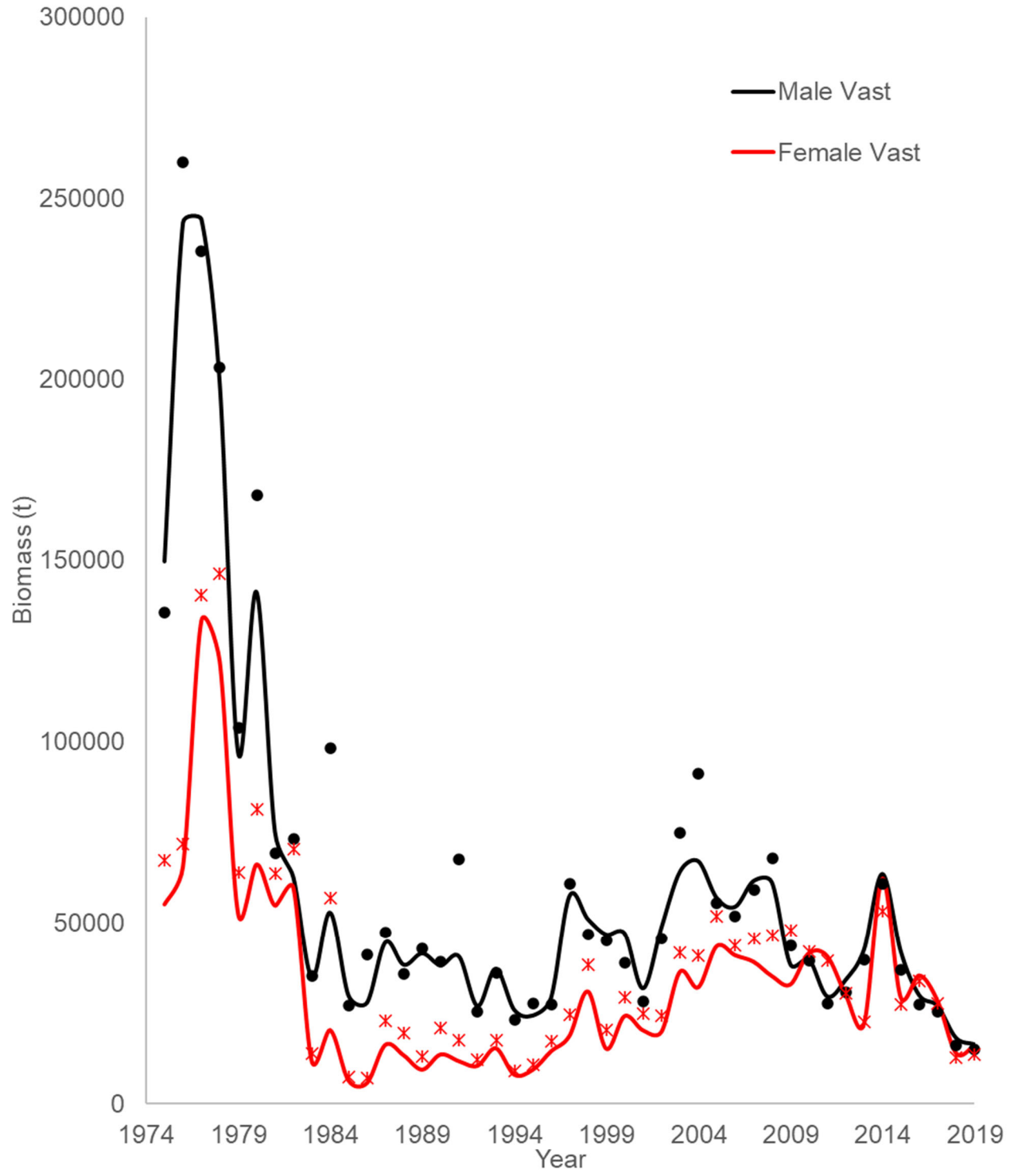


Figure 6. Comparison of area-swept and VAST-estimated survey biomasses for Bristol Bay red king crab from 1975 to 2019.

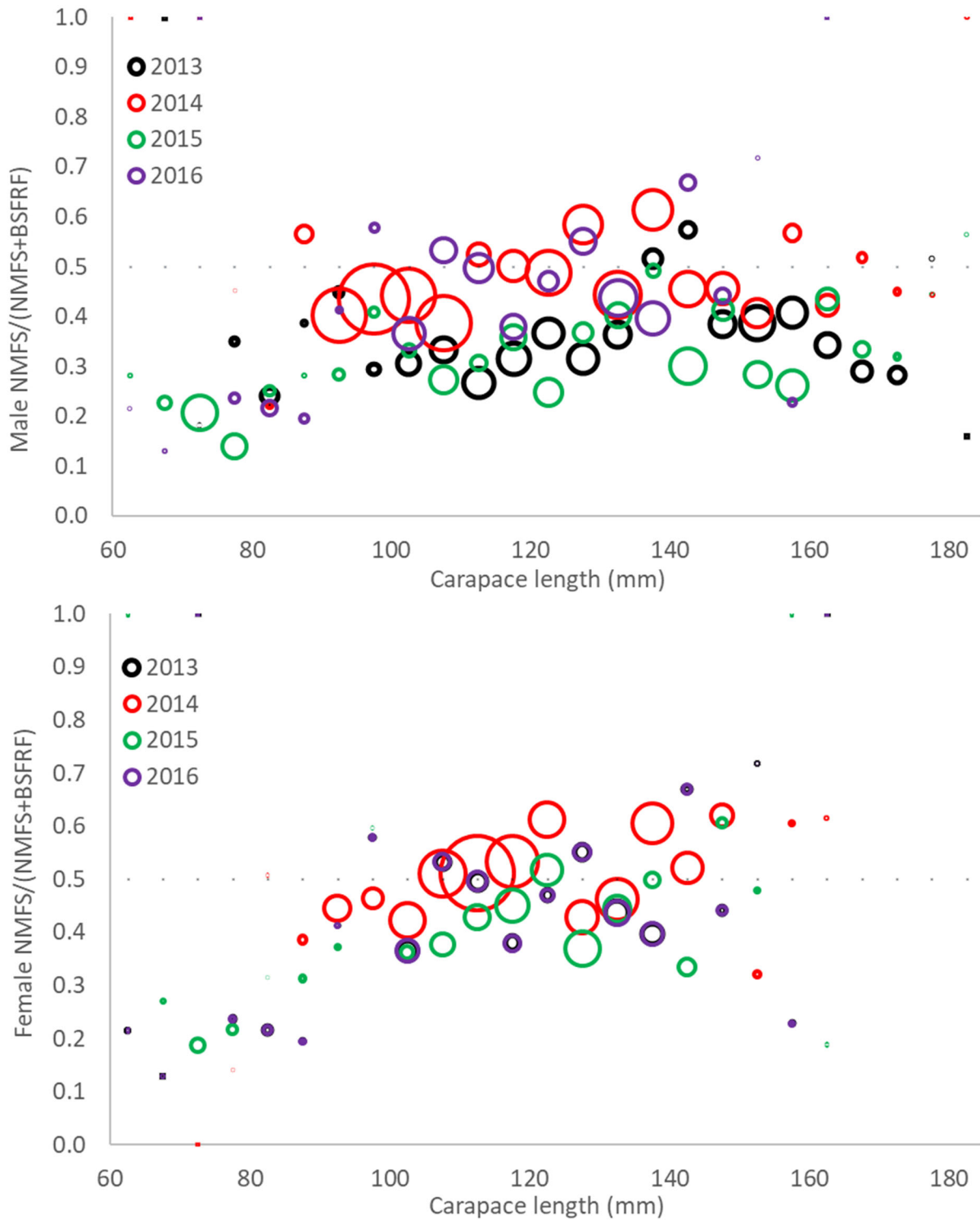


Figure 7a. Comparison of NMFS survey abundance proportions of total NMFS and BSFRF side-by-side trawl surveys during 2013-2016 for Bristol Bay red king crab. Sizes of circles are proportional to total abundances.

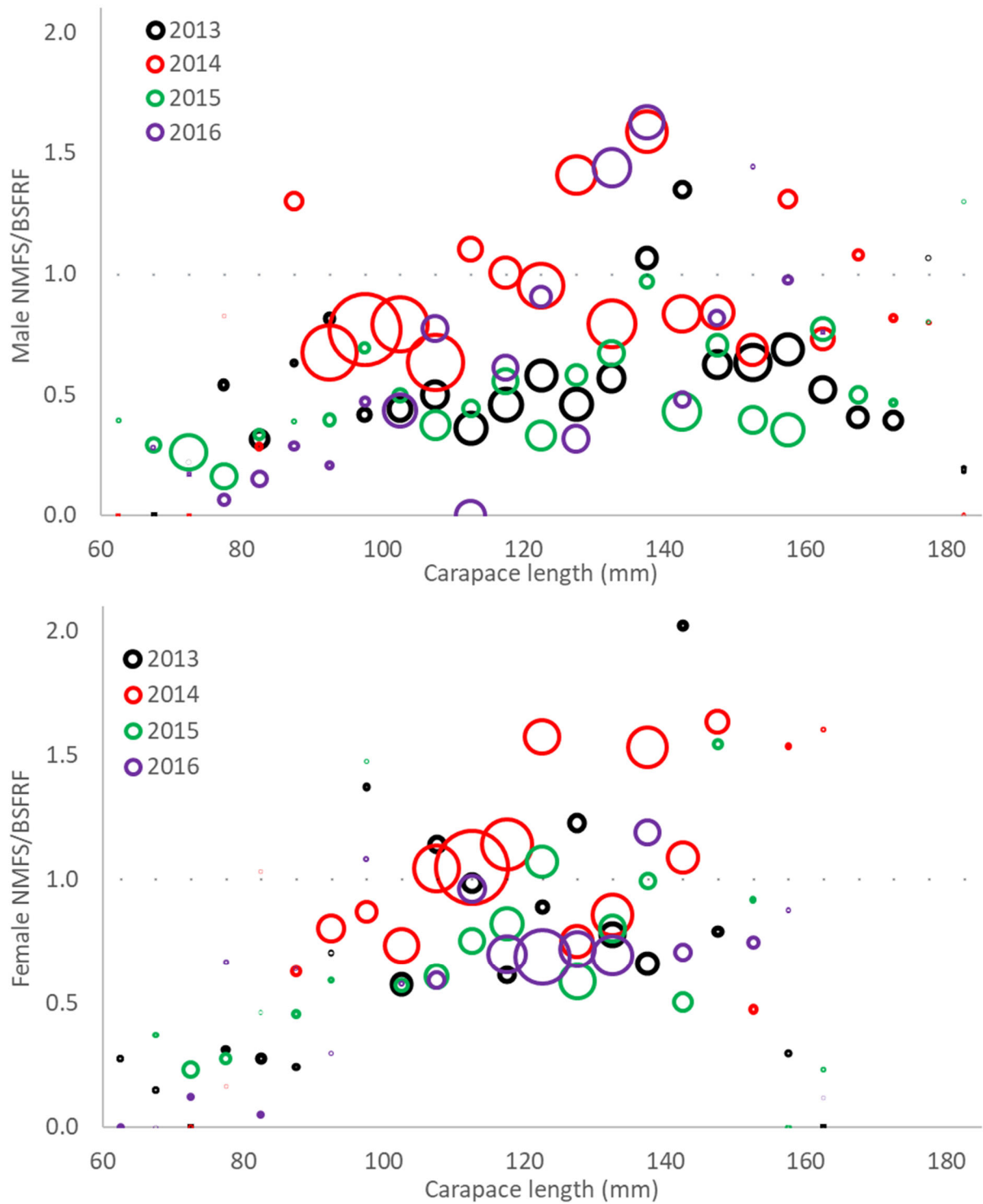


Figure 7b. Comparison of ratios of NMFS survey abundances to BSFRF side-by-side survey abundances during 2013-2016 for Bristol Bay red king crab. Sizes of circles are proportional to total abundances.

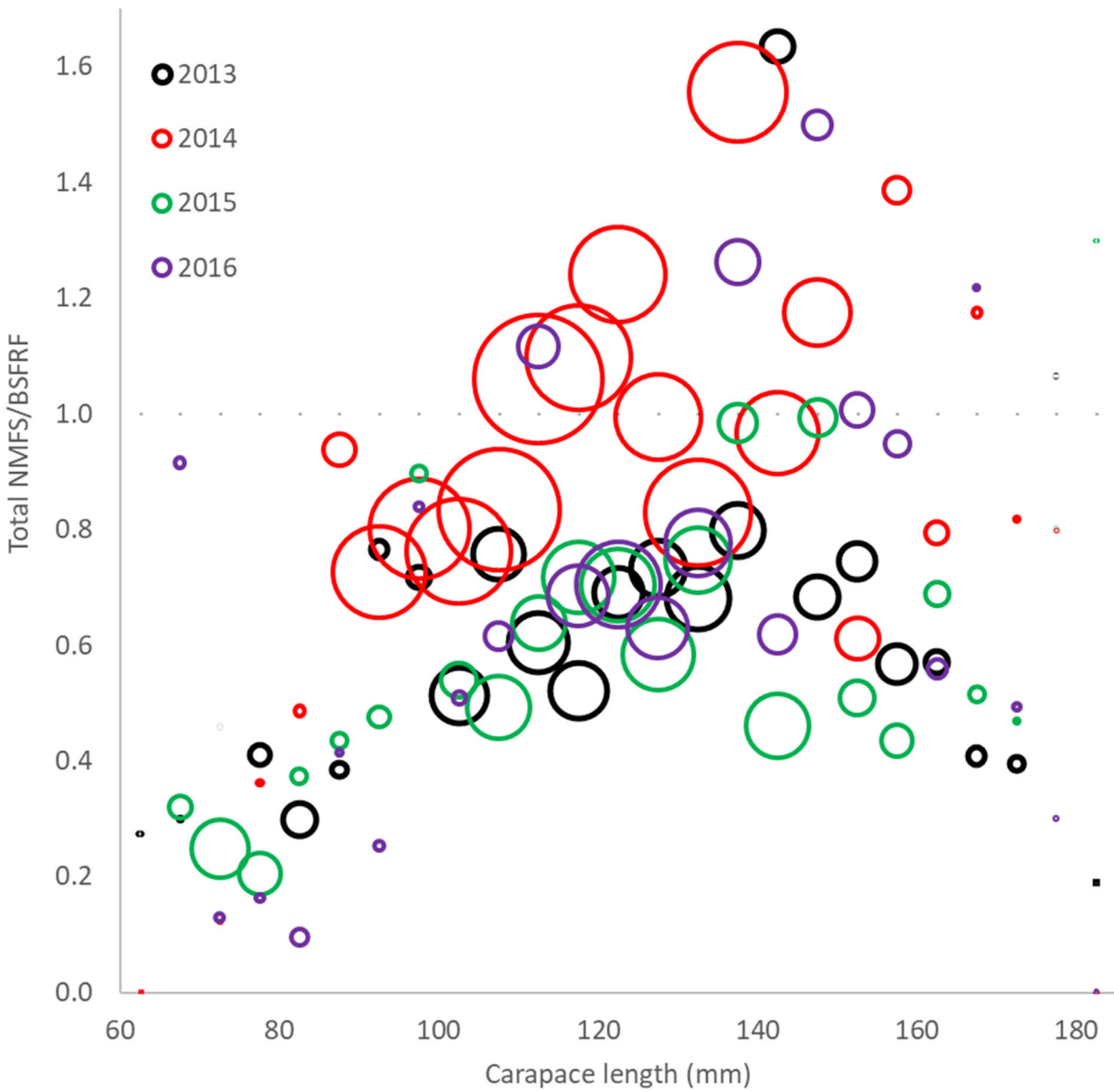


Figure 7c. Comparison of ratios of NMFS survey abundances to BSFRF side-by-side survey abundances during 2013-2016 for Bristol Bay red king crab. Sizes of circles are proportional to total abundances. The abundance-weighted average ratio is 0.891 for crab  $\geq 135$  mm carapace length from all four years of data. The approach to compute this overall ratio is documented in section D. Data, 4. Bering Sea Fisheries Research Foundation Survey Data.

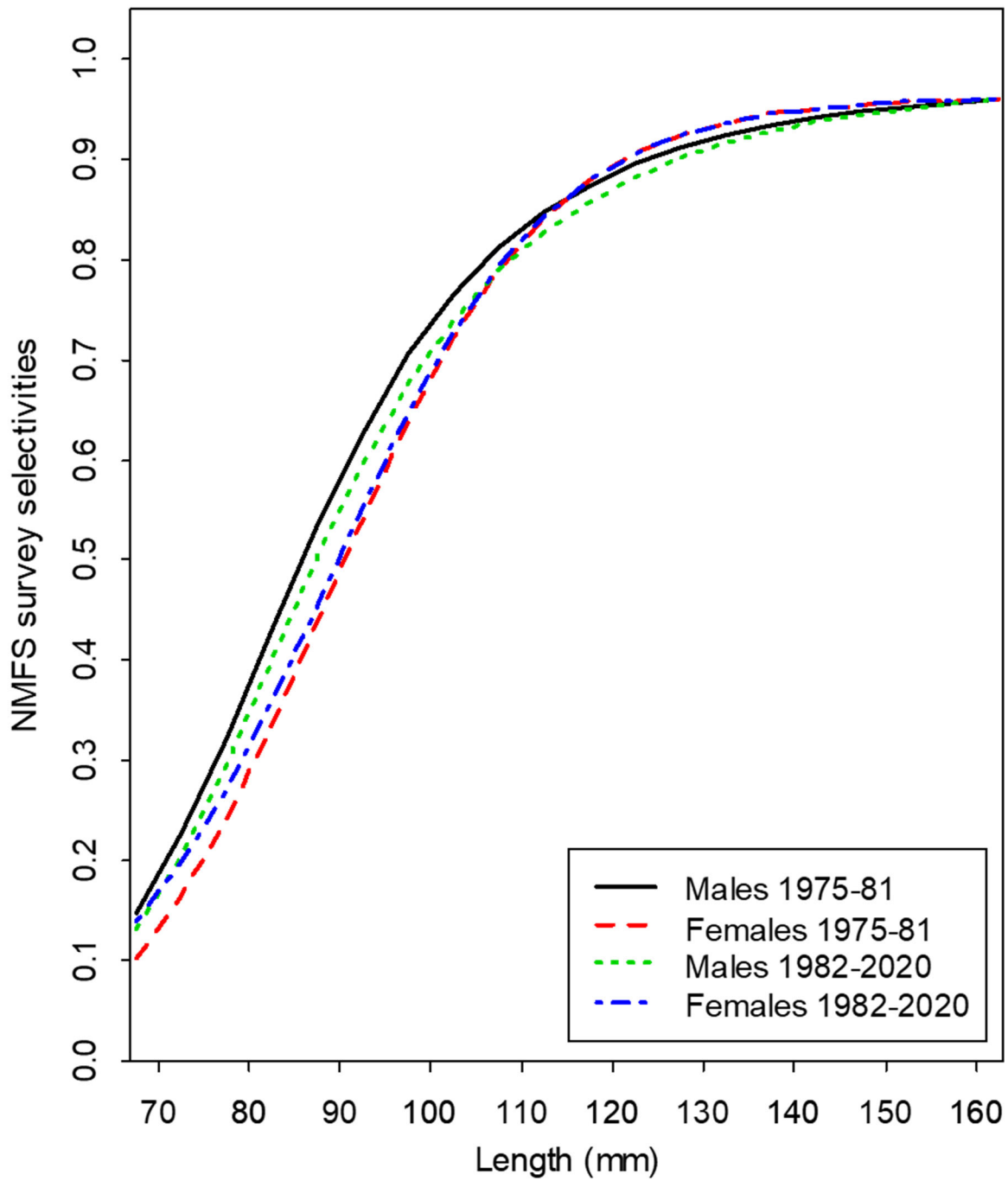


Figure 8a. Estimated NMFS trawl survey selectivities under model 19.3.

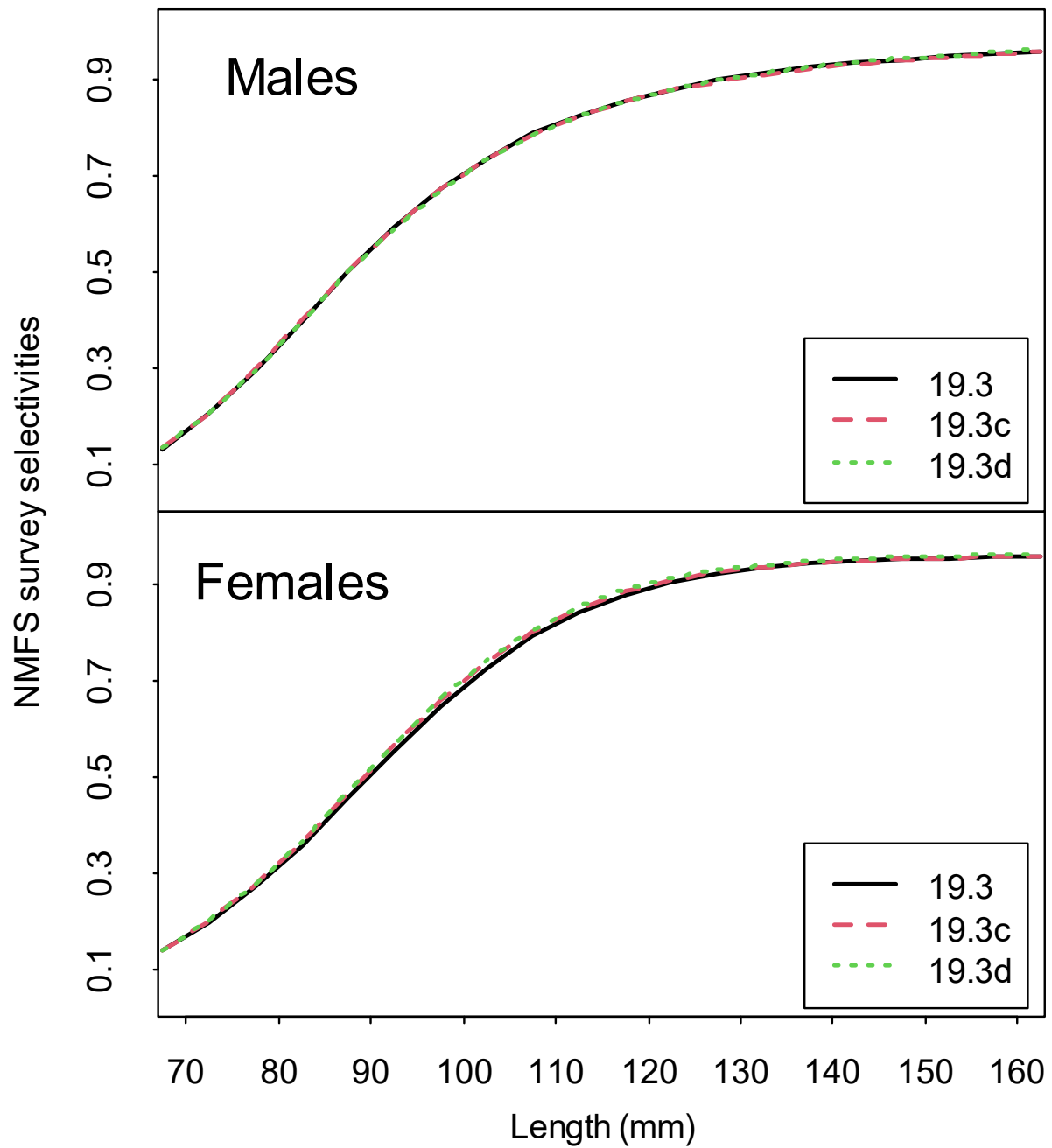


Figure 8b. Comparisons of estimated NMFS survey selectivities with models 19.3, 19.3c, and 19.3d during 1982-2020.



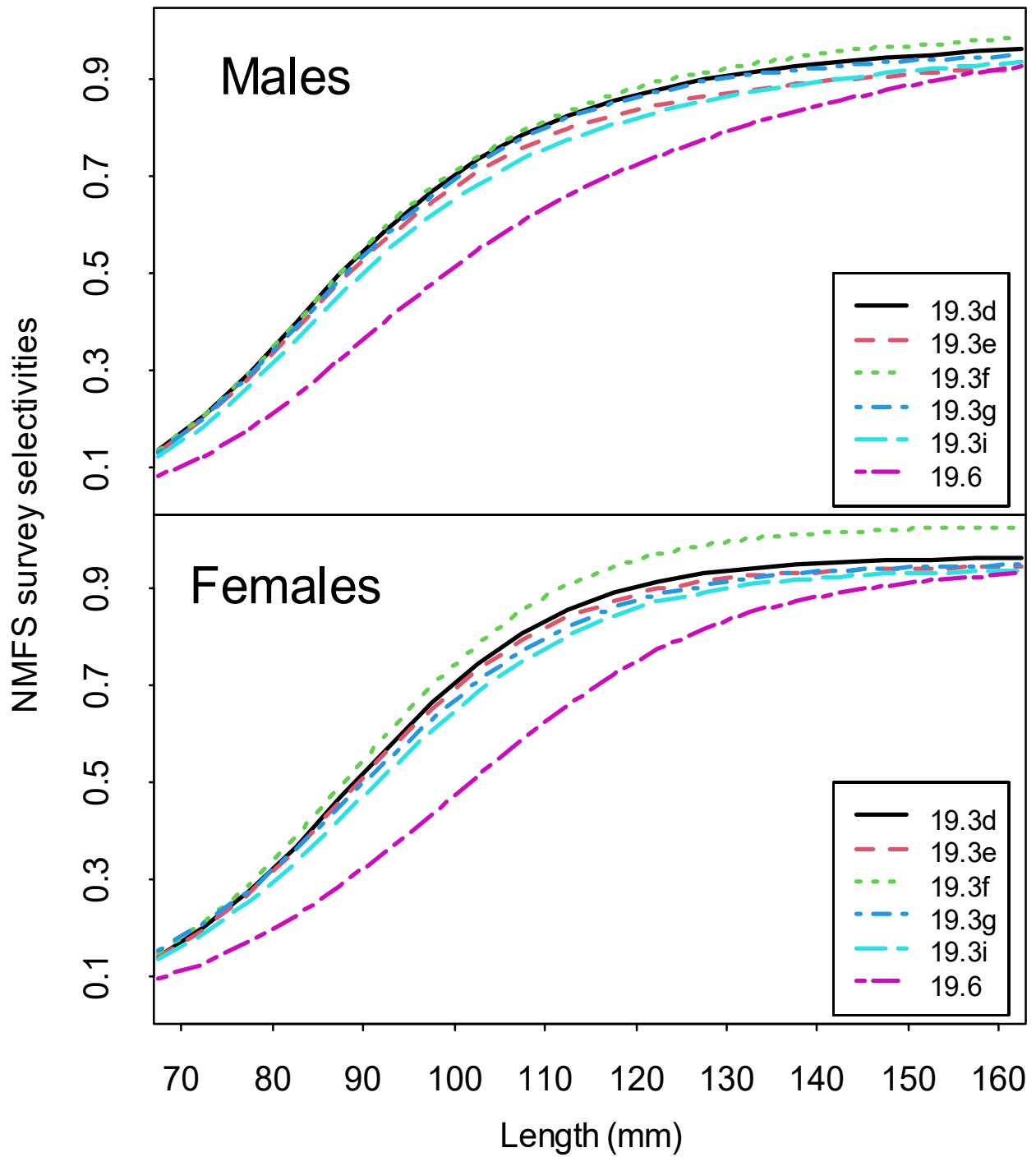


Figure 8c. Comparisons of estimated NMFS survey selectivities with models 19.3d, 19.3e, 19.3f, 19.3g, 19.3i, and 19.6 during 1982-2020.

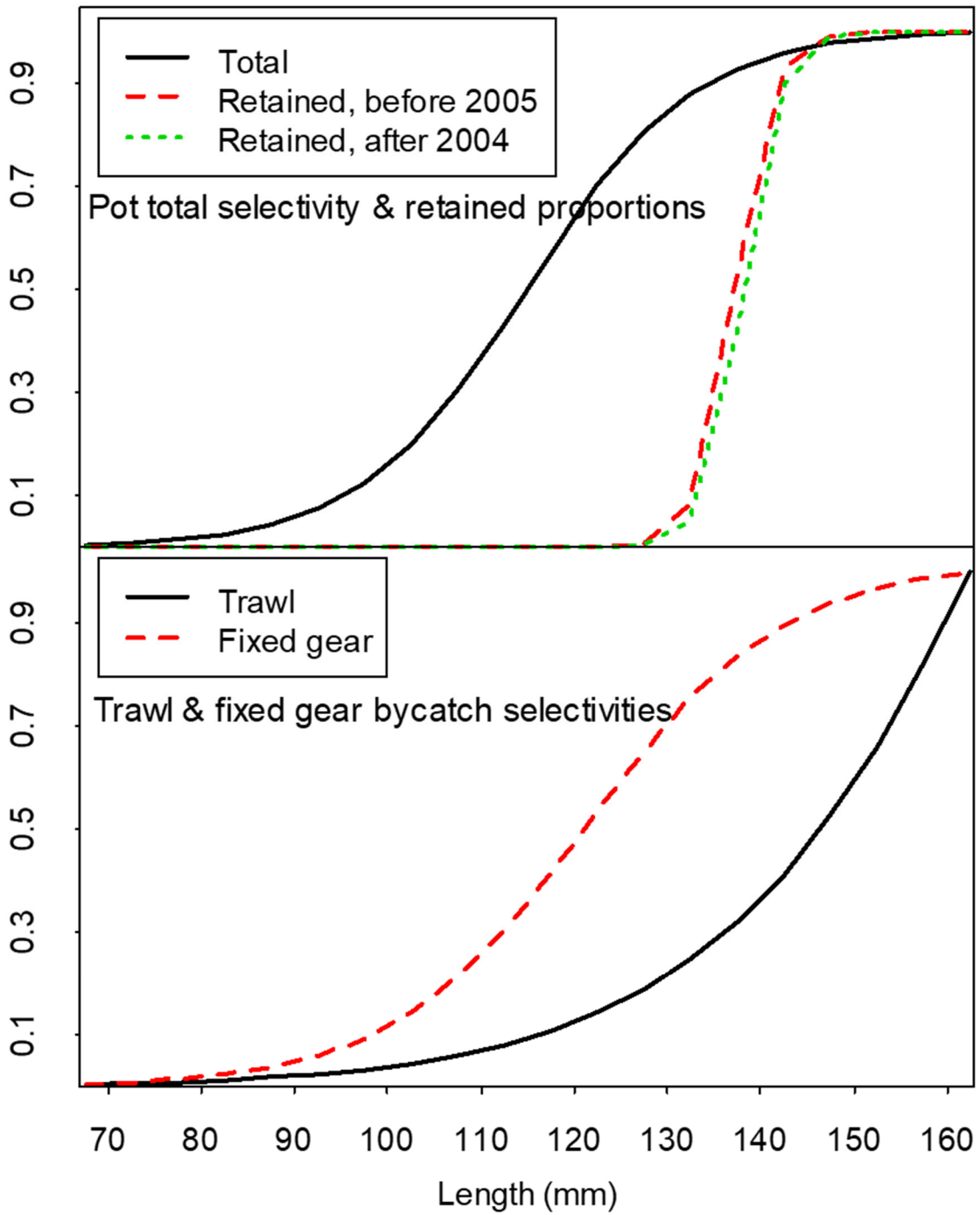


Figure 8d. Estimated total pot fishery selectivities and retained proportions and groundfish fisheries bycatch selectivities under model 19.3.

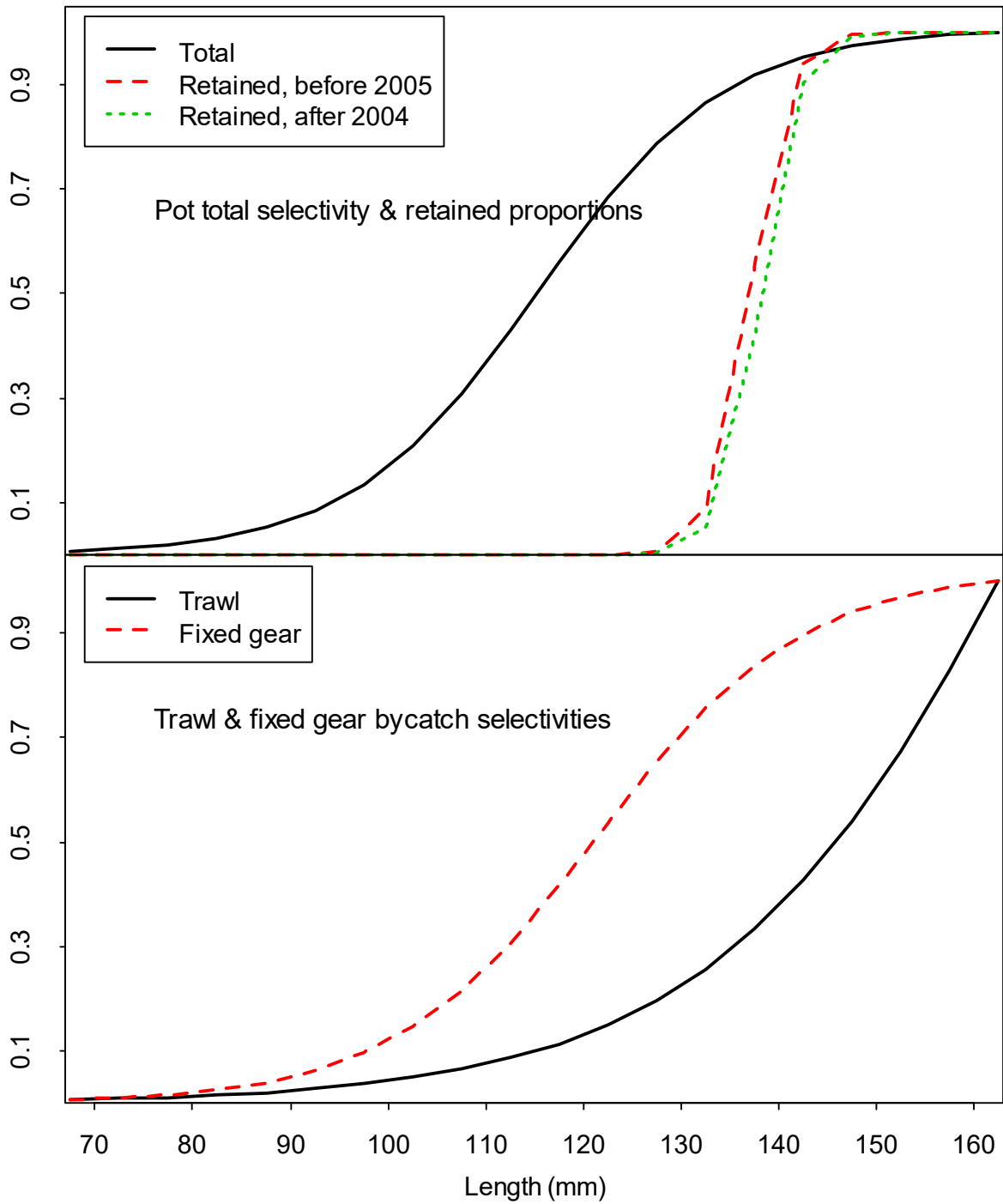


Figure 8e. Estimated total pot fishery selectivities and retained proportions and groundfish fisheries bycatch selectivities under model 19.3d.

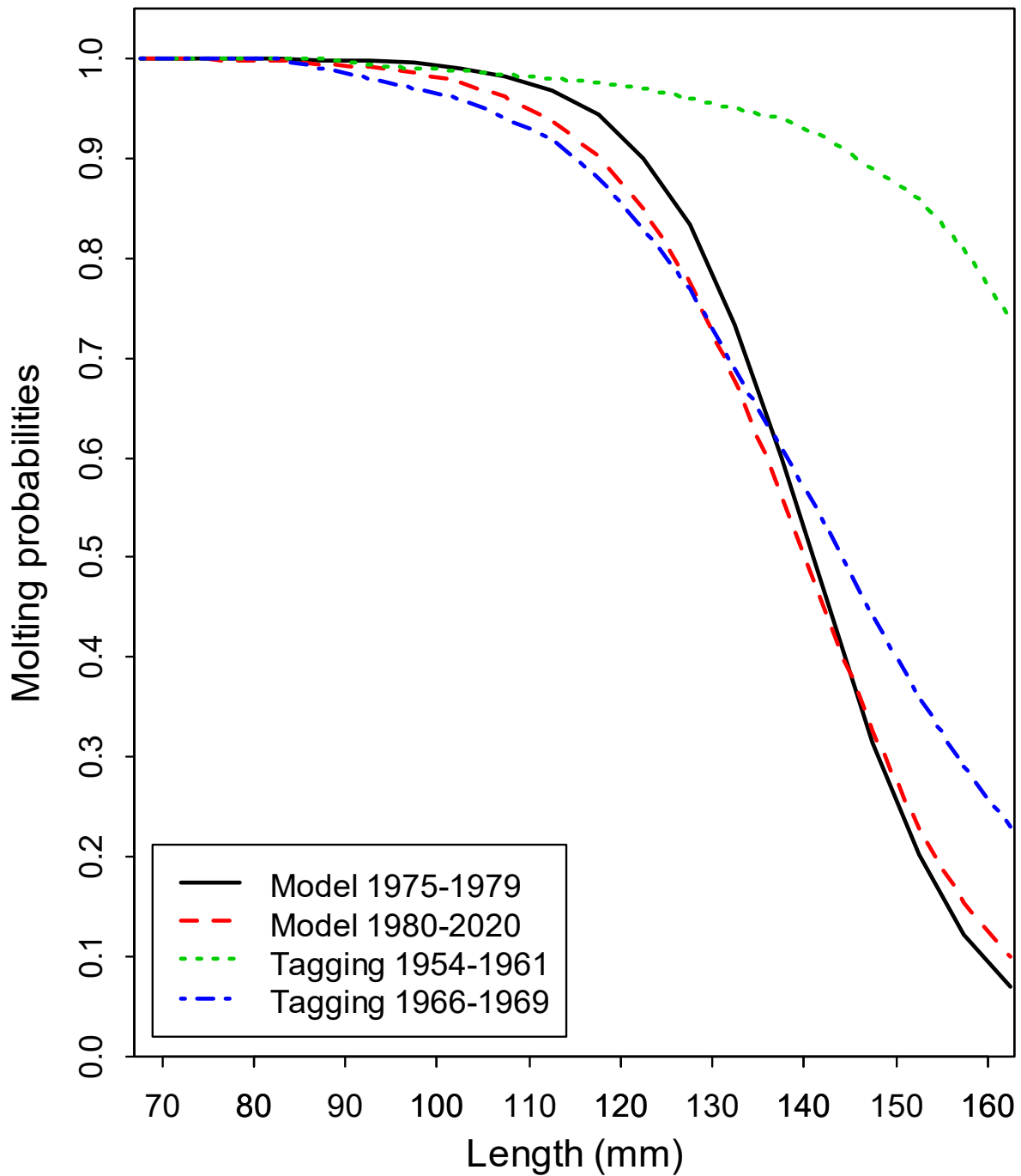


Figure 9a. Comparison of estimated probabilities of molting of male red king crab in Bristol Bay for different periods with model 19.3. Molting probabilities for periods 1954-1961 and 1966-1969 were estimated by Balsiger (1974) from tagging data. Molting probabilities for 1975-1979 and 1980-2020 were estimated with a length-based model.

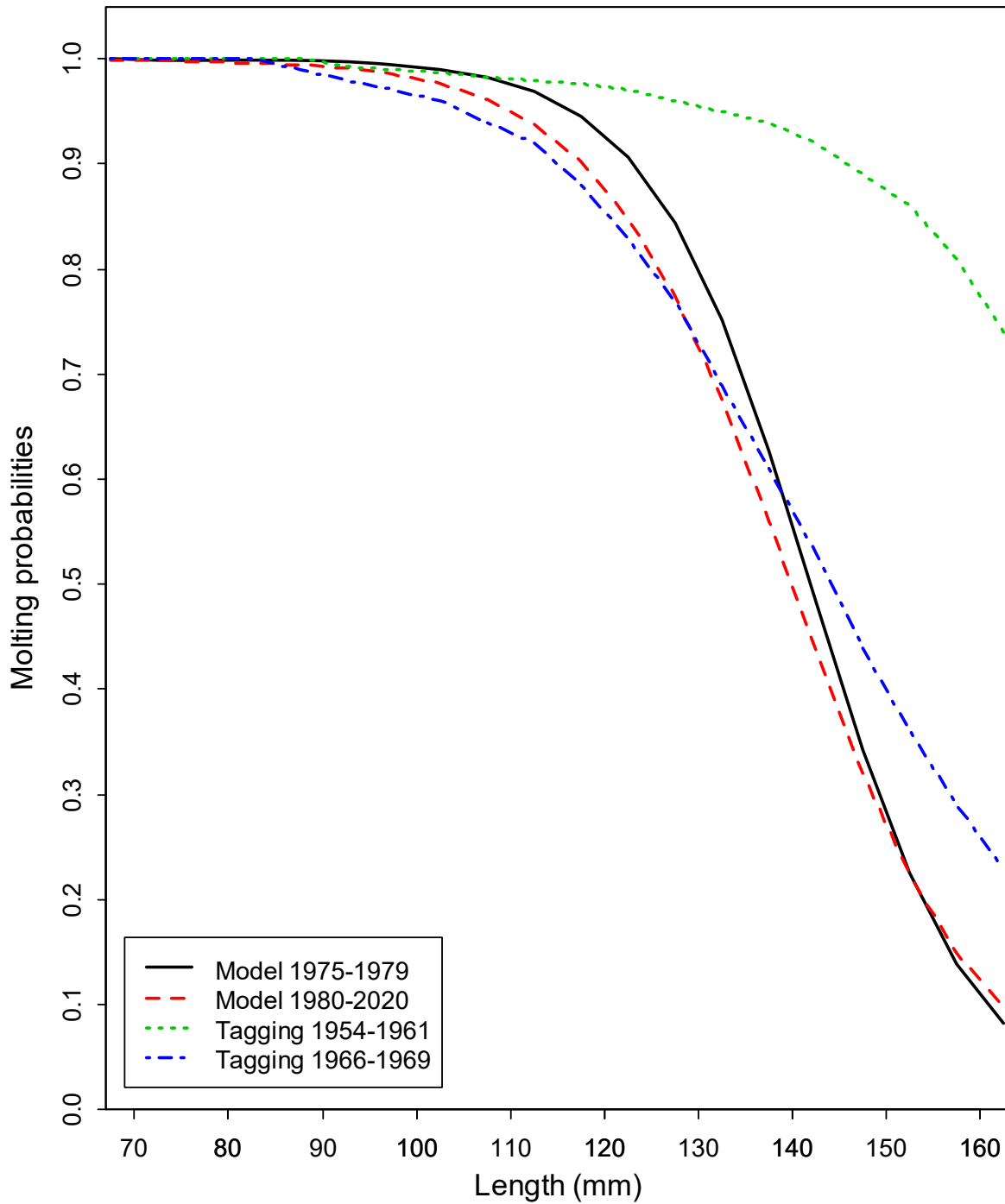


Figure 9b. Comparison of estimated probabilities of molting of male red king crab in Bristol Bay for different periods with model 19.3d. Molting probabilities for periods 1954-1961 and 1966-1969 were estimated by Balsiger (1974) from tagging data. Molting probabilities for 1975-1979 and 1980-2020 were estimated with a length-based model.

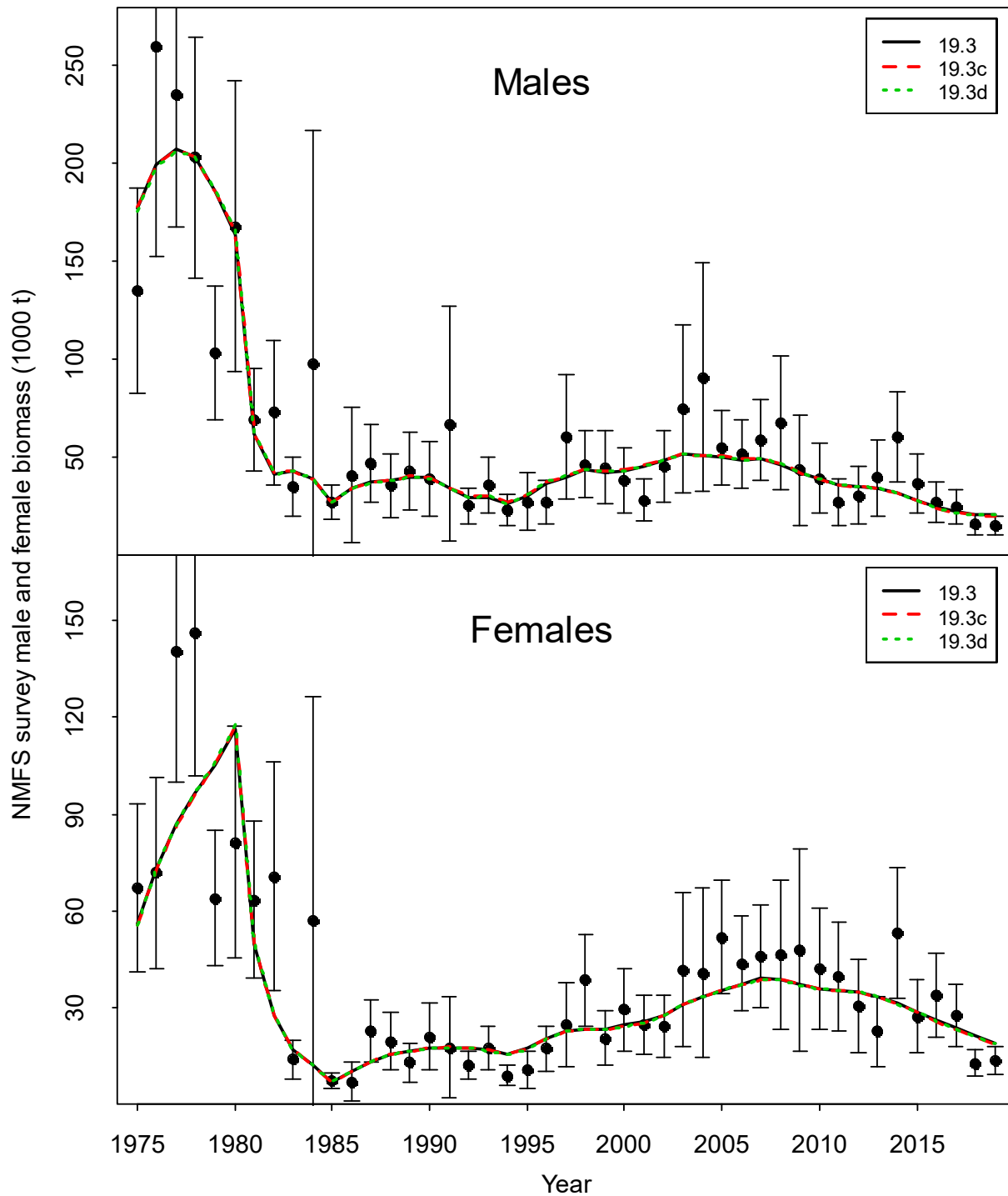


Figure 10a1. Comparisons of area-swept estimates of total NMFS survey biomass and model prediction for model estimates in 2020 under models 19.3, 19.3c, and 19.3d. The error bars are plus and minus 2 standard deviations of model 19.3.

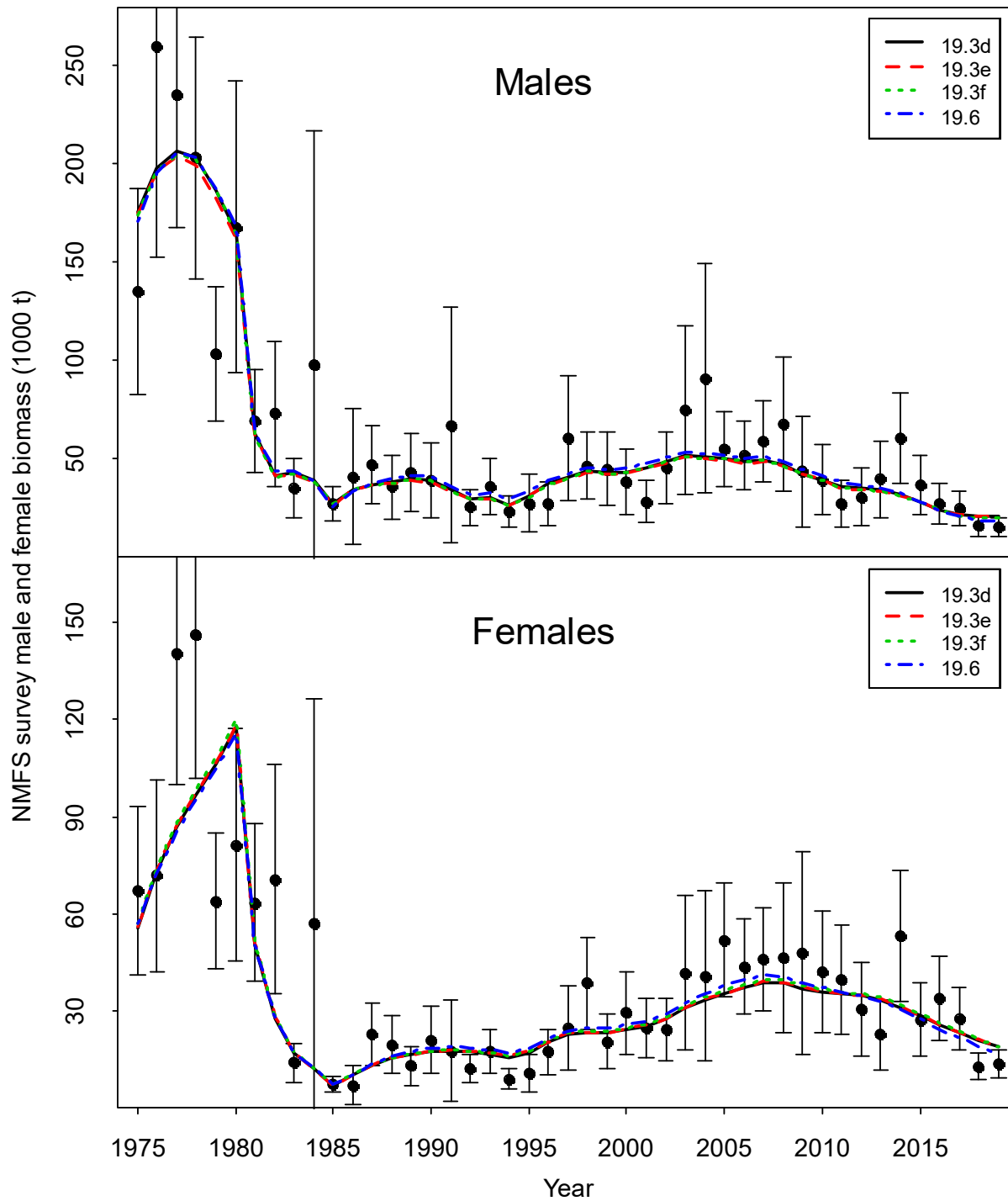


Figure 10a2. Comparisons of area-swept estimates of total NMFS survey biomass and model prediction for model estimates in 2020 under models 19.3d, 19.3e, 19.3f, and 19.6. The error bars are plus and minus 2 standard deviations of model 19.3d.

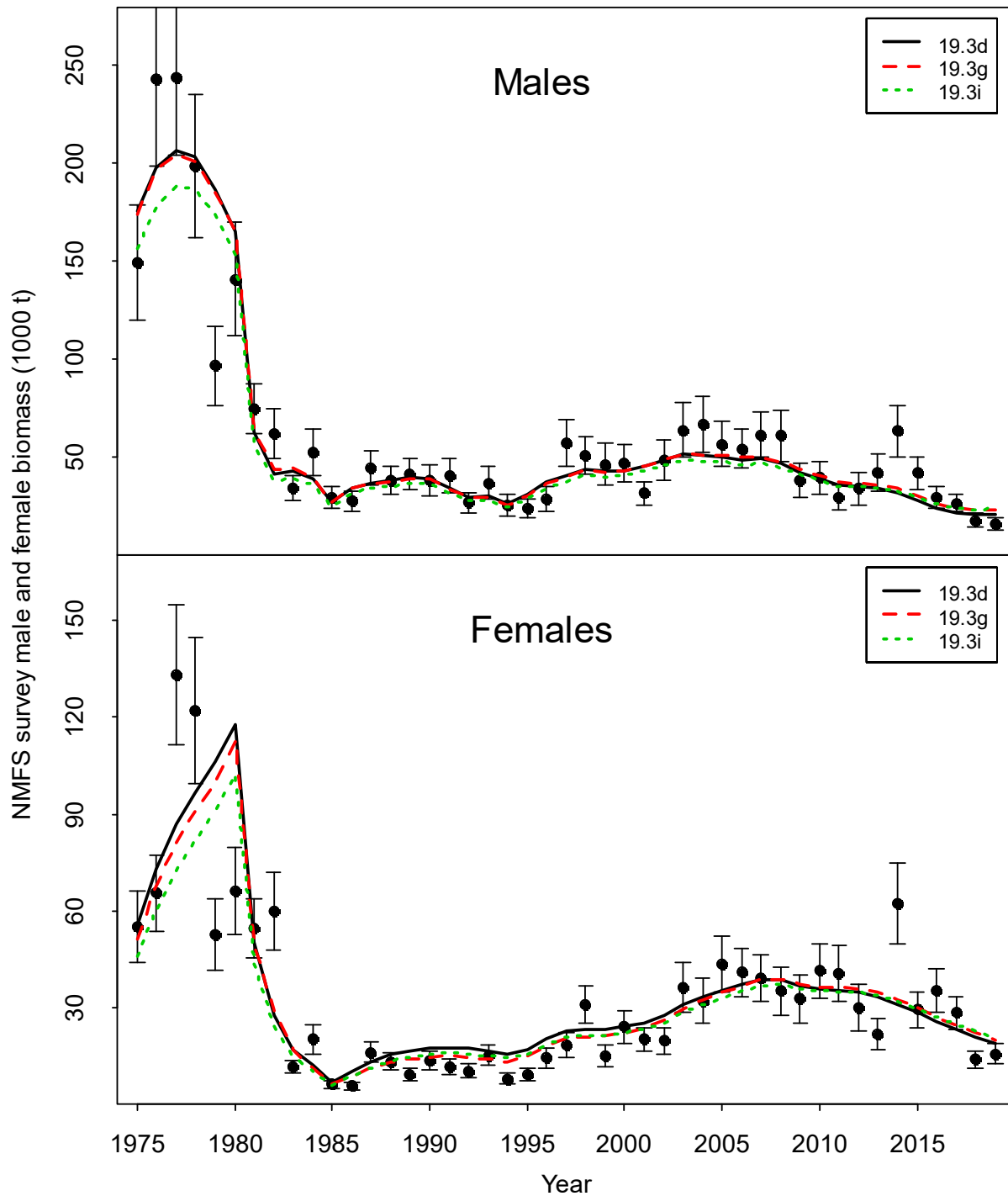


Figure 10a3. Comparisons of VAST estimates of total NMFS survey biomass and model prediction for model estimates in 2020 under models 19.3d, 19.3g, and 19.3i. The error bars are plus and minus 2 standard deviations of model 19.3g.



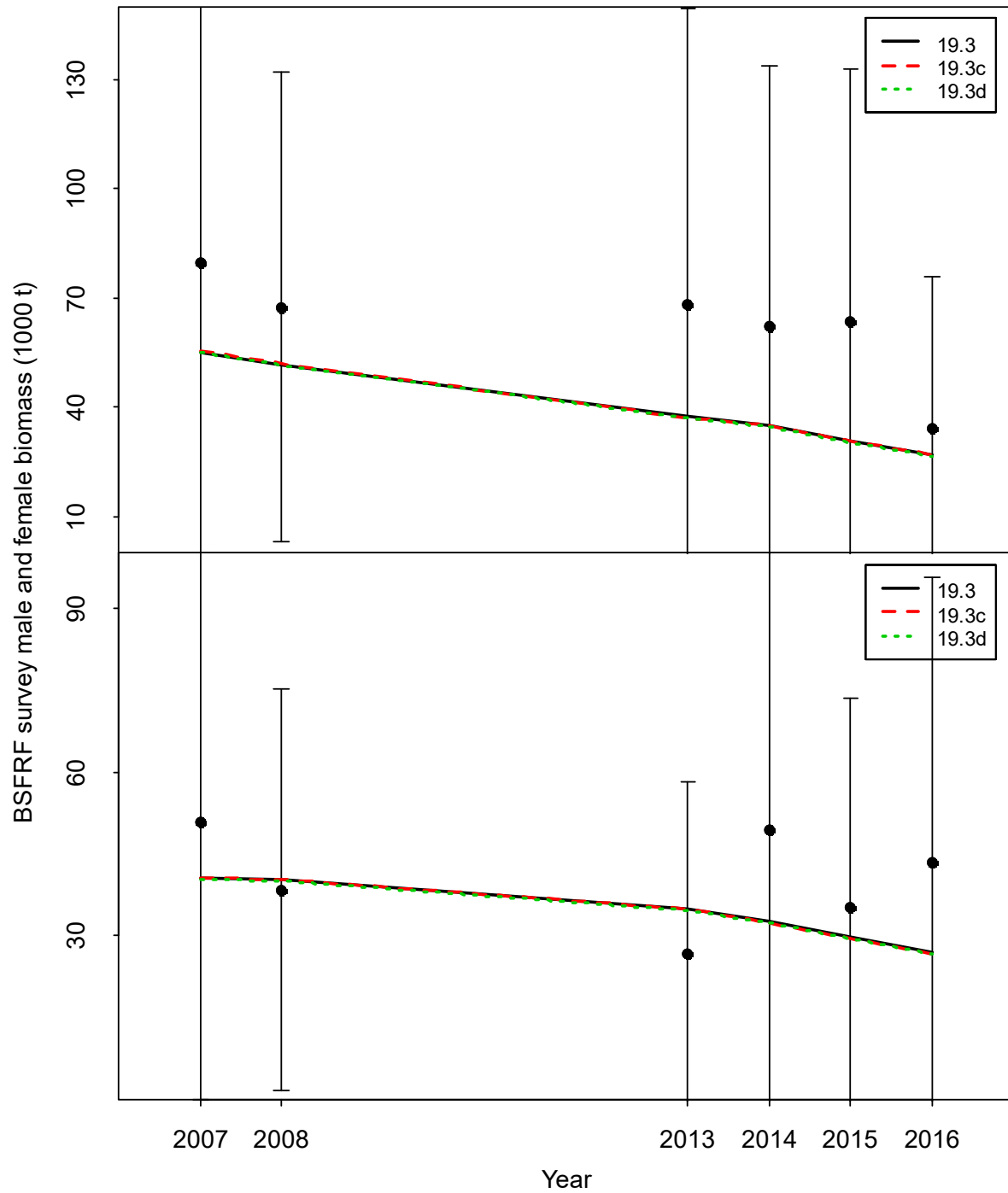


Figure 10b. Comparisons of survey biomass estimates by sex (upper plot for males and lower plot for females) by the BSFRF survey and the model for model estimates in 2020 (models 19.3, 19.3c, and 19.3d). The error bars are plus and minus 2 standard deviations of model 19.3.

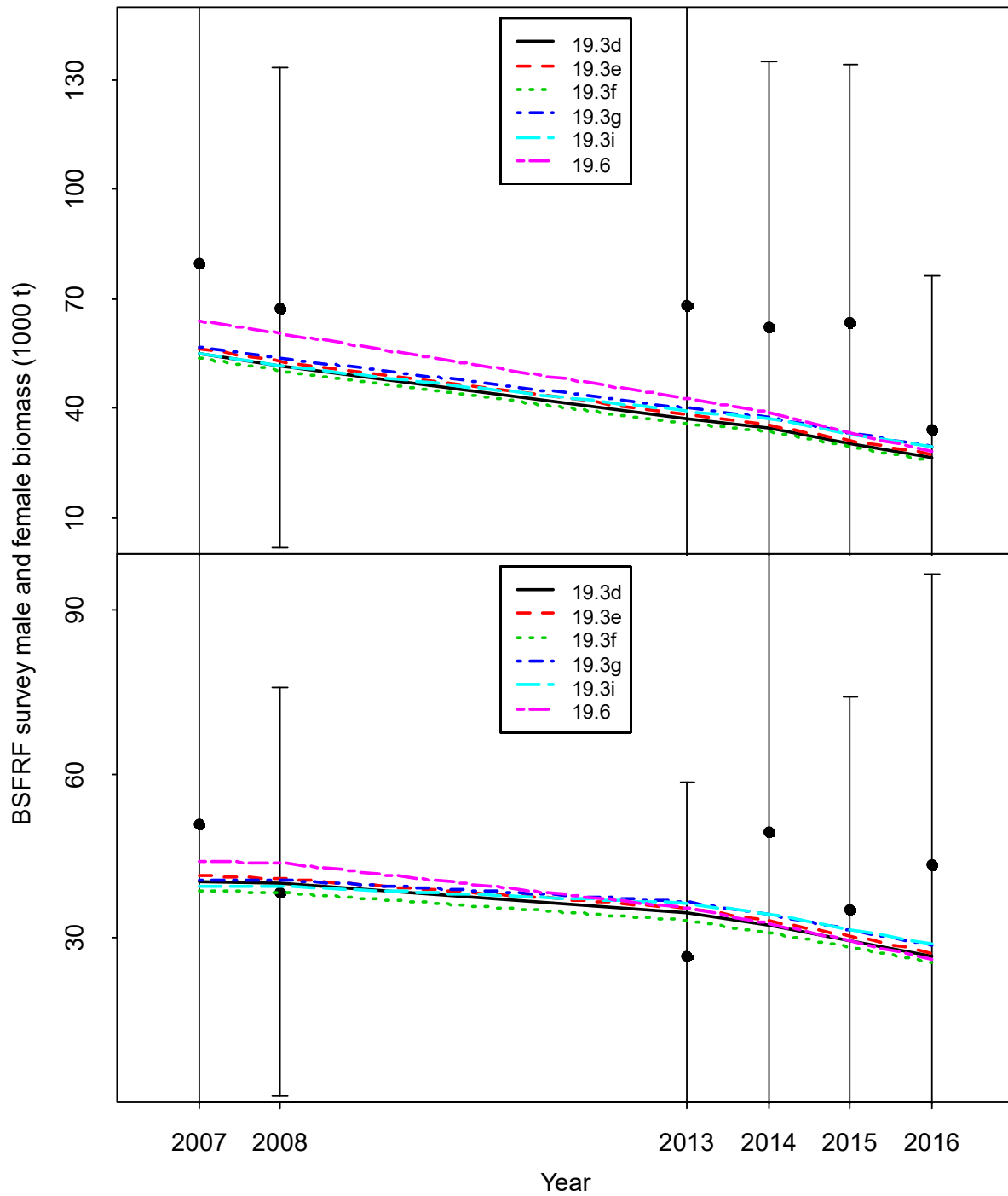


Figure 10c. Comparisons of survey biomass estimates by sex (upper plot for males and lower plot for females) by the BSFRF survey and the model for model estimates in 2020 (models 19.3d, 19.3e, 19.3f, 19.3g, 19.3i, and 19.6). The error bars are plus and minus 2 standard deviations of model 19.3d.

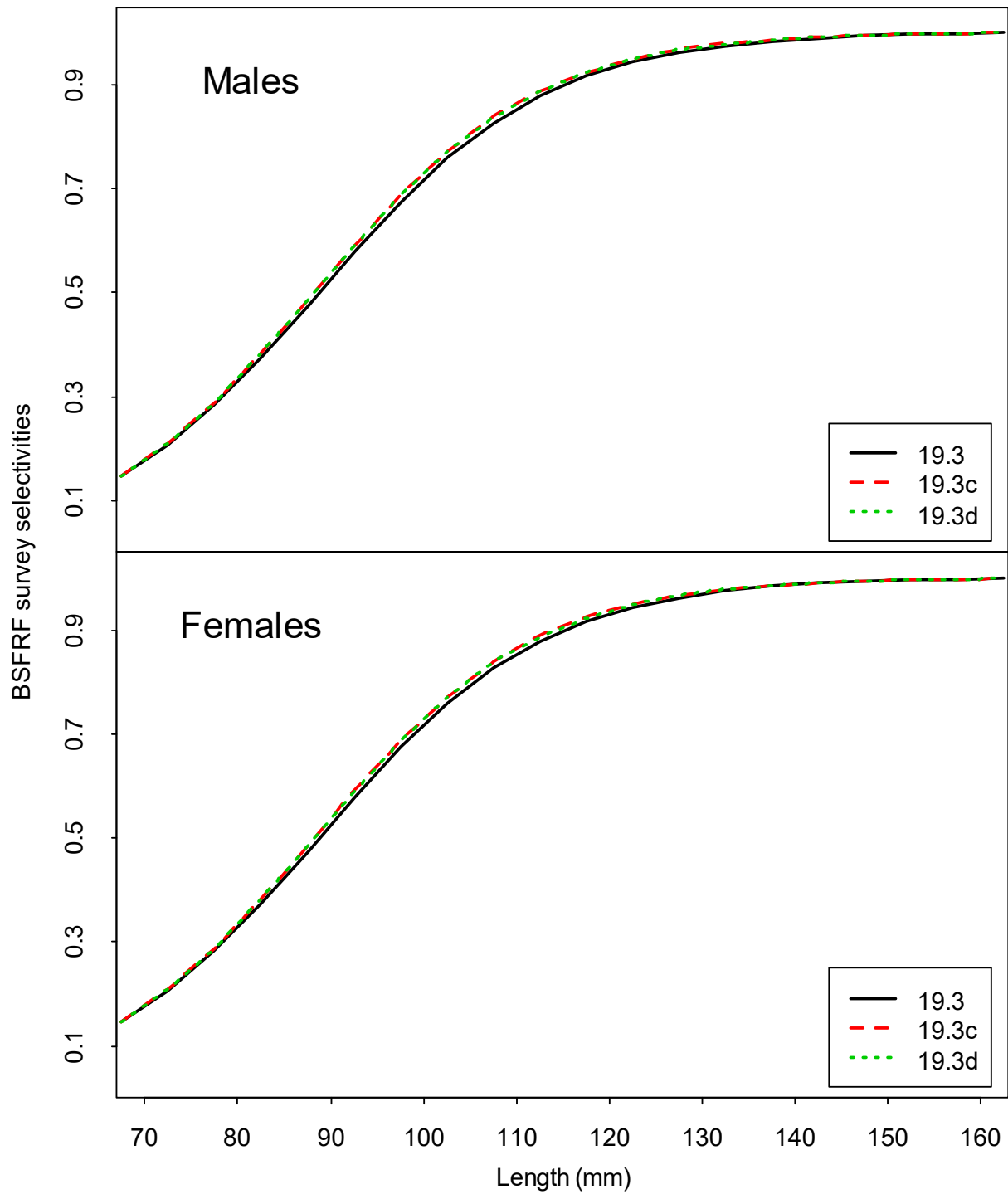


Figure 10d. Comparisons of estimated BSFRF survey selectivities with models 19.3, 19.3c, and 19.3d. The catchability is assumed to be 1.0.

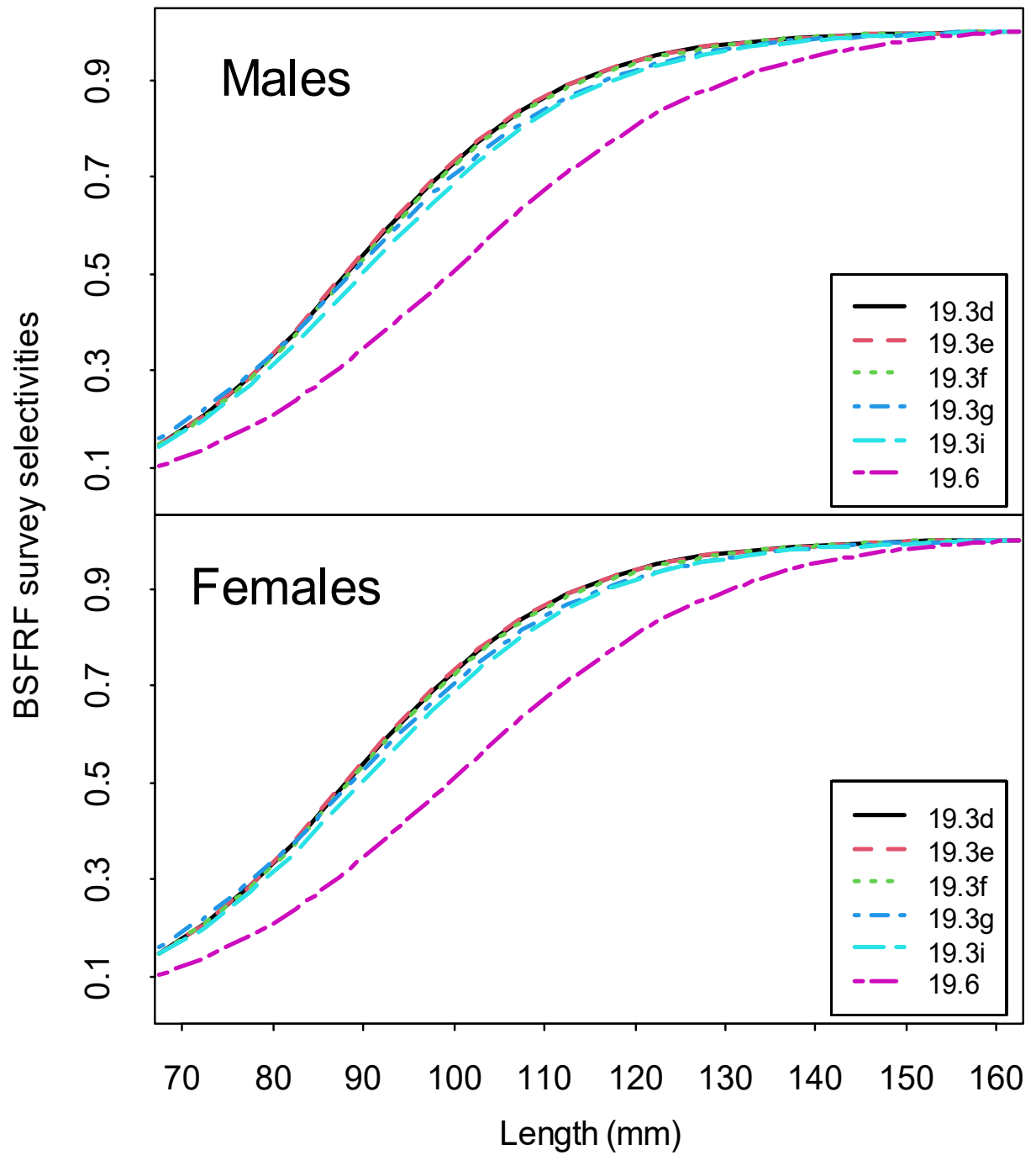


Figure 10d. Comparisons of estimated BSFRF survey selectivities with models 19.3d, 19.3e, 19.3f, 19.3g, 19.3i, and 19.6. The catchability is assumed to be 1.0.

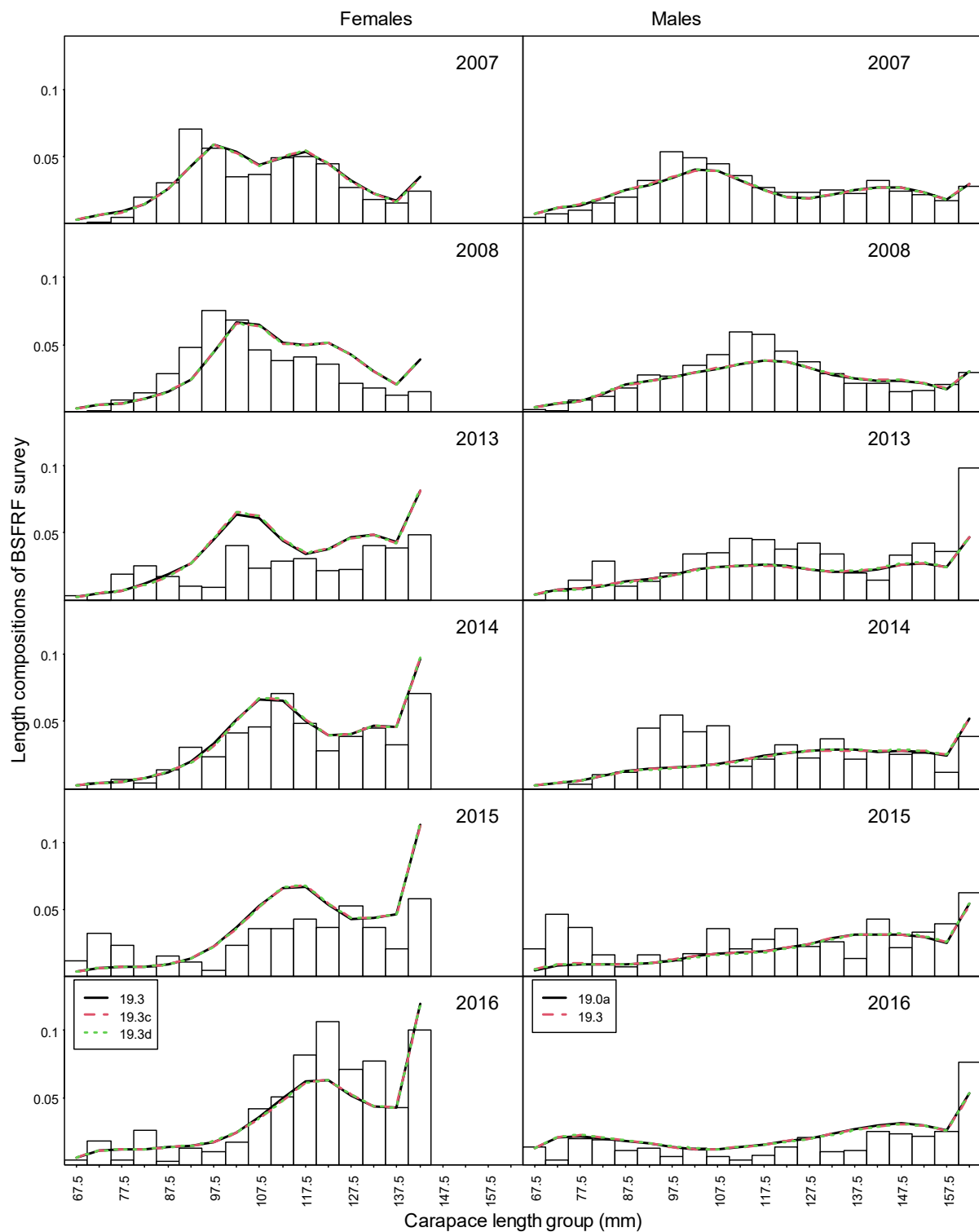


Figure 10f. Comparisons of length compositions by the BSFRF survey and the model estimates during 2007-2008 and 2013-2016 with models 19.3, 19.3c, and 19.3d.

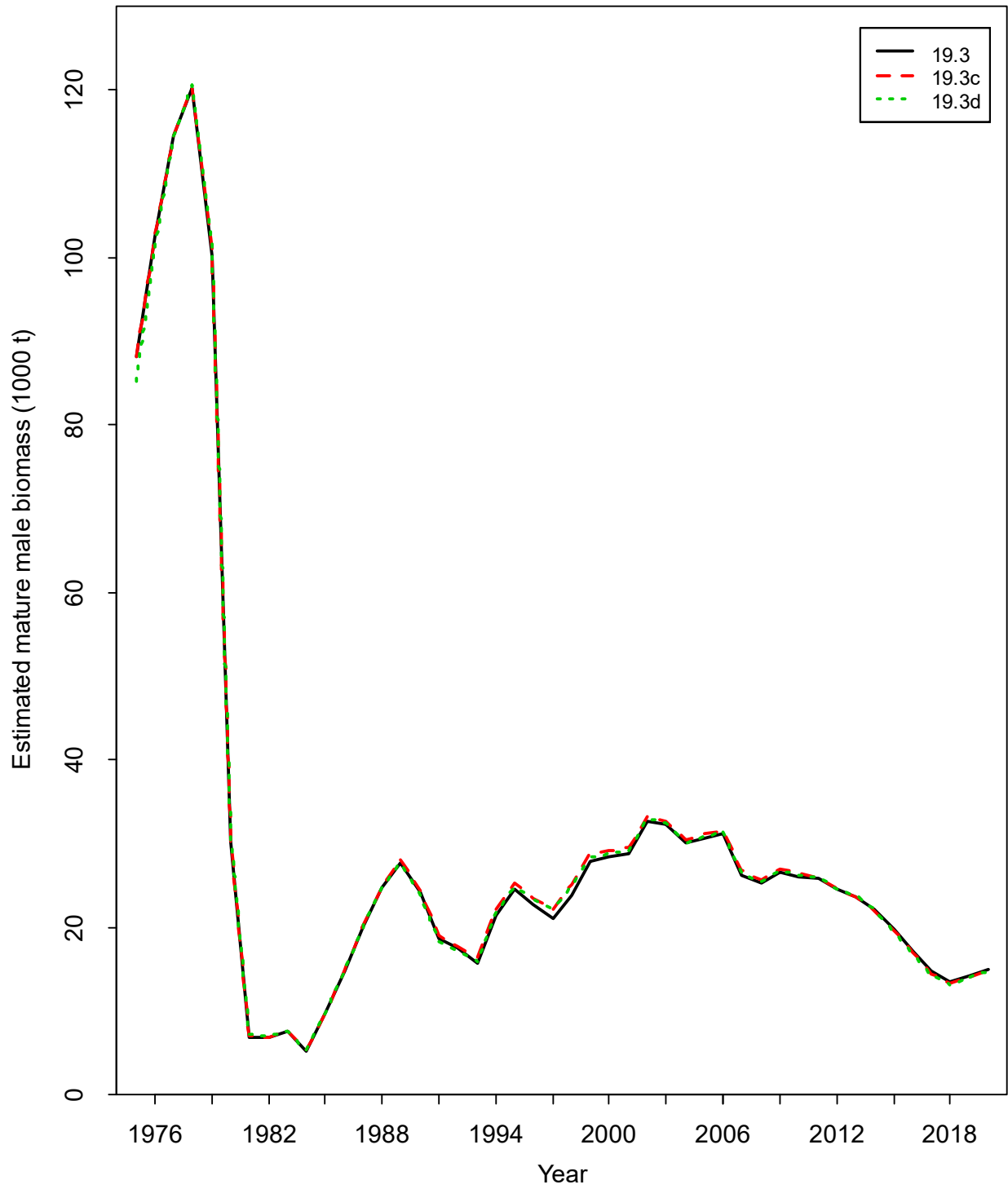


Figure 11a. Estimated absolute mature male biomasses during 1975-2020 for models 19.3, 19.3c, and 19.3d.

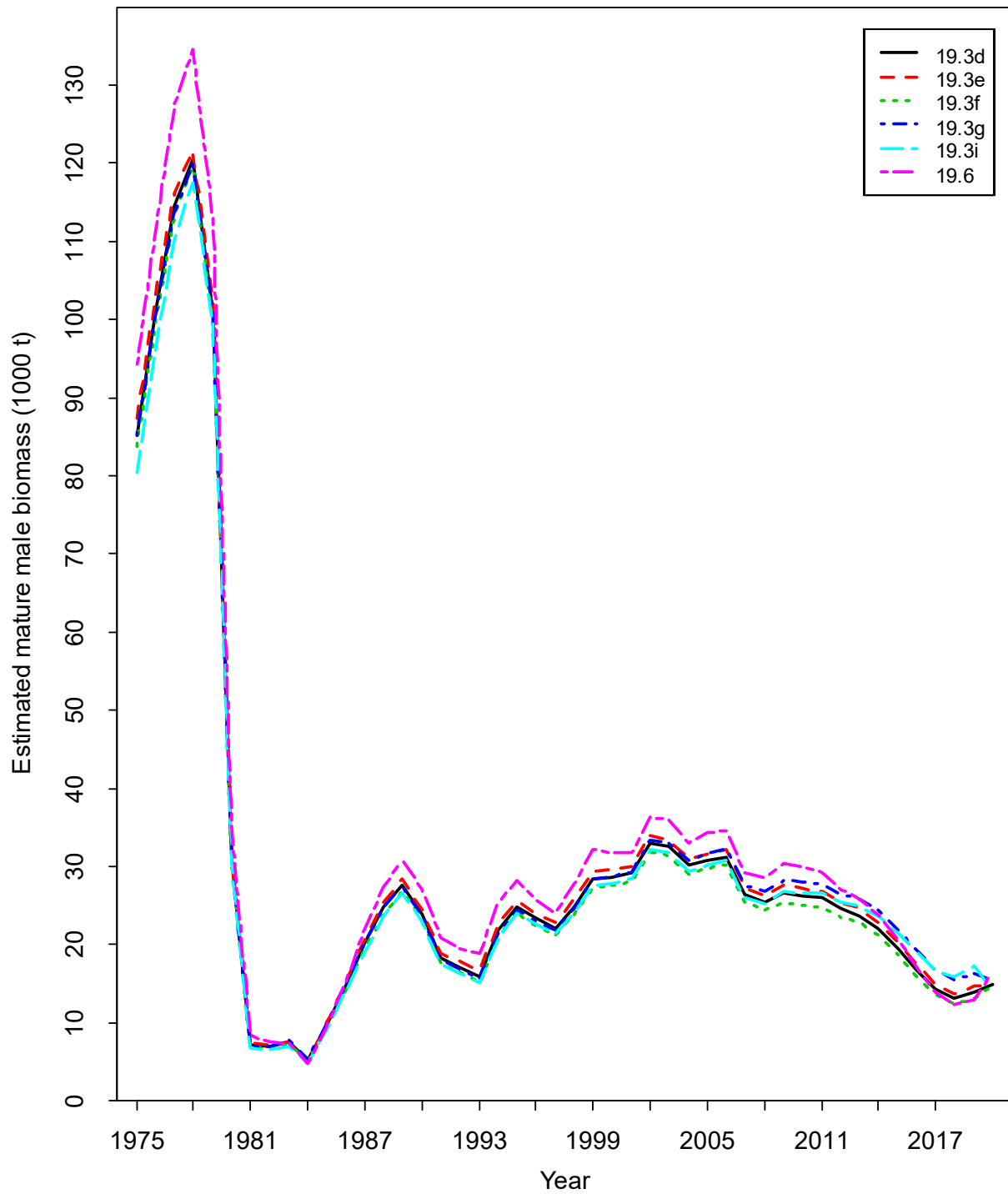


Figure 11b. Estimated absolute mature male biomasses during 1975-2020 for models 19.3d, 19.3e, 19.3f, 19.3g, 19.3i, and 19.6.

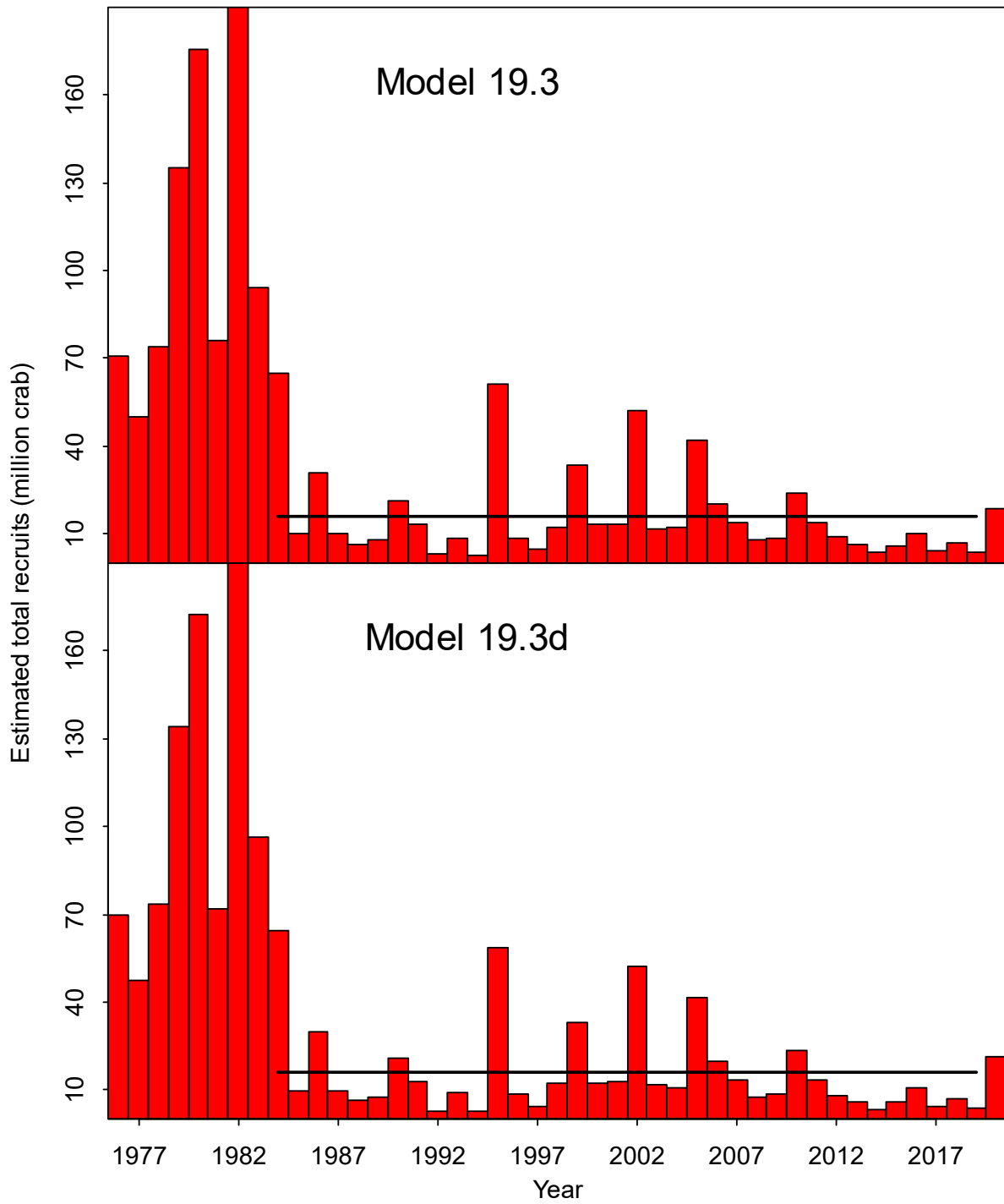


Figure 12a. Estimated recruitment time series during 1976-2020 with models 19.3 and 19.3d. Mean male recruits during 1984-2019 was used to estimate  $B_{35\%}$ .



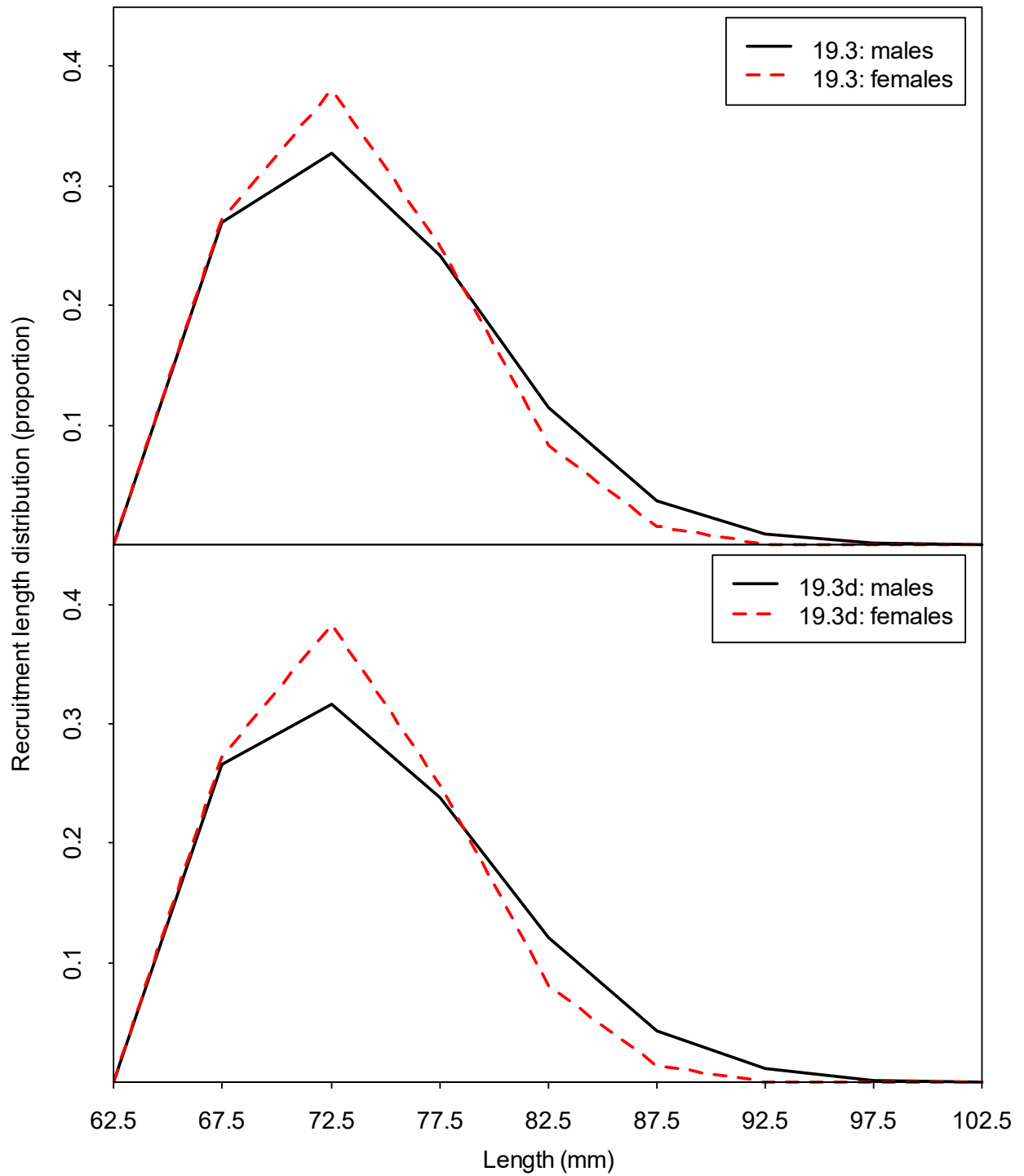


Figure 12b. Estimated recruitment length distributions with models 19.3 and 19.3d.

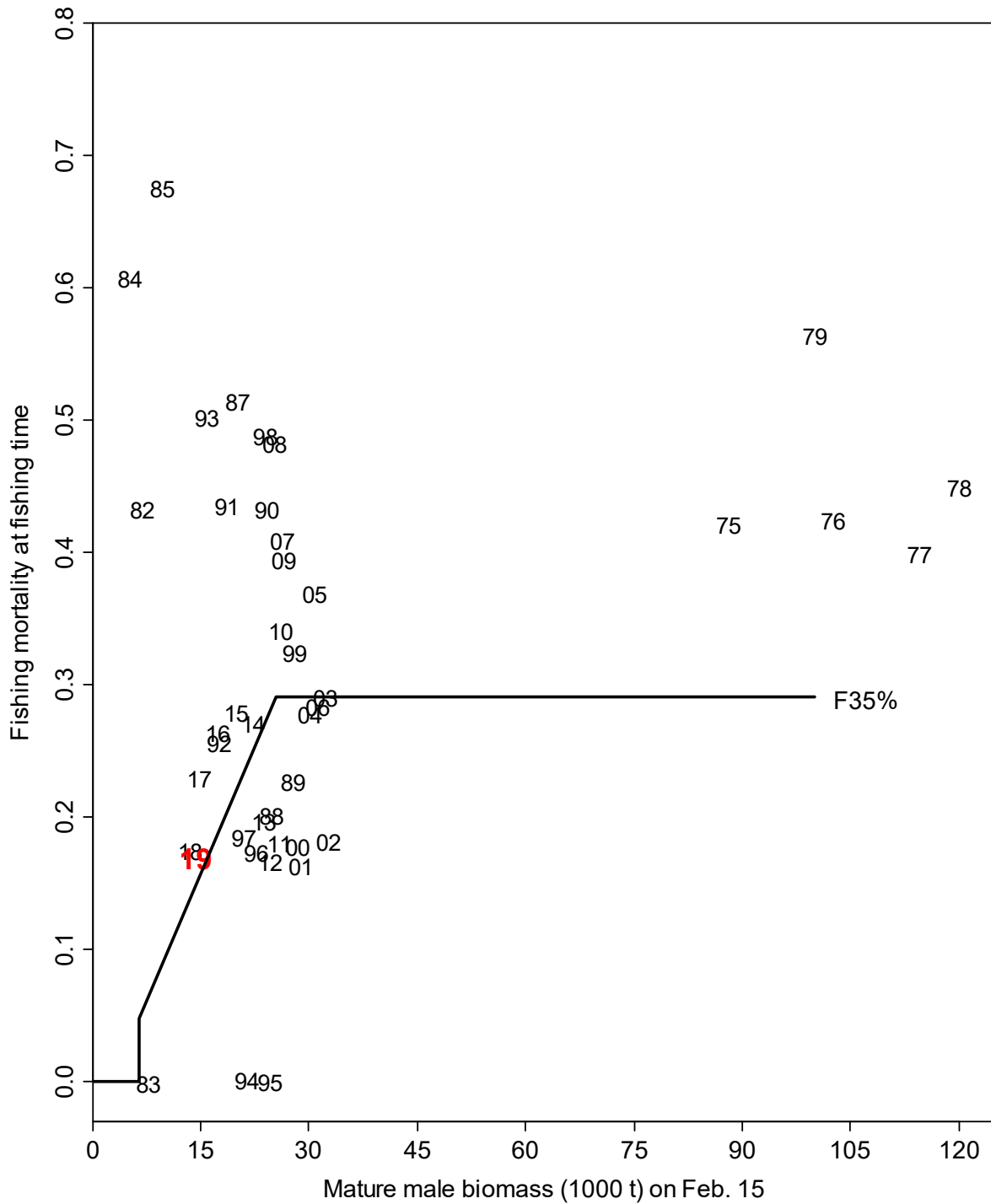


Figure 13a. Relationships between full fishing mortalities for the directed pot fishery and mature male biomass on Feb. 15 during 1975-2019 under model 19.3. Average of recruitment from 1984 to 2019 was used to estimate  $B_{35\%}$ .

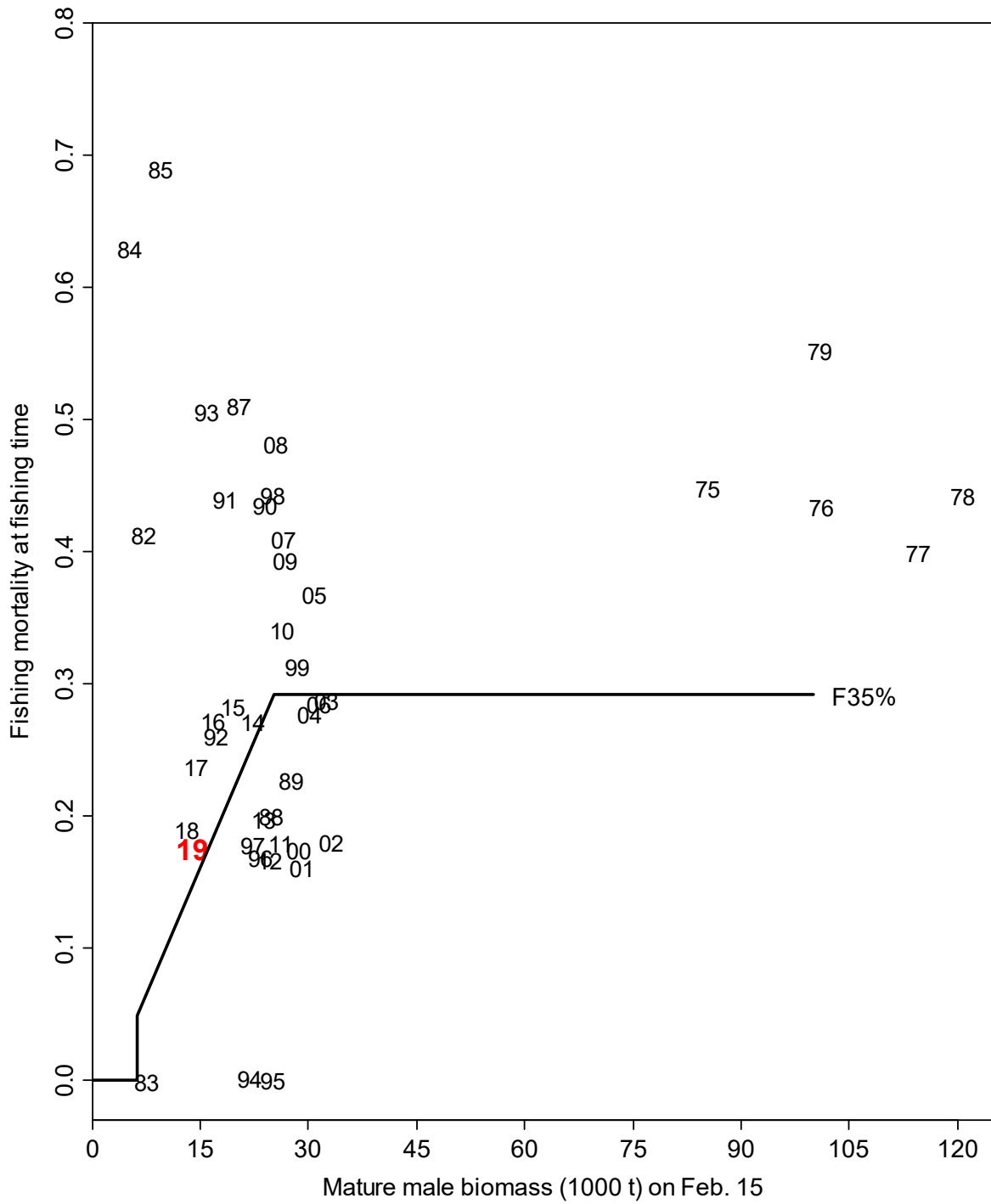


Figure 13b. Relationships between full fishing mortalities for the directed pot fishery and mature male biomass on Feb. 15 during 1975-2019 under model 19.3d. Average of recruitment from 1984 to 2019 was used to estimate  $B_{35\%}$ .

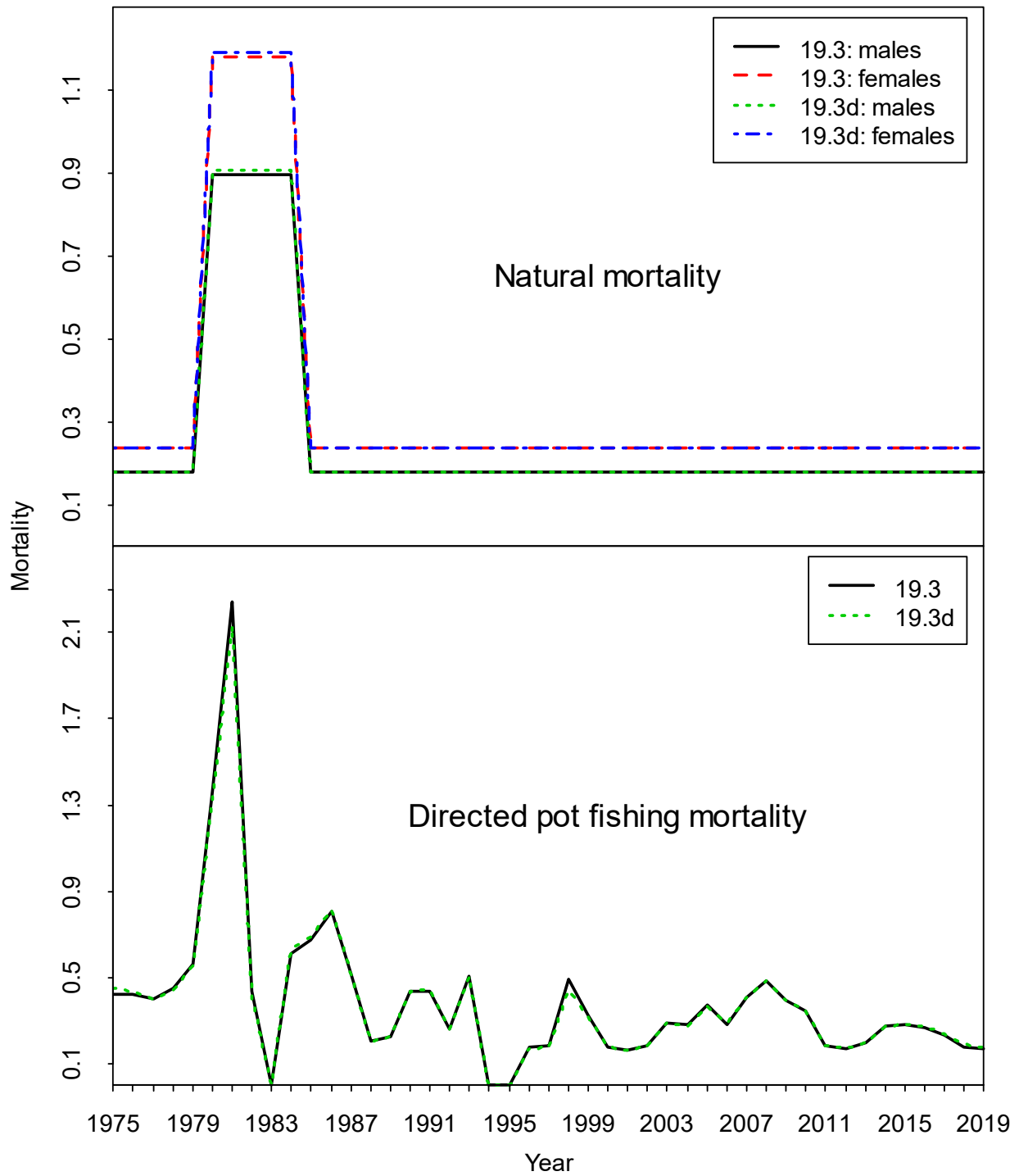


Figure 13c. Comparison of estimated natural mortality and directed pot fishing mortality for models 19.3 and 19.3d.

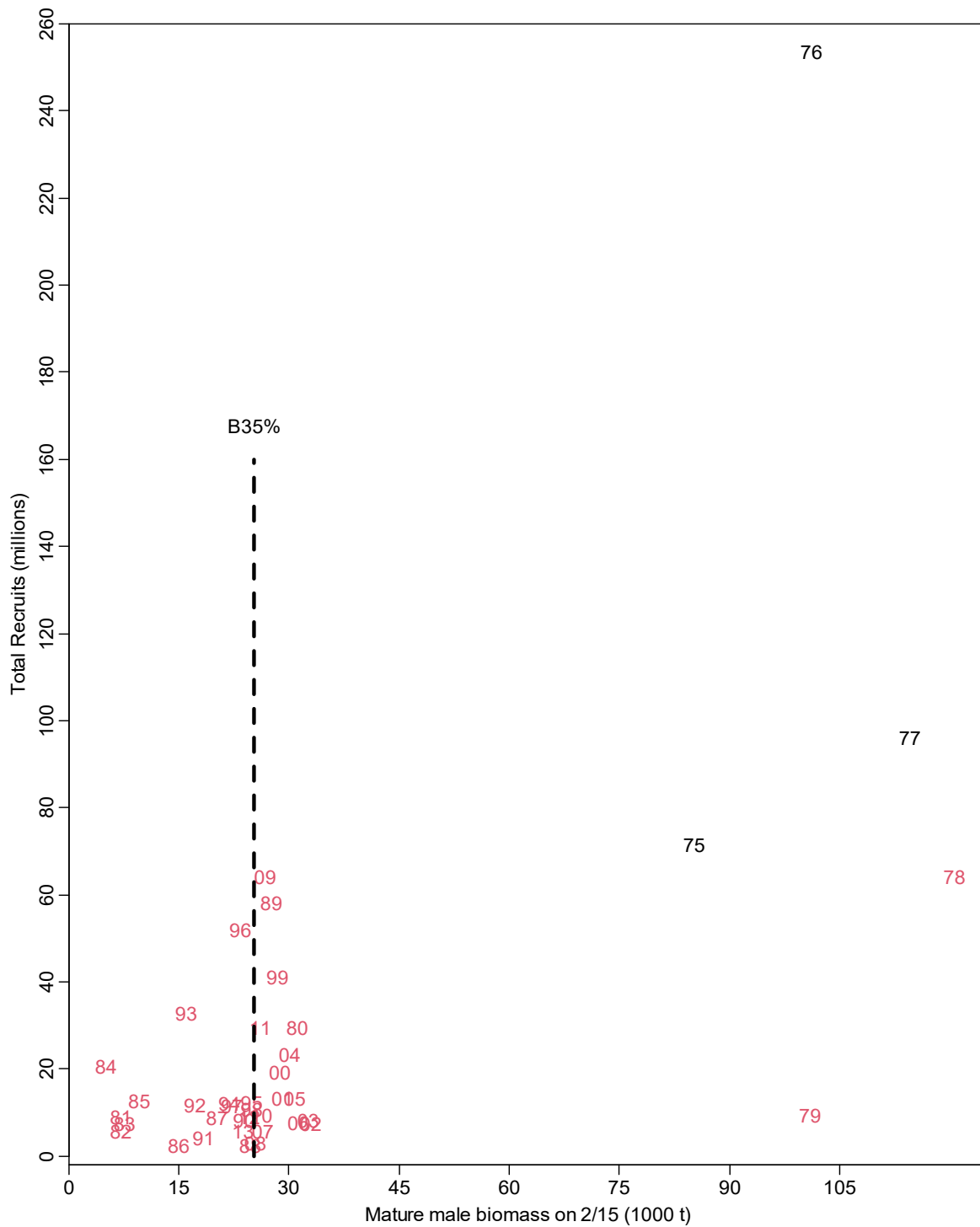


Figure 14a. Relationships between mature male biomass on Feb. 15 and total recruits at age 5 (i.e., 6-year time lag) for Bristol Bay red king crab under model 19.3d. Numerical labels are years of mating, and the vertical dotted line is the estimated  $B_{35\%}$  based on the mean recruitment level during 1984 to 2019.

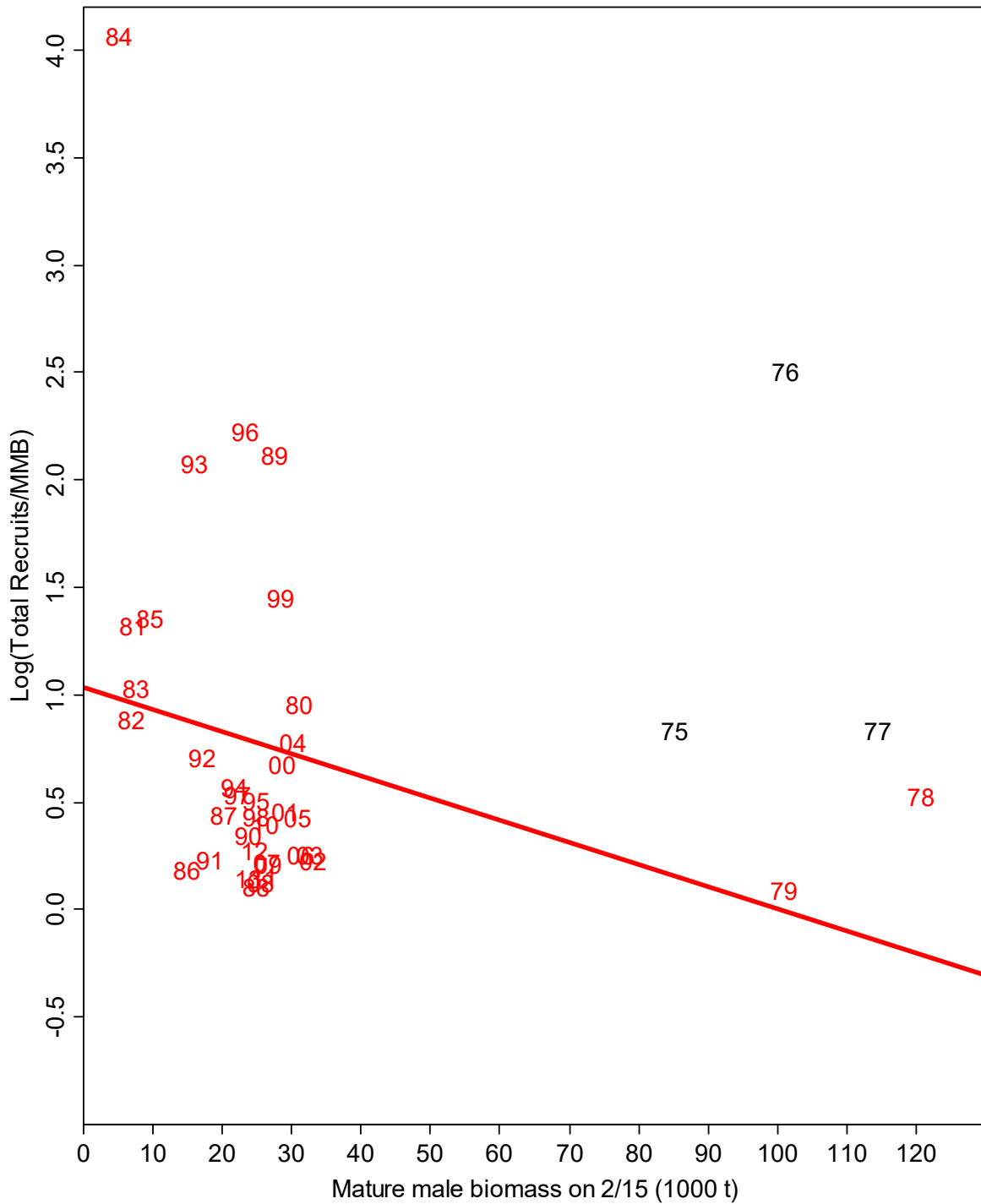


Figure 14b. Relationships between log recruitment per mature male biomass and mature male biomass on Feb. 15 for Bristol Bay red king crab under model 19.3d. Numerical labels are years of mating, and the line is the regression line for data of 1978-2013.

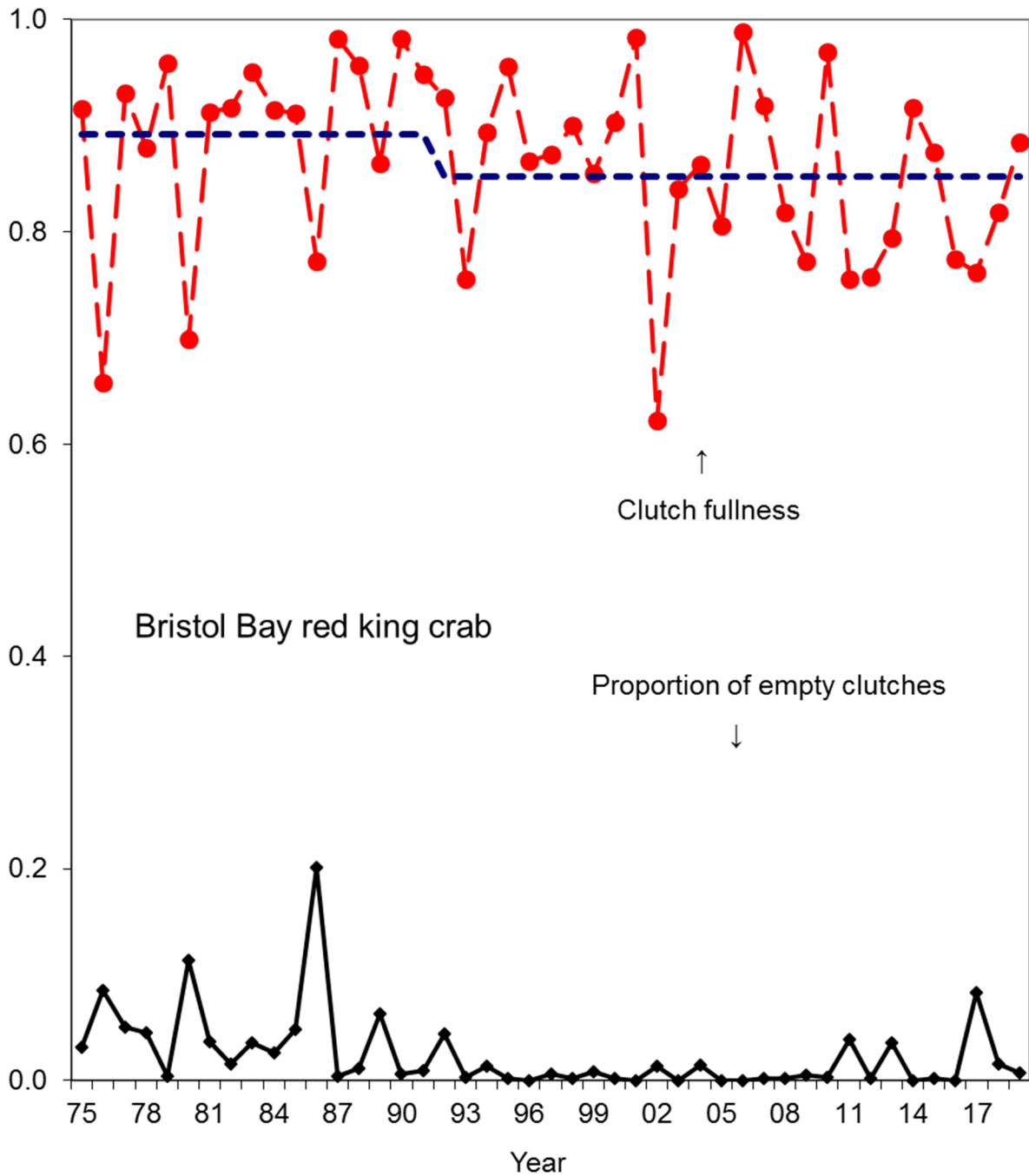


Figure 15. Average clutch fullness and proportion of empty clutches of newshell (shell conditions 1 and 2) mature female crab >89 mm CL from 1975 to 2019 from survey data. Oldshell females were excluded. The blue dashed line is the mean clutch fullness during two periods before 1992 and after 1991.

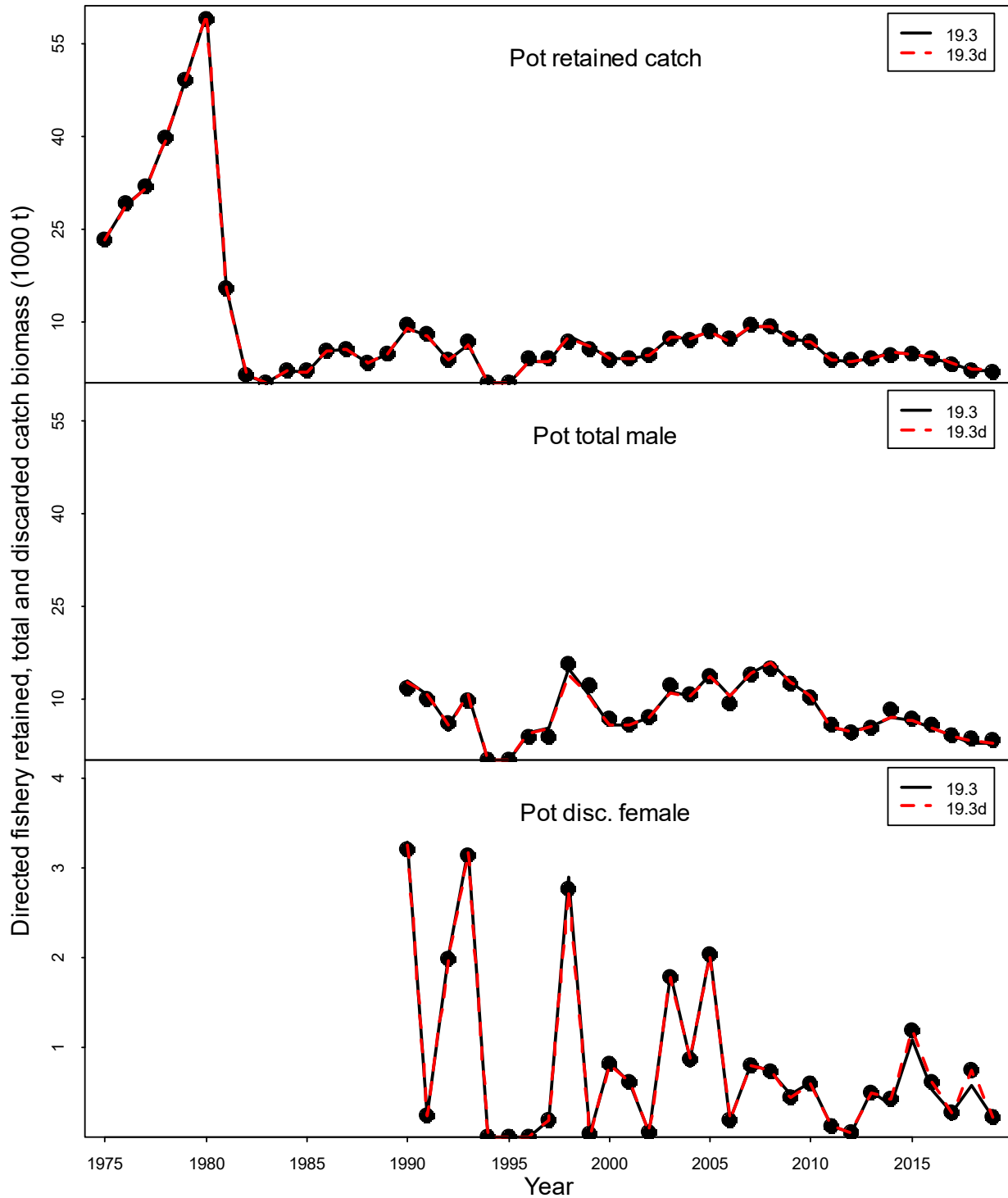


Figure 16a. Observed (dots) and predicted (lines) RKC catch and bycatch biomass under models 19.3 and 19.3d. The observed discarded catch biomass is from model 19.3d.



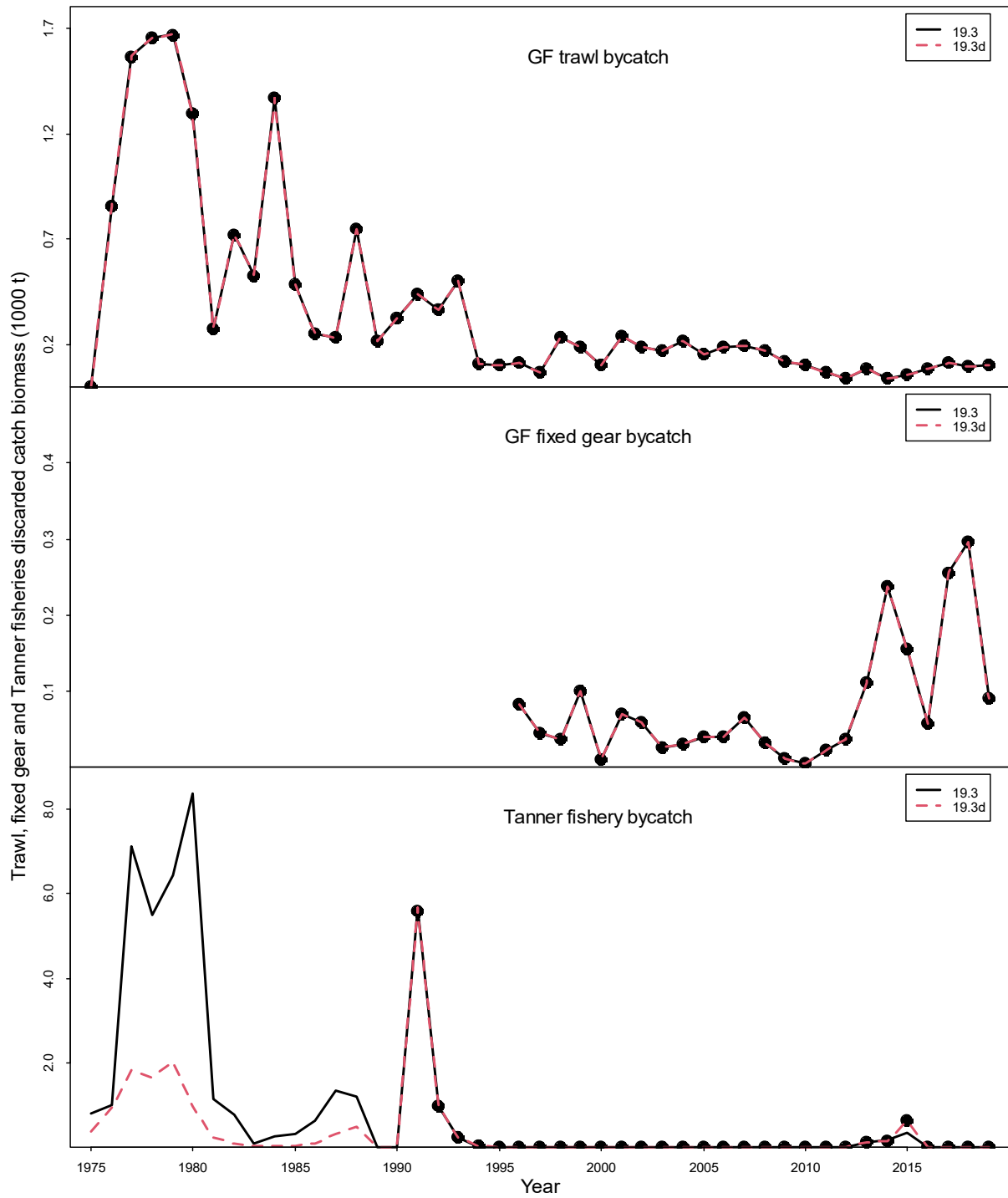


Figure 16b. Observed (dots) and predicted (lines) RKC bycatch biomass from groundfish fisheries and the Tanner crab fishery under models 19.3 and 19.3d. The observed discarded catch biomass from the Tanner crab fishery is from model 19.3d. Trawl bycatch biomass was 0 before 1976.

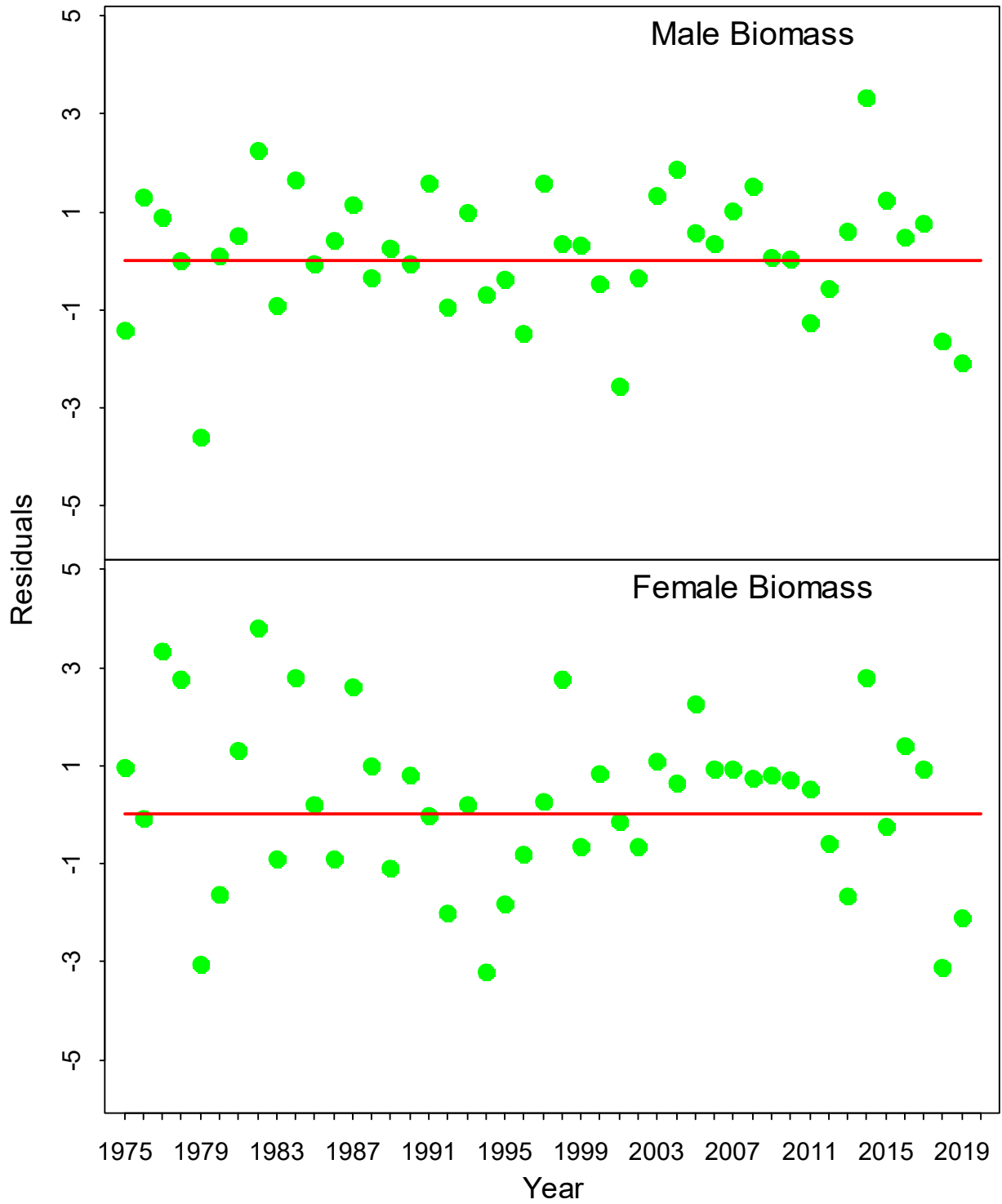


Figure 17a. Standardized residuals of NMFS survey biomass under model 19.3.

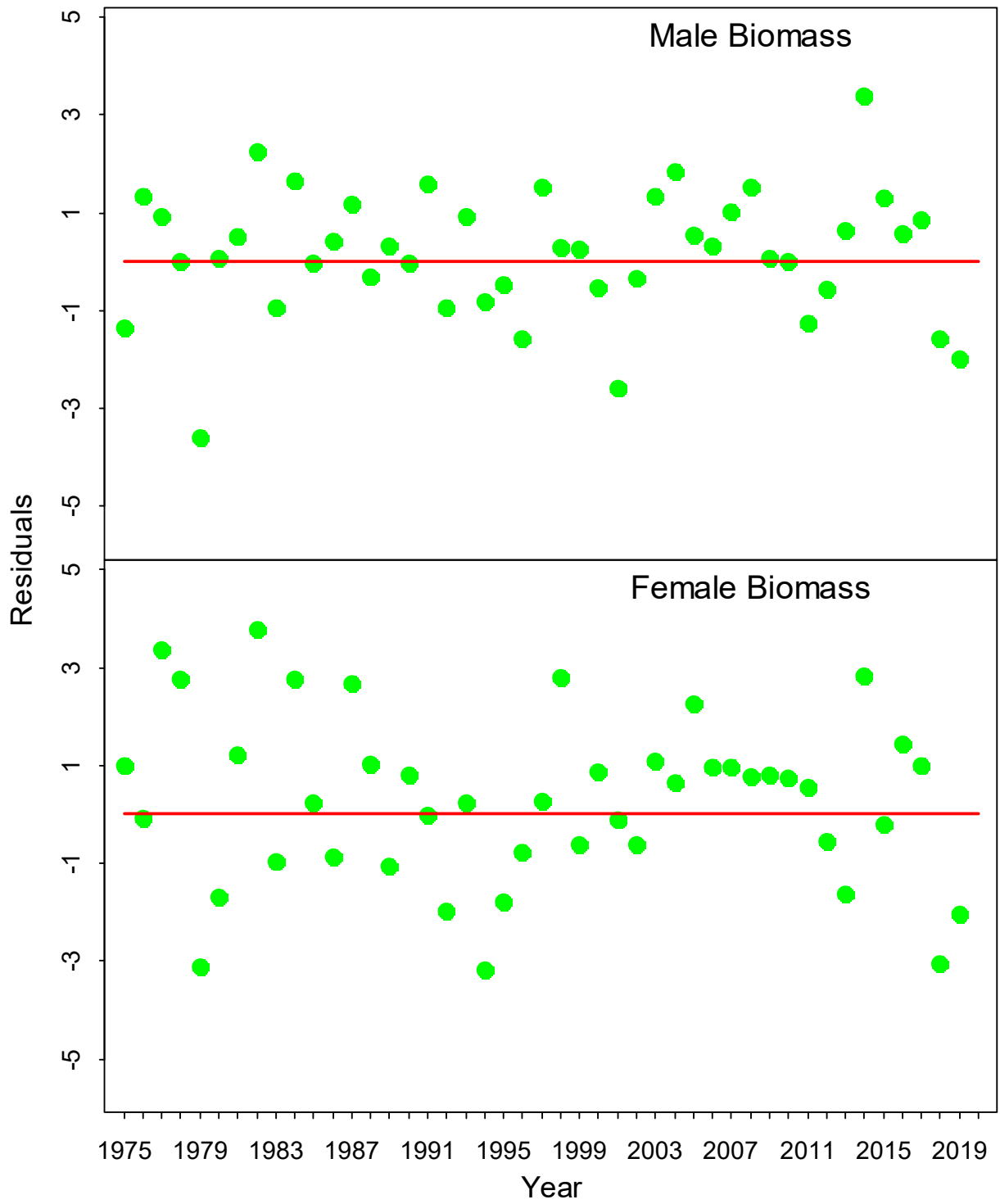


Figure 17b. Standardized residuals of NMFS survey biomass under model 19.3d.

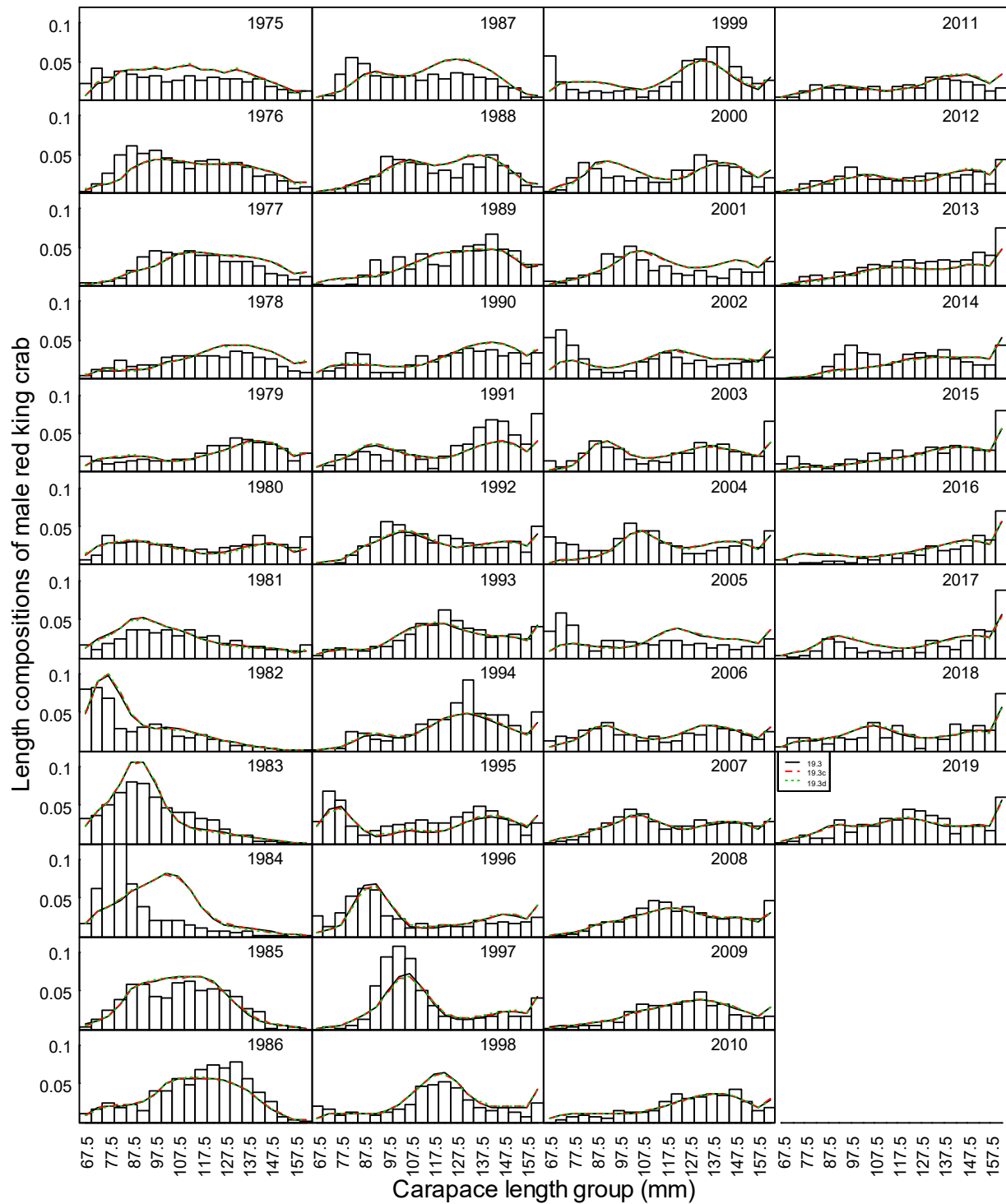


Figure 18. Comparison of area-swept and model estimated NMFS survey length frequencies of Bristol Bay male red king crab by year under models 19.3, 19.3c, and 19.3d.

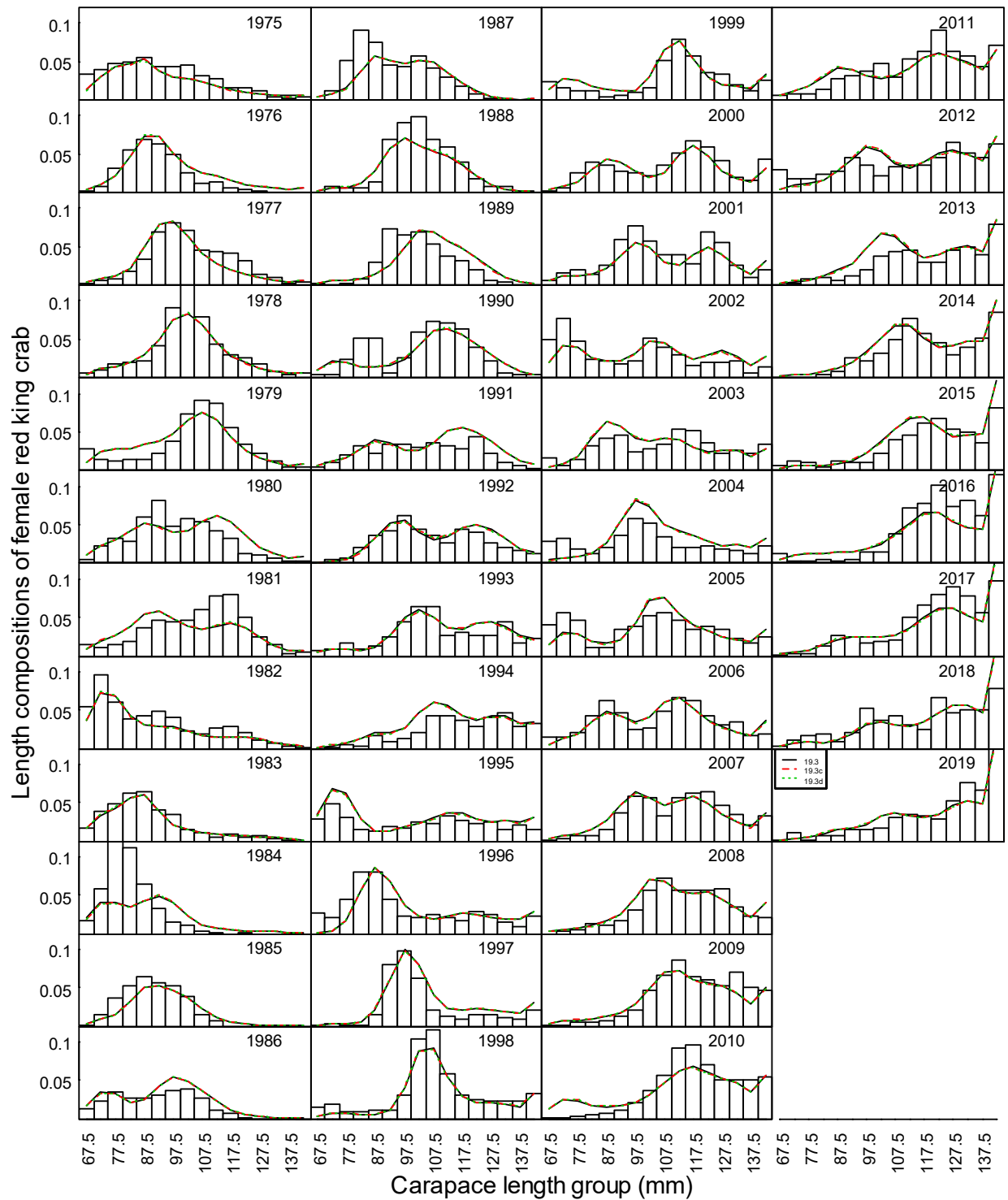


Figure 19. Comparison of area-swept and model estimated NMFS survey length frequencies of Bristol Bay female red king crab by year under models 19.3, 19.3c, and 19.3d.

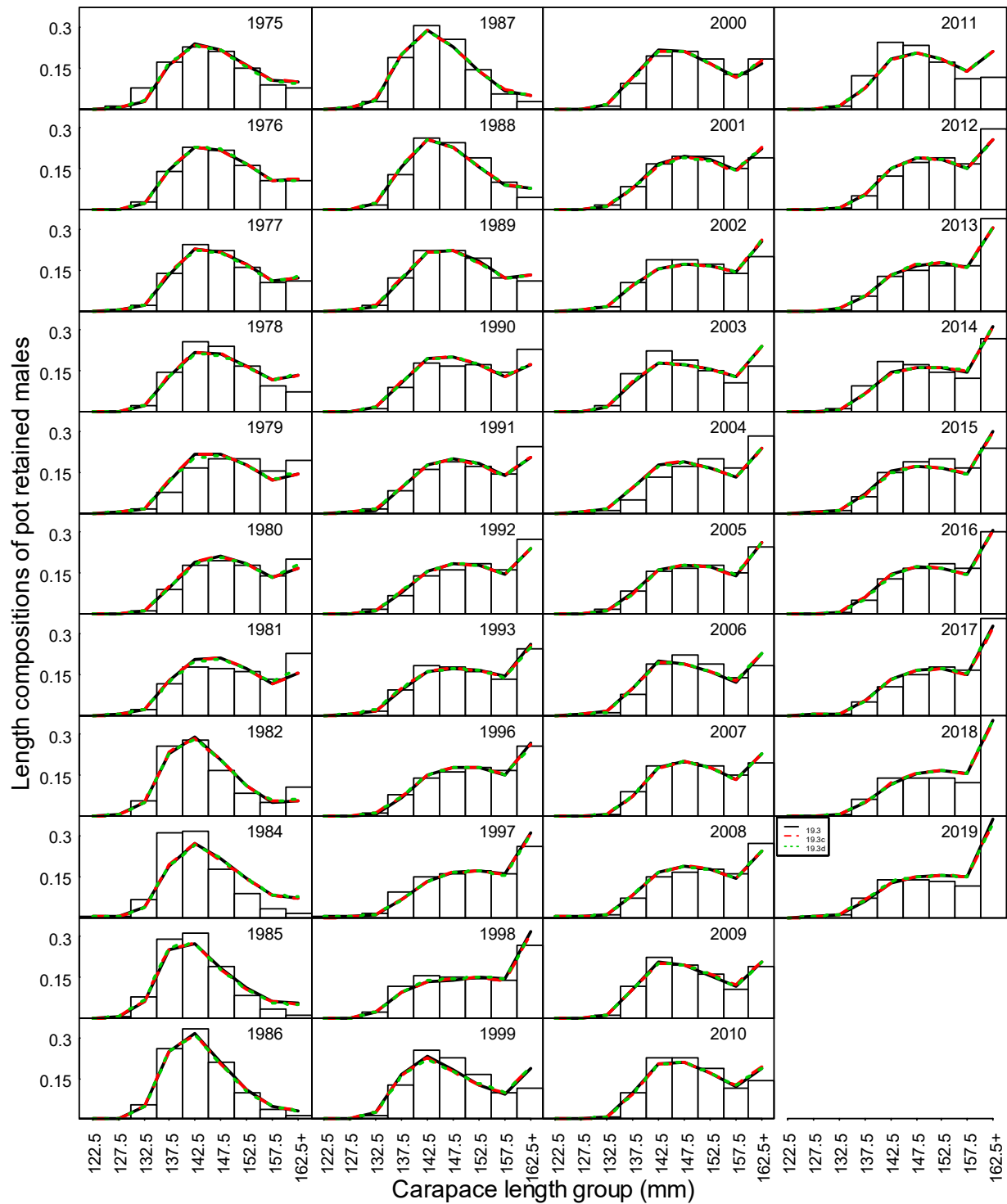


Figure 20. Comparison of observed and model estimated retained length frequencies of Bristol Bay male red king crab by year in the directed pot fishery under models 19.3, 19.3c, and 19.3d.

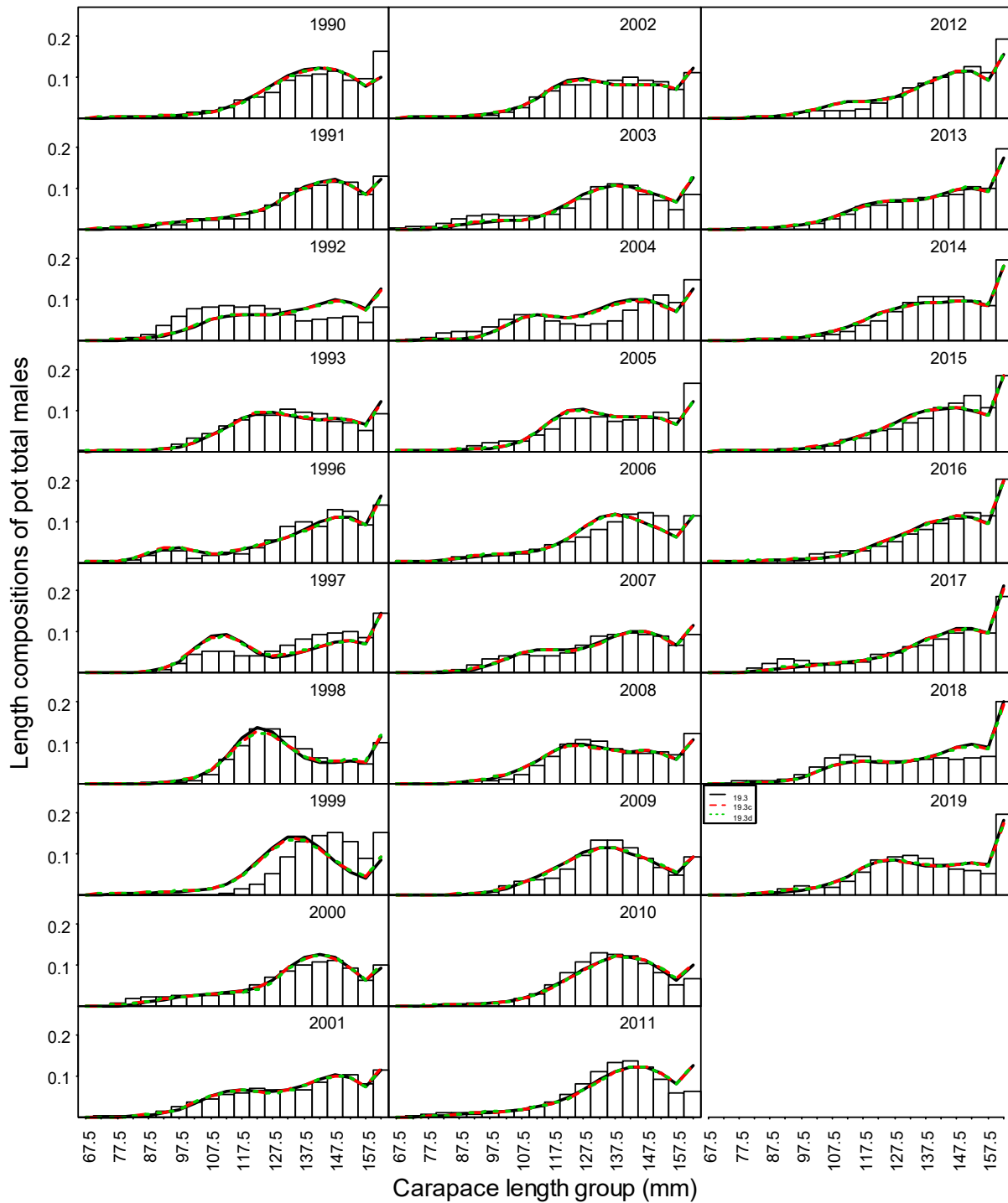


Figure 21. Comparison of observer and model estimated total observer length frequencies of Bristol Bay male red king crab by year in the directed pot fishery under models 19.3, 19.3c, and 19.3d. The observed length composition data are from model 19.3d.

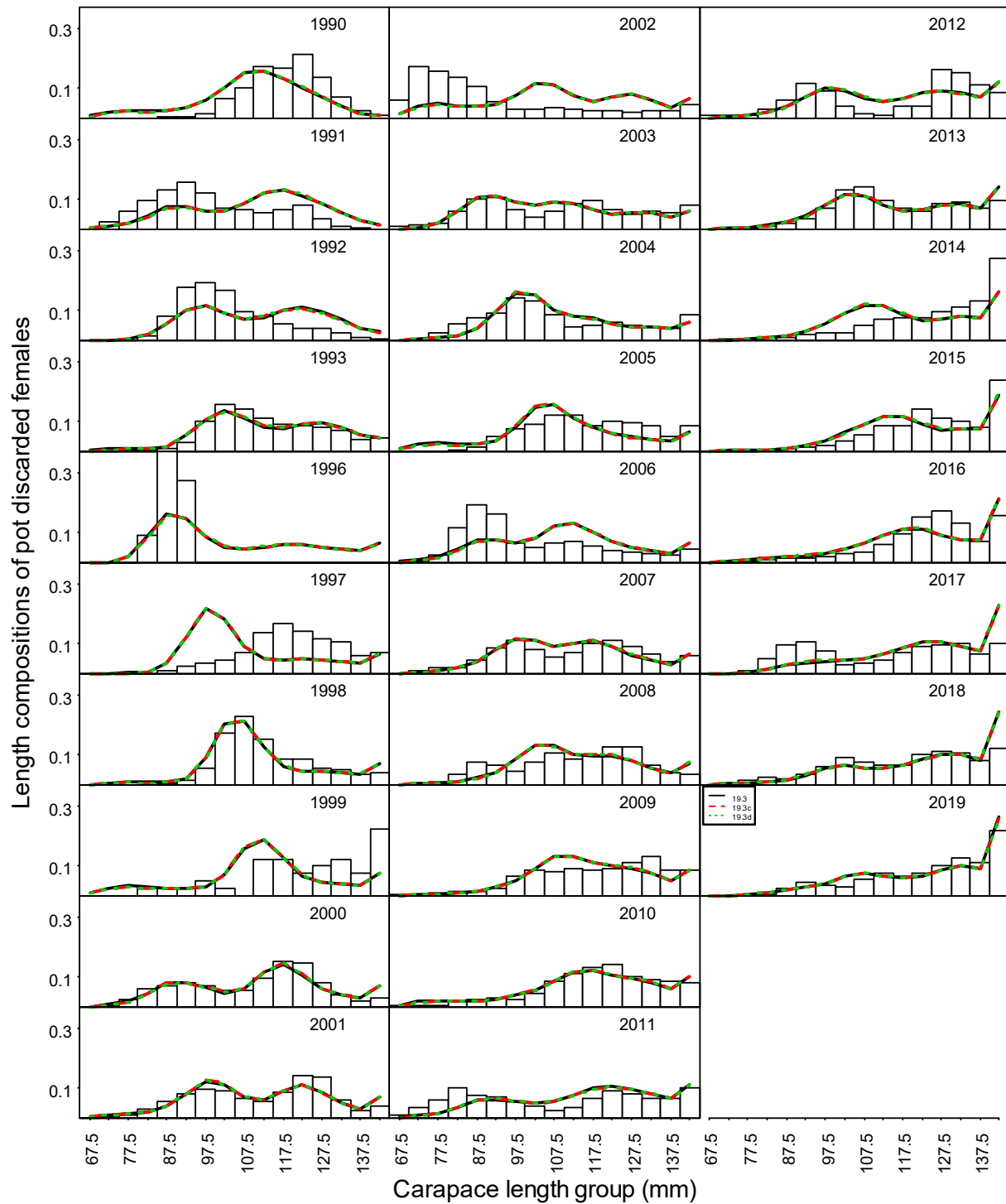


Figure 22. Comparison of observer and model estimated discarded length frequencies of Bristol Bay female red king crab by year in the directed pot fishery under models 19.3, 19.3c, and 19.3d. The observed length composition data are from model 19.3d.



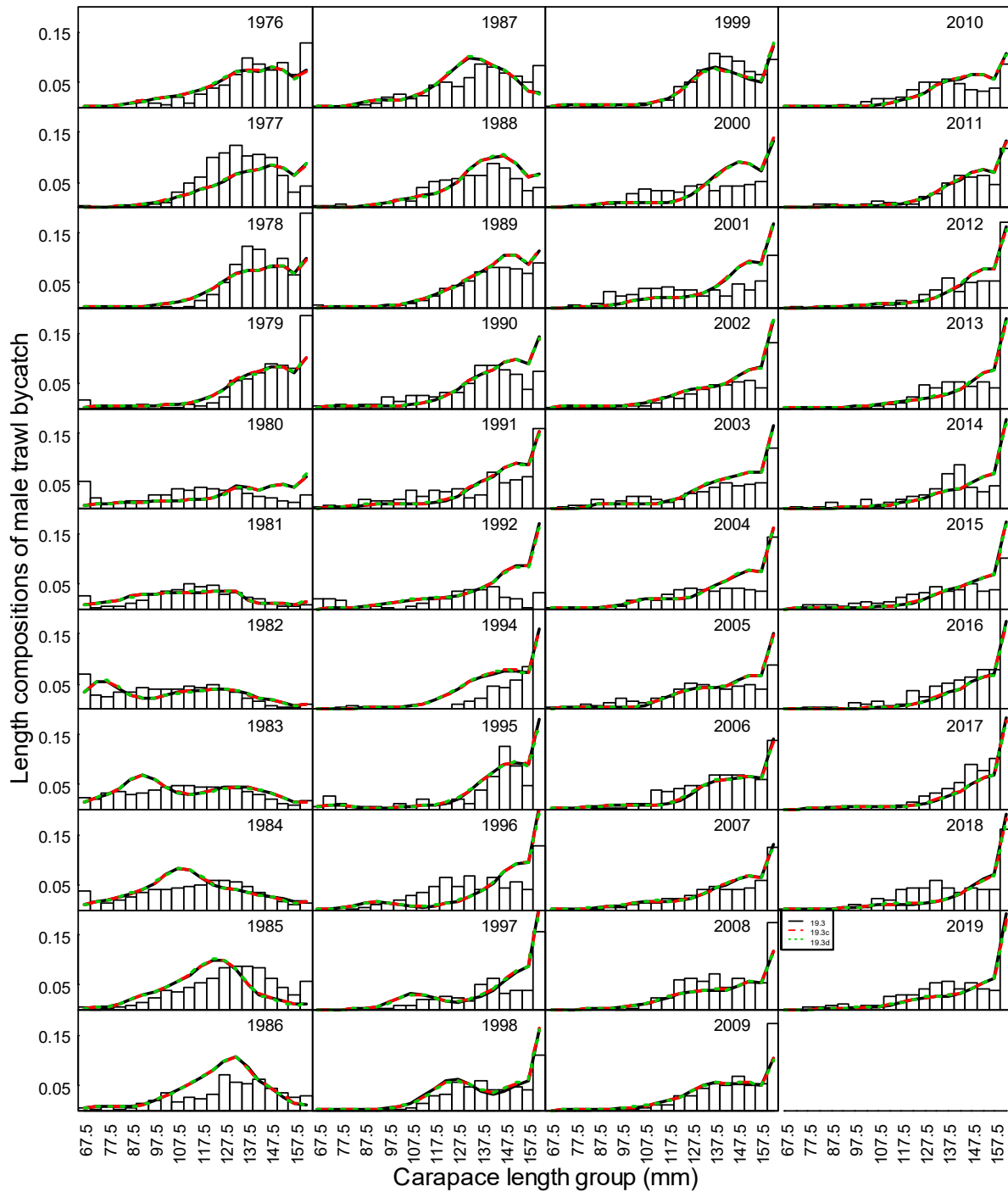


Figure 23a. Comparison of observer and model estimated discarded length frequencies of Bristol Bay male red king crab by year in the groundfish trawl fisheries under models 19.3, 19.3c, and 19.3d.

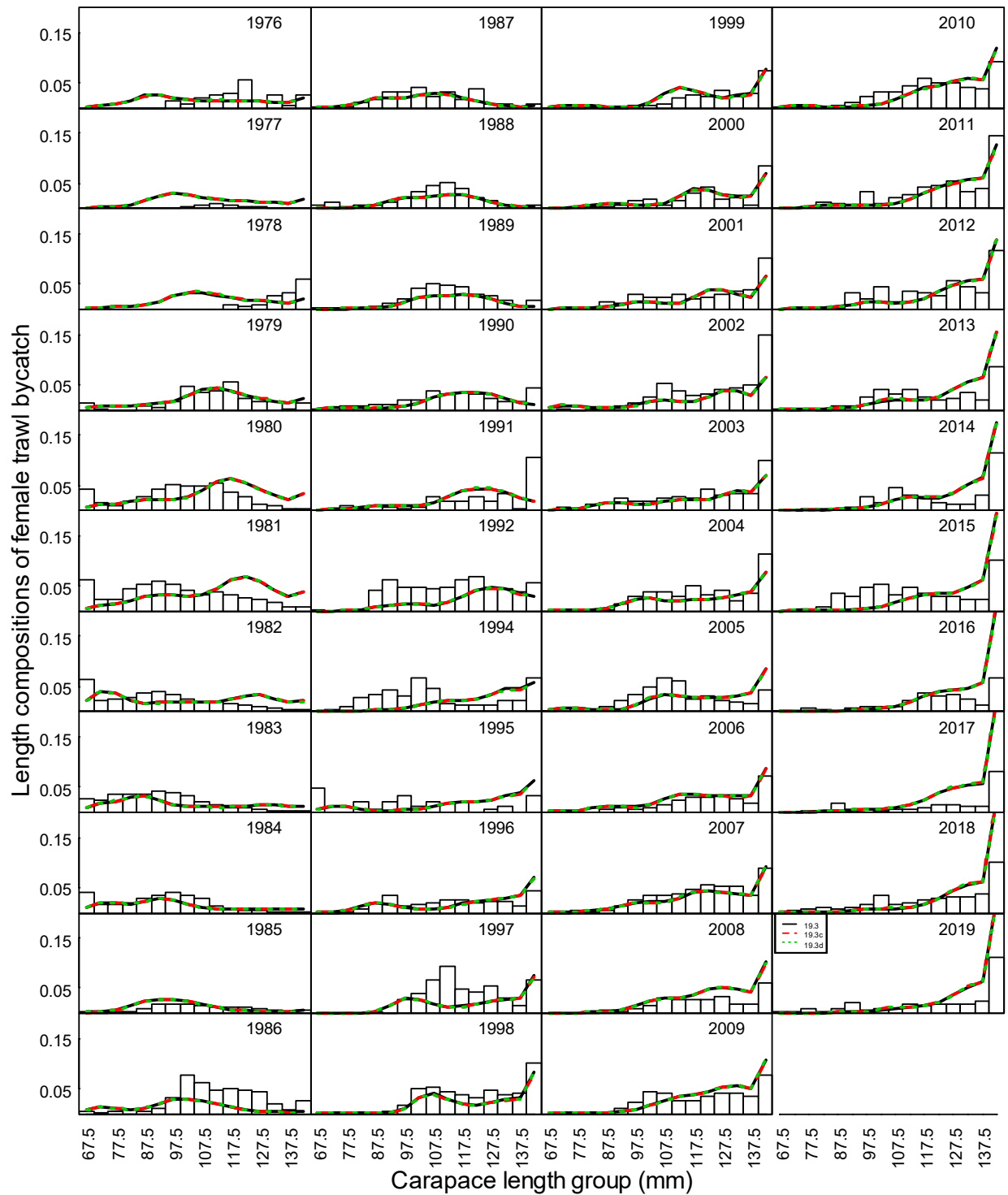


Figure 23b. Comparison of observer and model estimated discarded length frequencies of Bristol Bay female red king crab by year in the groundfish trawl fisheries under models 19.3, 19.3c, and 19.3d.

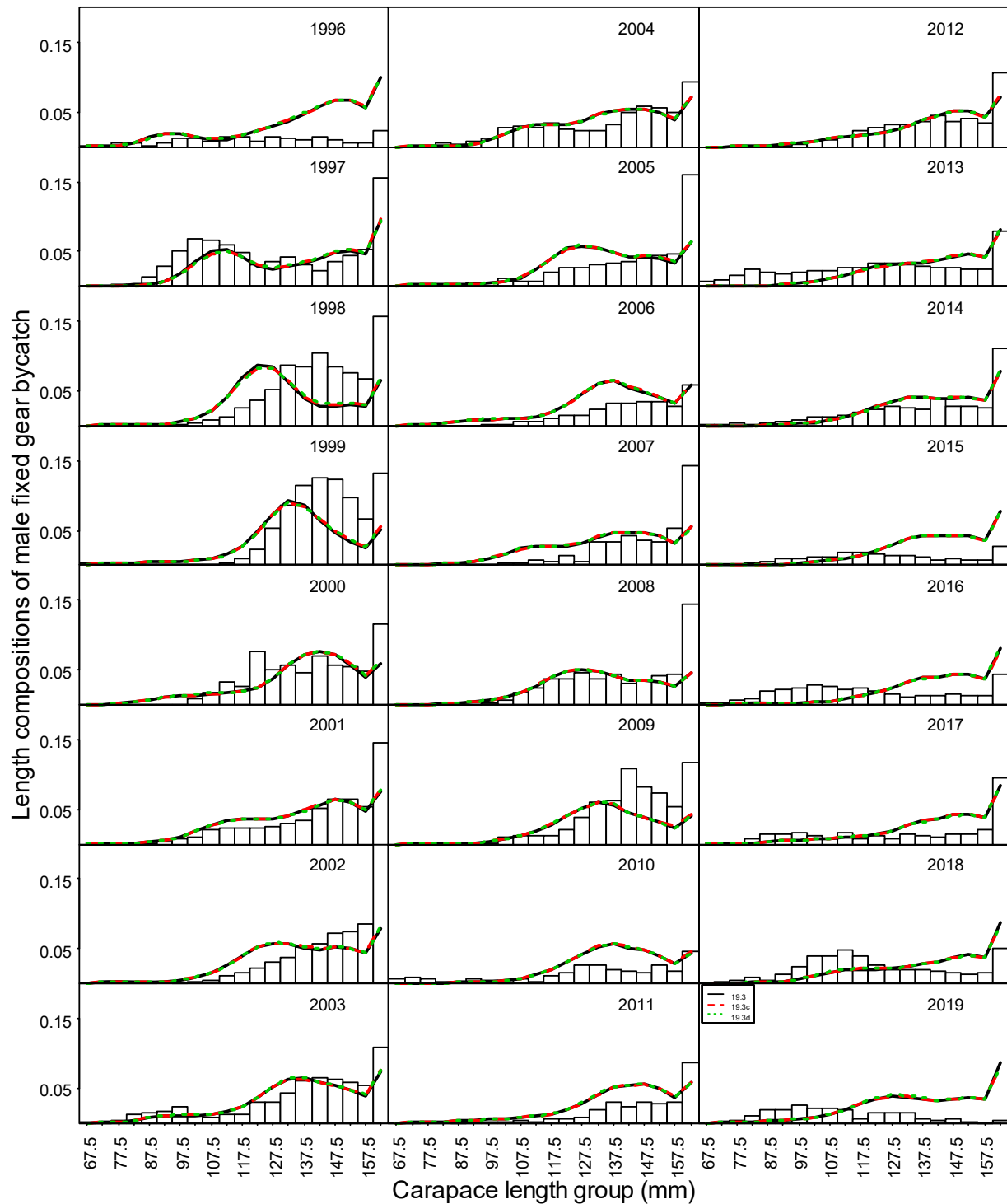


Figure 24a. Comparison of observer and model estimated discarded length frequencies of Bristol Bay male red king crab by year in the groundfish fixed gear fisheries under models 19.3, 19.3c, and 19.3d.

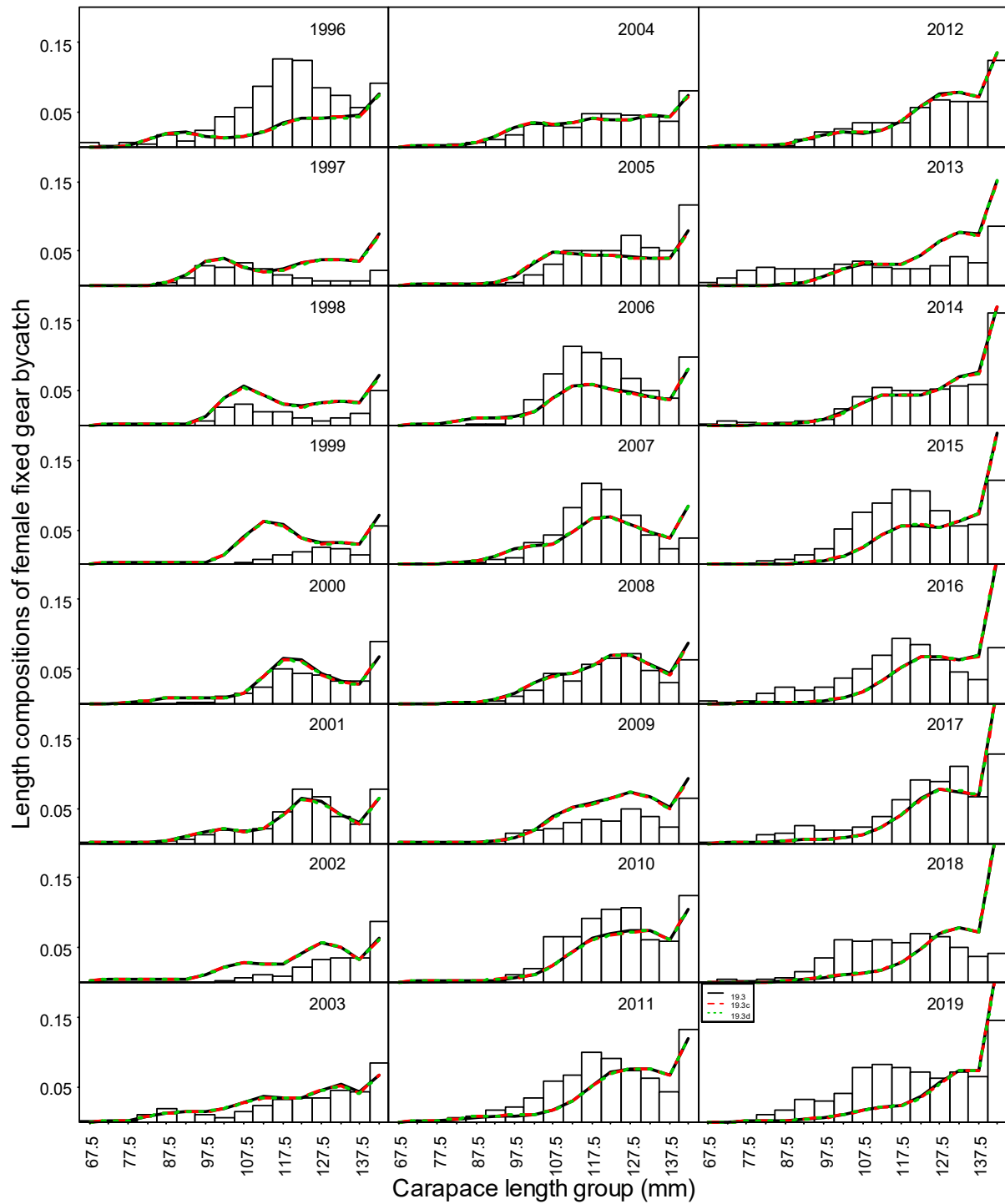


Figure 24b. Comparison of observer and model estimated discarded length frequencies of Bristol Bay female red king crab by year in the groundfish fixed gear fisheries under models 19.3, 19.3c, and 19.3d.

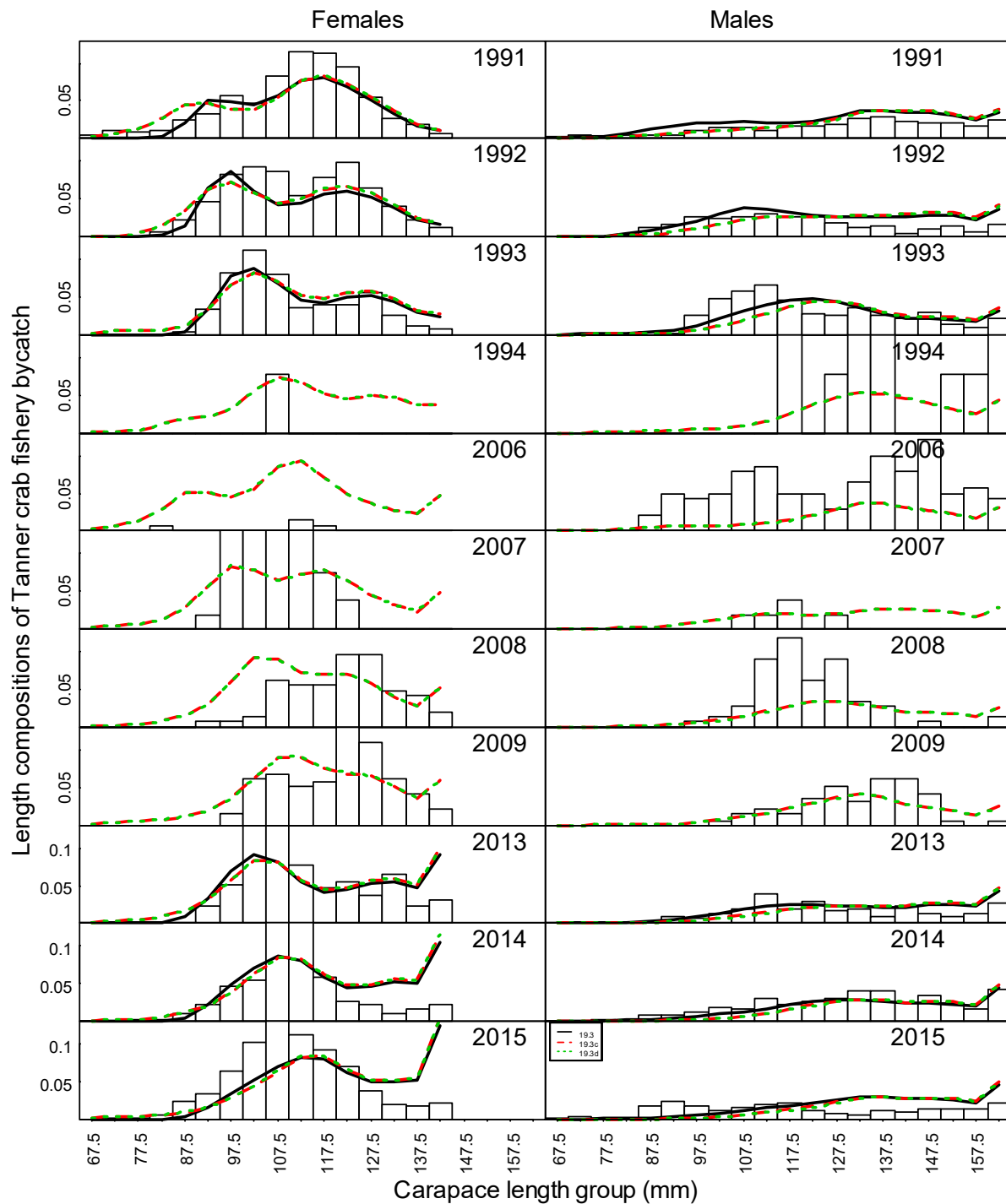


Figure 24c. Comparison of observer and model estimated discarded length frequencies of Bristol Bay red king crab by year in the Tanner crab fishery under models 19.3, 19.3c, and 19.3d. The observed length composition data are from model 19.3d. Length composition data during 1994-2009 are not used in model 19.3.

### Model 19.3, Survey Males

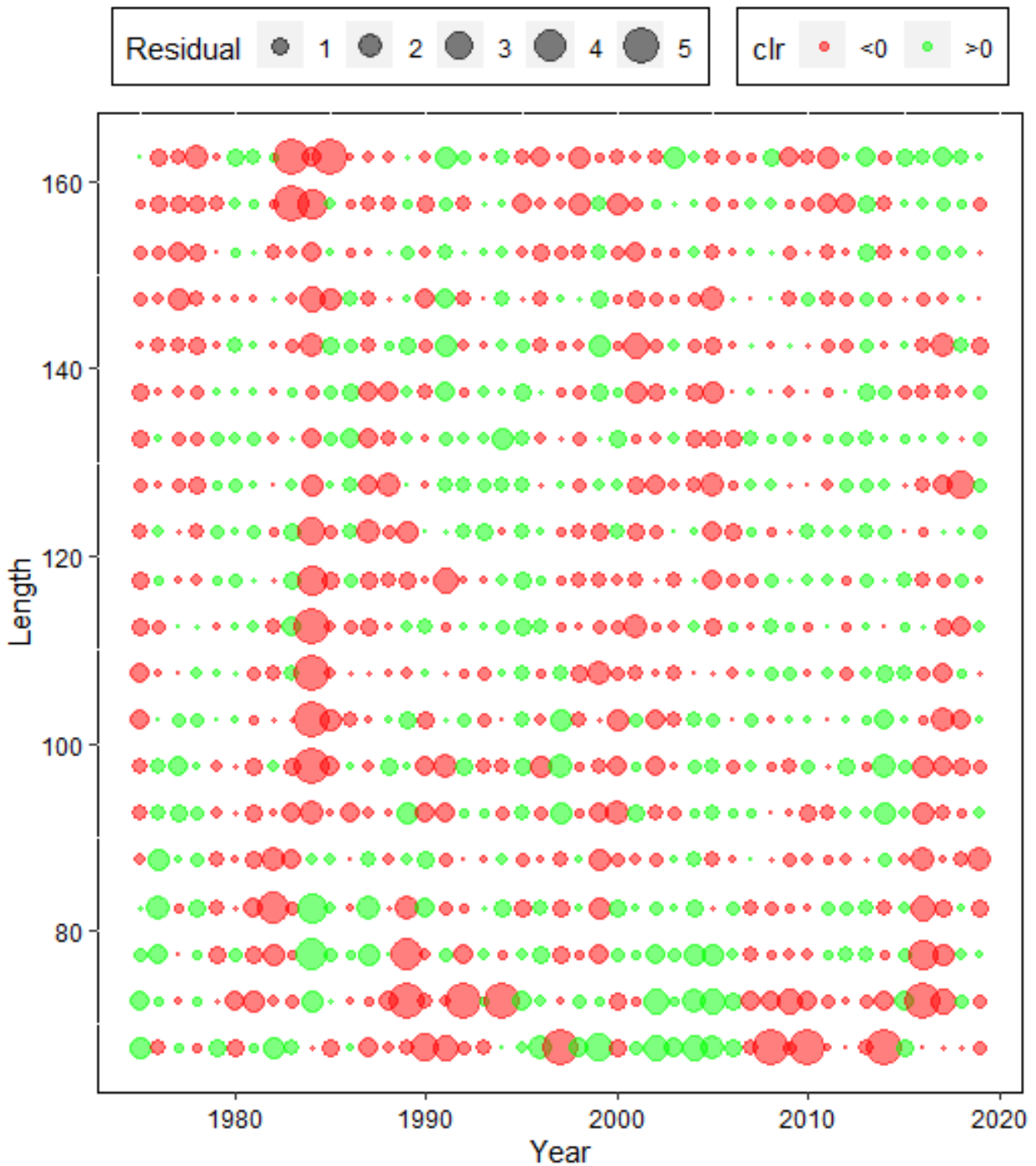


Figure 25a. Residuals of proportions of NMFS survey male red king crab by year and carapace length (mm) under model 19.3. Green circles are positive residuals, and red circles are negative residuals.

Model 19.3d, Survey Males

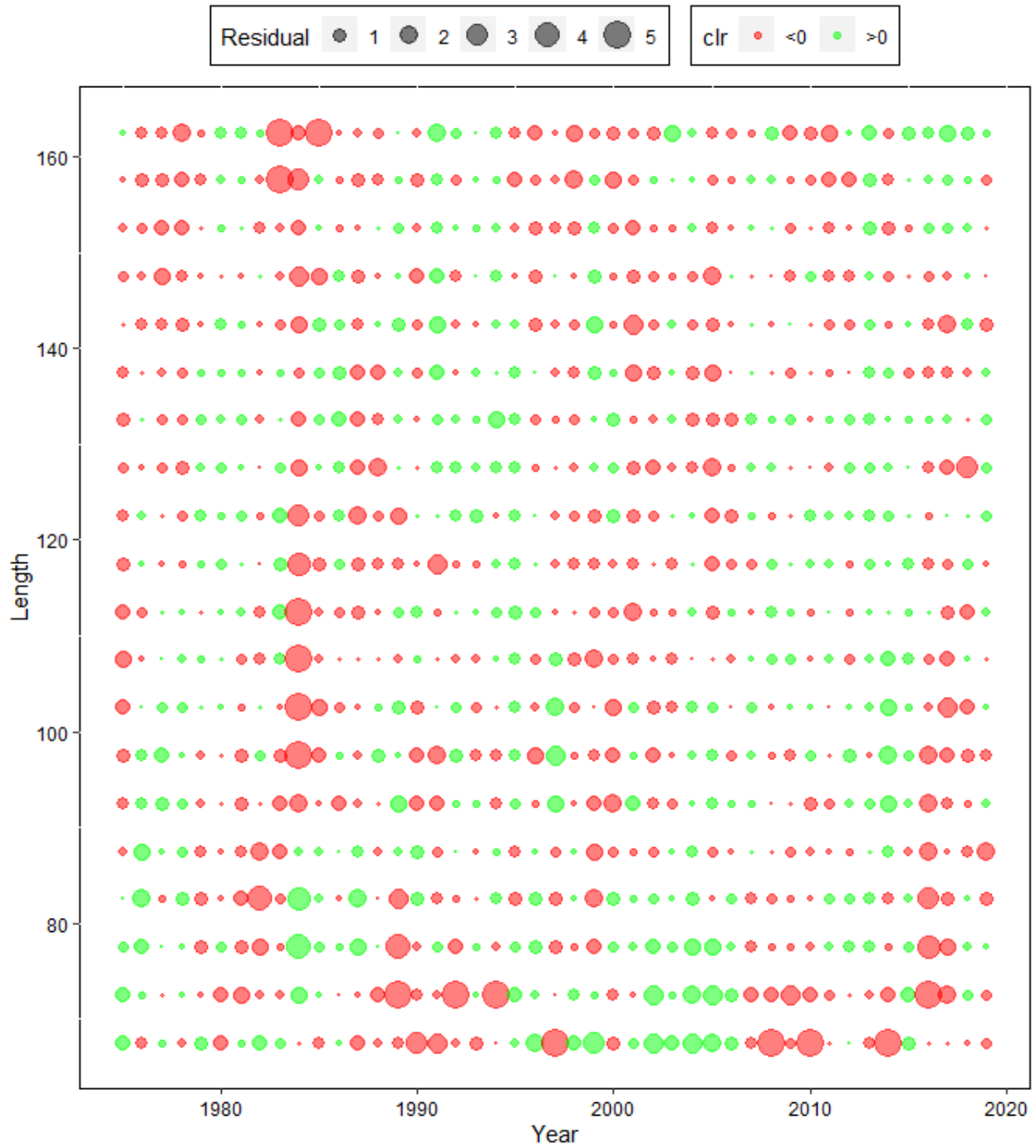


Figure 25b. Residuals of proportions of NMFS survey male red king crab by year and carapace length (mm) under model 19.3d. Green circles are positive residuals, and red circles are negative residuals.

### Model 19.3, Survey Females

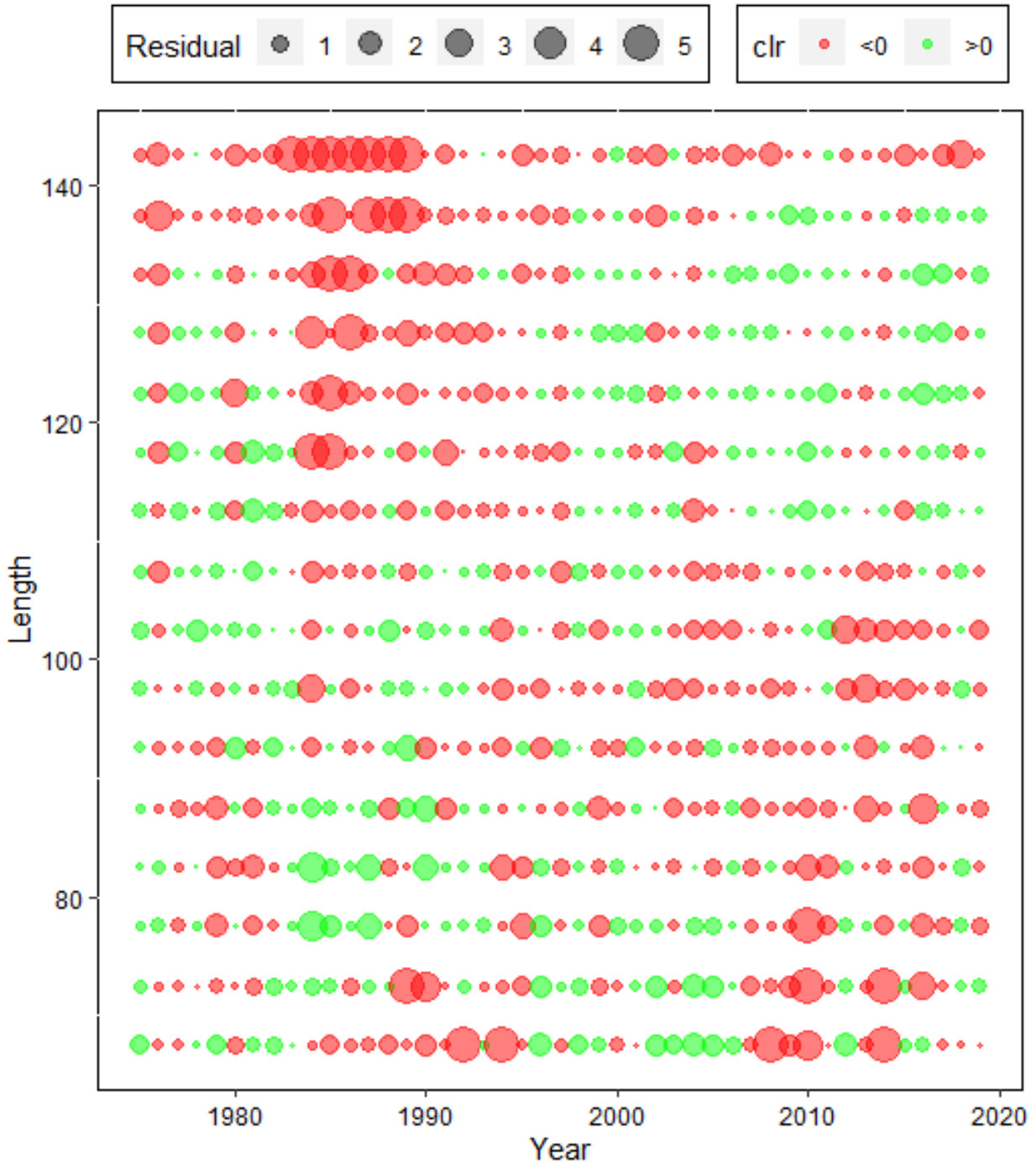


Figure 26a. Residuals of proportions of NMFS survey female red king crab by year and carapace length (mm) under model 19.3. Green circles are positive residuals, and red circles are negative residuals.



Model 19.3d, Survey Females

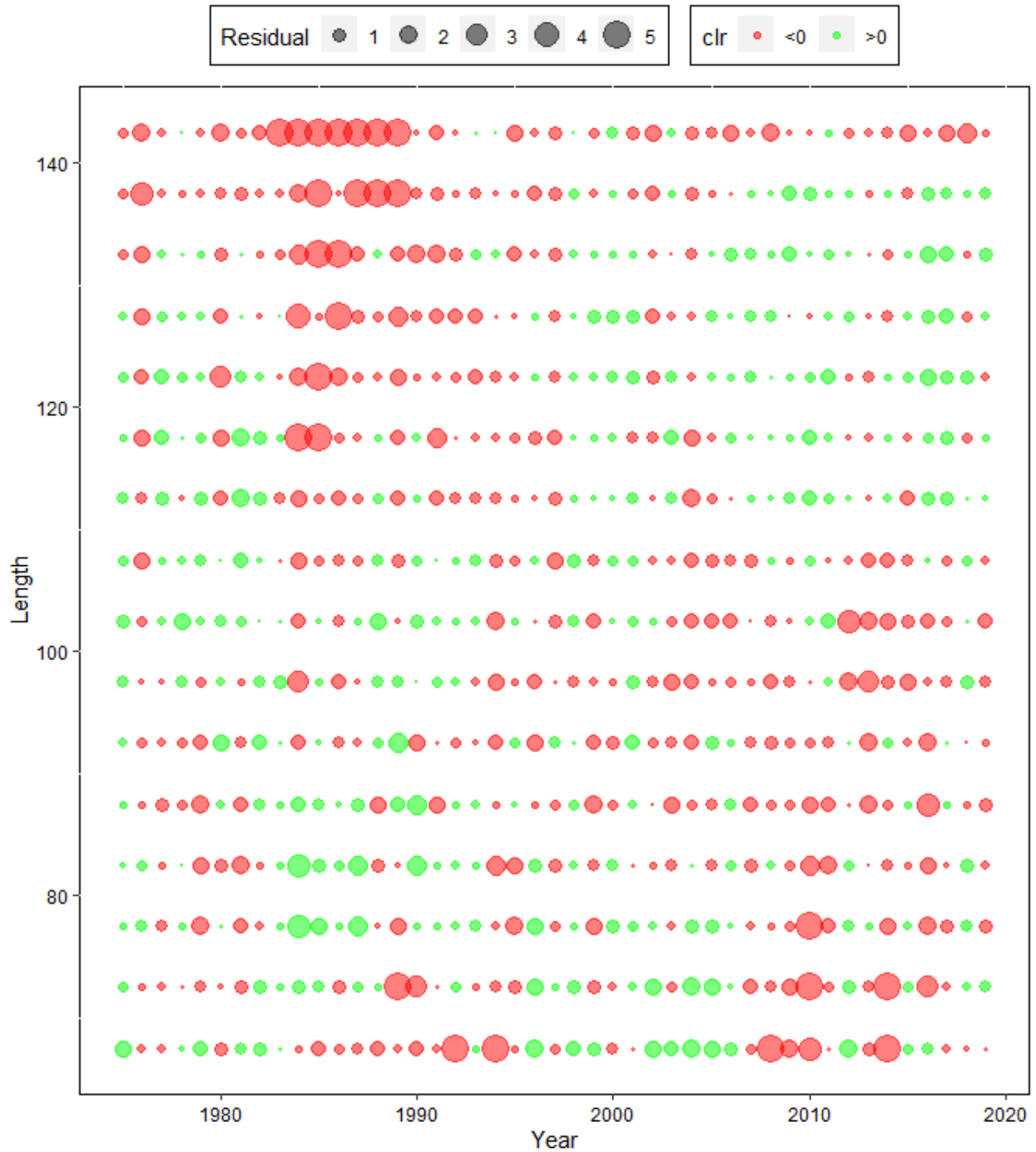


Figure 26b. Residuals of proportions of NMFS survey female red king crab by year and carapace length (mm) under model 19.3d. Green circles are positive residuals, and red circles are negative residuals.

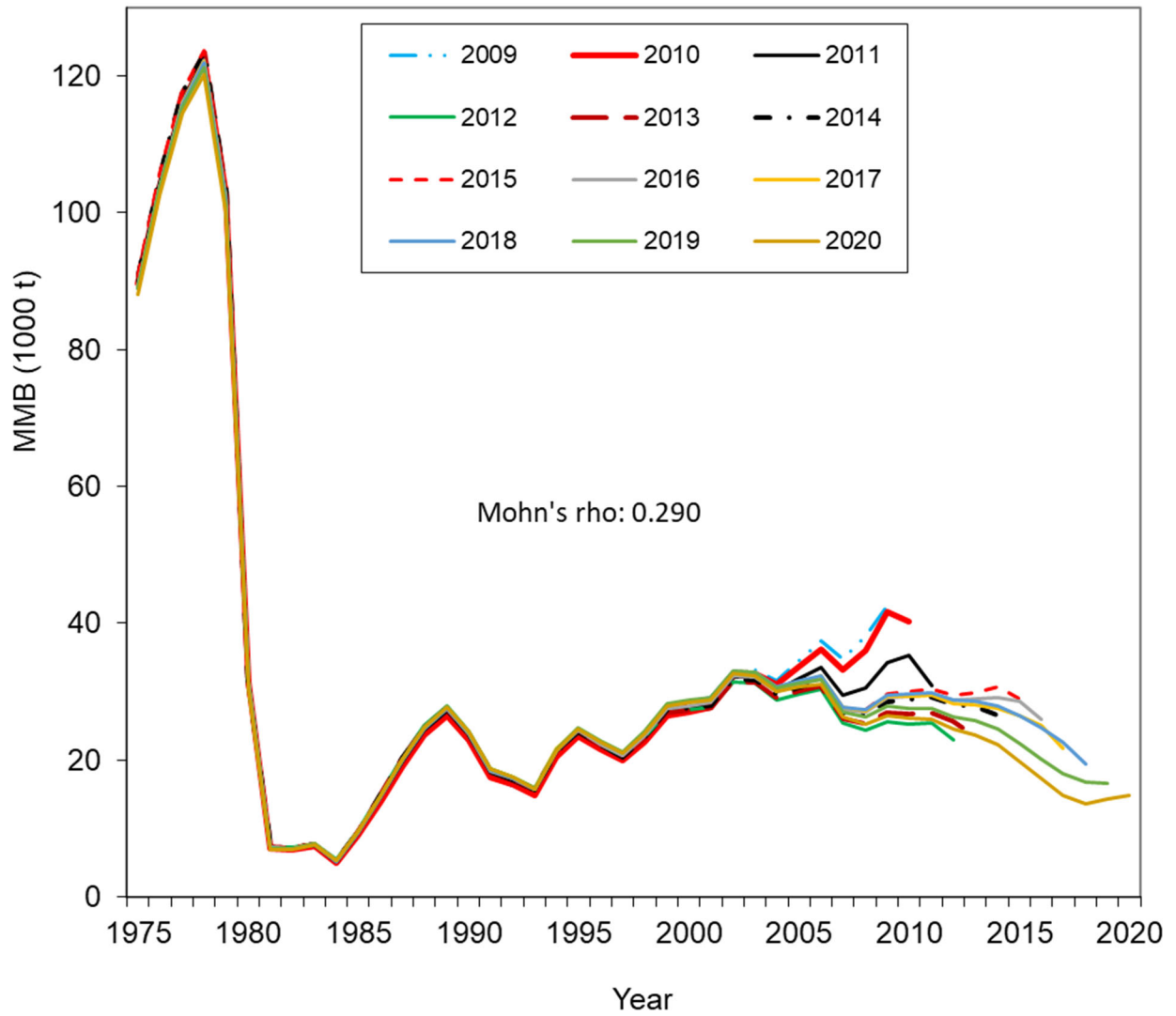


Figure 27a. Comparison of hindcast estimates of mature male biomass on Feb. 15 of Bristol Bay red king crab from 1975 to 2020 made with terminal years 2009-2020 with model 19.3. These are results of the 2020 model. Legend shows the terminal year.

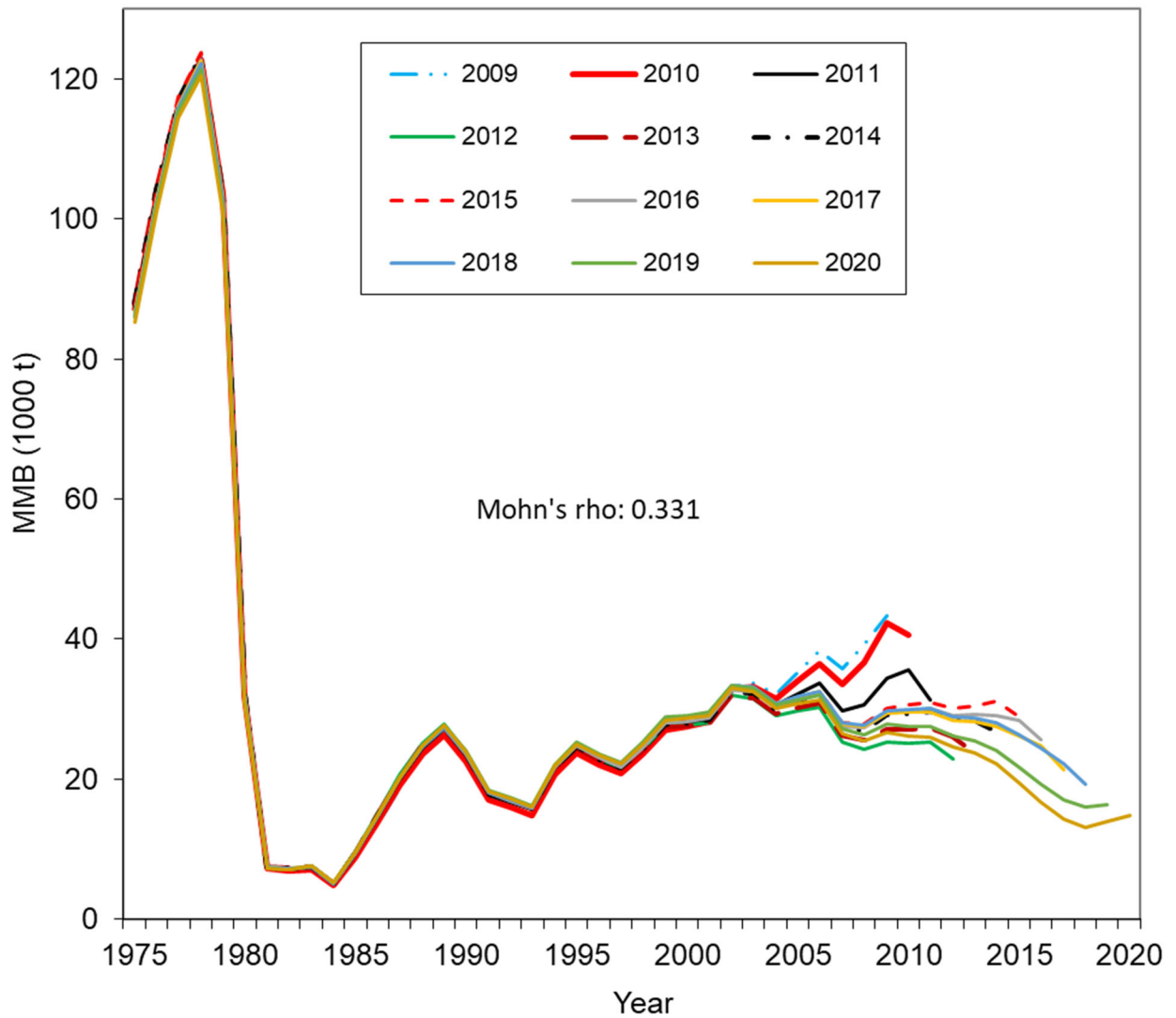


Figure 27b. Comparison of hindcast estimates of mature male biomass on Feb. 15 of Bristol Bay red king crab from 1975 to 2020 made with terminal years 2009-2020 with model 19.3d. These are results of the 2020 model. Legend shows the terminal year.

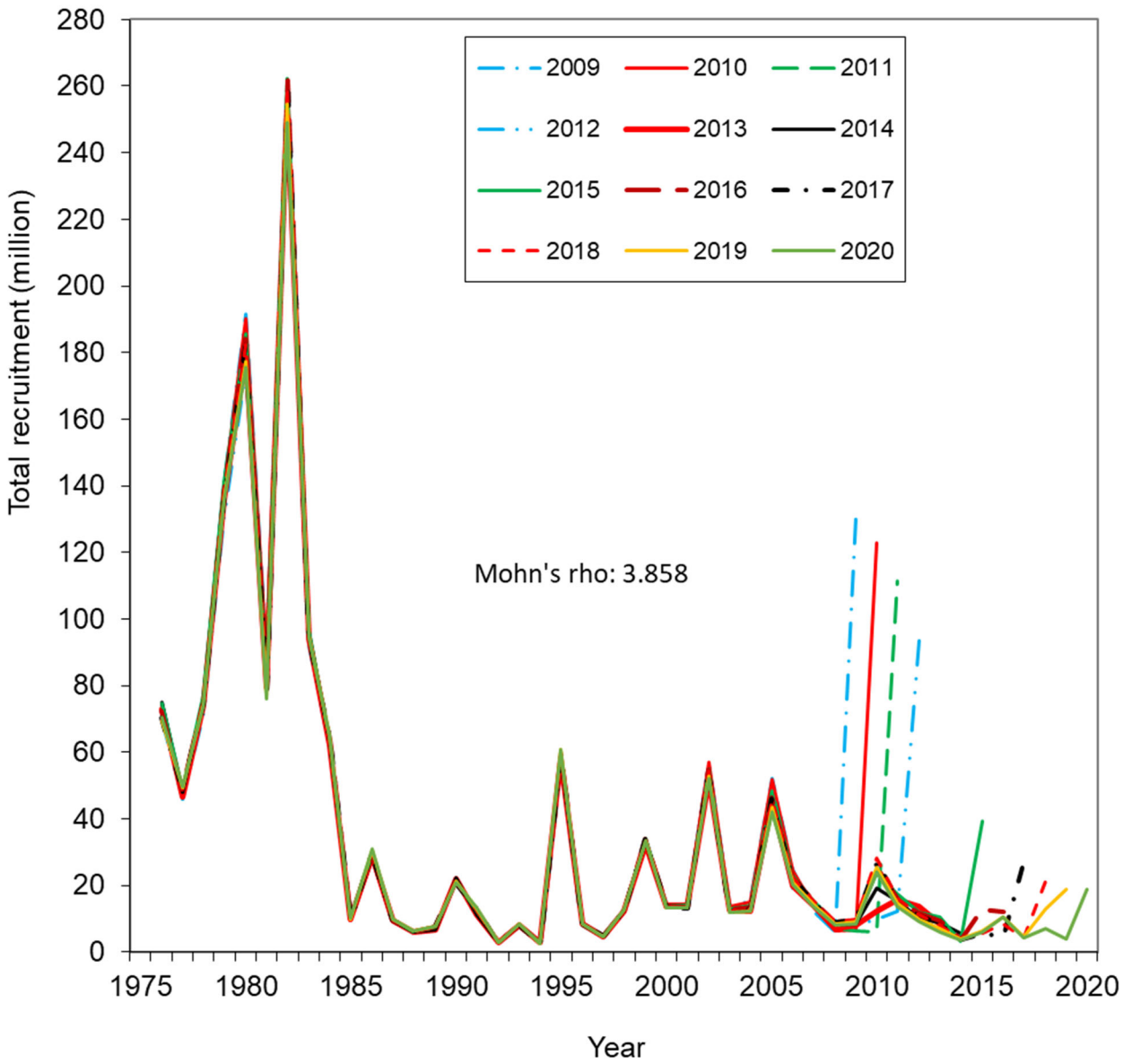


Figure 28a. Comparison of hindcast estimates of total recruitment for model 19.3 of Bristol Bay red king crab from 1976 to 2020 made with terminal years 2009-2020. These are results of the 2020 model. Legend shows the terminal year.

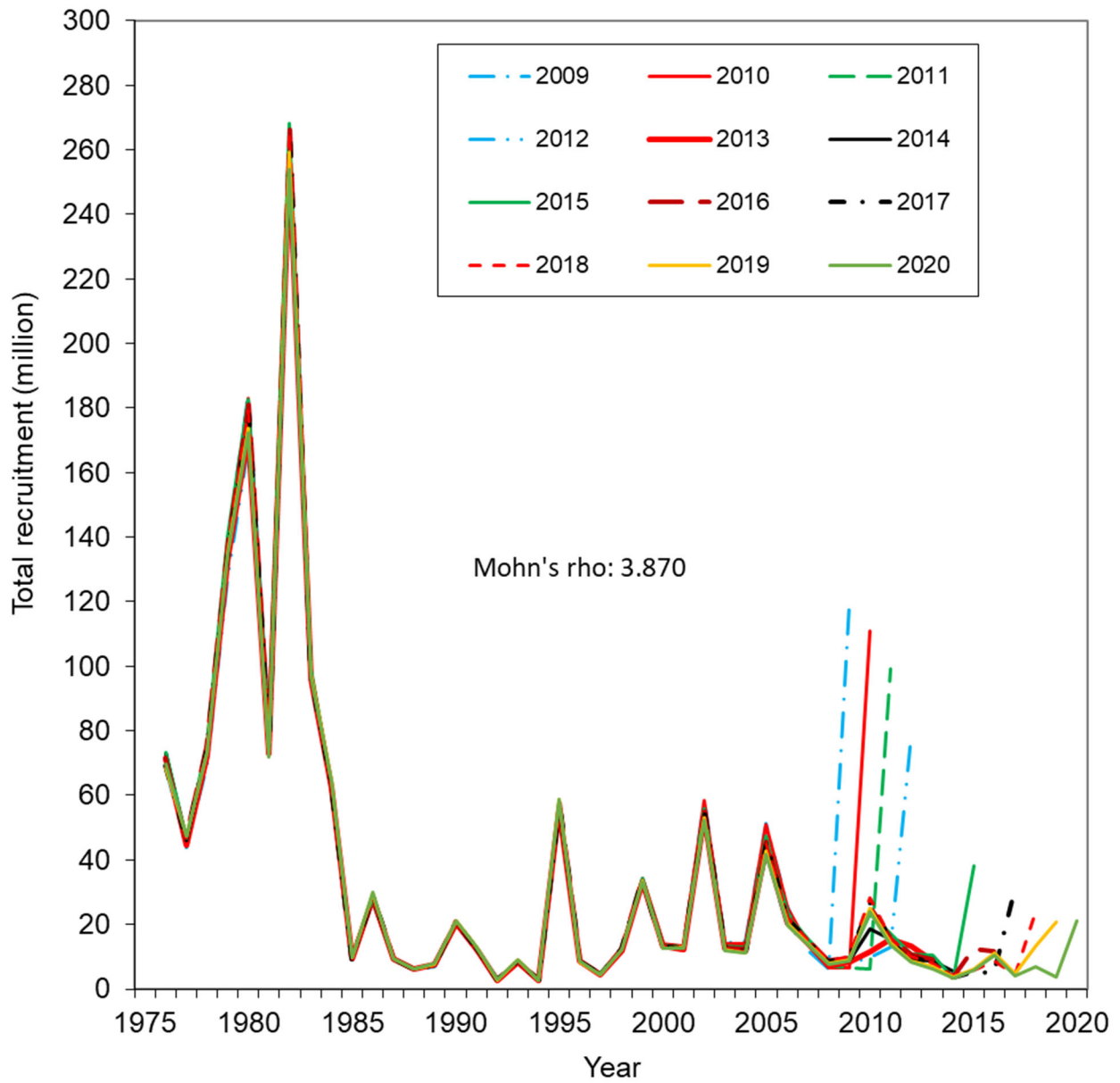


Figure 28b. Comparison of hindcast estimates of total recruitment for model 19.3d of Bristol Bay red king crab from 1976 to 2020 made with terminal years 2009-2020. These are results of the 2020 model. Legend shows the terminal year.

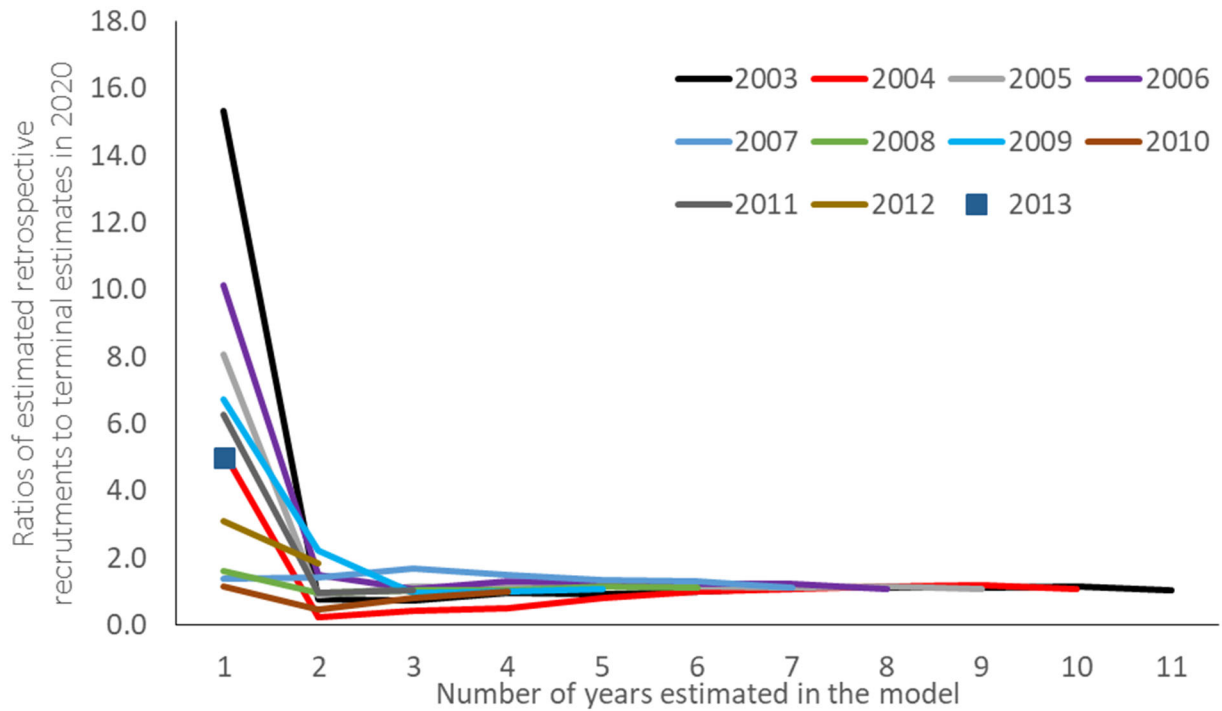


Figure 28c. Evaluation of Bristol Bay red king crab retrospective errors on recruitment estimates as a function of the number of years in the model for model 19.3.

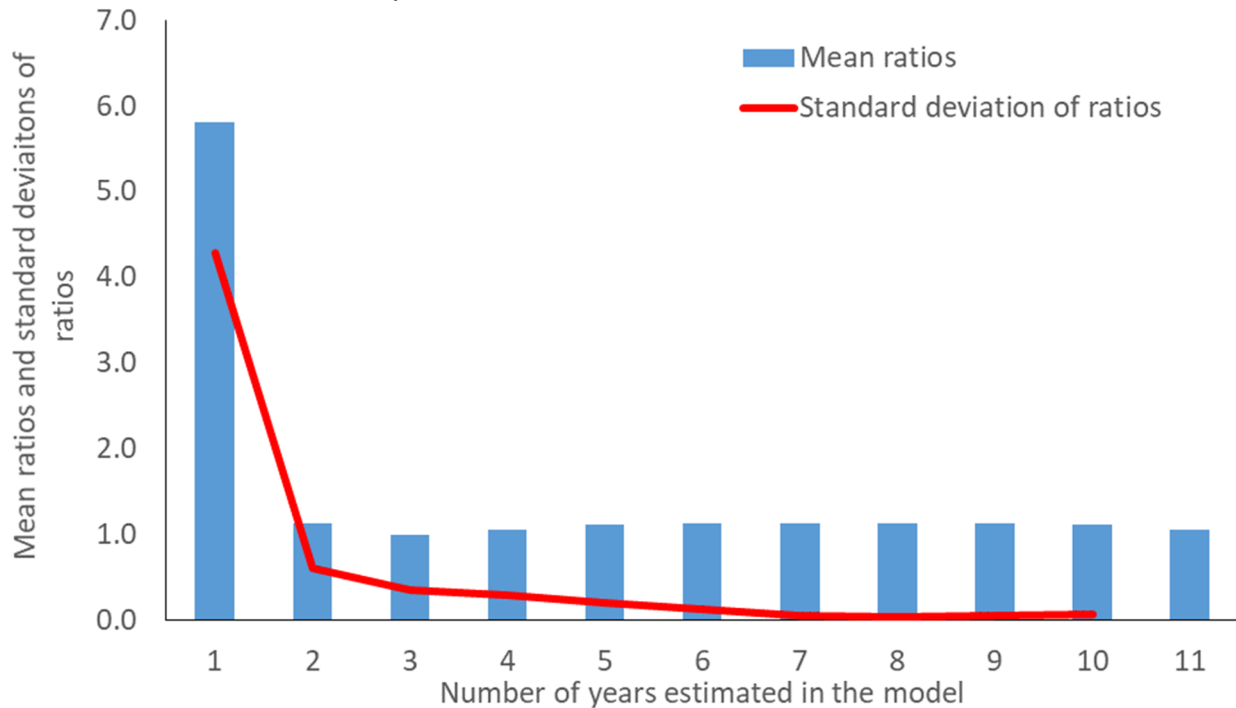


Figure 28d. Mean ratios of retrospective estimates of recruitments to those estimated in the most recent year (2020) and standard deviations of the ratios as a function of the number of years in the model for model 19.3.

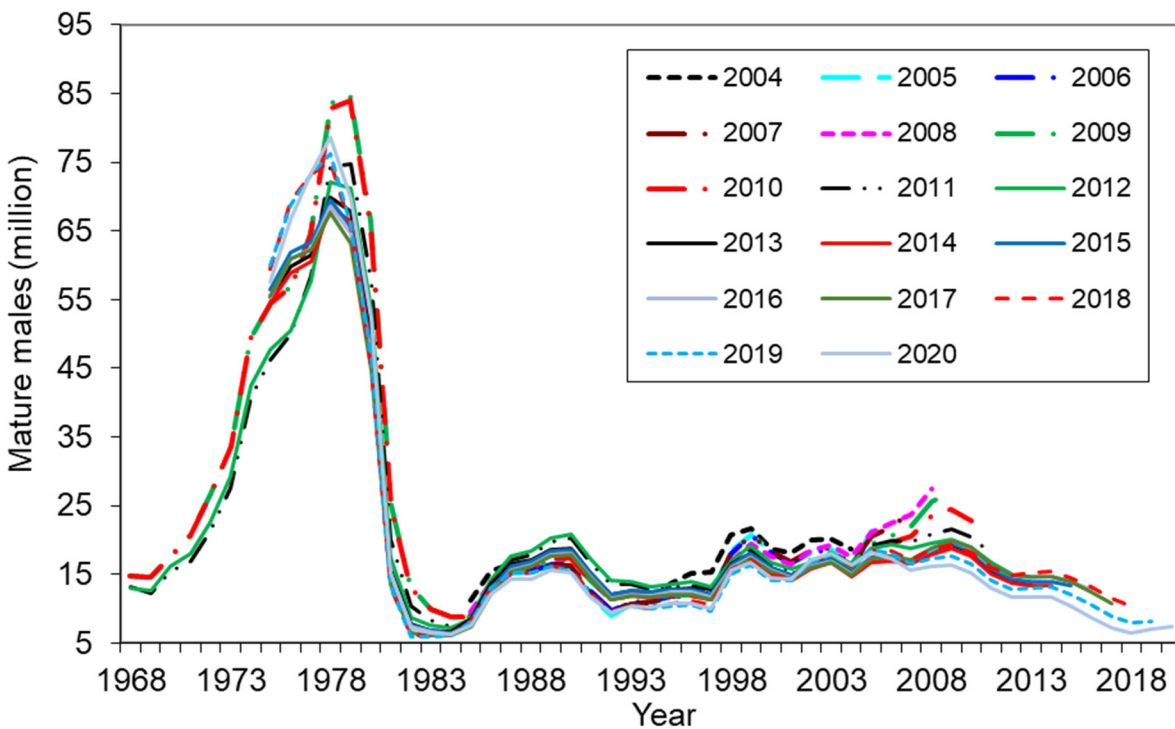
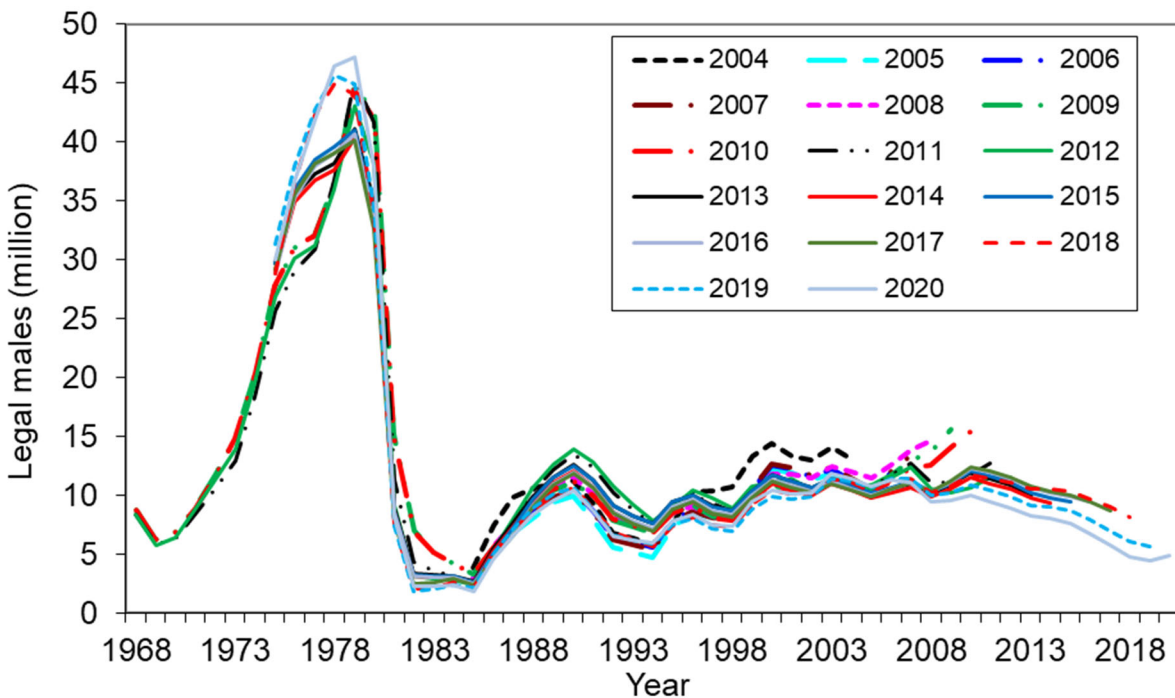


Figure 29. Comparison of estimates of legal male abundance (top) and mature males (bottom) of Bristol Bay red king crab from 1968 to 2020 made with terminal years 2004-2020 with the base models. Model 19.3 is used for 2020. These are results of historical assessments. Legend shows the year in which the assessment was conducted.

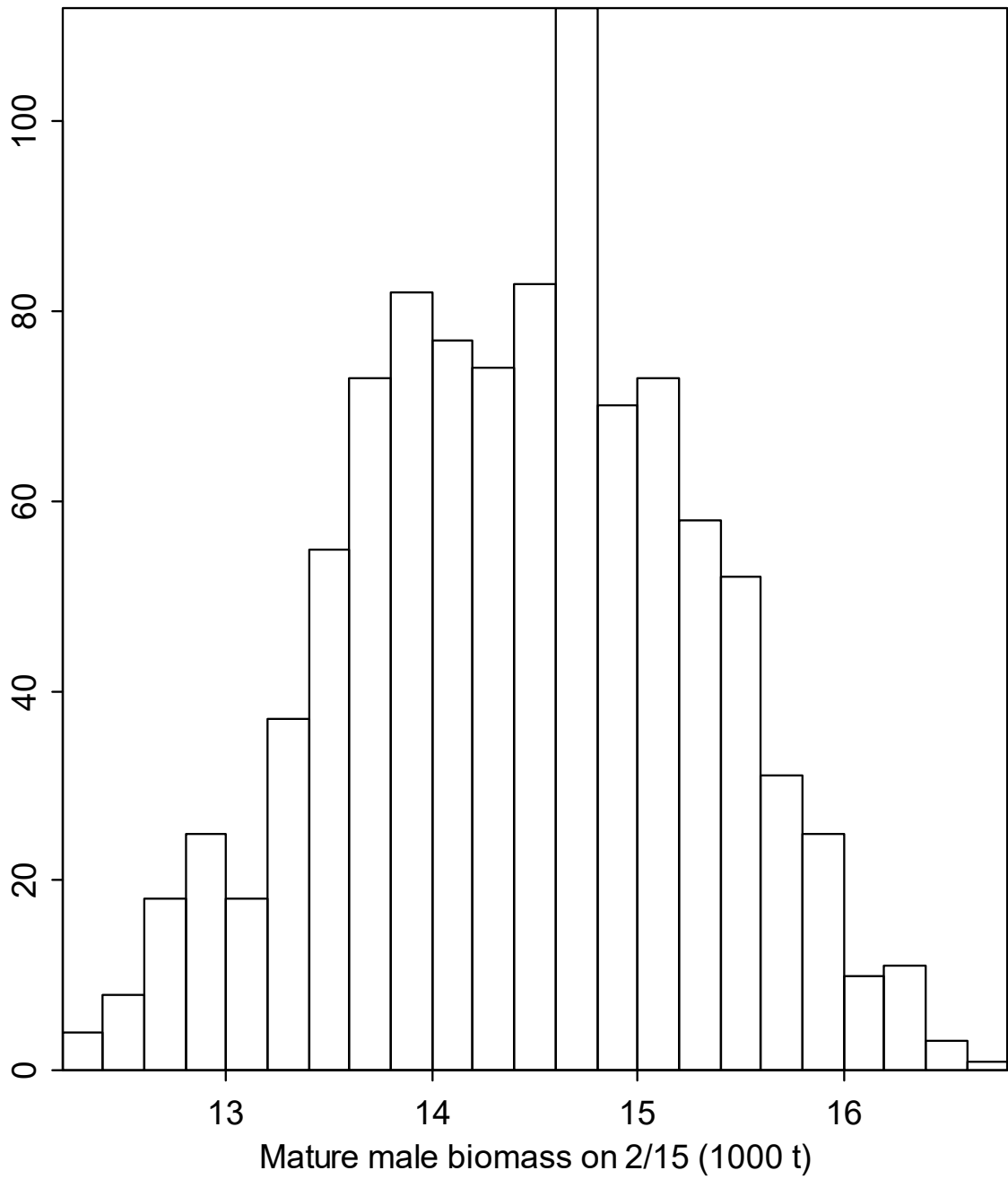


Figure 30a. Histogram of estimated mature male biomass on Feb. 15, 2021 under model 19.3 with the MCMC approach.



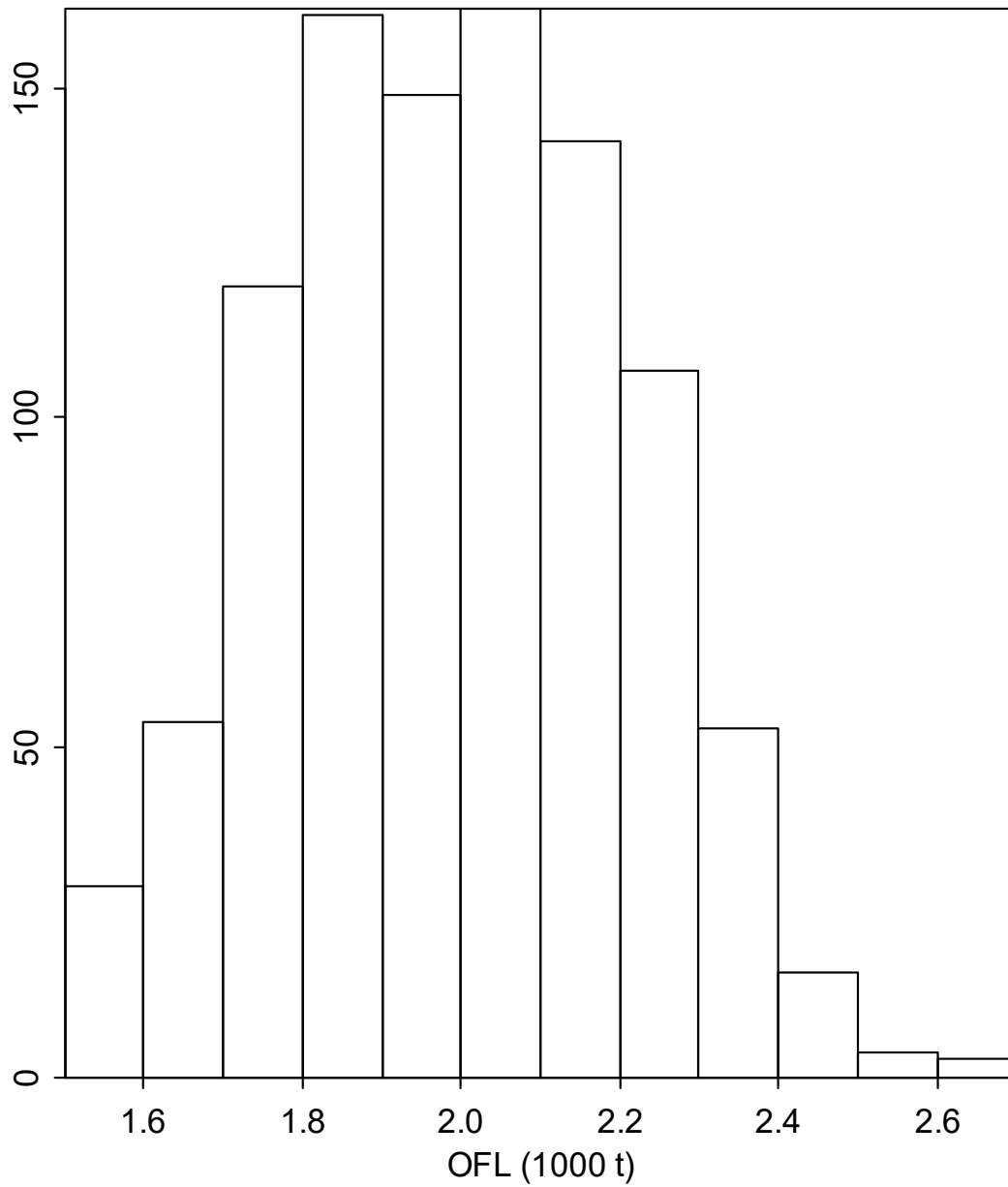


Figure 30b. Histogram of the 2020 estimated OFL under model 19.3 with the MCMC approach.

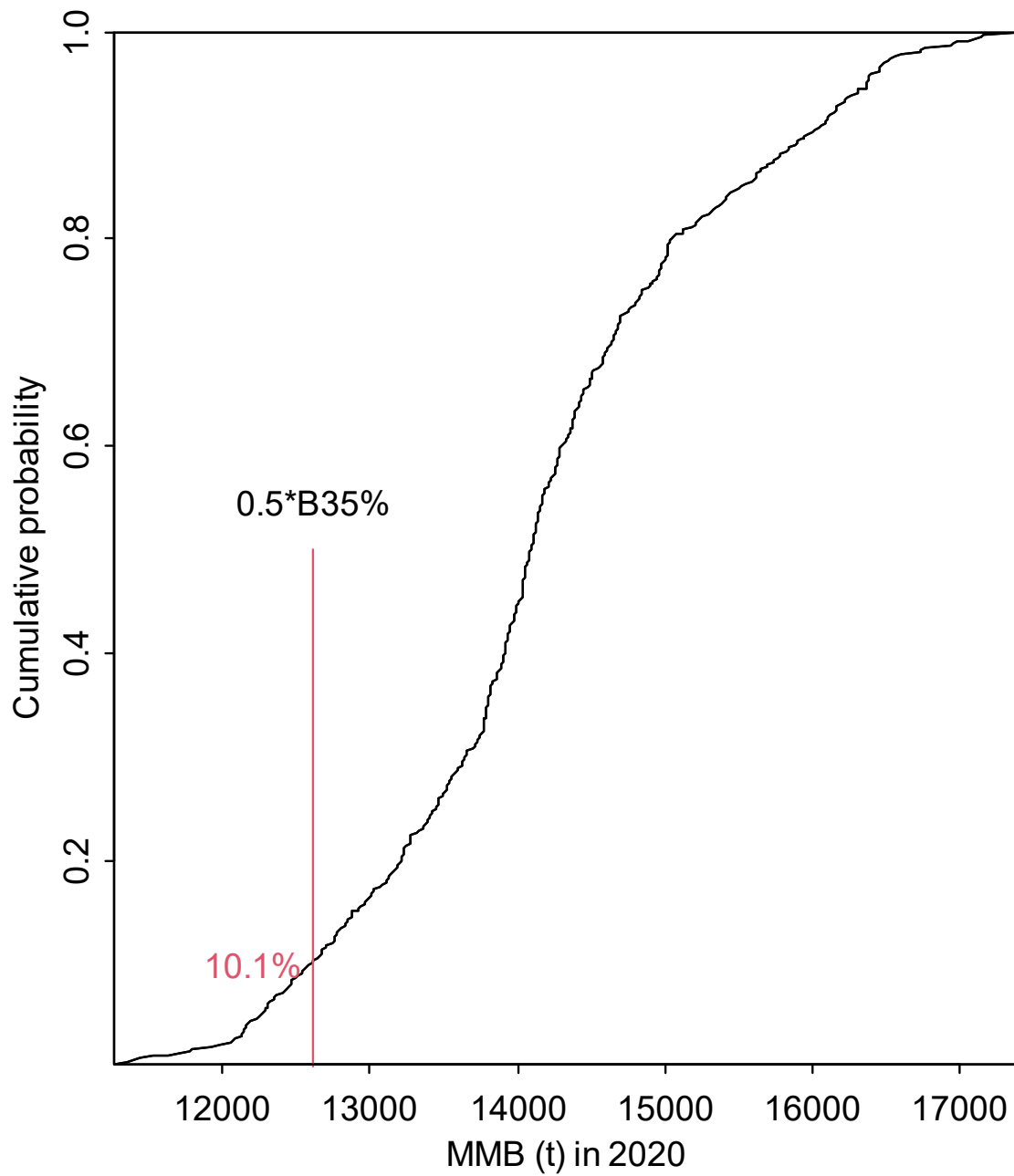


Figure 31. Cumulative probabilities of estimated mature male biomass (MMB) in 2020 under model 19.3d with the MCMC approach. About 10.1% probability is below the minimum threshold.

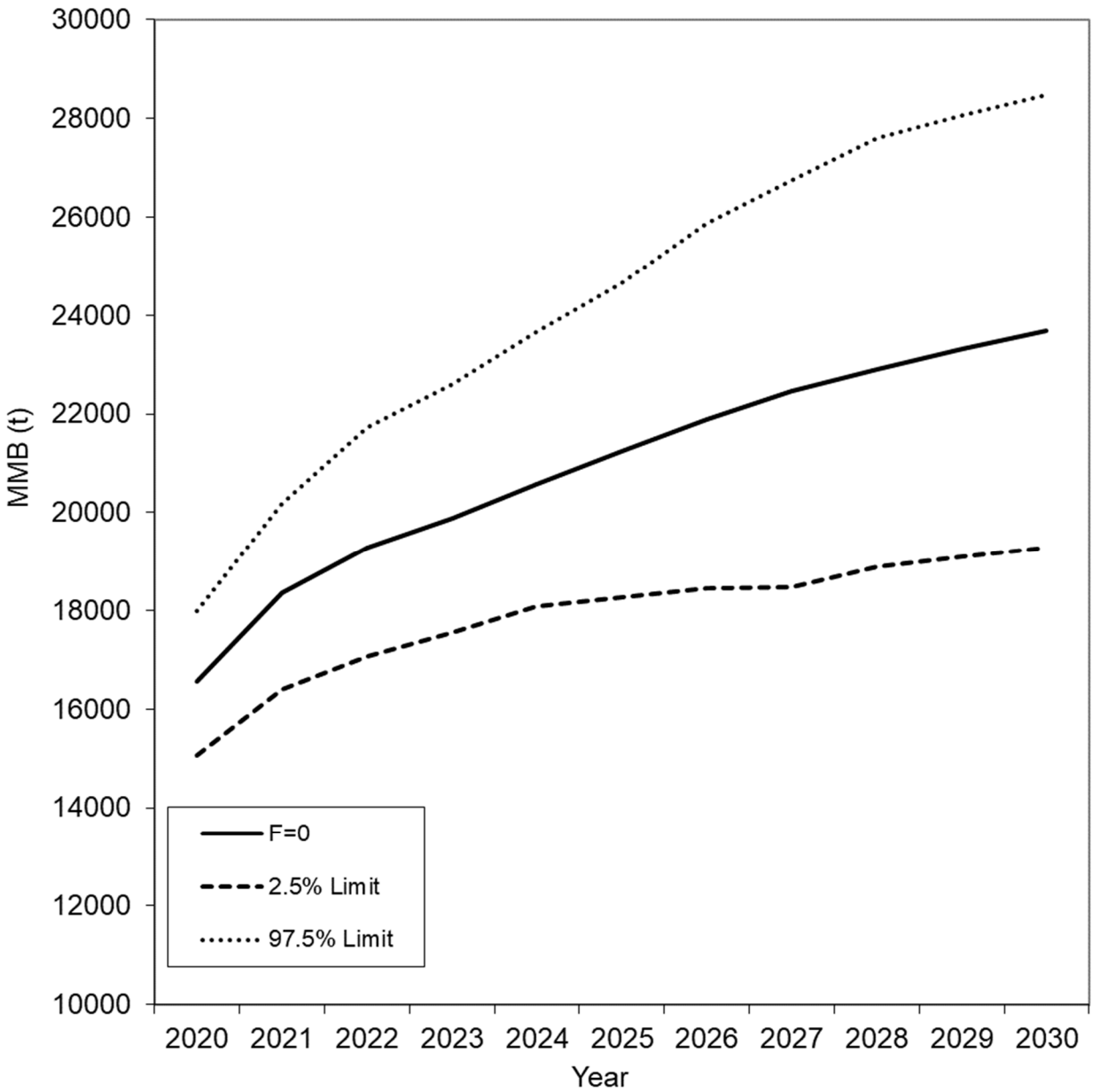


Figure 32a. Projected mature male biomass on Feb. 15 with  $F = 0$  harvest strategy during 2020-2030. Input parameter estimates are based on model 19.3a.

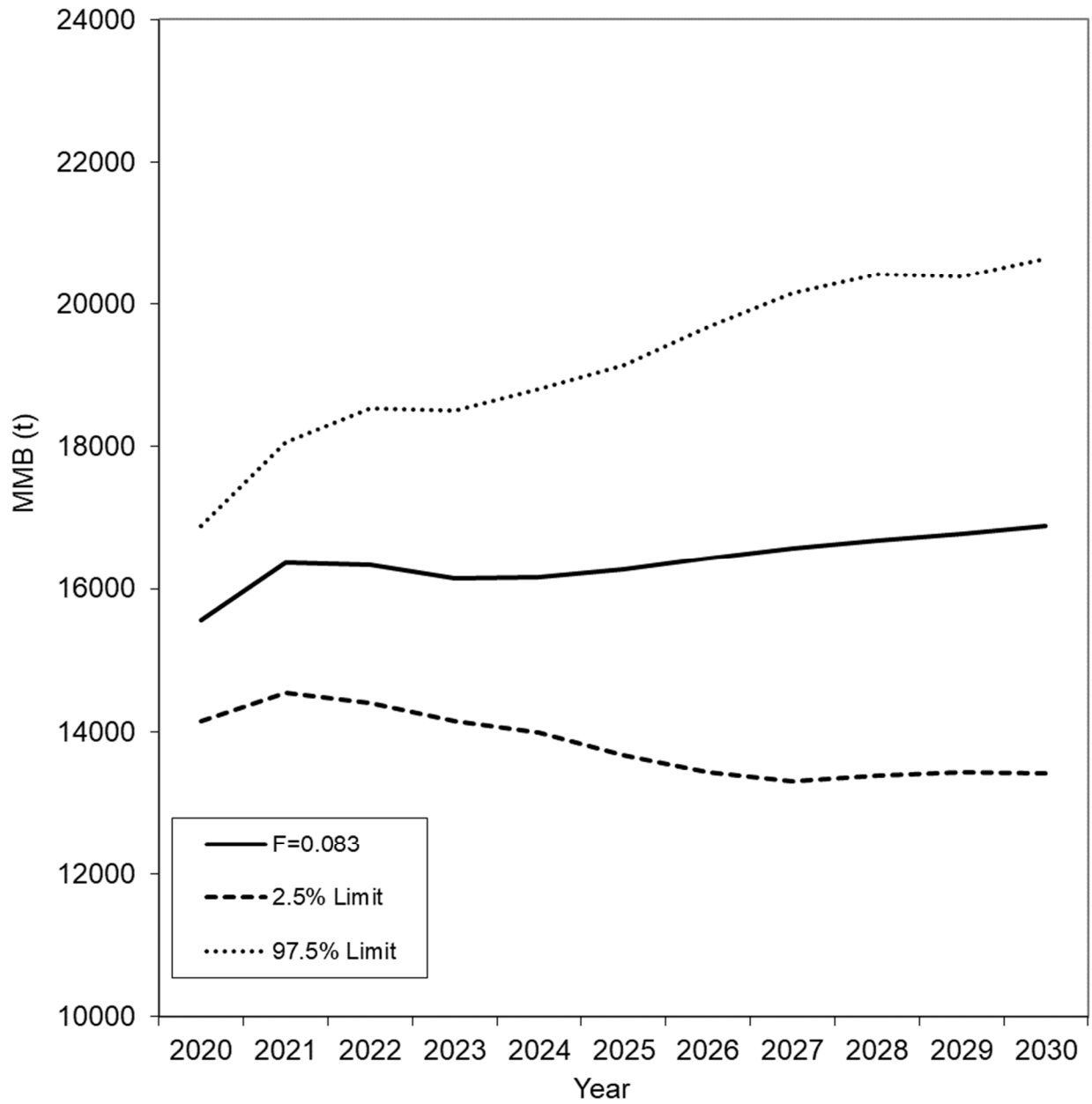


Figure 32b. Projected mature male biomass on Feb. 15 with  $F = 0.083$  harvest strategy during 2020-2030. Input parameter estimates are based on model 19.3a.

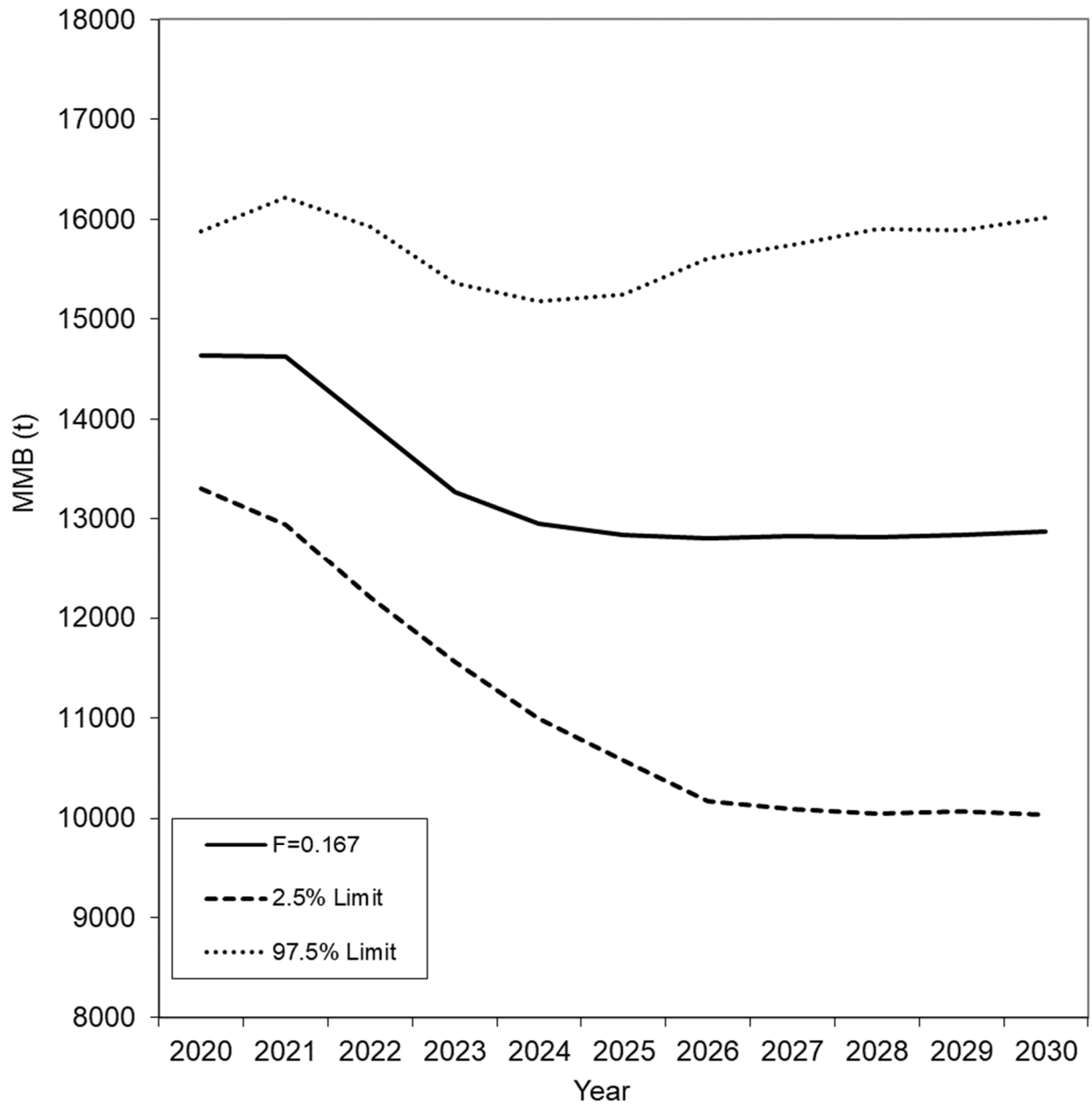


Figure 32c. Projected mature male biomass on Feb. 15 with  $F = 0.167$  harvest strategy during 2020-2030. Input parameter estimates are based on model 19.3a.

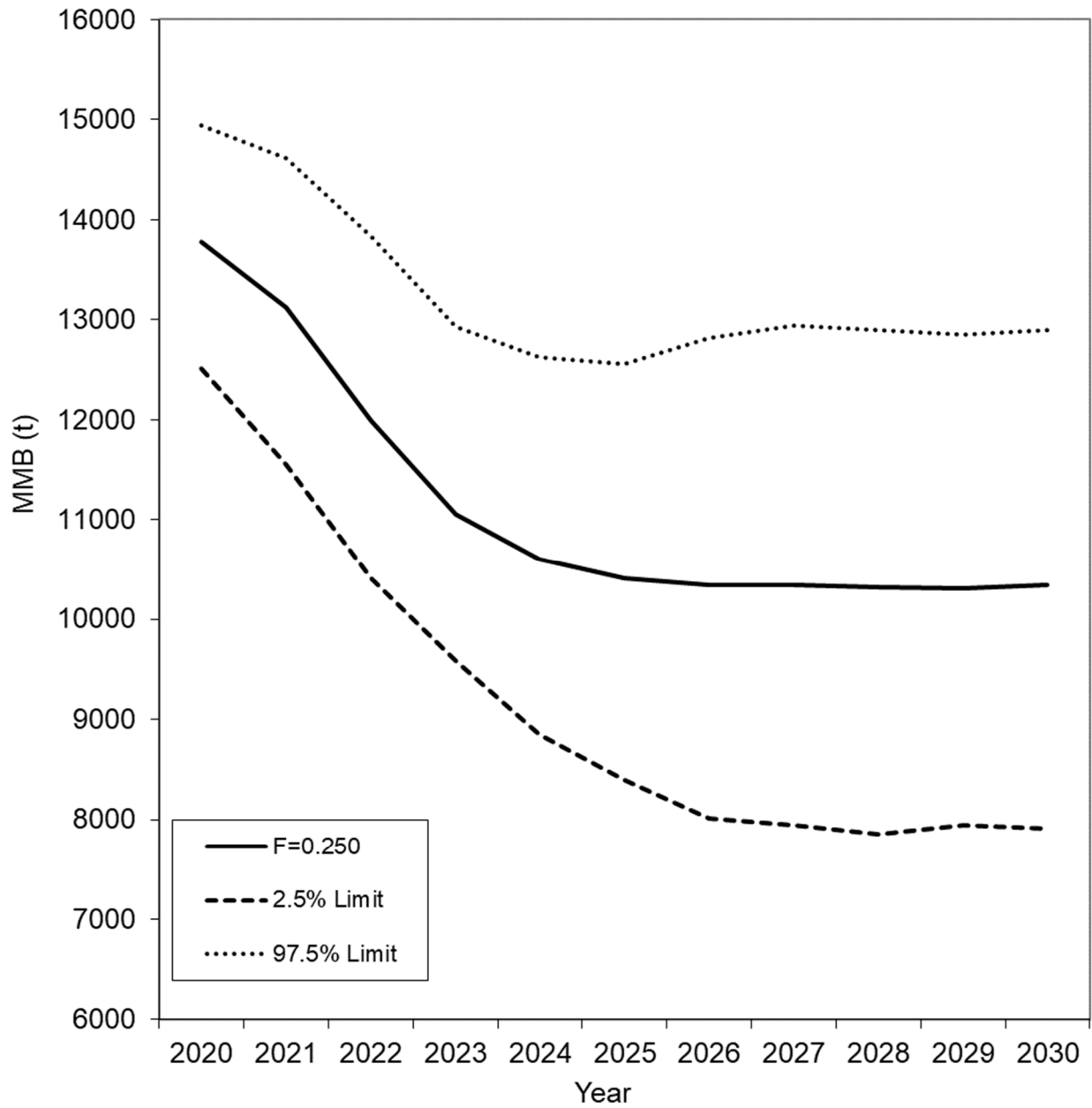


Figure 32d. Projected mature male biomass on Feb. 15 with  $F = 0.250$  harvest strategy during 2020-2030. Input parameter estimates are based on model 19.3a.

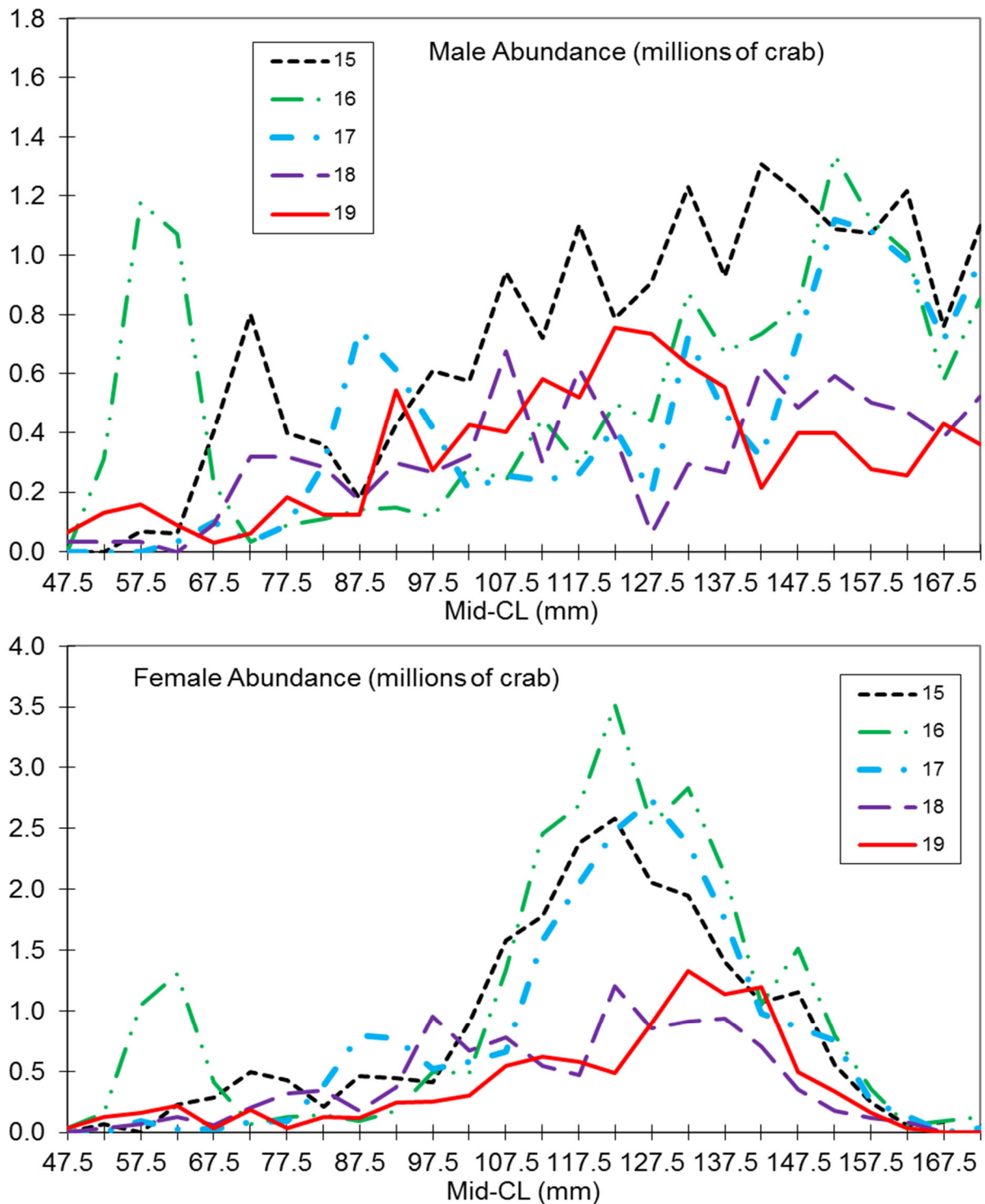


Figure 33. Length frequency distributions of male (top panel) and female (bottom panel) red king crab in Bristol Bay from NMFS trawl surveys during 2015-2019. For purposes of these graphs, abundance estimates are based on area-swept methods.

## ***Appendix A. Description of GMACS with Bristol Bay Red King Crab Options (mainly from the GMACS document)***

### **A. Model Description**

#### ***a. Population model***

The basic dynamics account for growth, mortality, maturity state and shell condition (although most of the equations below do not explicitly refer to maturity state and shell condition). For the case in which shell condition is not distinguished:

$$\underline{N}_{y,t}^g = ((\mathbf{I} - \mathbf{P}_{y,t-1}^g) + \mathbf{X}_{y,t-1}^g \mathbf{P}_{y,t-1}^g) \mathbf{S}_{y,t-1}^g \underline{N}_{y,t-1}^g + \tilde{\mathbf{R}}_{y,t}^g \quad (\text{A.1})$$

where  $\underline{N}_{y,t}^g$  is the number of animals by size-class of gender  $g$  at the start of season  $t$  of year  $y$ ,  $\mathbf{P}_{y,t}^g$  is a matrix with diagonals given by vector of molting probabilities for animals of gender  $g$  at the start of season  $t$  of year  $y$ ,  $\mathbf{S}_{y,t}^g$  is a matrix with diagonals given by the vector of probabilities of surviving for animals of gender  $g$  during time-step  $t$  of year  $y$  (which may be of zero duration):

$$S_{y,t,l}^g = \exp(-Z_{y,t,l}^g) \quad (\text{A.2})$$

$\mathbf{X}_{y,t}^g$  is the size-transition matrix (probability of growing from one size-class to each of the other size-classes or remains in the same size class) for animals of gender  $g$  during season  $t$  of year  $y$ ,  $\tilde{\mathbf{R}}_{y,t}^g$  is the recruitment (by size-class) to gear  $g$  during season  $t$  of year  $y$  (which will be zero except for one season – the recruitment season), and  $Z_{y,t,l}^g$  is the total mortality for animals of gender  $g$  in size-class  $l$  during season  $t$  of year  $y$ . Note that mortality is continuous across a time-step.

The initial conditions for the model (i.e., the numbers-at-size at the start of the first year,  $y_1$ ) is specified with an overall total recruitment multiplied by offsets for each size-class, i.e.:

$$N_{y_1,l}^g = R_{\text{mit}} e^{\delta_{y_1,l}^g} / \sum_{g'} \sum_{l'} e^{\delta_{y_1,l'}^{g'}} \quad (\text{A.3})$$

The minimum carapace length for both males and females is set at 65 mm, and crab abundance is modeled with a length-class interval of 5 mm. The last length class includes all crab  $\geq 160$ -mm CL for males and  $\geq 140$ -mm CL for females. Thus, length classes/groups are 20 for males and 16 for females.

#### ***b. Recruitment***

Recruitment occurs once during each year. Recruitment by sex and size-class is the product of total recruitment, the split of the total recruitment to sex and the assignment of sex-specific recruitment to size-classes, i.e.:



$$\tilde{R}_{y,t,l}^g = \bar{R} e^{\varepsilon_y} \begin{cases} (1 + e^{\phi_y})^{-1} p_l^{r,\text{mal}} & \text{if } g=\text{males} \\ \phi_y (1 + e^{\phi_y})^{-1} p_l^{r,\text{fem}} & \text{if } g=\text{females} \end{cases} \quad (\text{A.4})$$

where  $\bar{R}$  is median recruitment,  $\phi_y$  determines the sex ratio of recruitment during year  $y$ , and  $p_l^{r,g}$  is the proportion of the recruitment (by gender and year) that recruits to size-class  $l$ :

$$p_l^{r,g} = \int_{L_l^{\text{low}}}^{L_l^{\text{hi}}} \frac{1}{\Gamma(\alpha^{r,g}/\beta^{r,g})} (l/\beta^{r,g})^{(\alpha^{r,g}/\beta^{r,g})-1} e^{-l/\beta^{r,g}} dl \quad (\text{A.5})$$

where  $\alpha^{r,g}$  and  $\beta^{r,g}$  are the parameters that define a gamma function for the distribution of recruits to size-class. Equation A.5 can be restricted to a subset of size-classes, in which case the results from Equation A.5 are normalized to sum to 1 over the selected size-classes.

### c. Total mortality / probability of encountering the gear

Total mortality is the sum of fishing mortality and natural mortality, i.e.:

$$Z_{y,t,l}^g = \rho_{y,t}^M M_y^g \tilde{M}_l + \sum_f S_{y,t,l}^{f,g} (\lambda_{y,t,l}^{f,g} + \Omega_{y,t,l}^{f,g} (1 - \lambda_{y,t,l}^{f,g})) F_{y,t}^{f,g} \quad (\text{A.6})$$

where  $\rho_{y,t}^M$  is the proportion of natural mortality that occurs during season  $t$  for year  $y$ ,  $M_y^g$  is the rate of natural mortality for year  $y$  for animals of gender  $g$  (applies to animals for which  $\tilde{M}_l = 1$ ),  $\tilde{M}_l$  is the relative natural mortality for size-class  $l$ ,  $S_{y,t,l}^{f,g}$  is the (capture) selectivity for animals of gender  $g$  in size-class  $l$  by fleet  $f$  during season  $t$  of year  $y$ ,  $\lambda_{y,t,l}^{f,g}$  is the probability of retention for animals of gender  $g$  in size-class  $l$  by fleet  $f$  during season  $t$  of year  $y$ ,  $\Omega_{y,t,l}^{f,g}$  is the mortality rate for discards of gender  $g$  in size-class  $l$  by fleet  $f$  during season  $t$  of year  $y$ , and  $F_{y,t}^{f,g}$  is the fully-selected fishing mortality for animals of gender  $g$  by fleet  $f$  during season  $t$  of year  $y$ .

The probability of encountering the gear (occurs instantaneously) is given by:

$$\tilde{Z}_{y,t,l}^g = \sum_f S_{y,t,l}^{f,g} F_{y,t}^{f,g} \quad (\text{A.7})$$

Note that Equation A.7 is computed under the premise that fishing is instantaneous and hence that there is no natural mortality during season  $t$  of year  $y$ .

The logarithms of the fully-selected fishing mortalities by season are modelled as:

$$\ln F_{y,t}^{f,\text{mal}} = \ln F_{y,t}^{f,\text{mal}} + \xi_{y,t}^{f,\text{mal}} \quad (\text{A.8})$$

$$\ln F_{y,t}^{f,\text{fem}} = \ln F_{y,t}^{f,\text{mal}} + \phi^f + \xi_{y,t}^{f,\text{fem}} \quad (\text{A.9})$$

where  $F^{f,\text{mal}}$  is the reference fully-selected fishing mortality rate for fleet  $f$ ,  $\phi^f$  is the offset between female and male fully-selected fishing mortality for fleet  $f$ , and  $\xi_{y,t}^{f,g}$  are the annual deviation of fully-selected fishing mortality for fleet  $f$  (by gender).

Natural mortality can depend on time with blocked natural mortality (individual parameters). This option estimates natural mortality as parameters by block, i.e.:

$$M_y^g = e^{\psi_y^g} \quad (\text{A.10})$$

where  $M_{y_1}^g$  is the rate of natural mortality for gender  $g$  for the first year of the model, and  $\psi_y^g$  is the annual change in natural mortality and changes in blocks of years.

It is possible to ‘mirror’ the values for the  $\psi_y^g$  parameters (between genders and between blocks), which allows male and female natural mortality to be the same, and for natural mortality to be the same for discontinuous blocks (based on Equation A.10). It is also possible to estimate a ratio of natural mortality between genders. The deviations in natural mortality can also be penalized to avoid unrealistic changes in natural mortality to fit ‘quirks’ in the data.

#### **d. Landings, discards, total catch**

The model keeps track of (and can be fitted to) landings, discards, total catch by fleet in season with continuous mortality:

$$\text{Landed catch} \quad C_{y,t,l}^{\text{Land},f,g} = \frac{\lambda_{y,t,l}^{f,g} S_{y,t,l}^{f,g} F_{y,t}^{f,g}}{Z_{y,t,l}^g} N_{y,t,l}^{f,g} (1 - e^{-Z_{y,t,l}^g}) \quad (\text{A.11})$$

$$\text{Discards} \quad C_{y,t,l}^{\text{Disc},f,g} = \frac{(1 - \lambda_{y,t,l}^{f,g}) S_{y,t,l}^{f,g} F_{y,t}^{f,g}}{Z_{y,t,l}^g} N_{y,t,l}^{f,g} (1 - e^{-Z_{y,t,l}^g}) \quad (\text{A.12})$$

$$\text{Total catch} \quad C_{y,t,l}^{\text{Total},f,g} = \frac{S_{y,t,l}^{f,g} F_{y,t}^{f,g}}{Z_{y,t,l}^g} N_{y,t,l}^{f,g} (1 - e^{-Z_{y,t,l}^g}) \quad (\text{A.13})$$

Landings, discards, and total catches by fleet can be aggregated over gender (e.g., when fitting to removals reported as gender-combined). Equations A.11-13 are extended naturally for the case in which the population is represented by shell condition and/or maturity status (given the assumption that fishing mortality, retention and discard mortality depend on gender and time, but not on shell condition nor maturity status).

Landings, discards, and total catches by fleet can be reported in numbers (Equations A.11–13) or in terms of weight. For example, the landings, discards, and total catches by fleet, season, year, and gender for the total (over size-class) removals are computed as:

$$C_{y,t}^{\text{Land},g,f} = \sum_l C_{y,t,l}^{\text{Land},g,f} w_{y,l}^g; \quad C_{y,t}^{\text{Disc},g,f} = \sum_l C_{y,t,l}^{\text{Disc},g,f} w_{y,l}^g; \quad C_{y,t}^{\text{Total},g,f} = \sum_l C_{y,t,l}^{\text{Total},g,f} w_{y,l}^g \quad (\text{A.14})$$

where  $C_{y,t}^{\text{Land},g,f}$ ,  $C_{y,t}^{\text{Disc},g,f}$ , and  $C_{y,t}^{\text{Total},g,f}$  are respectively the landings, discards, and total catches in weight by fleet, season, year, and gender for the total (over size-class) removals, and  $w_{y,l}^g$  is the weight of an animal of gender  $g$  in size-class  $l$  during year  $y$ .

### ***e. Selectivity / retention***

Selectivity (the probability of encountering the gear) and retention (the probability of being landed given being captured) are logistic function:

$$S_l = 1 - \left( 1 + \frac{\exp((\bar{L}_l - S_{50}))}{\sigma^S} \right)^{-1} \quad (\text{A.15})$$

where  $S_{50}$  is the size corresponding to 50% selectivity,  $\sigma^S$  is the “standard deviation” of the selectivity curve, and  $\bar{L}_l$  is the midpoint of size-class  $l$ .

It is possible to assume that selectivity for one fleet is the product of two of the selectivity patterns. This option is used to model cases in which one survey (NMFS trawl survey) is located within the footprint of another survey (BSFRF trawl survey).

The options to model retention are the same as those for selectivity, except that it is possible to estimate an asymptotic parameter, which allows discard of animals that would be “fully retained” according to the standard options for (capture) selectivity.

Selectivity and retention can be defined for blocks of contiguous years. Two blocks are used for NMFS survey selectivity (before 1982 and after 1981) due to gear modifications and two blocks are used for the directed pot fishery retention (before 2005 and after 2004) due to the fishery rationalization.

### ***f. Growth***

Growth is a key component of any size-structured model. It is modelled in terms of molt probability and the size-transition matrix (the probability of growing from each size-class to each of the other size-classes, constrained to be zero for sizes less than the current size). Note that the size-transition matrix has entries on its diagonal, which represent animals that molt but do not change size-classes.

#### ***(1) Molt probability***

There are two options for modelling the probability of molting as a function of size,  $P_{l,l}$ :

- Constant probability (1 for females)
- Logistic probability (for males), i.e.:

$$P_{l,l} = 1 - (1 + \exp((\bar{L}_l - P_{50}) / \sigma^P))^{-1} \quad (\text{A.16})$$

where  $P_{50}$  is the size at which the probability of molting is 0.5, and  $\sigma^P$  is the “standard deviation” of the molt probability function.

Molt probability is specified by gender and can change in blocks (one block before 1981 and one block after 1980 for males).

## (2) Size-transition

The proportion of animals in size-class  $j$  that grow to be in size-class  $i$  ( $X_{i,j}$ ) can be pre-specified as gamma-distributed size-increments:

$$X_{i,j} = \int_{L_j^{\text{low}}}^{L_j^{\text{hi}}} \frac{1}{\Gamma(I_i/\tilde{\beta})} ((l - \bar{L}_i)/\tilde{\beta})^{(I_i/\tilde{\beta})-1} e^{-(l-\bar{L}_i)/\tilde{\beta}} dl \quad (\text{A.17})$$

where  $I_i$  is the ‘expected’ growth increment for an animal in size-class  $i$  (a linear function of the mid-point of size-class  $i$ ),  $\tilde{\beta}$  determines the variation in growth among individuals, and  $L_j^{\text{low}}$  and  $L_j^{\text{hi}}$  are respectively the lower and upper bounds of size-class  $j$ .

The size-transition matrix is specified by gender and can change in blocks (one block for males and three blocks for females (1975-1982, 1983-1993, and 1994-present based on changes in sizes at maturity)).

## B. Outputs, Projections and OFL Calculation

### a. Core model outputs

The core model outputs are the N-matrix, the matrix of fully-selected fishing mortalities, the time-series of spawning stock biomass, mature male biomass (SSB), the values for the model parameters, and the predictions related to the observations. The spawning stock biomass (and hence mature male biomass) is defined according to:

$$SSB_y = \sum_g p^{\text{SSB},g} \sum_l N_{y,t^*,l}^g \quad (\text{A.18})$$

where  $p^{\text{SSB},g}$  is the relative contribution of gender  $g$  to spawning biomass ( $p^{\text{SSB},\text{mal}} = 1$ ;  $p^{\text{SSB},\text{fem}} = 0$  corresponds to spawning stock biomass equating to mature male biomass), and  $t^*$  is the season in which spawning takes place (spawning occurs at the start of the season).

Definition of model outputs:

- (1) Biomass: two population biomass measurements are used in this report: total survey biomass (crab >64 mm CL) and mature male biomass (males >119 mm CL). Mating time is assumed to Feb. 15.
- (2) Recruitment: new entry of number of males in the 1st seven length classes (65- 99 mm CL) and new entry of number of females in the 1st five length classes (65-89 mm CL).
- (3) Fishing mortality: full-selected instantaneous annual fishing mortality rate at the time of fishery.

### ***b. Biological reference points***

The key biological reference points are the proxy for  $F_{MSY}$ , the proxy for  $B_{MSY}$  and the Overfishing Level (OFL).

#### *(1) The proxy for $F_{MSY}$*

The specification for the proxy for  $F_{MSY}$  depends on the tier in which the stock is placed. BBRKC belongs to Tier 3, and the proxy for  $F_{MSY}$  is  $F_{35\%}$ , the value of a multiplier on the fully-selected fishing mortality rates for directed fisheries in the final year of the assessment such that spawning biomass-per-recruit is 35% of the unfished level. The fully-selected fishing mortality rates for non-directed fisheries are set to recent averages (recent 5 years for BBRKC). The unfished spawning biomass-per-recruit,  $SSBPR(\underline{0})$ , is calculated by projecting the population model forward where fishing mortality is zero for all fleets, and recruitment is constant (and ideally equal to 1).  $F_{35\%}$  is then computed (using Newtons' method) such that:

$$SSBPR(\underline{\alpha}\bar{F}) = 0.35 SSBPR(\underline{0}) \quad (\text{A.19})$$

where  $\bar{F}$  is the vector of recent average fully-selected fishing mortalities, and  $\underline{\alpha}$  is a vector with 1 for the non-directed fisheries and a calculated constant for the directed fisheries.

#### *(2) The proxy for $B_{MSY}$*

The specification for the proxy for  $B_{MSY}$  depends on the tier in which the stock is placed. For stocks in Tier 4, the proxy for  $B_{MSY}$  is the average spawning stock biomass over a pre-specified number of years. For Tier 3, the proxy for  $B_{MSY}$  is  $0.35 SSBPR(\underline{0})$  multiplied by the mean recruitment over a pre-specified number of years. GMACS estimates annual recruitments by sex through estimating annual recruitment deviations and annual recruitment proportions by sex. Pre-specified numbers of years are needed in the control file for recruitment average and for mean recruitment sex ratio, respectively.

#### *(3) Calculating the OFL*

The OFL is the total catch (in weight) encountered by the gear that dies either due to being landed or due to being discarded when fully-selected fishing mortality is computed using the OFL control rule. The total catch

$$OFL = \sum_g \sum_t W_{y_2,t}^g \frac{S_{y_2,t,l}^{f,g} (\lambda_{y_2,t,l}^{f,g} + \Omega_{y_2,t,l}^{f,g} (1 - \lambda_{y_2,t,l}^{f,g})) S_{y_2,t,l}^{f,g} \alpha^{*,f} \bar{F}_t^{f,g}}{Z_{y_2+1,t,l}^g} N_{y_2+1,t,l}^{f,g} (1 - e^{-Z_{y_2+1,t,l}^g}) \quad (\text{A.20})$$

where  $y_2$  is the final year of the assessment,  $\alpha^{*,f}$  is the multiplier on average fully-selected fishing mortality for fleet  $f$  (1 for non-directed fisheries and a value computed from the OFL control rule for the directed fisheries),  $\bar{F}_t^{f,g}$  is recent average fully-selected fishing mortality for fleet  $f$  and

gender  $g$  during season  $t$ , and  $Z_{y_2+1,t,l}^g$  is the total mortality on animals of gender  $g$  in size-class  $l$  during season  $t$  of year  $y_2+1$ :

$$Z_{y_2+1,t,l}^g = \rho_{y_2,t}^M M_{y_2}^g \tilde{M}_l + \sum_f S_{y_2,t,l}^{f,g} (\lambda_{y_2,t,l}^{f,g} + \Omega_{y_2,t,l}^{f,g} (1 - \lambda_{y_2,t,l}^{f,g})) \alpha^{*,f} \bar{F}_t^{f,g} \quad (\text{A.21})$$

The values for entries of the vector  $\alpha^*$  for the directed fisheries are determined using the OFL control rule:

- If the projected spawning stock biomass in year  $y_2+1$  when  $\underline{\alpha}^* = \underline{\alpha}$  exceeds the proxy for  $B_{\text{MSY}}$ , then  $\alpha^{*,f} = \alpha^f$ .
- If the projected spawning stock biomass in year  $y_2+1$  when  $\underline{\alpha}^* = \underline{\alpha}$  is less than 25% of the proxy for  $B_{\text{MSY}}$ , then  $\alpha^{*,f} = 0$ .
- If the projected spawning stock biomass in year  $y_2+1$ ,  $SSB_{y_2}^*$  when  $\underline{\alpha}^* = \underline{\alpha}$  lies between less than 25% and 100% of the proxy for  $B_{\text{MSY}}$ , then  $\alpha^{*,f}$  is tuned according to

$$\alpha^{*,f} = \frac{\alpha^f \left( \frac{SSB_{y_2}^*}{B_{\text{MSY}}} - 0.1 \right)}{0.9} \text{ until convergence.}$$

### c. Projections

The specifications for the projections relate to:

- The duration of the projection.
- Whether the fully-selected fishing mortalities for the non-directed fisheries are set to zero or to recent averages by fleet.
- The way in which future recruitment is generated. The options available are:
  - Select a recruitment from a set of historical recruitments at random.
  - Generate a future recruitment from a Ricker stock-recruitment relationship, i.e.:

$$R_y^g = SSB_{y-a^*} / SSB_0 e^{-1.25 \ln h (SSB_{y-a^*} / SSB_0 - 1)} e^{\varepsilon_y - \sigma_R^2 / 2}; \quad \varepsilon_y \sim N(0; \sigma^2) \quad (\text{A.22})$$

where  $a^*$  is the time-lag between spawning and entering the first size-class in the model,  $SSB_0$  is unfished spawning stock biomass,  $h$  is the steepness of the stock-recruitment relationship,  $\sigma_R$  is the variation in recruitment about the stock-recruitment relationship.

- Generate a future recruitment from a Beverton-Holt stock-recruitment relationship, i.e.:

$$R_y^g = \frac{4R_0 SSB_{y-a^*} / SSB_0}{(1-h) + (5h-1)SSB_{y-a^*} / SSB_0} e^{\varepsilon_y - \sigma_R^2 / 2} \quad \varepsilon_y \sim N(0; \sigma^2) \quad (\text{A.23})$$

where  $R_0$  is unfished recruitment (i.e..  $SSB_0 / SSBPR(0)$ ).

- The control rule used to set fully-selected fishing mortality for the directed fisheries. The options are available

- Pre-specified values for fully-selected fishing mortality for each fishery.
- Pre-specified values subject to the dead catch not exceeding that corresponding to the OFL.
- Pre-specified values subject to the dead catch not exceeding that corresponding to the OFL and the landed catch not exceeding that corresponding to the State of Alaska harvest control rule.

The value for the steepness of the stock-recruitment relationship is computed such that the maximum sustainable yield occurs at  $F_{35\%}$ , i.e.:

$$\left. \frac{dC(\underline{F})}{dF} \right|_{F=\underline{\alpha}^* \bar{F}} \quad (\text{A.24})$$

where  $C(\underline{F})$  is the equilibrium landed catch when the population model is projected forward deterministically under one of the two stock-recruitment relationships.

### C. Parameter Estimation

#### a. *Estimating Bycatch Fishing Mortalities for Years without Observer Data*

Observer data are not available for the directed pot fishery before 1990 and the Tanner crab fishery before 1991. There are also extremely low observed bycatches in the Tanner crab fishery in 1994 and during 2006-2009. Bycatch fishing mortalities for male and females during 1975-1989 in the directed pot fishery were estimated as

$$F_t^{disc,s} = r^s F_t^{dir} \quad (\text{A.25})$$

where  $r^s$  is the mean ratio of estimated bycatch discard fishing mortalities to the estimated directed pot fishing mortalities during 1990-2004 for sex  $s$ . Directed pot fishing practice has changed after 2004 due to fishery rationalization.

We used pot fishing effort (potlifts) east of 163° W in the Tanner crab fishery to estimate red king crab bycatch discard fishing mortalities in that fishery when observer data are not available (1975-1990, 1994, 2006-2009):

$$F_t^{Tanner,s} = \alpha^s E_t \quad (\text{A.26})$$

where  $\alpha^s$  is the mean ratio of estimated Tanner crab fishery bycatch fishing mortalities to fishing efforts during 1991-1993 for sex  $s$ , and  $E_t$  is Tanner crab fishery fishing efforts east of 163° W in year  $t$ . Due to fishery rationalization after 2004, we used the data only during 1991-1993 to estimate the ratio.

#### b. Likelihood Components

A maximum likelihood approach was used to estimate parameters. For length compositions ( $p_{l,t,s,sh}$ ), the likelihood functions are :

$$Rf = \prod_{l=1}^L \prod_{t=1}^T \prod_{s=1}^2 \prod_{sh=1}^2 \frac{\left\{ \exp \left[ -\frac{(p_{l,t,s,sh} - \hat{p}_{l,t,s,sh})^2}{2\sigma_{l,t,s,sh}^2} \right] + 0.01 \right\}}{\sqrt{2\pi\sigma_{l,t,s,sh}^2}} \quad (\text{A.27})$$

$$\sigma_{l,t,s,sh}^2 = \frac{[p_{l,t,s,sh}(1-p_{l,t,s,sh}) + \frac{0.1}{L}]}{n_t}$$

where  $L$  is the number of length groups,  $T$  the number of years, and  $n_t$  the effective sample size in year  $t$ , which was estimated for trawl survey, pot retained catch, total directed pot male catch, directed pot female discard, groundfish trawl discard, groundfish fixed gear discard, and Tanner crab fishery discard length composition data.  $p_{l,t,s,sh}$  is the observed proportion of crab in length-class  $l$ , year  $t$ , sex  $s$  and shell condition  $sh$ , and  $\hat{p}_{l,t,s,sh}$  is the model-estimate corresponding to  $p_{l,t,s,sh}$ .

The weighted negative log likelihood functions are:

$$\begin{aligned} \text{Length compositions: } & - \sum \ln(Rf_i) \\ \text{Catch and bycatch biomasses: } & \sum \left[ \ln \left( \frac{C_t}{\hat{C}_t} \right)^2 / (2 \ln(cv_t^2 + 1)) \right] \\ \text{NMFS survey biomass: } & \sum \left[ \ln \left( \ln(CV_t^2 + 1) \right)^{0.5} + \frac{\ln \left( \frac{B_t}{\hat{B}_t} \right)^2}{(2 \ln(CV_t^2 + 1))} \right] \\ \text{BSFRF survey biomass: } & \sum \left[ \ln \left( \ln(CV_t^2 + AV^2 + 1) \right)^{0.5} + \frac{\ln \left( \frac{B_t}{\hat{B}_t} \right)^2}{(2 \ln(CV_t^2 + AV^2 + 1))} \right] \\ \text{R variation: } & \lambda_R \sum \left[ \ln \left( \frac{R_t}{\bar{R}} \right)^2 \right] \\ \text{R sex ratio: } & \lambda_S \sum \left[ \ln \left( \frac{\bar{R}_M}{\bar{R}_F} \right)^2 \right] \\ \text{Groundfish bycatch fishing mortalities: } & \lambda_t \sum \left[ \ln \left( \frac{F_{t,gf}}{\bar{F}_{gf}} \right)^2 \right] \\ \text{Pot female bycatch fishing mortalities: } & \lambda_p \sum \left[ \ln \left( \frac{F_{t,f}}{\bar{F}_f} \right)^2 \right] \\ \text{Trawl survey catchability: } & \frac{(Q - \hat{Q})^2}{2\sigma^2} \end{aligned} \quad (\text{A.28})$$

where  $R_t$  is the recruitment in year  $t$ ,  $\bar{R}$  the mean recruitment,  $\bar{R}_M$  the mean male recruitment,  $\bar{R}_F$  the mean female recruitment,  $AV$  is additional  $CV$  and estimated in the model,  $\bar{F}_{gf}$  the mean groundfish bycatch fishing mortality (this is separated into trawl and fixed gear fishery bycatch),  $\bar{F}_f$  the mean pot female bycatch fishing mortality,  $Q$  summer trawl survey catchability, and  $\sigma$  the estimated standard deviation of  $Q$  (all models).



Weights  $\lambda_j$  are assumed to be 2 for recruitment variation, 10 for recruitment sex ratio, 0.2 for pot female bycatch fishing mortality, and 0.1 for trawl bycatch fishing mortality. These  $\lambda_j$  values correspond to CV values of 0.53, 0.23, 3.34, and 12.14, respectively.

**c. Population State in Year 1.**

The total abundance and proportions for the first year are estimated in the model.

**d. Parameter estimation framework:**

*(1) Parameters estimated independently*

Basic natural mortality, length-weight relationships, and mean growth increments per molt were estimated independently outside of the model. Mean length of recruits to the model depends on growth and was assumed to be 72.5 for both males and females. Handling mortality rates were set to 0.2 for the directed pot fishery, 0.25 for the Tanner crab fishery, 0.5 for the groundfish fixed gear fishery, and 0.8 for the groundfish trawl fishery.

**i. Natural Mortality**

Based on an assumed maximum age of 25 years and the 1% rule (Zheng 2005), a base  $M$  was estimated to be 0.18 for males.

**ii. Length-weight Relationship**

Length-weight relationships for males and females were as follows:

$$\begin{aligned} \text{Immature Females: } & W = 0.000408 L^{3.127956} \\ \text{Ovigerous Females: } & W = 0.003593 L^{2.666076} \\ \text{Males: } & W = 0.0004031 L^{3.141334} \end{aligned} \tag{A.29}$$

where  $W$  is weight in grams, and  $L$  CL in mm.

**iii. Growth Increment per Molt**

A variety of data are available to estimate male mean growth increment per molt for Bristol Bay RKC. Tagging studies were conducted during the 1950s, 1960s and 1990s, and mean growth increment per molt data from these tagging studies in the 1950s and 1960s were analyzed by Weber and Miyahara (1962) and Balsiger (1974). Modal analyses were conducted for the data during 1957-1961 and the 1990s (Weber 1967; Loher et al. 2001). Mean growth increment per molt may be a function of body size and shell condition and vary over time (Balsiger 1974; McCaughran and Powell 1977); however, for simplicity, mean growth increment per molt was assumed to be only a function of body size in the models. Tagging data were used to estimate mean growth increment per molt as a function of pre-molt length for males (Figure A2). The results from modal analyses of 1957-1961 and the 1990s were used to estimate mean growth increment per molt for immature females during 1975-1993 and 1994-2020, respectively, and the data presented in Gray (1963) were used to estimate those for mature females for model scenarios (Figure A2). To make a

smooth transition of growth increment per molt from immature to mature females, weighted growth increment averages of 70% and 30% at 92.5 mm CL pre-molt length and 90% and 10% at 97.5 mm CL were used, respectively, for mature and immature females during 1983-1993. These percentages are roughly close to the composition of maturity. During 1975-1982, females matured at a smaller size, so the growth increment per molt as a function of length was shifted to smaller increments. Likewise, during 1994-2020, females matured at a slightly higher size, so the growth increment per molt was shifted to high increments for immature crab (Figure A2). Once mature, the growth increment per molt for male crab decreases slightly and annual molting probability decreases, whereas the growth increment for female crab decreases dramatically but annual molting probability remains constant at 1.0 (Powell 1967).

#### ***iv. Sizes at Maturity for Females***

The NMFS collected female reproductive condition data during the summer trawl surveys. Mature females are separated from immature females by a presence of egg clutches or egg cases. Proportions of mature females at 5-mm length intervals were summarized and a logistic curve was fitted to the data each year to estimate sizes at 50% maturity. Sizes at 50% maturity are illustrated in Figure A3 with mean values for three different periods (1975-82, 1983-93, and 1994-2020).

#### ***v. Sizes at Maturity for Males***

Although size at sexual maturity for Bristol Bay red king crab males has been estimated (Paul et al. 1991), there are no data for estimating size of functional maturity collected in the natural environment. Sizes at functional maturity for Bristol Bay male RKC have been assumed to be 120 mm CL (Schmidt and Pengilly 1990). This is based on mating pair data collected off Kodiak Island (Figure A4). Sizes at maturity for Bristol Bay female RKC are about 90 mm CL, about 15 mm CL less than Kodiak female RKC (Pengilly et al. 2002). The size ratio of mature males to females is 1.3333 at sizes at maturity for Bristol Bay RKC, and since mature males grow at much larger increments than mature females, the mean size ratio of mature males to females is most likely larger than this ratio. Size ratios of the large majority of Kodiak mating pairs were less than 1.3333, and in some bays, only a small proportion of mating pairs had size ratios above 1.3333 (Figure A4).

In the laboratory, male RKC as small as 80 mm CL from Kodiak and Southeast Alaska can successfully mate with females (Paul and Paul 1990). But few males less than 100 mm CL were observed to mate with females in the wild. Based on the size ratios of males to females in the Kodiak mating pair data, setting 120 mm CL as a minimum size of functional maturity for Bristol Bay male RKC is proper in terms of managing the fishery.

#### ***vi. Potential Reasons for High Mortality during the Early 1980s***

Bristol Bay red king crab abundance had declined sharply during the early 1980s. Many factors have been speculated for this decline: (i) completely wiped out by fishing: the directed pot fishery, the other directed pot fishery (Tanner crab fishery), and bottom trawling; and (ii) high fishing and natural mortality. With the survey abundance, harvest

rates in 1980 and 1981 were among the highest, thus the directed fishing definitely had a big impact on the stock decline, especially legal and mature males. However, for the sharp decline during 1980-1984 for males, 3 out of 5 years had low mature harvest rates. During the 1981-1984 decline for females, 3 out of 4 years had low mature harvest rates. Also pot catchability for females and immature males are generally much lower than for legal males, so the directed pot fishing alone cannot explain the sharp decline for all segments of the stock during the early 1980s.

Red king crab bycatch in the eastern Bering Sea Tanner crab fishery is another potential factor (Griffin et al. 1983). The main overlap between Tanner crab and Bristol Bay red king crab is east of 163° W. No absolute red king crab bycatch estimates are available until 1991. So there are insufficient data to fully evaluate the impact. Retained catch and potlifts from the eastern Bering Sea Tanner crab fishery are illustrated in Figure A5. The observed red king crab bycatch in the Tanner crab fishery during 1991-2015 and total potlifts east of 166° W during 1975 to 2015 were used to estimate the bycatch mortality in the current model. Because winter sea surface temperatures and air temperatures were warmer (which means a lower handling mortality rate) and there were fewer potlifts during the early 1980s than during the early 1990s, bycatch in the Tanner crab fishery is unlikely to have been a main factor for the sharp decline of Bristol Bay red king crab.

Several factors may have caused increases in natural mortality. Crab diseases in the early 1980s were documented by Sparks and Morado (1985), but inadequate data were collected to examine their effects on the stock. Stevens (1990) speculated that senescence may be a factor because many crab in the early 1980s were very old due to low temperatures in the 1960s and early 1970s. The biomass of the main crab predator, Pacific cod, increased about 10 times during the late 1970s and early 1980s. Yellowfin sole biomass also increased substantially during this period. Predation is primarily on juvenile and molting/softshell crab. But we lack stomach samples in shallow waters (juvenile habitat) and during the period when red king crab molt. Also cannibalism occurs during molting periods for red king crab. High crab abundance in the late 1970s and early 1980s may have increased the occurrence of cannibalism.

Overall, the likely causes for the sharp decline in the early 1980s are combinations of the above factors, such as pot fisheries on legal males, bycatch, and predation on females and juvenile and sublegal males, senescence for older crab, and disease for all crab. In our model, we estimated one mortality parameter for males and another for females during 1980-1984. We also estimated a mortality parameter for females during 1976-1979 and 1985-1993. These three mortality parameters are additional to the basic natural mortality of  $0.18\text{yr}^{-1}$ , all directed fishing mortality, and non-directed fishing mortality. These three mortality parameters could be attributed to natural mortality as well as undocumented non-directed fishing mortality. The model fit the data much better with these three parameters than without them.

## (2) Parameters estimated conditionally

The following model parameters were estimated for male and female crab: total recruits

for each year (year class strength  $R_t$  for  $t = 1976$  to  $2020$ ), total abundance in the first year (1975), growth parameter  $\beta$ , and recruitment parameter  $\beta_r$  for males and females separately. Molting probability parameters  $\beta$  and  $L_{50}$  were also estimated for male crab. Estimated parameters also include different sets of  $\beta$  and  $L_{50}$  for total selectivity and retained proportions,  $\beta$  and  $L_{50}$  for pot-discarded female selectivity,  $\beta$  and  $L_{50}$  for pot-discarded male and female selectivities from the eastern Bering Sea Tanner crab fishery,  $\beta$  and  $L_{50}$  for groundfish trawl and fixed gear discarded selectivities, and different sets of  $\beta$  and  $L_{50}$  for NMFS trawl survey male and female selectivities separately. The NMFS survey catchabilities  $Q$  for some models were also estimated. Different sets of  $\beta$  and  $L_{50}$  for selectivity parameters were estimated for the survey data from the Bering Fisheries Research Foundation. Annual fishing mortalities were also estimated for the directed pot fishery for males (1975-2019), pot-discarded females from the directed fishery (1990-2019), pot-discarded males and females from the eastern Bering Sea Tanner crab fishery (1991-93, 2013-15), groundfish trawl discarded males and females (1976-2019), and groundfish fixed gear discarded males and females (1996-2019). One additional mortality parameter for years 1980-1984 for males and a constant to multiply male natural mortality for estimating female natural mortality were also estimated. Some estimated parameters were constrained in the model. For example, male and female recruitment estimates were forced to be close to each other for a given year.

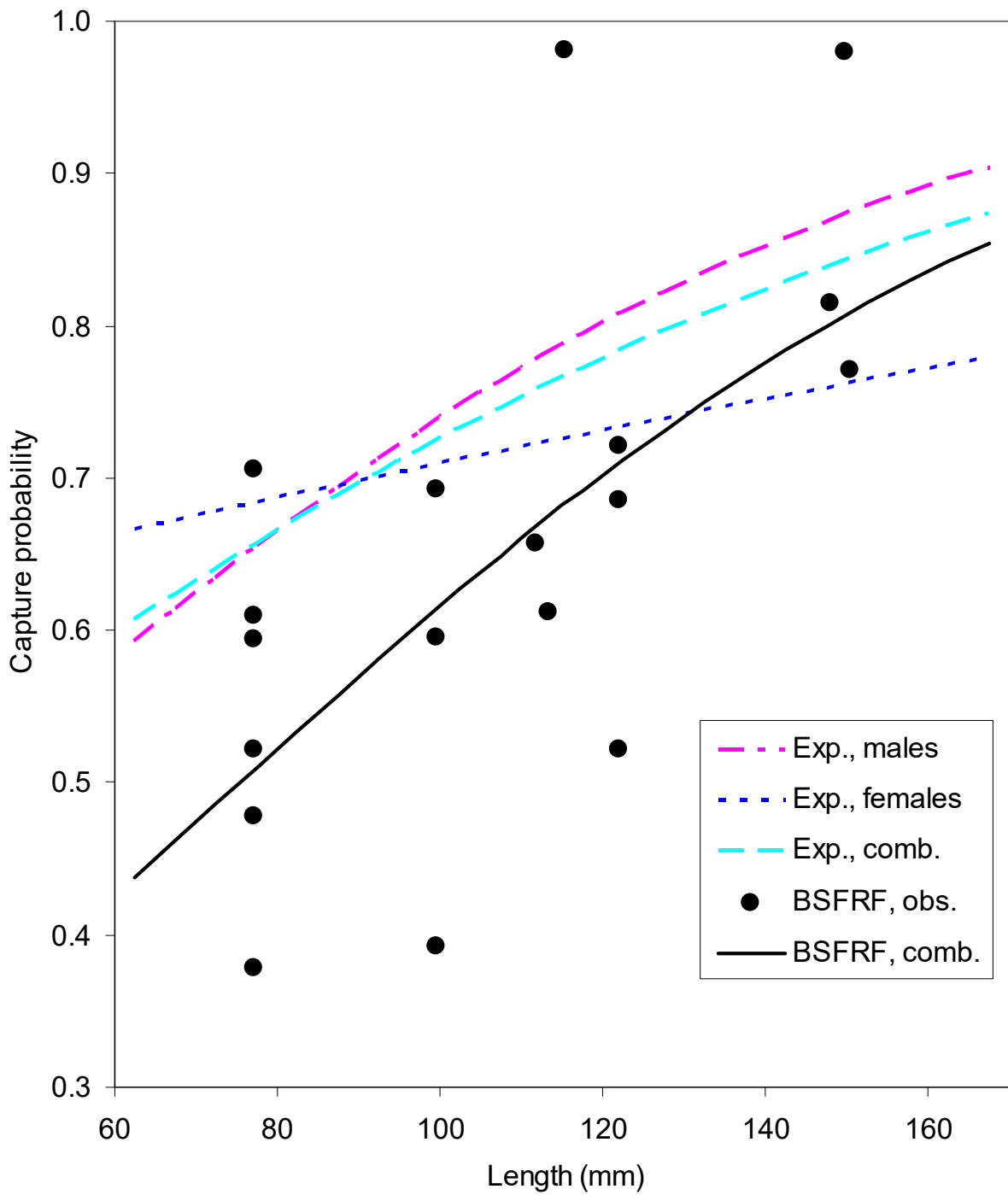


Figure A1. Estimated capture probabilities for NMFS Bristol Bay red king crab trawl surveys by Weinberg et al. (2004) and the Bering Sea Fisheries Research Foundation surveys.

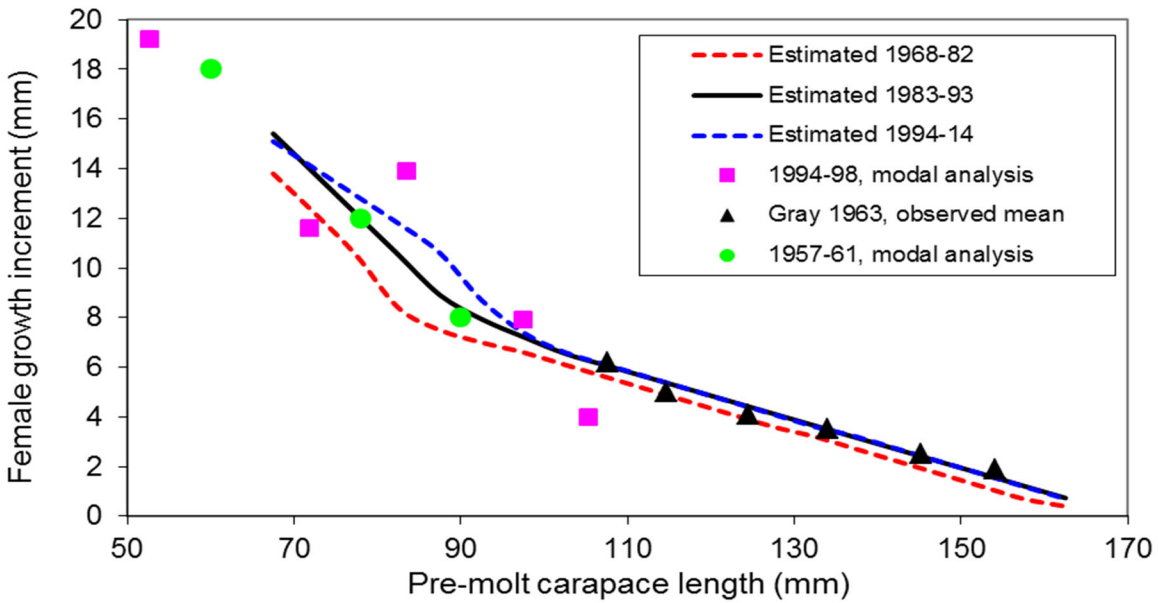
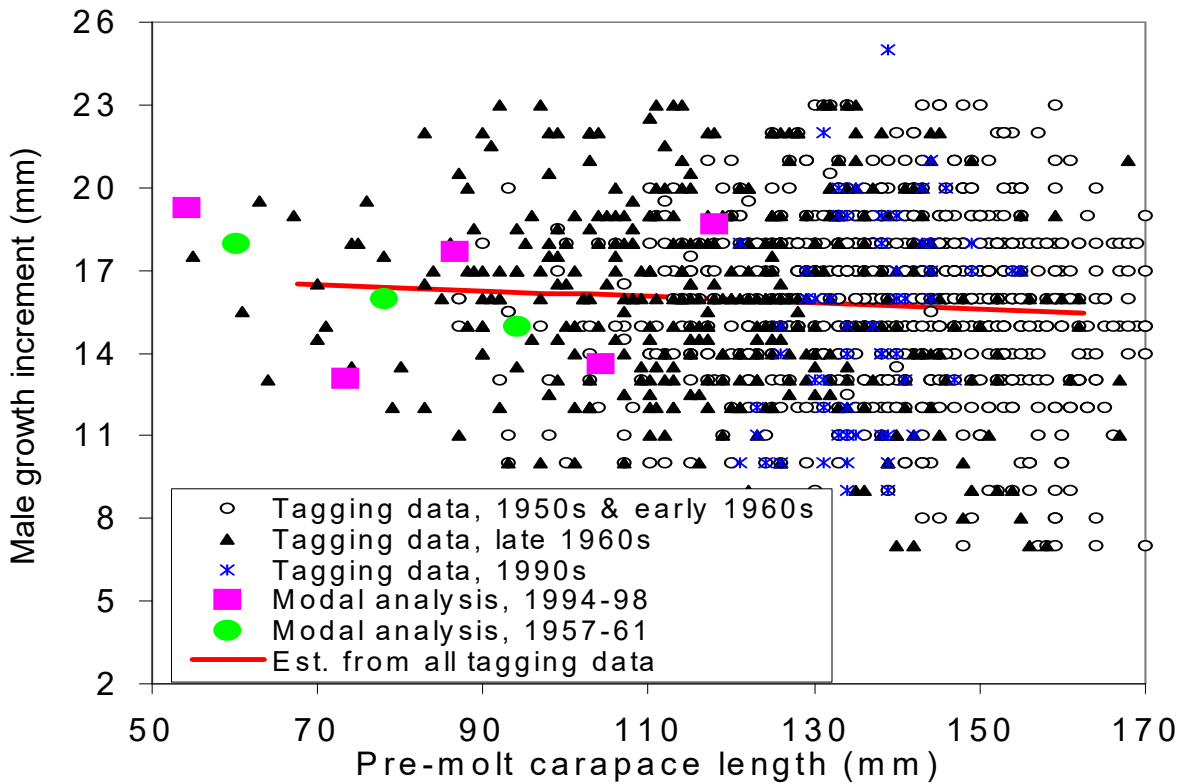


Figure A2. Mean growth increments per molt for Bristol Bay red king crab. Note: “tagging”---based on tagging data; “mode”---based on modal analysis. The female growth increments per molt are for different model scenarios.

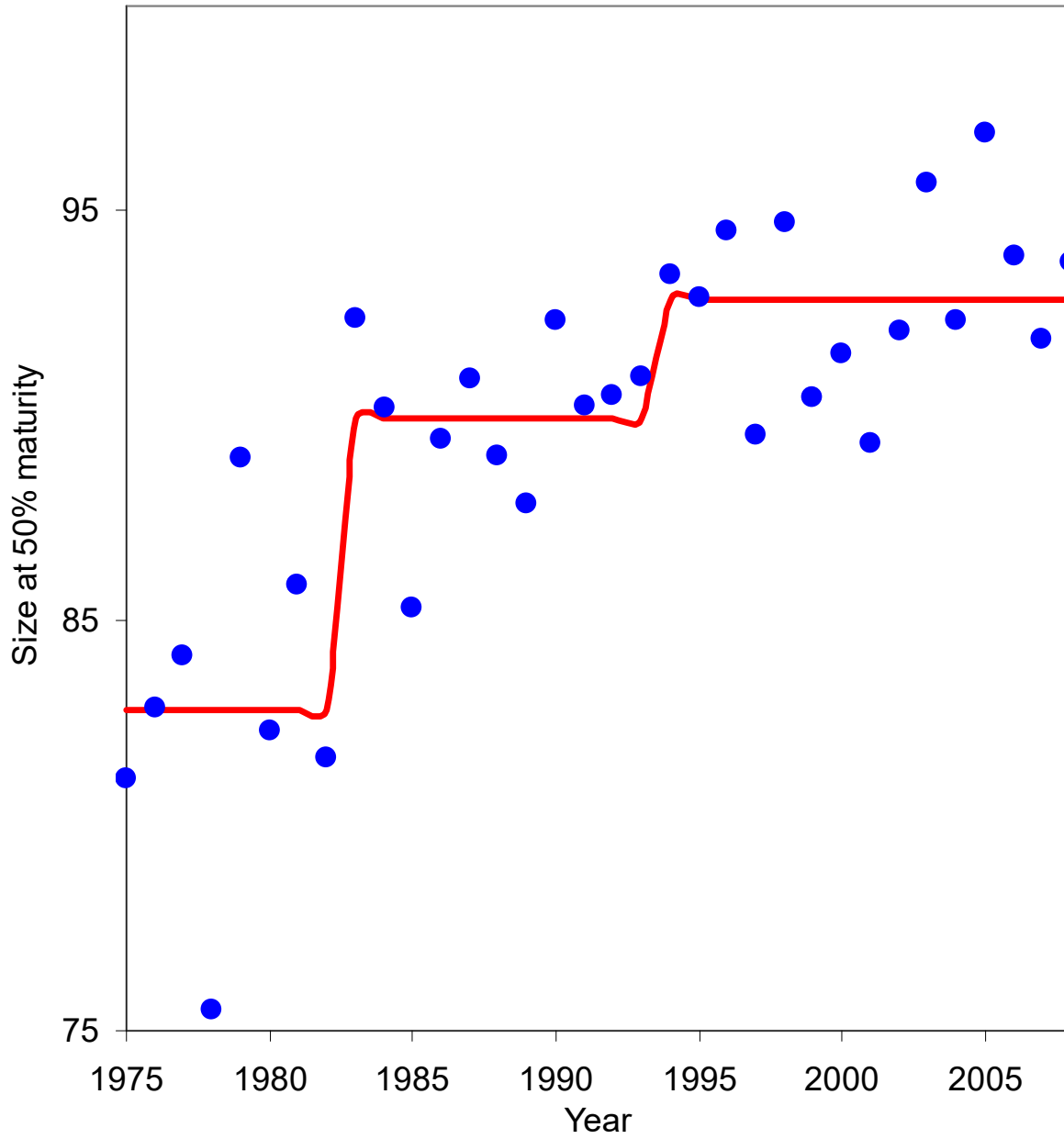


Figure A3. Estimated sizes at 50% maturity for Bristol Bay female red king crab from 1975 to 2008. Averages for three periods (1975-82, 1983-93, and 1994-08) are plotted with a line.

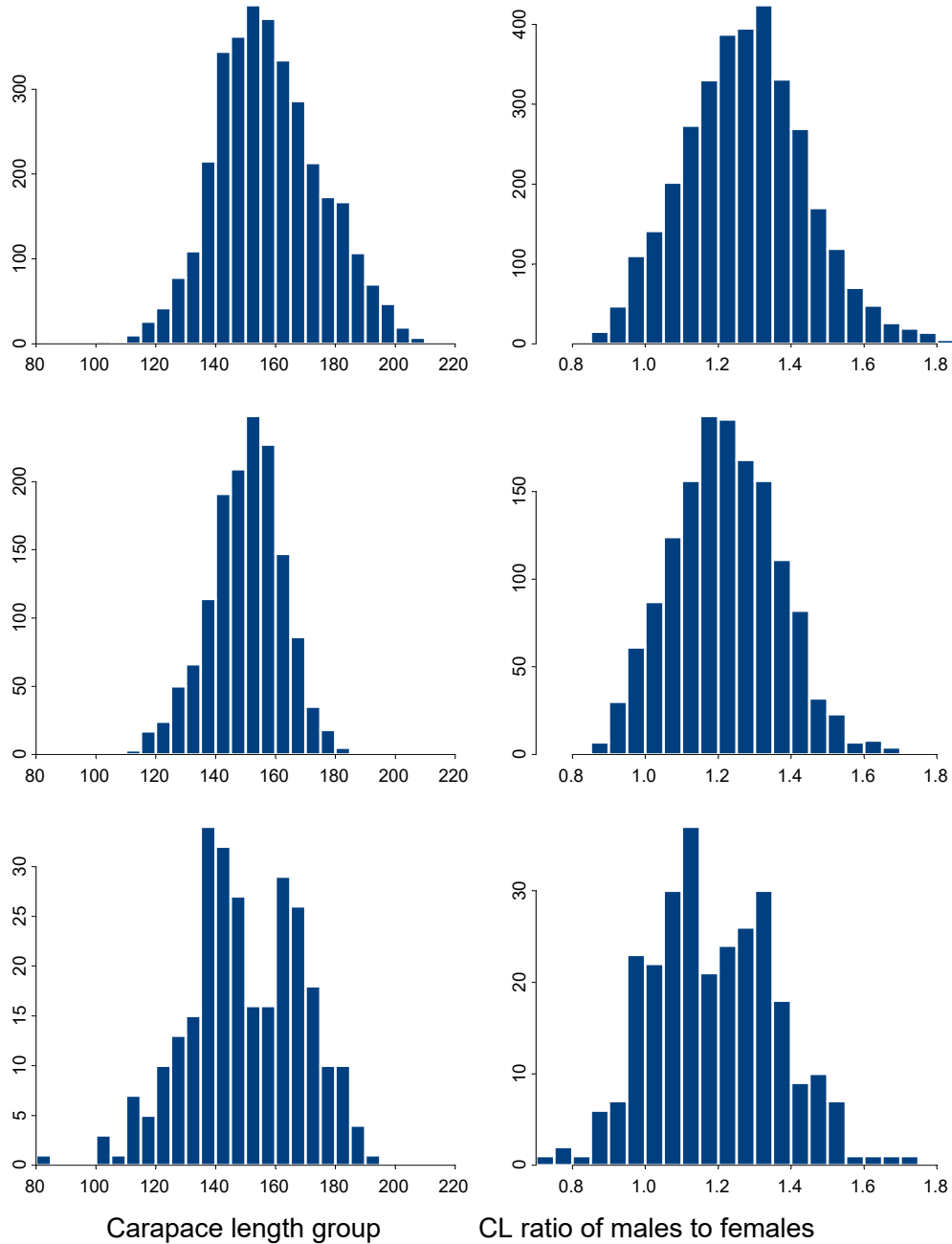


Figure A4. Histograms of carapace lengths (CL) and CL ratios of males to females for male shell ages  $\leq 13$  months of red king crab males in grasping pairs; Powell's Kodiak data. Upper plot: all locations and years pooled; middle plot: location 11; lower plot: locations 4 and 13. Sizes at maturity for Kodiak red king crab are about 15 mm larger than those for Bristol Bay red king crab. (Doug Pengilly, ADF&G, pers. comm.).



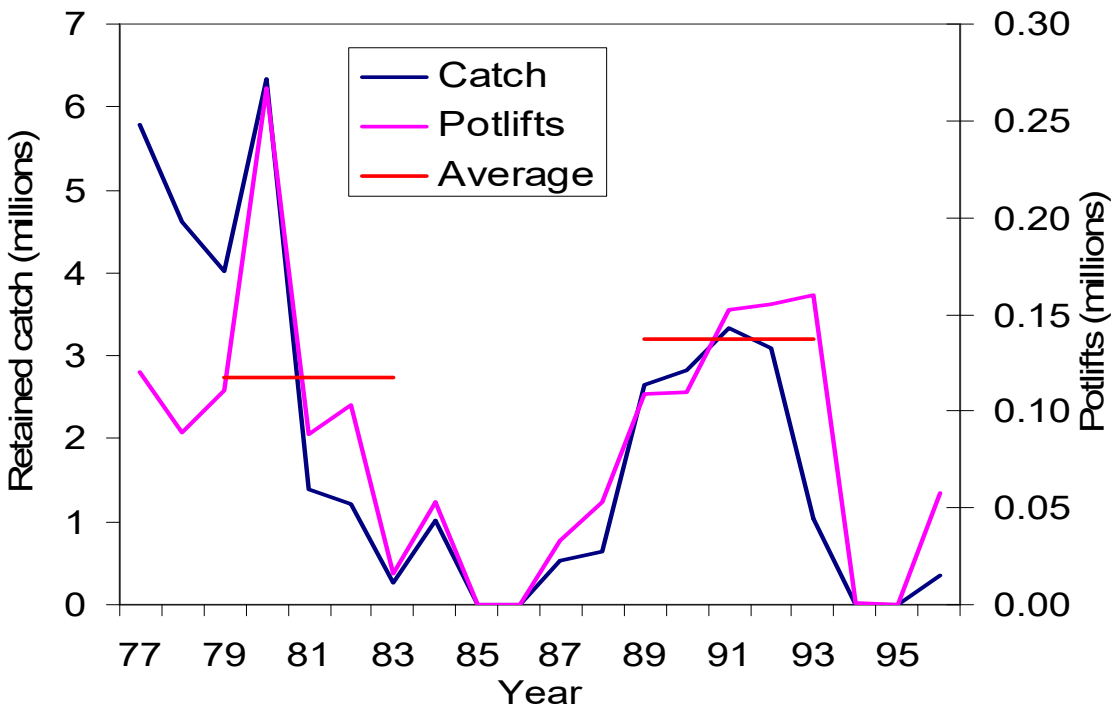
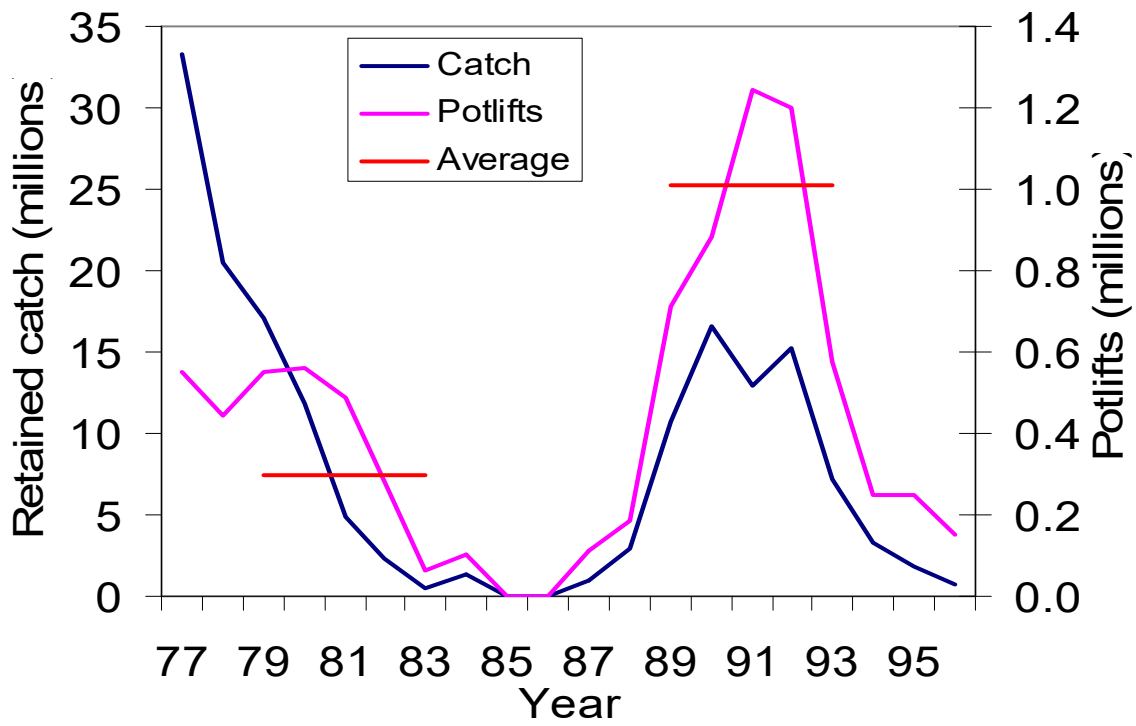


Figure A5. Retained catch and potlifts for total eastern Bering Sea Tanner crab fishery (upper plot) and the Tanner crab fishery east of 163° W (bottom).

**Appendix B. Input Data File for Model 19.3**

```

=====
#
#   Gmacs Main Data File Version      1.1:  BBRKC      Example
#   GEAR_INDEX      DESCRIPTION
#   1      :      Pot fishery retained catch.
#   1      :      Pot fishery with discarded catch.
#   2      :      Trawl bycatch
#   3      :      Trawl survey
#   Fisheries: 1 Pot Fishery, 2 Pot Discard, 3 Trawl
#             by-catch, 4 Tanner bycatch 5 fixed gear
#   Surveys: 6 NMFS Trawl Survey,7 BSFRFSurvey
=====
1975 # Start year
2019 # End year
7 # Number of seasons
6 # Number of fleets (fishing fleets and surveys)
2 # Number of sexes
2 # Number of shell condition types
1 # Number of maturity types
20 # Number of size-classes in the model
7 # Season recruitment occurs
7 # Season molting and growth occurs
6 # Season to calculate SSB
1 # Season for N output
# maximum size-class (males then females)
20 16
# size_breaks (a vector giving the break points between size intervals,
# dim=nclass+1)
65 70 75 80 85 90 95 100 105 110 115 120 125
# 130 135 140 145 150 155 160 165
# Natural mortality per season input type (1 = vector by season,
# 2 = matrix by season/year)
2
# Proportion of the total natural mortality to be applied each season
0.0000 0.2329 0.0000 0.2671 0.000 0.194 0.306 #1975
0.0000 0.2795 0.0000 0.2205 0.000 0.194 0.306 #1976
0.0000 0.3233 0.0000 0.1767 0.000 0.194 0.306 #1977
0.0000 0.2548 0.0000 0.2452 0.000 0.194 0.306 #1978
0.0000 0.2493 0.0000 0.2507 0.000 0.194 0.306 #1979
0.0000 0.2493 0.0000 0.2507 0.000 0.194 0.306 #1980
0.0000 0.2493 0.0000 0.2507 0.000 0.194 0.306 #1981

```

0.0000	0.2356	0.0000	0.2644	0.000	0.194	0.306	#1982
0.0000	0.2400	0.0000	0.2600	0.000	0.194	0.306	#1983
0.0000	0.2712	0.0000	0.2288	0.000	0.194	0.306	#1984
0.0000	0.2438	0.0000	0.2562	0.000	0.194	0.306	#1985
0.0000	0.2521	0.0000	0.2479	0.000	0.194	0.306	#1986
0.0000	0.2493	0.0000	0.2507	0.000	0.194	0.306	#1987
0.0000	0.2438	0.0000	0.2562	0.000	0.194	0.306	#1988
0.0000	0.2493	0.0000	0.2507	0.000	0.194	0.306	#1989
0.0000	0.3507	0.0000	0.1493	0.000	0.194	0.306	#1990
0.0000	0.3425	0.0000	0.1575	0.000	0.194	0.306	#1991
0.0000	0.3425	0.0000	0.1575	0.000	0.194	0.306	#1992
0.0000	0.3452	0.0000	0.1548	0.000	0.194	0.306	#1993
0.0000	0.3400	0.0000	0.1600	0.000	0.194	0.306	#1994
0.0000	0.3400	0.0000	0.1600	0.000	0.194	0.306	#1995
0.0000	0.3400	0.0000	0.1600	0.000	0.194	0.306	#1996
0.0000	0.3400	0.0000	0.1600	0.000	0.194	0.306	#1997
0.0000	0.3400	0.0000	0.1600	0.000	0.194	0.306	#1998
0.0000	0.3000	0.0000	0.2000	0.000	0.194	0.306	#1999
0.0000	0.3000	0.0000	0.2000	0.000	0.194	0.306	#2000
0.0000	0.3000	0.0000	0.2000	0.000	0.194	0.306	#2001
0.0000	0.3000	0.0000	0.2000	0.000	0.194	0.306	#2002
0.0000	0.3000	0.0000	0.2000	0.000	0.194	0.306	#2003
0.0000	0.3000	0.0000	0.2000	0.000	0.194	0.306	#2004
0.0000	0.3000	0.0000	0.2000	0.000	0.194	0.306	#2005
0.0000	0.3000	0.0000	0.2000	0.000	0.194	0.306	#2006
0.0000	0.3000	0.0000	0.2000	0.000	0.194	0.306	#2007
0.0000	0.3000	0.0000	0.2000	0.000	0.194	0.306	#2008
0.0000	0.3000	0.0000	0.2000	0.000	0.194	0.306	#2009
0.0000	0.3000	0.0000	0.2000	0.000	0.194	0.306	#2010
0.0000	0.3000	0.0000	0.2000	0.000	0.194	0.306	#2011
0.0000	0.3000	0.0000	0.2000	0.000	0.194	0.306	#2012
0.0000	0.3000	0.0000	0.2000	0.000	0.194	0.306	#2013
0.0000	0.3000	0.0000	0.2000	0.000	0.194	0.306	#2014
0.0000	0.3000	0.0000	0.2000	0.000	0.194	0.306	#2015
0.0000	0.3000	0.0000	0.2000	0.000	0.194	0.306	#2016
0.0000	0.3000	0.0000	0.2000	0.000	0.194	0.306	#2017
0.0000	0.3000	0.0000	0.2000	0.000	0.194	0.306	#2018
0.0000	0.3000	0.0000	0.2000	0.000	0.194	0.306	#2019

# Fishing fleet names (delimited with: no spaces in names)  
 Pot\_Fishery Trawl\_Bycatch Bairdi\_Fishery\_Bycatch Fixed\_Gear  
 # Survey names (delimited with: no spaces in names)  
 NMFS\_Trawl BSFRF

```

# Are the seasons instantaneous (0) or continuous (1)
1 1 1 1 1 1
# Number of catch data frames
7
# Number of rows in each data frame
45 30 30 44 25 25 24
##
## CATCH DATA
## Type of catch: 1 = retained, 2 = discard, 0 = total
## Units of catch: 1 = biomass, 2 = numbers
## for BBRKC Units are in 1000 mt for landed & discards.
##
## Male retained pot fishery (tonnes)
#year seas fleet sex obs cv type units mult effort discard_mortality
1975 3 1 1 23281.2 0.03 1 1 1 0 0.2
1976 3 1 1 28993.6 0.03 1 1 1 0 0.2
1977 3 1 1 31736.9 0.03 1 1 1 0 0.2
1978 3 1 1 39743 0.03 1 1 1 0 0.2
1979 3 1 1 48910 0.03 1 1 1 0 0.2
1980 3 1 1 58943.6 0.03 1 1 1 0 0.2
1981 3 1 1 15236.8 0.03 1 1 1 0 0.2
1982 3 1 1 1361.3 0.03 1 1 1 0 0.2
1983 3 1 1 0.1 0.03 1 1 1 0 0.2 #AEP
1984 3 1 1 1897.1 0.03 1 1 1 0 0.2
1985 3 1 1 1893.8 0.03 1 1 1 0 0.2
1986 3 1 1 5168.2 0.03 1 1 1 0 0.2
1987 3 1 1 5574.2 0.03 1 1 1 0 0.2
1988 3 1 1 3351.1 0.03 1 1 1 0 0.2
1989 3 1 1 4656 0.03 1 1 1 0 0.2
1990 3 1 1 9272.8 0.03 1 1 1 0 0.2
1991 3 1 1 7885.1 0.03 1 1 1 0 0.2
1992 3 1 1 3681.8 0.03 1 1 1 0 0.2
1993 3 1 1 6659.6 0.03 1 1 1 0 0.2
1994 3 1 1 42.3 0.03 1 1 1 0 0.2
1995 3 1 1 36.4 0.03 1 1 1 0 0.2
1996 3 1 1 3861.7 0.03 1 1 1 0 0.2
1997 3 1 1 4042.1 0.03 1 1 1 0 0.2
1998 3 1 1 6779.2 0.03 1 1 1 0 0.2
1999 3 1 1 5377.9 0.03 1 1 1 0 0.2
2000 3 1 1 3737.9 0.03 1 1 1 0 0.2
2001 3 1 1 3866.2 0.03 1 1 1 0 0.2
2002 3 1 1 4384.5 0.03 1 1 1 0 0.2
2003 3 1 1 7135.3 0.03 1 1 1 0 0.2

```

2004	3	1	1	7006.7	0.03	1	1	1	0	0.2
2005	3	1	1	8399.7	0.03	1	1	1	0	0.2
2006	3	1	1	7143.2	0.03	1	1	1	0	0.2
2007	3	1	1	9303.9	0.03	1	1	1	0	0.2
2008	3	1	1	9216.1	0.03	1	1	1	0	0.2
2009	3	1	1	7272.5	0.03	1	1	1	0	0.2
2010	3	1	1	6761.5	0.03	1	1	1	0	0.2
2011	3	1	1	3607.1	0.03	1	1	1	0	0.2
2012	3	1	1	3621.7	0.03	1	1	1	0	0.2
2013	3	1	1	3991	0.03	1	1	1	0	0.2
2014	3	1	1	4538.6	0.03	1	1	1	0	0.2
2015	3	1	1	4613.7	0.03	1	1	1	0	0.2
2016	3	1	1	3923.9	0.03	1	1	1	0	0.2
2017	3	1	1	3093.7	0.03	1	1	1	0	0.2
2018	3	1	1	2026.5	0.03	1	1	1	0	0.2
2019	3	1	1	1775.3	0.03	1	1	1	0	0.2

##	Total Male		pot	fishery (t)							
#year	seas	fleet	sex	obs	cv	type	units	mult	effort	discard	mortality
1990	3	1	1	11782.9		0.04	0	1	1	0	0.2
1991	3	1	1	9974	0.04	0	1	1	0	0.2	
1992	3	1	1	6013.7	0.04	0	1	1	0	0.2	
1993	3	1	1	9667.7	0.04	0	1	1	0	0.2	
1994	3	1	1	62.3	0.04	0	1	1	0	0.2	
1995	3	1	1	52.8	0.04	0	1	1	0	0.2	
1996	3	1	1	3902.3	0.04	0	1	1	0	0.2	
1997	3	1	1	3847.2	0.04	0	1	1	0	0.2	
1998	3	1	1	17681.4		0.04	0	1	1	0	0.2
1999	3	1	1	12245.2		0.04	0	1	1	0	0.2
2000	3	1	1	6672.3	0.04	0	1	1	0	0.2	
2001	3	1	1	5797	0.04	0	1	1	0	0.2	
2002	3	1	1	7065.3	0.04	0	1	1	0	0.2	
2003	3	1	1	12300.6		0.04	0	1	1	0	0.2
2004	3	1	1	10816.8		0.04	0	1	1	0	0.2
2005	3	1	1	13753.3		0.04	0	1	1	0	0.2
2006	3	1	1	9170.4	0.04	0	1	1	0	0.2	
2007	3	1	1	13956.6		0.04	0	1	1	0	0.2
2008	3	1	1	15068.7		0.04	0	1	1	0	0.2
2009	3	1	1	12300.3		0.04	0	1	1	0	0.2
2010	3	1	1	10087.4		0.04	0	1	1	0	0.2
2011	3	1	1	5732.6	0.04	0	1	1	0	0.2	
2012	3	1	1	4568.1	0.04	0	1	1	0	0.2	
2013	3	1	1	5260.7	0.04	0	1	1	0	0.2	

2014	3	1	1	8312.7	0.04	0	1	1	0	0.2	
2015	3	1	1	6706.4	0.04	0	1	1	0	0.2	
2016	3	1	1	5557.2	0.04	0	1	1	0	0.2	
2017	3	1	1	4075.76		0.04	0	1	1	0	0.2
2018	3	1	1	3060.34		0.04	0	1	1	0	0.2
2019	3	1	1	3143.25	0.04	0	1	1	0	0.2	
##	Female discards			Pot	fishery						

#year	seas	fleet	sex	obs	cv	type	units	mult	effort	discard	mortality
1990	3	1	2	3240.20	0.07	0	1	1	0	0.2	
1991	3	1	2	236.600		0.07	0	1	1	0	0.2
1992	3	1	2	2001.20	0.07	0	1	1	0	0.2	
1993	3	1	2	3174.40	0.07	0	1	1	0	0.2	
1994	3	1	2	1.877	0.07	0	1	1	0	0.2	
1995	3	1	2	1.612	0.07	0	1	1	0	0.2	
1996	3	1	2	5.200	0.07	0	1	1	0	0.2	
1997	3	1	2	184.800		0.07	0	1	1	0	0.2
1998	3	1	2	2897.10	0.07	0	1	1	0	0.2	
1999	3	1	2	28.200	0.07	0	1	1	0	0.2	
2000	3	1	2	833.700		0.07	0	1	1	0	0.2
2001	3	1	2	611.400		0.07	0	1	1	0	0.2
2002	3	1	2	46.100	0.07	0	1	1	0	0.2	
2003	3	1	2	1804.70	0.07	0	1	1	0	0.2	
2004	3	1	2	873.000		0.07	0	1	1	0	0.2
2005	3	1	2	2051.40	0.07	0	1	1	0	0.2	
2006	3	1	2	187.700		0.07	0	1	1	0	0.2
2007	3	1	2	816.700		0.07	0	1	1	0	0.2
2008	3	1	2	734.400		0.07	0	1	1	0	0.2
2009	3	1	2	468.500		0.07	0	1	1	0	0.2
2010	3	1	2	609.200		0.07	0	1	1	0	0.2
2011	3	1	2	123.400		0.07	0	1	1	0	0.2
2012	3	1	2	59.800	0.07	0	1	1	0	0.2	
2013	3	1	2	514.300		0.07	0	1	1	0	0.2
2014	3	1	2	362.200		0.07	0	1	1	0	0.2
2015	3	1	2	1081.60	0.07	0	1	1	0	0.2	
2016	3	1	2	527.000		0.07	0	1	1	0	0.2
2017	3	1	2	266.546		0.07	0	1	1	0	0.2
2018	3	1	2	574.047		0.07	0	1	1	0	0.2
2019	3	1	2	216.739	0.07	0	1	1	0	0.2	
##	Trawl fishery discards (t, without applying to handling mortality rate)										
#year	seas	fleet	sex	obs	cv	type	units	mult	effort	discard	mortality
1976	5	2	0	853.494		0.10	2	1	1	0	0.8
1977	5	2	0	1562.313		0.10	2	1	1	0	0.8

1978	5	2	0	1650.775	0.10	2	1	1	0	0.8
1979	5	2	0	1664.925	0.10	2	1	1	0	0.8
1980	5	2	0	1295.625	0.10	2	1	1	0	0.8
1981	5	2	0	274.229	0.10	2	1	1	0	0.8
1982	5	2	0	718.610	0.10	2	1	1	0	0.8
1983	5	2	0	525.554	0.10	2	1	1	0	0.8
1984	5	2	0	1367.550	0.10	2	1	1	0	0.8
1985	5	2	0	487.576	0.10	2	1	1	0	0.8
1986	5	2	0	250.758	0.10	2	1	1	0	0.8
1987	5	2	0	233.045	0.10	2	1	1	0	0.8
1988	5	2	0	747.996	0.10	2	1	1	0	0.8
1989	5	2	0	219.023	0.10	2	1	1	0	0.8
1990	5	2	0	324.883	0.10	2	1	1	0	0.8
1991	5	2	0	436.783	0.10	2	1	1	0	0.8
1992	5	2	0	366.816	0.10	2	1	1	0	0.8
1993	5	2	0	501.770	0.10	2	1	1	0	0.8
1994	5	2	0	109.129	0.10	2	1	1	0	0.8
1995	5	2	0	102.623	0.10	2	1	1	0	0.8
1996	5	2	0	113.495	0.10	2	1	1	0	0.8
1997	5	2	0	71.862	0.10	2	1	1	0	0.8
1998	5	2	0	232.580	0.10	2	1	1	0	0.8
1999	5	2	0	188.101	0.10	2	1	1	0	0.8
2000	5	2	0	102.161	0.10	2	1	1	0	0.8
2001	5	2	0	241.011	0.10	2	1	1	0	0.8
2002	5	2	0	189.018	0.10	2	1	1	0	0.8
2003	5	2	0	171.114	0.10	2	1	1	0	0.8
2004	5	2	0	216.889	0.10	2	1	1	0	0.8
2005	5	2	0	155.924	0.10	2	1	1	0	0.8
2006	5	2	0	189.660	0.10	2	1	1	0	0.8
2007	5	2	0	192.571	0.10	2	1	1	0	0.8
2008	5	2	0	170.561	0.10	2	1	1	0	0.8
2009	5	2	0	118.906	0.10	2	1	1	0	0.8
2010	5	2	0	104.086	0.10	2	1	1	0	0.8
2011	5	2	0	70.419	0.10	2	1	1	0	0.8
2012	5	2	0	42.786	0.10	2	1	1	0	0.8
2013	5	2	0	83.868	0.10	2	1	1	0	0.8
2014	5	2	0	43.460	0.10	2	1	1	0	0.8
2015	5	2	0	56.686	0.10	2	1	1	0	0.8
2016	5	2	0	84.127	0.10	2	1	1	0	0.8
2017	5	2	0	114.784	0.10	2	1	1	0	0.8
2018	5	2	0	97.891	0.10	2	1	1	0	0.8
2019	5	2	0	101.001	0.10	2	1	1	0	0.8

# Tanner crab fishery discards males

#year	seas	fleet	sex	obs	cv	type	units	mult	potlifts	discard	mortality
1975	5	3	1	0	0.07	2	1	1	20	0.25	
1976	5	3	1	0	0.07	2	1	1	20	0.25	
1977	5	3	1	0	0.07	2	1	1	120.031		0.25
1978	5	3	1	0	0.07	2	1	1	88.489	0.25	
1979	5	3	1	0	0.07	2	1	1	110.989		0.25
1980	5	3	1	0	0.07	2	1	1	267.154		0.25
1981	5	3	1	0	0.07	2	1	1	87.951	0.25	
1982	5	3	1	0	0.07	2	1	1	102.987		0.25
1983	5	3	1	0	0.07	2	1	1	16.239	0.25	
1984	5	3	1	0	0.07	2	1	1	52.598	0.25	
#1985	5	3	1	0	0.07	2	1	1	0.0001	0.25	
#1986	5	3	1	0	0.07	2	1	1	0.0001	0.25	
1987	5	3	1	0	0.07	2	1	1	32.75	0.25	
1988	5	3	1	0	0.07	2	1	1	53.203	0.25	
1989	5	3	1	0	0.07	2	1	1	108.519		0.25
1990	5	3	1	0	0.07	2	1	1	109.371		0.25
1991	5	3	1	1890.9	0.07	2	1	1	152.541		0.25
1992	5	3	1	269.526		0.07	2	1	1	154.976	0.25
1993	5	3	1	117.643		0.07	2	1	1	159.922	0.25
1994	5	3	1	0	0.07	2	1	1	1.042	0.25	
#1995	5	3	1	0	0.07	2	1	1	0.0001	0.25	
#1996	5	3	1	0	0.07	2	1	1	0.0001	0.25	
#1997	5	3	1	0	0.07	2	1	1	0.0001	0.25	
#1998	5	3	1	0	0.07	2	1	1	0.0001	0.25	
#1999	5	3	1	0	0.07	2	1	1	0.0001	0.25	
#2000	5	3	1	0	0.07	2	1	1	0.0001	0.25	
#2001	5	3	1	0	0.07	2	1	1	0.0001	0.25	
#2002	5	3	1	0	0.07	2	1	1	0.0001	0.25	
#2003	5	3	1	0	0.07	2	1	1	0.0001	0.25	
#2004	5	3	1	0	0.07	2	1	1	0.0001	0.25	
#2005	5	3	1	0	0.07	2	1	1	0.0001	0.25	
2006	5	3	1	0	0.07	2	1	1	0.4	0.25	
2007	5	3	1	0	0.07	2	1	1	0.5	0.25	
2008	5	3	1	0	0.07	2	1	1	0.5	0.25	
2009	5	3	1	0	0.07	2	1	1	0.2	0.25	
#2010	5	3	1	0	0.07	2	1	1	0.0001	0.25	
#2011	5	3	1	0	0.07	2	1	1	0.0001	0.25	
#2012	5	3	1	0	0.07	2	1	1	0.0001	0.25	
2013	5	3	1	37.4687		0.07	2	1	1	2	0.25
2014	5	3	1	83.5014		0.07	2	1	1	2	0.25
2015	5	3	1	116.404		0.07	2	1	1	139.171	0.25



#	Tanner crab	fishery	discards	sex	obs	cv	type	units	mult	potlifts	discard	mortality
#2016	5	3	1	0	0.07	2	1	1	0.0001	0.25		
#2017	5	3	1	0	0.07	2	1	1	0.0001	0.25		
#1975	5	3	2	0	0.07	2	1	1	20	0.25		
#1976	5	3	2	0	0.07	2	1	1	20	0.25		
#1977	5	3	2	0	0.07	2	1	1	120.031		0.25	
#1978	5	3	2	0	0.07	2	1	1	88.489	0.25		
#1979	5	3	2	0	0.07	2	1	1	110.989		0.25	
#1980	5	3	2	0	0.07	2	1	1	267.154		0.25	
#1981	5	3	2	0	0.07	2	1	1	87.951	0.25		
#1982	5	3	2	0	0.07	2	1	1	102.987		0.25	
#1983	5	3	2	0	0.07	2	1	1	16.239	0.25		
#1984	5	3	2	0	0.07	2	1	1	52.598	0.25		
#1985	5	3	2	0	0.07	2	1	1	0.0001	0.25		
#1986	5	3	2	0	0.07	2	1	1	0.0001	0.25		
#1987	5	3	2	0	0.07	2	1	1	32.75	0.25		
#1988	5	3	2	0	0.07	2	1	1	53.203	0.25		
#1989	5	3	2	0	0.07	2	1	1	108.519		0.25	
#1990	5	3	2	0	0.07	2	1	1	109.371		0.25	
#1991	5	3	2	3716.45		0.07	2	1	1	152.541		0.25
#1992	5	3	2	708.223		0.07	2	1	1	154.976		0.25
#1993	5	3	2	100.927		0.07	2	1	1	159.922		0.25
#1994	5	3	2	0	0.07	2	1	1	1.042	0.25		
#1995	5	3	2	0	0.07	2	1	1	0.0001	0.25		
#1996	5	3	2	0	0.07	2	1	1	0.0001	0.25		
#1997	5	3	2	0	0.07	2	1	1	0.0001	0.25		
#1998	5	3	2	0	0.07	2	1	1	0.0001	0.25		
#1999	5	3	2	0	0.07	2	1	1	0.0001	0.25		
#2000	5	3	2	0	0.07	2	1	1	0.0001	0.25		
#2001	5	3	2	0	0.07	2	1	1	0.0001	0.25		
#2002	5	3	2	0	0.07	2	1	1	0.0001	0.25		
#2003	5	3	2	0	0.07	2	1	1	0.0001	0.25		
#2004	5	3	2	0	0.07	2	1	1	0.0001	0.25		
#2005	5	3	2	0	0.07	2	1	1	0.0001	0.25		
2006	5	3	2	0	0.07	2	1	1	0.4	0.25		
2007	5	3	2	0	0.07	2	1	1	0.5	0.25		
2008	5	3	2	0	0.07	2	1	1	0.5	0.25		
2009	5	3	2	0	0.07	2	1	1	0.2	0.25		
#2010	5	3	2	0	0.07	2	1	1	0.0001	0.25		
#2011	5	3	2	0	0.07	2	1	1	0.0001	0.25		
#2012	5	3	2	0	0.07	2	1	1	0.0001	0.25		
2013	5	3	2	76.3798		0.07	2	1	1	2		0.25

2014	5	3	2	84.5793	0.07	2	1	1	2	0.25
2015	5	3	2	220.311	0.07	2	1	1	139.171	0.25
#2016	5	3	2	0	0.07	2	1	1	0.0001	0.25
#2017	5	3	1	0	0.07	2	1	1	0.0001	0.25
##	Fixed gear	crab	fishery discards	(t, without applying to handling mortality rate)						
1996	5	4	0	82.859	0.10	2	1	1	0	0.5
1997	5	4	0	44.979	0.10	2	1	1	0	0.5
1998	5	4	0	36.916	0.10	2	1	1	0	0.5
1999	5	4	0	100.242	0.10	2	1	1	0	0.5
2000	5	4	0	9.446	0.10	2	1	1	0	0.5
2001	5	4	0	70.553	0.10	2	1	1	0	0.5
2002	5	4	0	58.382	0.10	2	1	1	0	0.5
2003	5	4	0	25.351	0.10	2	1	1	0	0.5
2004	5	4	0	30.422	0.10	2	1	1	0	0.5
2005	5	4	0	39.802	0.10	2	1	1	0	0.5
2006	5	4	0	39.134	0.10	2	1	1	0	0.5
2007	5	4	0	64.655	0.10	2	1	1	0	0.5
2008	5	4	0	31.158	0.10	2	1	1	0	0.5
2009	5	4	0	11.616	0.10	2	1	1	0	0.5
2010	5	4	0	4.736	0.10	2	1	1	0	0.5
2011	5	4	0	21.706	0.10	2	1	1	0	0.5
2012	5	4	0	36.895	0.10	2	1	1	0	0.5
2013	5	4	0	110.970	0.10	2	1	1	0	0.5
2014	5	4	0	237.651	0.10	2	1	1	0	0.5
2015	5	4	0	154.810	0.10	2	1	1	0	0.5
2016	5	4	0	57.896	0.10	2	1	1	0	0.5
2017	5	4	0	255.155	0.10	2	1	1	0	0.5
2018	5	4	0	295.916	0.10	2	1	1	0	0.5
2019	5	4	0	90.109	0.10	2	1	1	0	0.5

## \_\_\_\_\_ ##

## RELATIVE ABUNDANCE DATA ##

## Units of Abundance: 1 = biomass, 2 = numbers

## TODO:add column for maturity for terminal molt life-histories

## for BBRKC Units are in 1000 mt.

## \_\_\_\_\_ ##

## Number of relative abundance indices

2

## Number of rows in each index

102

# Survey data (abundance indices, units are 1000 mt)

#Index Year Season Fleet Sex Abundance CV Units

1	1975	1	5	1	0	135463.3	0.193	1
1	1976	1	5	1	0	260149.5	0.207	1
1	1977	1	5	1	0	235411.4	0.144	1
1	1978	1	5	1	0	203192.7	0.152	1
1	1979	1	5	1	0	103715.0	0.164	1
1	1980	1	5	1	0	168047.2	0.221	1
1	1981	1	5	1	0	69161.2	0.190	1
1	1982	1	5	1	0	73232.9	0.251	1
1	1983	1	5	1	0	35368.0	0.214	1
1	1984	1	5	1	0	98281.5	0.606	1
1	1985	1	5	1	0	27203.7	0.159	1
1	1986	1	5	1	0	41113.6	0.420	1
1	1987	1	5	1	0	47410.5	0.209	1
1	1988	1	5	1	0	35852.6	0.228	1
1	1989	1	5	1	0	42967.7	0.232	1
1	1990	1	5	1	0	39271.6	0.242	1
1	1991	1	5	1	0	67458.4	0.443	1
1	1992	1	5	1	0	25442.5	0.176	1
1	1993	1	5	1	0	36217.5	0.198	1
1	1994	1	5	1	0	23285.5	0.174	1
1	1995	1	5	1	0	27670.5	0.266	1
1	1996	1	5	1	0	27277.5	0.203	1
1	1997	1	5	1	0	60719.6	0.264	1
1	1998	1	5	1	0	46693.7	0.182	1
1	1999	1	5	1	0	45126.5	0.204	1
1	2000	1	5	1	0	38787.8	0.216	1
1	2001	1	5	1	0	28367.5	0.187	1
1	2002	1	5	1	0	45597.0	0.202	1
1	2003	1	5	1	0	74997.9	0.283	1
1	2004	1	5	1	0	91090.1	0.321	1
1	2005	1	5	1	0	55471.4	0.171	1
1	2006	1	5	1	0	51948.6	0.169	1
1	2007	1	5	1	0	59064.2	0.174	1
1	2008	1	5	1	0	67945.7	0.249	1
1	2009	1	5	1	0	43692.8	0.326	1
1	2010	1	5	1	0	39555.6	0.223	1
1	2011	1	5	1	0	27529.9	0.213	1
1	2012	1	5	1	0	30830.4	0.237	1
1	2013	1	5	1	0	39833.2	0.244	1
1	2014	1	5	1	0	60859.1	0.191	1
1	2015	1	5	1	0	36919.3	0.208	1
1	2016	1	5	1	0	27302.6	0.194	1
1	2017	1	5	1	0	25344.0	0.173	1

1	2018	1	5	1	0	16064.2	0.161	1
1	2019	1	5	1	0	15127.4	0.157	1
1	1975	1	5	2	0	67267.3	0.193	1
1	1976	1	5	2	0	71718.0	0.207	1
1	1977	1	5	2	0	140249.6	0.144	1
1	1978	1	5	2	0	146351.8	0.152	1
1	1979	1	5	2	0	63911.7	0.164	1
1	1980	1	5	2	0	81275.0	0.221	1
1	1981	1	5	2	0	63507.9	0.190	1
1	1982	1	5	2	0	70506.7	0.251	1
1	1983	1	5	2	0	13951.7	0.214	1
1	1984	1	5	2	0	57030.0	0.606	1
1	1985	1	5	2	0	7330.8	0.159	1
1	1986	1	5	2	0	7044.8	0.420	1
1	1987	1	5	2	0	22852.7	0.209	1
1	1988	1	5	2	0	19519.6	0.228	1
1	1989	1	5	2	0	12973.6	0.232	1
1	1990	1	5	2	0	21049.2	0.242	1
1	1991	1	5	2	0	17596.5	0.443	1
1	1992	1	5	2	0	12244.8	0.176	1
1	1993	1	5	2	0	17485.5	0.198	1
1	1994	1	5	2	0	9049.4	0.174	1
1	1995	1	5	2	0	10725.7	0.266	1
1	1996	1	5	2	0	17371.1	0.203	1
1	1997	1	5	2	0	24557.1	0.264	1
1	1998	1	5	2	0	38482.0	0.182	1
1	1999	1	5	2	0	20477.3	0.204	1
1	2000	1	5	2	0	29314.2	0.216	1
1	2001	1	5	2	0	24820.6	0.187	1
1	2002	1	5	2	0	24188.9	0.202	1
1	2003	1	5	2	0	41796.1	0.283	1
1	2004	1	5	2	0	40819.8	0.321	1
1	2005	1	5	2	0	51869.8	0.171	1
1	2006	1	5	2	0	43727.8	0.169	1
1	2007	1	5	2	0	45777.1	0.174	1
1	2008	1	5	2	0	46484.5	0.249	1
1	2009	1	5	2	0	47980.0	0.326	1
1	2010	1	5	2	0	42086.5	0.223	1
1	2011	1	5	2	0	39523.3	0.213	1
1	2012	1	5	2	0	30417.8	0.237	1
1	2013	1	5	2	0	22576.6	0.244	1
1	2014	1	5	2	0	53243.9	0.191	1
1	2015	1	5	2	0	27320.8	0.208	1

1	2016	1	5	2	0	33928.4	0.194	1
1	2017	1	5	2	0	27577.5	0.173	1
1	2018	1	5	2	0	12868.2	0.161	1
1	2019	1	5	2	0	13616.4	0.157	1

	#	BSFRF										
2	2007	1	6	1	0	79542	0.116	1				
2	2008	1	6	1	0	67569	0.094	1				
2	2013	1	6	1	0	68384	0.209	1				
2	2014	1	6	1	0	62327	0.192	1				
2	2015	1	6	1	0	63709	0.161	1				
2	2016	1	6	1	0	34417	0.22	1				
2	2007	1	6	2	0	50811	0.116	1				
2	2008	1	6	2	0	38472	0.094	1				
2	2013	1	6	2	0	26633	0.209	1				
2	2014	1	6	2	0	49414	0.192	1				
2	2015	1	6	2	0	35244	0.161	1				
2	2016	1	6	2	0	43399	0.22	1				

## Number of length frequency matrices  
13

## Number of rows in each matrix

42	28	28	43	43	6	6	24	24	45	45	6	6
##	Number		of	bins	in	each	matrix	(columns	of	size		data)
20	20	16	20	16	20	16	20	16	20	16	20	16

## SIZE COMPOSITION DATA FOR ALL FLEETS

## \_\_\_\_\_ ##

## SIZE COMP LEGEND

## Sex: 1 = male, 2 = female, 0 = both sexes combined

## Type of composition: 1 = retained, 2 = discard, 0 = total composition

## Maturity state: 1 = immature, 2 = mature, 0 = both states combined

## Shell condition: 1 = new shell, 2 = old shell, 0 = both shell types combined

## \_\_\_\_\_ ##

#Retained males

#Year	Season	Fleet	Sex	Type	Shell	Maturity	Nsamp	Data	Vec			
1975	3	1	1	1	0	0	100	0	0	0	0	0
			0	0	0	0	0	0	0.0071	0.0741	0.1721	0.2239
			0.2122	0.1464	0.0858	0.0785						
1976	3	1	1	1	0	0	100	0	0	0	0	0

		0	0	0	0	0	0	0.0016	0.029	0.1418	0.2316	
		0.2199	0.1635	0.1071	0.1055							
1977	3	1	1	1	0	0	100	0	0	0	0	
		0	0	0	0	0	0	0	0.0017	0.0192	0.1382	0.2442
		0.2226	0.1605	0.104	0.1096							
1978	3	1	1	1	0	0	100	0	0	0	0	
		0	0	0	0	0	0	0	0.0012	0.0209	0.1441	0.2588
		0.2401	0.1673	0.0966	0.0711							
1979	3	1	1	1	0	0	100	0	0	0	0	
		0	0	0	0	0	0	0	0.0013	0.0119	0.0747	0.1649
		0.1998	0.2004	0.1556	0.1914							
1980	3	1	1	1	0	0	100	0	0	0	0	
		0	0	0	0	0	0	0	0.0008	0.0138	0.0919	0.1771
		0.195	0.1792	0.1404	0.2019							
1981	3	1	1	1	0	0	100	0	0	0	0	
		0	0	0	0	0	0	0	0.0006	0.0225	0.1164	0.1743
		0.1711	0.1584	0.1284	0.2283							
1982	3	1	1	1	0	0	100	0	0	0	0	
		0	0	0	0	0	0	0	0	0.0544	0.2576	0.2802
		0.1667	0.0837	0.0508	0.1067							
1984	3	1	1	1	0	0	100	0	0	0	0	
		0	0	0	0	0	0	0.0003	0.0023	0.0654	0.311	0.3135
		0.1763	0.0846	0.0321	0.0145							
1985	3	1	1	1	0	0	100	0	0	0	0	
		0	0	0	0	0	0	0.0005	0.0044	0.079	0.2869	0.3098
		0.1898	0.086	0.0306	0.0129							
1986	3	1	1	1	0	0	100	0	0	0	0	
		0	0	0	0	0	0	0	0.0016	0.0531	0.2613	0.3289
		0.2084	0.0978	0.0352	0.0137							
1987	3	1	1	1	0	0	100	0	0	0	0	
		0	0	0	0	0	0	0	0.0013	0.0284	0.1895	0.3045
		0.2522	0.1421	0.0565	0.0255							
1988	3	1	1	1	0	0	100	0	0	0	0	
		0	0	0	0	0	0	0	0	0.0202	0.1294	0.2646
		0.2471	0.1876	0.1033	0.0477							
1989	3	1	1	1	0	0	100	0	0	0	0	
		0	0	0	0	0	0	0	0.0005	0.0187	0.1211	0.2209
		0.219	0.1908	0.1197	0.1094							
1990	3	1	1	1	0	0	100	0	0	0	0	
		0	0	0	0	0	0	0.0003	0	0.0146	0.0887	0.1801
		0.1707	0.1728	0.1431	0.2297							
1991	3	1	1	1	0	0	100	0	0	0	0	
		0	0	0	0	0	0	0.0001	0.0005	0.0141	0.0848	0.1651

			0.179	0.1739	0.1432	0.2392							
1992	3	1	1	1	0	0	100	0	0	0	0	0	
		0	0	0	0	0	0.0003	0.0002	0.0005	0.0095	0.0638	0.1317	
			0.1673	0.1747	0.1636	0.2886							
1993	3	1	1	1	0	0	100	0	0	0	0	0	
		0	0	0	0	0	0	0	0.0014	0.0138	0.094	0.1789	
			0.1739	0.1596	0.1331	0.2453							
1996	3	1	1	1	0	0	100	0	0	0	0	0	
		0	0	0	0	0	0	0.0006	0.0006	0.0129	0.0779	0.1407	
			0.162	0.1771	0.1671	0.2612							
1997	3	1	1	1	0	0	100	0	0	0	0	0	
		0	0	0	0	0	0	0.0004	0.0003	0.0138	0.0899	0.1486	
			0.1603	0.1699	0.1588	0.258							
1998	3	1	1	1	0	0	100	0	0	0	0	0	
		0	0.0001	0.0001	0.0001	0.0001	0.0004	0.0002	0.0008	0.0225	0.1187	0.1596	
			0.149	0.1432	0.1394	0.266							
1999	3	1	1	1	0	0	100	0	0	0	0	0	
		0	0	0	0	0	0.0001	0	0.0001	0.0147	0.1313	0.2575	
			0.2292	0.1624	0.0961	0.1087							
2000	3	1	1	1	0	0	100	0	0	0	0	0	
		0	0	0	0.0001	0.0001	0	0.0001	0.0003	0.0111	0.0931	0.1945	
			0.2111	0.1822	0.1247	0.1826							
2001	3	1	1	1	0	0	100	0	0	0	0	0	
		0	0	0.0001	0.0001	0.0001	0.0002	0.0002	0.0012	0.0181	0.0836	0.1681	
			0.1986	0.1953	0.1506	0.1838							
2002	3	1	1	1	0	0	100	0	0	0	0	0	
		0	0.0001	0	0.0001	0.0001	0.0001	0	0.0002	0.0151	0.108	0.1884	
			0.1915	0.1683	0.1334	0.1948							
2003	3	1	1	1	0	0	100	0	0	0	0	0	
		0	0	0	0	0.0001	0.0001	0.0002	0.0009	0.0243	0.1464	0.232	
			0.1871	0.1497	0.0994	0.1597							
2004	3	1	1	1	0	0	100	0	0	0	0	0	
		0	0	0	0	0	0	0	0.0002	0.0064	0.0514	0.1302	
			0.1702	0.1971	0.1632	0.2812							
2005	3	1	1	1	0	0	100	0	0	0	0	0	
		0	0.0001	0	0	0	0.0001	0.0001	0.0008	0.015	0.0859	0.1543	
			0.1661	0.1783	0.1516	0.2475							
2006	3	1	1	1	0	0	100	0	0	0	0	0	
		0	0	0	0	0	0.0001	0.0001	0.0004	0.0102	0.0739	0.1905	
			0.2203	0.1887	0.137	0.1787							
2007	3	1	1	1	0	0	100	0	0	0	0	0	
		0	0	0	0	0	0	0.0002	0.0003	0.0067	0.0871	0.1833	
			0.1934	0.1846	0.1472	0.1973							

2008	3	1	1	1	0	0	100	0	0	0	0	0
		0	0	0	0	0	0	0.0001	0.0002	0.01	0.0746	0.1457
		0.1619	0.179	0.1625	0.2659							
2009	3	1	1	1	0	0	100	0	0	0	0	0
		0	0	0	0	0	0	0	0.0002	0.0108	0.1152	0.2215
		0.1968	0.1588	0.1084	0.1882							
2010	3	1	1	1	0	0	100	0	0	0	0	0
		0	0	0	0	0	0	0	0.0003	0.0091	0.0986	0.2244
		0.2238	0.1861	0.1144	0.1433							
2011	3	1	1	1	0	0	100	0	0	0	0	0
		0	0	0	0	0	0.0003	0.0001	0.0003	0.0114	0.118	0.2436
		0.2292	0.1725	0.1077	0.1169							
2012	3	1	1	1	0	0	100	0	0	0	0	0
		0	0	0	0.0001	0	0.0001	0	0	0.0044	0.0499	0.1249
		0.173	0.1886	0.1654	0.2937							
2013	3	1	1	1	0	0	100	0	0	0	0	0
		0	0	0.0001	0.0001	0	0	0.0001	0.0001	0.0054	0.0525	0.1271
		0.1484	0.1657	0.1632	0.3374							
2014	3	1	1	1	0	0	100	0	0	0	0	0
		0	0	0	0	0	0	0	0.0004	0.0117	0.0964	0.1831
		0.1696	0.1454	0.1246	0.2689							
2015	3	1	1	1	0	0	100	0	0	0	0	0
		0	0	0	0	0	0	0.0001	0.0003	0.0067	0.0616	0.1473
		0.1864	0.1947	0.1634	0.2397							
2016	3	1	1	1	0	0	100	0	0	0	0	0
		0	0	0	0	0	0	0	0.0002	0.0062	0.0489	0.127
		0.166	0.1822	0.1689	0.3006							
2017	3	1	1	1	0	0	100	0	0	0	0	0
		0	0	0	0.0001	0.0001	0	0	0	0.0044	0.0453	0.1055
		0.1441	0.1781	0.1664	0.356							
2018	3	1	1	1	0	0	100	0	0	0	0	0
		0	0	0	0	0	0	0	0.0001	0.0052	0.0593	0.1370
		0.1406	0.1386	0.1239	0.3951							
2019	3	1	1	1	0	0	100	0	0	0	0	0
		0	0	0	0	0	0.0000	0.0004	0.0086	0.0678	0.1360	0.1338
		0.1276	0.1139	0.4119								

#Total males

#Year	Season	Fleet	Sex	Type	Shell	Maturity	Nsamp	Data	Vec				
1990	3	1	1	0	0	0	100	0	0	0.0004	0.0028	0.0016	
			0.0043	0.0024	0.013	0.0173	0.0263	0.0421	0.0523	0.0641	0.0943	0.1018	0.1108
			0.1156	0.0924	0.0971	0.1616							
1991	3	1	1	0	0	0	100	0.0009	0.0038	0.0075	0.0081	0.0092	



			0.0149	0.0124	0.0241	0.0236	0.0262	0.0243	0.0428	0.0605	0.0884	0.1014	0.1069
			0.1152	0.1161	0.085	0.129							
1992	3	1	1	0	0	0	100	0	0.0006	0.0008	0.0075	0.0151	
			0.0375	0.0591	0.0777	0.0806	0.0838	0.0806	0.0852	0.0756	0.0603	0.0477	0.0503
			0.0538	0.0578	0.0448	0.081							
1993	3	1	1	0	0	0	100	0.0008	0.0024	0.0031	0.003	0.004	
			0.0073	0.0176	0.0325	0.0455	0.062	0.0745	0.0854	0.0832	0.0991	0.0909	0.0898
			0.0749	0.0725	0.0567	0.0946							
1996	3	1	1	0	0	0	100	0	0	0	0.0047	0.0187	
			0.0296	0.0265	0.0109	0.0171	0.0249	0.0218	0.0358	0.053	0.0872	0.0981	0.0888
			0.1277	0.1246	0.0903	0.1402							
1997	3	1	1	0	0	0	100	0	0.0001	0.0002	0.0003	0.0006	
			0.0081	0.0227	0.0446	0.0519	0.0534	0.0422	0.041	0.0522	0.0701	0.0832	0.0938
			0.0967	0.1035	0.0886	0.1467							
1998	3	1	1	0	0	0	100	0.0001	0.0002	0.0004	0.0021	0.0037	
			0.0054	0.0056	0.0104	0.0246	0.0588	0.0946	0.1362	0.1335	0.1122	0.0476	0.0117
			0.0386	0.0565	0.0525	0.2052							
1999	3	1	1	0	0	0	100	0	0	0	0.0013	0.0013	
			0.0006	0.0017	0.0013	0.0025	0.006	0.0138	0.0264	0.0537	0.0923	0.1302	0.1444
			0.1518	0.1301	0.091	0.1515							
2000	3	1	1	0	0	0	100	0.0002	0.002	0.0071	0.0185	0.0234	
			0.0242	0.0256	0.0262	0.0254	0.0291	0.0349	0.0507	0.0718	0.0843	0.1001	0.1083
			0.1114	0.0943	0.0638	0.0988							
2001	3	1	1	0	0	0	100	0.0004	0.0023	0.0037	0.005	0.0066	
			0.0139	0.0249	0.0381	0.0447	0.0539	0.0605	0.0696	0.0659	0.0647	0.0652	0.0843
			0.0982	0.1023	0.0824	0.1133							
2002	3	1	1	0	0	0	100	0.0017	0.0046	0.0044	0.0051	0.0043	
			0.0054	0.0066	0.0151	0.0272	0.0504	0.0684	0.0822	0.083	0.0901	0.0939	0.0985
			0.0913	0.0881	0.0689	0.1108							
2003	3	1	1	0	0	0	100	0.0034	0.0053	0.0065	0.0144	0.0257	
			0.0323	0.0355	0.0335	0.0315	0.0322	0.036	0.0526	0.0756	0.1021	0.1115	0.108
			0.0867	0.0715	0.0494	0.0863							
2004	3	1	1	0	0	0	100	0.0001	0.0019	0.0061	0.016	0.021	
			0.0231	0.0316	0.0519	0.0613	0.0616	0.0486	0.0411	0.035	0.0389	0.0474	0.0731
			0.0927	0.1087	0.0917	0.1482							
2005	3	1	1	0	0	0	100	0.0001	0.0005	0.0008	0.0017	0.0044	
			0.0128	0.0199	0.0243	0.0264	0.0383	0.0556	0.0801	0.0806	0.0849	0.0723	0.0769
			0.0794	0.0949	0.0818	0.1643							
2006	3	1	1	0	0	0	100	0.0001	0.0006	0.0019	0.0065	0.014	
			0.0171	0.0166	0.0154	0.02	0.0334	0.0412	0.0506	0.0611	0.0815	0.098	0.1153
			0.1191	0.113	0.0806	0.1138							
2007	3	1	1	0	0	0	100	0.0006	0.0021	0.0034	0.0051	0.0089	
			0.0191	0.0341	0.044	0.0477	0.044	0.0423	0.0513	0.0676	0.0899	0.0952	0.0974

2008	3	1	1	0	0	0	100	0.0001	0.0002	0.0007	0.0025	0.0059
		0.0078	0.0088	0.0118	0.0242	0.0444	0.0697	0.0985	0.1095	0.1038	0.0868	0.0768
		0.0766	0.0772	0.0703	0.1244							
2009	3	1	1	0	0	0	100	0.0002	0.0005	0.0009	0.0016	0.0021
		0.0038	0.0093	0.0213	0.033	0.0371	0.0428	0.0638	0.0978	0.1348	0.1354	0.1172
		0.0895	0.0659	0.0499	0.0931							
2010	3	1	1	0	0	0	100	0.0004	0.0006	0.0013	0.0028	0.0044
		0.0061	0.0077	0.0113	0.0179	0.0286	0.0504	0.0807	0.107	0.1302	0.1264	0.121
		0.1031	0.0821	0.0512	0.067							
2011	3	1	1	0	0	0	100	0.0008	0.0031	0.0055	0.0096	0.0099
		0.0089	0.0128	0.0147	0.0192	0.0264	0.0358	0.0564	0.0822	0.1114	0.1321	0.1357
		0.1212	0.0926	0.0583	0.0633							
2012	3	1	1	0	0	0	100	0.0002	0.0003	0.0008	0.0014	0.0037
		0.0088	0.014	0.0188	0.0178	0.0192	0.0236	0.0359	0.0519	0.0746	0.0861	0.099
		0.112	0.1276	0.1127	0.1915							
2013	3	1	1	0	0	0	100	0.0001	0.0007	0.0017	0.0022	0.0047
		0.0059	0.0097	0.0152	0.0261	0.0381	0.0546	0.0609	0.0673	0.0742	0.0761	0.0826
		0.0842	0.1033	0.0981	0.1944							
2014	3	1	1	0	0	0	100	0.0003	0.0006	0.0008	0.0012	0.0017
		0.0038	0.0063	0.0111	0.0155	0.0206	0.0345	0.0474	0.0701	0.0902	0.1051	0.108
		0.1051	0.0972	0.0846	0.196							
2015	3	1	1	0	0	0	100	0.0001	0.0002	0.0008	0.0017	0.0038
		0.0059	0.0063	0.007	0.012	0.0272	0.0337	0.0492	0.0541	0.0675	0.0799	0.107
		0.117	0.137	0.1056	0.1841							
2016	3	1	1	0	0	0	100	0.0001	0.0002	0.0015	0.0034	0.0046
		0.0064	0.0111	0.0188	0.0225	0.028	0.0295	0.04	0.0509	0.0675	0.0814	0.0938
		0.1068	0.1214	0.1118	0.2005							
2017	3	1	1	0	0	0	100	0.0003	0.0006	0.0034	0.012	0.0258
		0.0362	0.0313	0.0248	0.0207	0.0259	0.0306	0.047	0.0505	0.0641	0.0671	0.0809
		0.097	0.1032	0.0949	0.1839							
2018	3	1	1	0	0	0	100	0.0004	0.0017	0.0065	0.0074	0.0060
		0.0100	0.0217	0.0402	0.0630	0.0704	0.0659	0.0551	0.0560	0.0565	0.0621	0.0649
		0.0632	0.0669	0.0698	0.2124							
2019	3	1	1	0	0	0	100	0.0000	0.0001	0.0002	0.0021	0.0094
		0.0186	0.0241	0.0214	0.0212	0.0383	0.0591	0.0896	0.0975	0.0981	0.0889	0.0736
		0.0608	0.0588	0.0503	0.1879							

#Total females

#Year	Season	Fleet	Sex	Type	Shell	Maturity	Nsamp	Data	Vec			
1990	3	1	2	0	0	0	50	0	0.0014	0.0029	0.0029	0.0057
									0.0072	0.0143	0.0672	0.1016
									0.1731	0.1688	0.2132	0.1359
									0.0715	0.0243	0.01	
1991	3	1	2	0	0	0	37.5	0.0027	0.024	0.0613	0.096	0.1333

1992	3	0.16	0.1227	0.072	0.0693	0.056	0.0693	0.08	0.0347	0.0107	0.0053	0.0027
		1	2	0	0	0	50	0	0.0013	0.0029	0.0177	0.0803
		0.1765	0.195	0.1698	0.0958	0.0815	0.0572	0.0404	0.0395	0.0256	0.0118	0.0046
1993	3	1	2	0	0	0	50	0.0013	0.0023	0.0047	0.006	0.0137
		0.033	0.1017	0.1606	0.1446	0.1136	0.09	0.0849	0.0829	0.0735	0.043	0.0442
1996	3	1	2	0	0	0	1.1	0	0	0	0.0909	0.6364
		0.2727	0	0	0	0	0	0	0	0	0	0
1997	3	1	2	0	0	0	50	0	0	0.0011	0.0011	0.0099
		0.0265	0.0364	0.0464	0.0695	0.1391	0.1667	0.1435	0.117	0.1082	0.0607	0.074
1998	3	1	2	0	0	0	50	0.0002	0.0004	0.0009	0.0024	0.0062
		0.0165	0.0519	0.168	0.2191	0.1527	0.0862	0.0853	0.0578	0.0533	0.0362	0.0628
1999	3	1	2	0	0	0	3.6	0	0	0	0.025	0.025
		0.025	0.05	0.025	0	0.125	0.125	0.075	0.1	0.125	0.075	0.225
2000	3	1	2	0	0	0	50	0	0.0044	0.0256	0.0607	0.0744
		0.0816	0.0701	0.0543	0.055	0.0998	0.1541	0.146	0.0799	0.042	0.0224	0.0296
2001	3	1	2	0	0	0	50	0.0007	0.0042	0.0129	0.0307	0.0568
		0.0844	0.0986	0.0909	0.0646	0.0568	0.0883	0.1407	0.14	0.0638	0.0269	0.0396
2002	3	1	2	0	0	0	30.2	0.0595	0.1714	0.1601	0.1388	0.1091
		0.0581	0.0297	0.0326	0.0382	0.0326	0.0241	0.0241	0.0198	0.0269	0.0283	0.0467
2003	3	1	2	0	0	0	50	0.012	0.0164	0.0231	0.0635	0.102
		0.1075	0.0682	0.043	0.06	0.0866	0.0984	0.0675	0.054	0.0596	0.0572	0.0811
2004	3	1	2	0	0	0	50	0.0003	0.0056	0.0258	0.0575	0.0774
		0.0918	0.1413	0.1308	0.0876	0.0449	0.0503	0.0611	0.0531	0.0446	0.0431	0.0851
2005	3	1	2	0	0	0	50	0.0004	0.0013	0.0022	0.005	0.0146
		0.05	0.0788	0.0931	0.1233	0.1212	0.0871	0.1021	0.0958	0.0885	0.0519	0.0848
2006	3	1	2	0	0	0	50	0.0003	0.004	0.0256	0.1183	0.1939
		0.1616	0.0692	0.0519	0.0672	0.0704	0.0576	0.0403	0.0358	0.0323	0.0256	0.0461
2007	3	1	2	0	0	0	50	0.0029	0.0124	0.0214	0.0235	0.0461
		0.0886	0.1116	0.0832	0.0556	0.0739	0.1005	0.1146	0.0942	0.0671	0.0437	0.0604
2008	3	1	2	0	0	0	50	0.0004	0.0018	0.0097	0.0362	0.0775
		0.0662	0.0472	0.0772	0.1071	0.0871	0.0954	0.126	0.1254	0.067	0.0391	0.0368
2009	3	1	2	0	0	0	50	0.0036	0.0083	0.0099	0.0144	0.0164
		0.0282	0.0652	0.0867	0.0803	0.0912	0.0857	0.09	0.1141	0.1308	0.0875	0.0877
2010	3	1	2	0	0	0	50	0.0036	0.0051	0.0052	0.0199	0.0276
		0.0292	0.0269	0.0444	0.0882	0.1135	0.1315	0.1423	0.1011	0.0917	0.0879	0.0816
2011	3	1	2	0	0	0	50	0.013	0.037	0.0604	0.101	0.076
		0.0698	0.0583	0.0411	0.0266	0.0359	0.0693	0.0911	0.0823	0.0667	0.0672	0.1042
2012	3	1	2	0	0	0	50	0.0089	0.0107	0.0124	0.0337	0.0604
		0.1155	0.0941	0.0391	0.0178	0.0124	0.0409	0.0426	0.1652	0.151	0.1101	0.0853
2013	3	1	2	0	0	0	50	0.0005	0.0017	0.0083	0.0109	0.0187
		0.037	0.0716	0.1327	0.1428	0.0967	0.0716	0.0637	0.0851	0.0904	0.0731	0.0952
2014	3	1	2	0	0	0	50	0.0011	0.0053	0.0068	0.0086	0.0086
		0.021	0.0282	0.0274	0.0526	0.0713	0.0755	0.0762	0.0965	0.1142	0.1303	0.2764

2015	3	1	2	0	0	0	50	0	0.0011	0.0018	0.0051	0.012
									0.0164	0.0197	0.0354	0.0556
2016	3	1	2	0	0	0	50	0	0.0003	0.0073	0.0122	0.0187
									0.0181	0.0213	0.0312	0.0377
2017	3	1	2	0	0	0	50	0.0005	0.003	0.0137	0.0526	0.0983
									0.1093	0.0806	0.0333	0.0371
2018	3	1	2	0	0	0	50	0.0003	0.0046	0.0171	0.0233	0.0221
									0.0338	0.0542	0.0839	0.0766
2019	3	1	2	0	0	0	50	0.0000	0.0000	0.0018	0.0053	0.0263
									0.0458	0.0362	0.0337	0.0564
									0.0777	0.0702	0.0770	0.1057
									0.1302	0.1153	0.2185	

#Trawl bycatch			male									
#Year	Season	Fleet	Sex	Type	Shell	Maturity		Nsamp	Data	Vec		
1976	5	2	1	0.0	0	0	50	0.0000	0.0000	0.0000	0.0000	0.0000
									0.0130	0.0087	0.0043	0.0216
									0.0736	0.0909	0.0649	0.1299
1977	5	2	1	0.0	0	0	50	0.0036	0.0009	0.0009	0.0009	0.0026
									0.0035	0.0079	0.0097	0.0317
									0.1004	0.0634	0.0326	0.0441
1978	5	2	1	0.0	0	0	50	0.0000	0.0000	0.0000	0.0000	0.0000
									0.0000	0.0000	0.0025	0.0012
									0.0797	0.0984	0.0672	0.1880
1979	5	2	1	0.0	0	0	50	0.0178	0.0013	0.0025	0.0013	0.0025
									0.0076	0.0038	0.0025	0.0013
									0.0898	0.0860	0.0809	0.1858
1980	5	2	1	0.0	0	0	50	0.0531	0.0207	0.0096	0.0135	0.0142
									0.0163	0.0274	0.0263	0.0380
									0.0207	0.0142	0.0131	0.0265
1981	5	2	1	0.0	0	0	50	0.0262	0.0028	0.0045	0.0066	0.0112
									0.0175	0.0279	0.0349	0.0386
									0.0112	0.0064	0.0051	0.0087
1982	5	2	1	0.0	0	0	50	0.0701	0.0268	0.0247	0.0326	0.0356
									0.0443	0.0409	0.0403	0.0401
									0.0084	0.0052	0.0038	0.0099
1983	5	2	1	0.0	0	0	50	0.0231	0.0214	0.0336	0.0344	0.0311
									0.0319	0.0377	0.0445	0.0473
									0.0204	0.0129	0.0096	0.0180
1984	5	2	1	0.0	0	0	50	0.0366	0.0156	0.0147	0.0199	0.0270
									0.0342	0.0399	0.0407	0.0431
									0.0264	0.0170	0.0109	0.0146
1985	5	2	1	0.0	0	0	50	0.0051	0.0014	0.0034	0.0059	0.0100
									0.0164	0.0256	0.0396	0.0357
									0.0638	0.0455	0.0299	0.0578

1986	5	2	1	0.0	0	0	50	0.0038	0.0019	0.0085	0.0019	0.0056	0.0136	0.0193	0.0357	0.0160	0.0249	0.0221	0.0320	0.0710	0.0555	0.0527	0.0635	0.0456	0.0362	0.0259	0.0282
1987	5	2	1	0.0	0	0	49.9	0.0020	0.0000	0.0010	0.0020	0.0050	0.0080	0.0190	0.0271	0.0170	0.0220	0.0441	0.0491	0.0401	0.0581	0.0852	0.0812	0.0671	0.0611	0.0511	0.0842
1988	5	2	1	0.0	0	0	31.55	0.0048	0.0048	0.0063	0.0016	0.0032	0.0000	0.0095	0.0175	0.0127	0.0397	0.0524	0.0540	0.0571	0.0635	0.0651	0.0889	0.0794	0.0587	0.0349	0.0397
1989	5	2	1	0.0	0	0	50	0.0047	0.0026	0.0019	0.0006	0.0019	0.0019	0.0045	0.0047	0.0097	0.0142	0.0237	0.0379	0.0439	0.0534	0.0710	0.0809	0.0798	0.0783	0.0678	0.0897
1990	5	2	1	0.0	0	0	50	0.0051	0.0041	0.0071	0.0020	0.0081	0.0071	0.0234	0.0142	0.0244	0.0264	0.0224	0.0305	0.0325	0.0508	0.0843	0.0843	0.0772	0.0681	0.0376	0.0742
1991	5	2	1	0.0	0	0	16.3	0.0036	0.0072	0.0036	0.0072	0.0181	0.0144	0.0144	0.0181	0.0361	0.0253	0.0361	0.0325	0.0397	0.0217	0.0289	0.0722	0.0505	0.0578	0.0650	0.1588
1992	5	2	1	0.0	0	0	22	0.0210	0.0210	0.0180	0.0000	0.0060	0.0060	0.0030	0.0000	0.0060	0.0120	0.0240	0.0210	0.0360	0.0390	0.0390	0.0450	0.0240	0.0210	0.0030	0.0330
1994	5	2	1	0.0	0	0	28.6	0.0000	0.0000	0.0035	0.0070	0.0018	0.0000	0.0000	0.0000	0.0000	0.0000	0.0018	0.0018	0.0088	0.0158	0.0210	0.0473	0.0438	0.0578	0.0841	0.2785
1995	5	2	1	0.0	0	0	8	0.0067	0.0267	0.0133	0.0067	0.0067	0.0067	0.0000	0.0133	0.0067	0.0200	0.0000	0.0133	0.0200	0.0133	0.0400	0.0667	0.1267	0.0867	0.0467	0.2467
1996	5	2	1	0.0	0	0	50	0.0000	0.0008	0.0000	0.0016	0.0049	0.0114	0.0147	0.0188	0.0294	0.0343	0.0474	0.0662	0.0466	0.0686	0.0392	0.0645	0.0425	0.0564	0.0417	0.1266
1997	5	2	1	0.0	0	0	17.45	0.0000	0.0000	0.0000	0.0000	0.0029	0.0029	0.0029	0.0088	0.0088	0.0206	0.0206	0.0265	0.0235	0.0176	0.0500	0.0647	0.0324	0.0382	0.0382	0.1559
1998	5	2	1	0.0	0	0	50	0.0007	0.0007	0.0007	0.0000	0.0000	0.0000	0.0035	0.0028	0.0056	0.0133	0.0280	0.0314	0.0566	0.0475	0.0580	0.0419	0.0419	0.0475	0.0405	0.1097
1999	5	2	1	0.0	0	0	32.15	0.0016	0.0016	0.0000	0.0016	0.0031	0.0000	0.0063	0.0031	0.0079	0.0126	0.0142	0.0409	0.0504	0.0756	0.1071	0.1008	0.0913	0.0709	0.0661	0.0945
2000	5	2	1	0.0	0	0	36.7	0.0000	0.0000	0.0014	0.0014	0.0014	0.0068	0.0095	0.0286	0.0368	0.0327	0.0354	0.0313	0.0422	0.0463	0.0354	0.0422	0.0436	0.0463	0.0518	0.2262
2001	5	2	1	0.0	0	0	40.1	0.0000	0.0000	0.0050	0.0025	0.0100	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000

			0.0339	0.0226	0.0263	0.0402	0.0376	0.0427	0.0351	0.0351	0.0251	0.0351	0.0226
			0.0477	0.0351	0.0527	0.1041							
2002	5	2	1	0.0	0	0	50	0.0009	0.0009	0.0009	0.0009	0.0009	0.0018
				0.0026	0.0061	0.0044	0.0061	0.0105	0.0219	0.0193	0.0280	0.0368	0.0464
				0.0517	0.0569	0.0412	0.1322						
2003	5	2	1	0.0	0	0	26.25	0.0019	0.0039	0.0058	0.0077	0.0193	
				0.0097	0.0154	0.0232	0.0251	0.0174	0.0135	0.0193	0.0309	0.0347	0.0425
				0.0463	0.0483	0.0521	0.1216						
2004	5	2	1	0.0	0	0	33.3	0.0015	0.0000	0.0000	0.0015	0.0015	
				0.0045	0.0060	0.0166	0.0211	0.0166	0.0302	0.0392	0.0407	0.0377	0.0347
				0.0422	0.0392	0.0347	0.1448						
2005	5	2	1	0.0	0	0	50	0.0029	0.0038	0.0019	0.0086	0.0077	
				0.0134	0.0211	0.0154	0.0125	0.0230	0.0259	0.0393	0.0509	0.0480	0.0422
				0.0461	0.0480	0.0403	0.0883						
2006	5	2	1	0.0	0	0	50	0.0000	0.0000	0.0000	0.0000	0.0017	
				0.0025	0.0025	0.0127	0.0110	0.0391	0.0365	0.0425	0.0484	0.0467	0.0688
				0.0688	0.0671	0.0586	0.1393						
2007	5	2	1	0.0	0	0	50	0.0000	0.0000	0.0000	0.0016	0.0024	
				0.0032	0.0048	0.0112	0.0128	0.0136	0.0233	0.0217	0.0289	0.0393	0.0457
				0.0393	0.0425	0.0586	0.1252						
2008	5	2	1	0.0	0	0	50	0.0000	0.0000	0.0006	0.0000	0.0025	
				0.0025	0.0019	0.0025	0.0131	0.0255	0.0255	0.0597	0.0622	0.0566	0.0715
				0.0646	0.0547	0.0541	0.1753						
2009	5	2	1	0.0	0	0	50	0.0000	0.0000	0.0000	0.0000	0.0008	
				0.0025	0.0025	0.0033	0.0066	0.0108	0.0116	0.0298	0.0298	0.0431	0.0547
				0.0671	0.0497	0.0530	0.1740						
2010	5	2	1	0.0	0	0	45.95	0.0000	0.0000	0.0022	0.0022	0.0022	
				0.0054	0.0033	0.0120	0.0185	0.0174	0.0196	0.0348	0.0490	0.0501	0.0566
				0.0359	0.0337	0.0370	0.0860						
2011	5	2	1	0.0	0	0	22.3	0.0000	0.0000	0.0022	0.0067	0.0067	
				0.0022	0.0022	0.0067	0.0135	0.0090	0.0067	0.0067	0.0224	0.0269	0.0493
				0.0605	0.0628	0.0448	0.1188						
2012	5	2	1	0.0	0	0	14.15	0.0000	0.0035	0.0000	0.0000	0.0000	
				0.0035	0.0071	0.0071	0.0035	0.0071	0.0141	0.0106	0.0283	0.0353	0.0601
				0.0495	0.0530	0.0530	0.1696						
2013	5	2	1	0.0	0	0	24.2	0.0000	0.0021	0.0000	0.0021	0.0021	
				0.0000	0.0000	0.0021	0.0041	0.0083	0.0103	0.0227	0.0455	0.0393	0.0517
				0.0434	0.0517	0.0393	0.2624						
2014	5	2	1	0.0	0	0	13.05	0.0000	0.0038	0.0000	0.0038	0.0115	
				0.0038	0.0000	0.0192	0.0038	0.0115	0.0192	0.0230	0.0268	0.0383	0.0690
				0.0421	0.0345	0.0460	0.2069						
2015	5	2	1	0.0	0	0	20.45	0.0000	0.0000	0.0073	0.0073	0.0073	
				0.0049	0.0122	0.0147	0.0122	0.0147	0.0220	0.0293	0.0318	0.0440	0.0342
													0.0391

2016	5	2	1	0.0	0	0	30.85	0.0000	0.0016	0.0032	0.0049	0.0032		
				0.0016	0.0130	0.0097	0.0162	0.0065	0.0113	0.0357	0.0243	0.0470	0.0519	0.0583
				0.0632	0.0794	0.0778	0.2107							
2017	5	2	1	0.0	0	0	35.9	0.0000	0.0000	0.0000	0.0000	0.0056		
				0.0042	0.0056	0.0056	0.0070	0.0056	0.0084	0.0153	0.0265	0.0320	0.0418	0.0529
				0.0891	0.0766	0.1017	0.3231							
2018	5	2	1	0.0	0	0	44.65	0.0011	0.0000	0.0022	0.0000	0.0022		
				0.0045	0.0112	0.0045	0.0213	0.0202	0.0403	0.0426	0.0437	0.0594	0.0448	0.0336
				0.0448	0.0403	0.0403	0.1601							
2019	5	2	1	0.0	0	0	38.0	0.0013	0.0013	0.0053	0.0079	0.0092		
				0.0118	0.0053	0.0092	0.0092	0.0276	0.0303	0.0316	0.0434	0.0553	0.0566	0.0434
				0.0539	0.0421	0.0395	0.2132							

#Trawl bycatch			female															
#Year	Season	Fleet	Sex	Type	Shell	Maturity		Nsamp	Data	Vec								
1976	5	2	2	0	0	0	0	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000					
								0.0000	0.0130	0.0087	0.0216	0.0260	0.0303	0.0563	0.0130	0.0260	0.0043	0.0260
1977	5	2	2	0	0	0	0	0.0000	0.0009	0.0009	0.0000	0.0000	0.0000					
								0.0009	0.0026	0.0053	0.0070	0.0088	0.0062	0.0053	0.0044	0.0026	0.0009	0.0009
1978	5	2	2	0	0	0	0	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000					
								0.0000	0.0000	0.0000	0.0000	0.0000	0.0075	0.0050	0.0075	0.0262	0.0324	0.0610
1979	5	2	2	0	0	0	0	0.0130	0.0013	0.0000	0.0000	0.0000	0.0063					
								0.0038	0.0152	0.0468	0.0354	0.0392	0.0544	0.0215	0.0164	0.0177	0.0013	0.0139
1980	5	2	2	0	0	0	0	0.0433	0.0160	0.0096	0.0189	0.0281						
								0.0409	0.0497	0.0472	0.0489	0.0525	0.0362	0.0265	0.0134	0.0081	0.0039	0.0040
1981	5	2	2	0	0	0	0	0.0612	0.0245	0.0245	0.0437	0.0540						
								0.0608	0.0525	0.0425	0.0315	0.0383	0.0312	0.0267	0.0240	0.0158	0.0093	0.0086
1982	5	2	2	0	0	0	0	0.0631	0.0235	0.0237	0.0285	0.0379						
								0.0413	0.0332	0.0246	0.0190	0.0177	0.0156	0.0144	0.0104	0.0080	0.0034	0.0049
1983	5	2	2	0	0	0	0	0.0281	0.0233	0.0351	0.0363	0.0358						
								0.0407	0.0392	0.0316	0.0222	0.0154	0.0100	0.0087	0.0065	0.0042	0.0030	0.0041
1984	5	2	2	0	0	0	0	0.0400	0.0156	0.0155	0.0211	0.0298						
								0.0344	0.0399	0.0359	0.0287	0.0151	0.0085	0.0060	0.0042	0.0031	0.0019	0.0029
1985	5	2	2	0	0	0	0	0.0034	0.0013	0.0024	0.0046	0.0096						
								0.0171	0.0195	0.0193	0.0163	0.0128	0.0119	0.0111	0.0108	0.0057	0.0025	0.0066
1986	5	2	2	0	0	0	0	0.0038	0.0014	0.0038	0.0000	0.0038						
								0.0099	0.0329	0.0762	0.0630	0.0470	0.0494	0.0466	0.0428	0.0202	0.0085	0.0268
1987	5	2	2	0	0	0	0	0.0020	0.0020	0.0030	0.0100	0.0180						
								0.0311	0.0331	0.0401	0.0220	0.0311	0.0160	0.0391	0.0080	0.0080	0.0030	0.0090
1988	5	2	2	0	0	0	0	0.0079	0.0143	0.0032	0.0079	0.0063						
								0.0127	0.0222	0.0333	0.0476	0.0524	0.0397	0.0222	0.0175	0.0079	0.0048	0.0063
1989	5	2	2	0	0	0	0	0.0028	0.0024	0.0015	0.0022	0.0065						

1990	5	0.0108	0.0204	0.0430	0.0504	0.0480	0.0435	0.0295	0.0256	0.0170	0.0065	0.0168
		2	2	0	0	0	0	0.0020	0.0041	0.0071	0.0081	0.0112
		0.0112	0.0183	0.0203	0.0366	0.0305	0.0335	0.0325	0.0234	0.0173	0.0152	0.0447
1991	5	2	2	0	0	0	0	0.0000	0.0036	0.0108	0.0036	0.0000
		0.0072	0.0036	0.0072	0.0289	0.0181	0.0181	0.0289	0.0181	0.0325	0.0036	0.1047
1992	5	2	2	0	0	0	0	0.0030	0.0000	0.0000	0.0030	0.0420
		0.0631	0.0480	0.0480	0.0450	0.0480	0.0631	0.0691	0.0480	0.0450	0.0390	0.0571
1994	5	2	2	0	0	0	0	0.0000	0.0035	0.0088	0.0280	0.0333
		0.0438	0.0298	0.0665	0.0455	0.0175	0.0140	0.0123	0.0140	0.0210	0.0210	0.0683
1995	5	2	2	0	0	0	0	0.0467	0.0000	0.0000	0.0200	0.0067
		0.0200	0.0333	0.0133	0.0200	0.0000	0.0200	0.0000	0.0067	0.0133	0.0000	0.0333
1996	5	2	2	0	0	0	0	0.0000	0.0000	0.0008	0.0090	0.0204
		0.0335	0.0147	0.0163	0.0188	0.0253	0.0253	0.0188	0.0237	0.0212	0.0139	0.0425
1997	5	2	2	0	0	0	0	0.0000	0.0000	0.0000	0.0000	0.0029
		0.0000	0.0265	0.0382	0.0676	0.0941	0.0471	0.0412	0.0559	0.0294	0.0147	0.0676
1998	5	2	2	0	0	0	0	0.0000	0.0000	0.0000	0.0007	0.0014
		0.0042	0.0182	0.0503	0.0545	0.0440	0.0391	0.0321	0.0468	0.0370	0.0398	0.1013
1999	5	2	2	0	0	0	0	0.0000	0.0000	0.0000	0.0016	0.0000
		0.0000	0.0047	0.0047	0.0079	0.0205	0.0252	0.0220	0.0346	0.0236	0.0299	0.0756
2000	5	2	2	0	0	0	0	0.0000	0.0000	0.0000	0.0027	0.0041
		0.0082	0.0150	0.0191	0.0082	0.0163	0.0313	0.0422	0.0177	0.0232	0.0082	0.0845
2001	5	2	2	0	0	0	0	0.0000	0.0000	0.0025	0.0025	0.0138
		0.0125	0.0289	0.0226	0.0251	0.0301	0.0201	0.0238	0.0301	0.0351	0.0376	0.1016
2002	5	2	2	0	0	0	0	0.0000	0.0009	0.0000	0.0018	0.0035
		0.0079	0.0149	0.0271	0.0525	0.0368	0.0280	0.0315	0.0394	0.0438	0.0490	0.1480
2003	5	2	2	0	0	0	0	0.0000	0.0058	0.0039	0.0116	0.0154
		0.0232	0.0174	0.0193	0.0232	0.0270	0.0251	0.0425	0.0309	0.0328	0.0328	0.0985
2004	5	2	2	0	0	0	0	0.0000	0.0000	0.0000	0.0015	0.0015
		0.0136	0.0287	0.0377	0.0392	0.0287	0.0513	0.0332	0.0407	0.0211	0.0362	0.1131
2005	5	2	2	0	0	0	0	0.0010	0.0058	0.0077	0.0048	0.0086
		0.0211	0.0355	0.0499	0.0672	0.0605	0.0259	0.0307	0.0221	0.0192	0.0154	0.0441
2006	5	2	2	0	0	0	0	0.0000	0.0000	0.0008	0.0008	0.0051
		0.0093	0.0068	0.0102	0.0153	0.0229	0.0297	0.0306	0.0340	0.0272	0.0178	0.0731
2007	5	2	2	0	0	0	0	0.0000	0.0000	0.0032	0.0016	0.0032
		0.0144	0.0265	0.0353	0.0353	0.0369	0.0457	0.0554	0.0514	0.0514	0.0353	0.0899
2008	5	2	2	0	0	0	0	0.0000	0.0000	0.0000	0.0006	0.0068
		0.0044	0.0081	0.0168	0.0305	0.0267	0.0267	0.0267	0.0342	0.0199	0.0186	0.0609
2009	5	2	2	0	0	0	0	0.0000	0.0000	0.0000	0.0000	0.0017
		0.0116	0.0232	0.0456	0.0414	0.0257	0.0273	0.0348	0.0423	0.0414	0.0365	0.0779
2010	5	2	2	0	0	0	0	0.0011	0.0011	0.0011	0.0011	0.0044
		0.0120	0.0239	0.0316	0.0326	0.0435	0.0598	0.0511	0.0501	0.0424	0.0392	0.0914
2011	5	2	2	0	0	0	0	0.0000	0.0000	0.0045	0.0135	0.0090
		0.0067	0.0336	0.0090	0.0224	0.0269	0.0426	0.0448	0.0538	0.0336	0.0404	0.1457



2012	5	2	2	0	0	0	0	0.0000	0.0000	0.0000	0.0000	0.0035
								0.0318	0.0212	0.0459	0.0141	0.0353
2013	5	2	2	0	0	0	0	0.0021	0.0000	0.0021	0.0000	0.0083
								0.0062	0.0248	0.0413	0.0331	0.0393
2014	5	2	2	0	0	0	0	0.0000	0.0000	0.0038	0.0038	0.0038
								0.0077	0.0268	0.0153	0.0460	0.0307
2015	5	2	2	0	0	0	0	0.0000	0.0024	0.0024	0.0073	0.0342
								0.0293	0.0465	0.0538	0.0318	0.0465
2016	5	2	2	0	0	0	0	0.0000	0.0000	0.0065	0.0049	0.0016
								0.0081	0.0097	0.0097	0.0097	0.0227
2017	5	2	2	0	0	0	0	0.0000	0.0000	0.0028	0.0028	0.0181
								0.0056	0.0070	0.0028	0.0056	0.0070
2018	5	2	2	0	0	0	0	0.0000	0.0045	0.0067	0.0112	0.0078
								0.0112	0.0157	0.0347	0.0168	0.0202
2019	5	2	2	0	0	0	0	0.0026	0.0026	0.0105	0.0039	0.0092
								0.0211	0.0079	0.0105	0.0105	0.0171

#Tanner crab bycatch Male (male and female combined compositons are normalized to be 1)

#Year Season Fleet Sex Type Shell Maturity Nsamp DataVec

1991	5	3	1	0.000	0	0	50	0.0026	0.0049	0.0029	0.0042	0.0052
								0.0042	0.0104	0.0143	0.0146	0.0110
								0.0211	0.0201	0.0169	0.0249	
1992	5	3	1	0.000	0	0	48.25	0.0000	0.0000	0.0010	0.0031	0.0114
								0.0166	0.0259	0.0238	0.0259	0.0301
								0.0104	0.0135	0.0073	0.0166	
1993	5	3	1	0.000	0	0	24.85	0.0000	0.0000	0.0000	0.0000	0.0040
								0.0020	0.0261	0.0483	0.0584	0.0664
								0.0302	0.0141	0.0101	0.0221	
2013	5	3	1	0.000	0	0	40.7	0.0000	0.0012	0.0000	0.0000	0.0000
								0.0086	0.0074	0.0135	0.0184	0.0393
								0.0123	0.0098	0.0135	0.0270	
2014	5	3	1	0.000	0	0	31.85	0.0000	0.0000	0.0016	0.0000	0.0078
								0.0078	0.0126	0.0188	0.0157	0.0314
								0.0345	0.0251	0.0173	0.0424	
2015	5	3	1	0.000	0	0	50	0.0017	0.0038	0.0017	0.0024	0.0180
								0.0246	0.0176	0.0114	0.0152	0.0201
								0.0135	0.0142	0.0149	0.0211	

#Tanner crab bycatch female

#Year Season Fleet Sex Type Shell Maturity Nsamp DataVec

1991	5	3	2	0	0	0	0	0.0052	0.0107	0.0097	0.0103	0.0243
		0.0331	0.0567	0.0463	0.0839	0.1160	0.1134	0.0956	0.0548	0.0269	0.0188	0.0071
1992	5	3	2	0	0	0	0	0.0000	0.0000	0.0011	0.0062	0.0228
		0.0456	0.0818	0.0933	0.0870	0.0539	0.0777	0.0995	0.0653	0.0404	0.0228	0.0124
1993	5	3	2	0	0	0	0	0.0000	0.0000	0.0000	0.0000	0.0040
		0.0342	0.0825	0.1127	0.0805	0.0362	0.0403	0.0403	0.0564	0.0262	0.0121	0.0081
2013	5	3	2	0	0	0	0	0.0000	0.0000	0.0000	0.0000	0.0000
		0.0221	0.0504	0.1806	0.1437	0.0774	0.0467	0.0553	0.0368	0.0651	0.0234	0.0307
2014	5	3	2	0	0	0	0	0.0000	0.0000	0.0016	0.0031	0.0110
		0.0220	0.0471	0.0550	0.1428	0.1586	0.0581	0.0267	0.0220	0.0110	0.0173	0.0220
2015	5	3	2	0	0	0	0	0.0004	0.0013	0.0028	0.0052	0.0239
		0.0346	0.0637	0.1032	0.1440	0.1115	0.0921	0.0689	0.0374	0.0201	0.0170	0.0228
# Fixed gear	crab	bycatch										
#Year	season	Fleet	Sex	Type	Male	Shell	Maturity		Nsamp	Data	Vec	
1996	5	4	1	0	0	0	39	0.0026	0.0013	0.0066	0.0053	0.0026
		0.0053	0.0132	0.0132	0.0079	0.0146	0.0146	0.0079	0.0146	0.0132	0.0106	0.0146
		0.0106	0.0066	0.0066	0.0238							
1997	5	4	1	0	0	0	50	0.0000	0.0000	0.0024	0.0024	0.0134
		0.0284	0.0504	0.0686	0.0654	0.0607	0.0496	0.0315	0.0347	0.0418	0.0315	0.0221
		0.0362	0.0441	0.0528	0.1560							
1998	5	4	1	0	0	0	50	0.0000	0.0000	0.0000	0.0000	0.0000
		0.0019	0.0019	0.0039	0.0077	0.0125	0.0251	0.0367	0.0521	0.0869	0.0849	0.1052
		0.0840	0.0772	0.0666	0.1564							
1999	5	4	1	0	0	0	50	0.0031	0.0006	0.0019	0.0000	0.0000
		0.0000	0.0000	0.0000	0.0000	0.0025	0.0094	0.0218	0.0524	0.0868	0.1142	0.1255
		0.1242	0.0980	0.0674	0.1311							
2000	5	4	1	0	0	0	44.2	0.0000	0.0000	0.0000	0.0000	0.0000
		0.0000	0.0000	0.0085	0.0169	0.0321	0.0271	0.0761	0.0508	0.0575	0.0457	0.0694
		0.0558	0.0541	0.0474	0.1151							
2001	5	4	1	0	0	0	50	0.0000	0.0002	0.0006	0.0004	0.0016
		0.0044	0.0074	0.0111	0.0201	0.0221	0.0239	0.0233	0.0257	0.0298	0.0340	0.0513
		0.0652	0.0638	0.0547	0.1456							
2002	5	4	1	0	0	0	50	0.0000	0.0000	0.0000	0.0003	0.0009
		0.0017	0.0003	0.0020	0.0049	0.0111	0.0151	0.0220	0.0305	0.0365	0.0520	0.0582
		0.0722	0.0748	0.0854	0.2880							
2003	5	4	1	0	0	0	50	0.0011	0.0000	0.0032	0.0117	0.0149
		0.0171	0.0235	0.0107	0.0075	0.0117	0.0128	0.0299	0.0309	0.0421	0.0597	0.0645



2018	5	4	1	0	0	0	50	0.0009	0.0021	0.0040	0.0081	0.0045
								0.0126	0.0241	0.0396	0.0406	0.0475
								0.0390	0.0258	0.0204	0.0206	0.0207
								0.0153	0.0141	0.0164	0.0507	0.0181
2019	5	4	1	0	0	0	43.1	0.0000	0.0023	0.0046	0.0104	0.0186
								0.0197	0.0255	0.0209	0.0209	0.0197
								0.0070	0.0139	0.0139	0.0139	0.0058
								0.0058	0.0012	0.0000	0.0046	0.0035

# Fixed gear	crab	bycatch	female									
#Year	Season	Fleet	Sex	Type	Shell	Maturity	Nsamp	Data	Vec			
# ERROR CHECK												
1996	5	4	2	0	0	0	0	0.0066	0.0013	0.0053	0.0040	0.0159
								0.0079	0.0238	0.0423	0.0556	0.0860
								0.1270	0.1230	0.0847	0.0741	0.0556
								0.0913	0.0000	0.0000	0.0008	0.0008
1997	5	4	2	0	0	0	0	0.0126	0.0299	0.0260	0.0339	0.0252
								0.0165	0.0126	0.0071	0.0071	0.0079
								0.0229	0.0000	0.0000	0.0010	0.0000
1998	5	4	2	0	0	0	0	0.0000	0.0000	0.0010	0.0000	0.0000
								0.0000	0.0068	0.0251	0.0309	0.0193
								0.0203	0.0097	0.0058	0.0106	0.0174
								0.0502	0.0000	0.0000	0.0000	0.0000
1999	5	4	2	0	0	0	0	0.0000	0.0000	0.0000	0.0000	0.0000
								0.0000	0.0000	0.0000	0.0031	0.0075
								0.0131	0.0194	0.0256	0.0237	0.0137
								0.0549	0.0000	0.0000	0.0000	0.0000
2000	5	4	2	0	0	0	0	0.0017	0.0017	0.0102	0.0152	0.0237
								0.0508	0.0440	0.0423	0.0321	0.0321
								0.0897	0.0004	0.0002	0.0000	0.0016
								0.0028	0.0066	0.0127	0.0195	0.0177
								0.0205	0.0441	0.0787	0.0678	0.0380
								0.0266	0.0000	0.0003	0.0009	0.0000
2002	5	4	2	0	0	0	0	0.0006	0.0000	0.0029	0.0060	0.0106
								0.0086	0.0226	0.0340	0.0348	0.0354
								0.0876	0.0011	0.0005	0.0011	0.0101
								0.0197	0.0155	0.0096	0.0069	0.0149
								0.0240	0.0331	0.0336	0.0341	0.0443
								0.0427	0.0005	0.0005	0.0023	0.0032
2004	5	4	2	0	0	0	0	0.0114	0.0173	0.0328	0.0292	0.0282
								0.0474	0.0483	0.0456	0.0428	0.0374
								0.0811	0.0000	0.0000	0.0000	0.0005
								0.0005	0.0023	0.0056	0.0149	0.0322
								0.0503	0.0499	0.0517	0.0718	0.0555
								0.0499	0.0517	0.0718	0.0555	0.0499
								0.1174	0.0000	0.0000	0.0000	0.0000
2006	5	4	2	0	0	0	0	0.0016	0.0122	0.0371	0.0736	0.1128
								0.1053	0.0969	0.0667	0.0492	0.0392
								0.0979	0.0000	0.0012	0.0012	0.0012
								0.0025	0.0074	0.0099	0.0321	0.0432
								0.0827	0.1173	0.1086	0.0704	0.0420
								0.0222	0.0000	0.0000	0.0000	0.0000
2008	5	4	2	0	0	0	0	0.0043	0.0120	0.0198	0.0438	0.0335
								0.0576	0.0653	0.0730	0.0490	0.0301
								0.0644	0.0000	0.0000	0.0000	0.0000
								0.0000	0.0028	0.0147	0.0184	0.0220
								0.0294	0.0340	0.0312	0.0487	0.0395
								0.0239	0.0000	0.0000	0.0000	0.0036
								0.0036	0.0109	0.0201	0.0657	0.0657
								0.0912	0.1058	0.1077	0.0620	0.0584
								0.1241	0.0000	0.0025	0.0008	0.0067
								0.0076	0.0176	0.0202	0.0336	0.0579
								0.0663	0.0999	0.0907	0.0739	0.0638
								0.0428	0.0000	0.0000	0.0010	0.0027
								0.0020	0.0000	0.0000	0.0010	0.0027
								0.0020	0.0000	0.0000	0.0010	0.0020

2013	5	4	2	0	0	0	0	0.0104	0.0215	0.0262	0.0339	0.0346	0.0339	0.0571	0.0668	0.0648	0.0658	0.1236					
								0.0056	0.0108	0.0224	0.0266	0.0243	0.0245	0.0249	0.0316	0.0354	0.0272	0.0251	0.0241	0.0296	0.0412	0.0334	0.0853
2014	5	4	2	0	0	0	0	0.0023	0.0061	0.0049	0.0014	0.0042	0.0056	0.0084	0.0229	0.0422	0.0537	0.0497	0.0502	0.0511	0.0560	0.0597	0.1624
2015	5	4	2	0	0	0	0	0.0002	0.0002	0.0002	0.0045	0.0072	0.0132	0.0228	0.0512	0.0745	0.0879	0.1082	0.1064	0.0767	0.0557	0.0586	0.1216
2016	5	4	2	0	0	0	0	0.0037	0.0028	0.0044	0.0162	0.0245	0.0208	0.0231	0.0370	0.0499	0.0695	0.0931	0.0845	0.0640	0.0464	0.0342	0.0815
2017	5	4	2	0	0	0	0	0.0007	0.0007	0.0021	0.0127	0.0155	0.0261	0.0184	0.0184	0.0240	0.0382	0.0615	0.0912	0.0876	0.1110	0.0671	0.1272
2018	5	4	2	0	0	0	0	0.0006	0.0040	0.0026	0.0049	0.0066	0.0164	0.0349	0.0621	0.0592	0.0605	0.0573	0.0711	0.0654	0.0507	0.0366	0.0417
2019	5	4	2	0	0	0	0	0.0000	0.0000	0.0012	0.0104	0.0174	0.0313	0.0290	0.0406	0.0789	0.0824	0.0789	0.0719	0.0638	0.0708	0.0650	0.1462

#NMFS		males	combined																								
#Year	Season	Fleet	Sex	Type	Shell	Maturity		Nsamp	Data	Vec																	
1975	1	5	1	0.000	0	0	200	0.0222	0.0411	0.0299	0.0379	0.0342	0.0299	0.0309	0.0246	0.0264	0.0314	0.0268	0.0292	0.0284	0.0273	0.0244	0.0270	0.0183	0.0134	0.0097	0.0113
1976	1	5	1	0.000	0	0	200	0.0025	0.0127	0.0268	0.0503	0.0623	0.0522	0.0559	0.0449	0.0392	0.0329	0.0409	0.0438	0.0369	0.0392	0.0335	0.0221	0.0236	0.0154	0.0070	0.0077
1977	1	5	1	0.000	0	0	200	0.0040	0.0043	0.0065	0.0102	0.0199	0.0376	0.0453	0.0441	0.0414	0.0450	0.0409	0.0409	0.0311	0.0324	0.0322	0.0259	0.0166	0.0140	0.0084	0.0121
1978	1	5	1	0.000	0	0	200	0.0043	0.0120	0.0136	0.0240	0.0172	0.0191	0.0178	0.0279	0.0296	0.0297	0.0300	0.0304	0.0291	0.0367	0.0346	0.0283	0.0260	0.0173	0.0108	0.0091
1979	1	5	1	0.000	0	0	200	0.0206	0.0154	0.0103	0.0123	0.0144	0.0163	0.0137	0.0155	0.0164	0.0157	0.0235	0.0338	0.0333	0.0432	0.0415	0.0378	0.0359	0.0298	0.0136	0.0235
1980	1	5	1	0.000	0	0	200	0.0067	0.0133	0.0376	0.0287	0.0295	0.0296	0.0265	0.0262	0.0224	0.0192	0.0208	0.0165	0.0231	0.0251	0.0264	0.0378	0.0266	0.0268	0.0216	0.0357
1981	1	5	1	0.000	0	0	200	0.0160	0.0113	0.0182	0.0240	0.0366	0.0362	0.0331	0.0367	0.0291	0.0356	0.0261	0.0285	0.0194	0.0221	0.0156	0.0145	0.0112	0.0106	0.0085	0.0176
1982	1	5	1	0.000	0	0	200	0.0792	0.0811	0.0682	0.0287	0.0240	0.0310	0.0353	0.0287	0.0197	0.0171	0.0198	0.0141	0.0131	0.0079	0.0066	0.0043	0.0039	0.0005	0.0004	0.0018
1983	1	5	1	0.000	0	0	200	0.0325	0.0356	0.0497	0.0665	0.0801															

			0.0783	0.0598	0.0468	0.0402	0.0398	0.0320	0.0309	0.0190	0.0119	0.0107	0.0037
			0.0025	0.0012	0.0000	0.0000							
1984	1	5	1	0.000	0	0	200	0.0161	0.0626	0.1229	0.1327	0.0682	
				0.0389	0.0206	0.0202	0.0208	0.0154	0.0119	0.0072	0.0063	0.0050	0.0065
				0.0009	0.0009	0.0001	0.0003						
1985	1	5	1	0.000	0	0	200	0.0026	0.0128	0.0244	0.0395	0.0589	
				0.0582	0.0424	0.0403	0.0602	0.0614	0.0513	0.0523	0.0497	0.0418	0.0279
				0.0018	0.0051	0.0042	0.0000						
1986	1	5	1	0.000	0	0	200	0.0112	0.0179	0.0248	0.0201	0.0232	
				0.0156	0.0408	0.0400	0.0559	0.0485	0.0675	0.0734	0.0700	0.0788	0.0563
				0.0275	0.0073	0.0029	0.0023						
1987	1	5	1	0.000	0	0	200	0.0012	0.0071	0.0340	0.0546	0.0469	
				0.0317	0.0290	0.0291	0.0310	0.0253	0.0332	0.0270	0.0363	0.0345	0.0290
				0.0183	0.0154	0.0038	0.0039						
1988	1	5	1	0.000	0	0	200	0.0013	0.0013	0.0066	0.0110	0.0133	
				0.0215	0.0469	0.0430	0.0405	0.0374	0.0262	0.0308	0.0210	0.0371	0.0331
				0.0368	0.0268	0.0094	0.0093						
1989	1	5	1	0.000	0	0	200	0.0017	0.0000	0.0009	0.0024	0.0149	
				0.0348	0.0184	0.0376	0.0232	0.0412	0.0288	0.0253	0.0450	0.0523	0.0535
				0.0483	0.0466	0.0283	0.0278						
1990	1	5	1	0.000	0	0	200	0.0013	0.0106	0.0151	0.0348	0.0329	
				0.0094	0.0080	0.0084	0.0182	0.0296	0.0219	0.0298	0.0341	0.0401	0.0369
				0.0299	0.0344	0.0196	0.0342						
1991	1	5	1	0.000	0	0	200	0.0011	0.0090	0.0224	0.0168	0.0265	
				0.0217	0.0137	0.0274	0.0221	0.0172	0.0053	0.0198	0.0347	0.0364	0.0588
				0.0658	0.0482	0.0369	0.0757						
1992	1	5	1	0.000	0	0	200	0.0010	0.0000	0.0020	0.0127	0.0252	
				0.0355	0.0552	0.0528	0.0382	0.0399	0.0291	0.0378	0.0348	0.0280	0.0234
				0.0219	0.0307	0.0169	0.0496						
1993	1	5	1	0.000	0	0	200	0.0021	0.0110	0.0137	0.0105	0.0095	
				0.0157	0.0142	0.0235	0.0309	0.0443	0.0417	0.0627	0.0479	0.0390	0.0371
				0.0288	0.0298	0.0242	0.0411						
1994	1	5	1	0.000	0	0	163.75	0.0016	0.0000	0.0031	0.0237	0.0235	
				0.0152	0.0124	0.0173	0.0213	0.0354	0.0412	0.0403	0.0627	0.0907	0.0474
				0.0468	0.0327	0.0229	0.0504						
1995	1	5	1	0.000	0	0	200	0.0283	0.0683	0.0557	0.0220	0.0110	
				0.0169	0.0222	0.0255	0.0275	0.0305	0.0263	0.0268	0.0343	0.0402	0.0490
				0.0323	0.0238	0.0108	0.0262						
1996	1	5	1	0.000	0	0	200	0.0278	0.0135	0.0298	0.0529	0.0632	
				0.0594	0.0276	0.0225	0.0117	0.0179	0.0140	0.0150	0.0139	0.0130	0.0218
				0.0190	0.0171	0.0183	0.0252						
1997	1	5	1	0.000	0	0	200	0.0000	0.0036	0.0022	0.0052	0.0127	
				0.0564	0.0943	0.1070	0.0910	0.0515	0.0301	0.0162	0.0149	0.0132	0.0142
													0.0168

1998	1	5	1	0.000	0	0	200	0.0209	0.0174	0.0103	0.0127	0.0120
				0.0101	0.0135	0.0169	0.0226	0.0467	0.0485	0.0523	0.0451	0.0291
				0.0183	0.0153	0.0196	0.0135	0.0080	0.0245			
1999	1	5	1	0.000	0	0	200	0.0583	0.0244	0.0134	0.0104	0.0120
				0.0110	0.0121	0.0148	0.0047	0.0132	0.0182	0.0233	0.0520	0.0536
				0.0700	0.0688	0.0435	0.0303	0.0221	0.0252			
2000	1	5	1	0.000	0	0	200	0.0018	0.0047	0.0195	0.0396	0.0310
				0.0200	0.0228	0.0163	0.0201	0.0147	0.0134	0.0296	0.0294	0.0489
				0.0416	0.0360	0.0343	0.0229	0.0085	0.0196			
2001	1	5	1	0.000	0	0	200	0.0069	0.0050	0.0106	0.0149	0.0156
				0.0421	0.0372	0.0523	0.0346	0.0200	0.0253	0.0166	0.0140	0.0202
				0.0132	0.0112	0.0219	0.0191	0.0192	0.0327			
2002	1	5	1	0.000	0	0	200	0.0534	0.0638	0.0436	0.0272	0.0119
				0.0091	0.0076	0.0106	0.0229	0.0266	0.0347	0.0290	0.0203	0.0252
				0.0170	0.0193	0.0195	0.0222	0.0242	0.0274			
2003	1	5	1	0.000	0	0	200	0.0149	0.0069	0.0142	0.0236	0.0392
				0.0320	0.0301	0.0165	0.0112	0.0143	0.0133	0.0251	0.0236	0.0386
				0.0348	0.0364	0.0254	0.0216	0.0212	0.0666			
2004	1	5	1	0.000	0	0	200	0.0371	0.0289	0.0268	0.0195	0.0187
				0.0187	0.0350	0.0535	0.0436	0.0445	0.0293	0.0238	0.0142	0.0150
				0.0179	0.0232	0.0240	0.0327	0.0232	0.0447			
2005	1	5	1	0.000	0	0	200	0.0353	0.0586	0.0419	0.0160	0.0098
				0.0228	0.0234	0.0215	0.0184	0.0171	0.0219	0.0233	0.0159	0.0189
				0.0125	0.0158	0.0103	0.0155	0.0144	0.0252			
2006	1	5	1	0.000	0	0	200	0.0133	0.0197	0.0173	0.0276	0.0291
				0.0369	0.0210	0.0208	0.0129	0.0188	0.0116	0.0128	0.0236	0.0205
				0.0329	0.0280	0.0271	0.0200	0.0144	0.0246			
2007	1	5	1	0.000	0	0	200	0.0017	0.0025	0.0053	0.0084	0.0196
				0.0271	0.0345	0.0436	0.0386	0.0288	0.0187	0.0233	0.0236	0.0315
				0.0273	0.0288	0.0277	0.0262	0.0229	0.0290			
2008	1	5	1	0.000	0	0	200	0.0000	0.0008	0.0038	0.0068	0.0149
				0.0188	0.0194	0.0239	0.0372	0.0470	0.0453	0.0328	0.0382	0.0317
				0.0249	0.0226	0.0242	0.0236	0.0222	0.0467			
2009	1	5	1	0.000	0	0	200	0.0010	0.0005	0.0037	0.0053	0.0053
				0.0104	0.0096	0.0225	0.0330	0.0301	0.0315	0.0328	0.0363	0.0479
				0.0312	0.0329	0.0198	0.0163	0.0148	0.0169			
2010	1	5	1	0.000	0	0	200	0.0000	0.0033	0.0080	0.0094	0.0077
				0.0054	0.0161	0.0134	0.0130	0.0153	0.0270	0.0363	0.0302	0.0325
				0.0367	0.0348	0.0423	0.0262	0.0145	0.0200			
2011	1	5	1	0.000	0	0	200	0.0036	0.0044	0.0125	0.0204	0.0169
				0.0138	0.0168	0.0151	0.0182	0.0132	0.0181	0.0203	0.0161	0.0295
				0.0275	0.0257	0.0242	0.0204	0.0115	0.0165			

2012	1	5	1	0.000	0	0	200	0.0025	0.0040	0.0120	0.0159	0.0128
				0.0227	0.0336	0.0247	0.0174	0.0174	0.0153	0.0196	0.0217	0.0264
				0.0232	0.0281	0.0132	0.0434					
2013	1	5	1	0.000	0	0	200	0.0008	0.0025	0.0123	0.0145	0.0101
				0.0174	0.0134	0.0235	0.0280	0.0261	0.0323	0.0348	0.0303	0.0319
				0.0340	0.0431	0.0395	0.0749					
2014	1	5	1	0.000	0	0	200	0.0000	0.0005	0.0026	0.0030	0.0160
				0.0313	0.0437	0.0348	0.0313	0.0192	0.0231	0.0326	0.0336	0.0309
				0.0224	0.0189	0.0180	0.0439					
2015	1	5	1	0.000	0	0	200	0.0105	0.0207	0.0103	0.0093	0.0047
				0.0110	0.0158	0.0149	0.0244	0.0187	0.0285	0.0203	0.0235	0.0318
				0.0313	0.0282	0.0278	0.0796					
2016	1	5	1	0.000	0	0	200	0.0066	0.0009	0.0026	0.0032	0.0041
				0.0043	0.0034	0.0083	0.0069	0.0129	0.0085	0.0145	0.0127	0.0254
				0.0241	0.0389	0.0324	0.0709					
2017	1	5	1	0.000	0	0	200	0.0032	0.0011	0.0029	0.0095	0.0243
				0.0199	0.0135	0.0068	0.0083	0.0077	0.0086	0.0134	0.0064	0.0234
				0.0233	0.0363	0.0351	0.0868					
2018	1	5	1	0.000	0	0	161	0.0051	0.0173	0.0173	0.0153	0.0093
				0.0161	0.0144	0.0174	0.0367	0.0160	0.0334	0.0210	0.0033	0.0160
				0.0262	0.0321	0.0272	0.0746					
2019	1	5	1	0.000	0	0	143	0.0017	0.0036	0.0106	0.0071	0.0071
				0.0314	0.0157	0.0244	0.0231	0.0336	0.0299	0.0436	0.0424	0.0363
				0.0229	0.0230	0.0160	0.0602					

#Year	Season	Fleet	Sex	Type	Shell	Maturity	Nsamp	Data	Vec			
1975	1	5	2	0.000	0	0	0	0.0331	0.0401	0.0481	0.0494	0.0564
				0.0439	0.0444	0.0454	0.0326	0.0289	0.0162	0.0158	0.0116	0.0035
				0.0029	0.0092	0.0313	0.0563	0.0688				
1976	1	5	2	0.000	0	0	0	0.0029	0.0092	0.0313	0.0563	0.0688
				0.0628	0.0494	0.0269	0.0121	0.0137	0.0066	0.0049	0.0023	0.0015
				0.0003	0.0011							
1977	1	5	2	0.000	0	0	0	0.0026	0.0068	0.0079	0.0193	0.0337
				0.0701	0.0808	0.0715	0.0453	0.0435	0.0415	0.0316	0.0151	0.0100
				0.0033	0.0046							
1978	1	5	2	0.000	0	0	0	0.0060	0.0111	0.0187	0.0201	0.0233
				0.0418	0.0920	0.1212	0.0791	0.0440	0.0301	0.0267	0.0176	0.0089
				0.0045	0.0075							
1979	1	5	2	0.000	0	0	0	0.0286	0.0154	0.0121	0.0147	0.0148
				0.0230	0.0381	0.0734	0.0922	0.0876	0.0565	0.0336	0.0215	0.0123
				0.0043	0.0057							
1980	1	5	2	0.000	0	0	0	0.0048	0.0219	0.0322	0.0292	0.0597
				0.0820	0.0487	0.0581	0.0540	0.0424	0.0315	0.0130	0.0110	0.0059
				0.0035	0.0020							
1981	1	5	2	0.000	0	0	0	0.0152	0.0113	0.0151	0.0190	0.0366
				0.0456	0.0443	0.0472	0.0600	0.0774	0.0804	0.0510	0.0252	0.0143
				0.0028	0.0042							
1982	1	5	2	0.000	0	0	0	0.0536	0.0954	0.0603	0.0378	0.0423
				0.0482	0.0398	0.0232	0.0190	0.0257	0.0281	0.0203	0.0114	0.0063
				0.0024	0.0009							



1983	1	5	2	0.000	0	0	0	0.0174	0.0383	0.0475	0.0629	0.0647
				0.0398	0.0341	0.0152	0.0107	0.0042	0.0090	0.0056	0.0061	0.0022
1984	1	5	2	0.000	0	0	0	0.0174	0.0585	0.1229	0.1105	0.0647
				0.0325	0.0159	0.0119	0.0038	0.0017	0.0000	0.0004	0.0001	0.0002
1985	1	5	2	0.000	0	0	0	0.0009	0.0155	0.0377	0.0521	0.0643
				0.0555	0.0516	0.0397	0.0161	0.0068	0.0000	0.0000	0.0015	0.0000
1986	1	5	2	0.000	0	0	0	0.0124	0.0224	0.0355	0.0274	0.0263
				0.0313	0.0362	0.0388	0.0274	0.0113	0.0072	0.0008	0.0000	0.0008
1987	1	5	2	0.000	0	0	0	0.0013	0.0124	0.0525	0.0918	0.0761
				0.0462	0.0445	0.0569	0.0414	0.0292	0.0179	0.0079	0.0018	0.0004
1988	1	5	2	0.000	0	0	0	0.0006	0.0076	0.0064	0.0062	0.0139
				0.0695	0.0910	0.0979	0.0697	0.0600	0.0407	0.0184	0.0077	0.0077
1989	1	5	2	0.000	0	0	0	0.0017	0.0000	0.0017	0.0082	0.0310
				0.0740	0.0646	0.0692	0.0531	0.0376	0.0315	0.0194	0.0064	0.0041
1990	1	5	2	0.000	0	0	0	0.0041	0.0052	0.0235	0.0513	0.0525
				0.0071	0.0256	0.0601	0.0732	0.0708	0.0633	0.0410	0.0215	0.0062
1991	1	5	2	0.000	0	0	0	0.0042	0.0115	0.0196	0.0320	0.0218
				0.0344	0.0343	0.0310	0.0366	0.0329	0.0281	0.0431	0.0232	0.0110
1992	1	5	2	0.000	0	0	0	0.0000	0.0053	0.0074	0.0197	0.0364
				0.0414	0.0625	0.0448	0.0353	0.0273	0.0450	0.0407	0.0265	0.0212
1993	1	5	2	0.000	0	0	0	0.0066	0.0080	0.0175	0.0085	0.0131
				0.0248	0.0437	0.0647	0.0639	0.0269	0.0300	0.0268	0.0271	0.0445
1994	1	5	2	0.000	0	0	0	0.0000	0.0016	0.0044	0.0030	0.0169
				0.0092	0.0124	0.0213	0.0431	0.0416	0.0362	0.0280	0.0395	0.0469
1995	1	5	2	0.000	0	0	0	0.0294	0.0482	0.0316	0.0145	0.0139
				0.0182	0.0163	0.0254	0.0234	0.0334	0.0272	0.0234	0.0240	0.0145
1996	1	5	2	0.000	0	0	0	0.0260	0.0219	0.0436	0.0794	0.0796
				0.0436	0.0226	0.0218	0.0245	0.0202	0.0161	0.0285	0.0244	0.0156
1997	1	5	2	0.000	0	0	0	0.0004	0.0037	0.0016	0.0020	0.0146
				0.0791	0.0969	0.0616	0.0212	0.0137	0.0095	0.0146	0.0143	0.0109
1998	1	5	2	0.000	0	0	0	0.0145	0.0196	0.0101	0.0088	0.0111
				0.0116	0.0303	0.1040	0.1153	0.0594	0.0303	0.0252	0.0225	0.0235
1999	1	5	2	0.000	0	0	0	0.0243	0.0169	0.0125	0.0115	0.0044
				0.0055	0.0093	0.0164	0.0512	0.0800	0.0583	0.0358	0.0340	0.0199
2000	1	5	2	0.000	0	0	0	0.0018	0.0067	0.0269	0.0403	0.0357
				0.0272	0.0255	0.0226	0.0358	0.0524	0.0676	0.0603	0.0419	0.0208
2001	1	5	2	0.000	0	0	0	0.0056	0.0168	0.0195	0.0136	0.0259
				0.0598	0.0779	0.0579	0.0395	0.0398	0.0291	0.0691	0.0560	0.0262
2002	1	5	2	0.000	0	0	0	0.0506	0.0769	0.0485	0.0247	0.0222
				0.0176	0.0225	0.0520	0.0399	0.0296	0.0163	0.0206	0.0205	0.0221
2003	1	5	2	0.000	0	0	0	0.0163	0.0059	0.0143	0.0314	0.0414
				0.0464	0.0239	0.0292	0.0351	0.0533	0.0526	0.0356	0.0219	0.0265
2004	1	5	2	0.000	0	0	0	0.0279	0.0327	0.0194	0.0132	0.0199

2005	1	5	2	0.000	0	0	0	0.0405	0.0561	0.0457	0.0116	0.0099
				0.0336	0.0386	0.0521	0.0567	0.0468	0.0336	0.0383	0.0347	0.0227
				0.0165	0.0246							
2006	1	5	2	0.000	0	0	0	0.0143	0.0139	0.0198	0.0425	0.0615
				0.0462	0.0254	0.0259	0.0481	0.0656	0.0619	0.0415	0.0301	0.0352
				0.0167	0.0167							
2007	1	5	2	0.000	0	0	0	0.0015	0.0023	0.0064	0.0078	0.0155
				0.0356	0.0574	0.0560	0.0325	0.0570	0.0614	0.0641	0.0459	0.0343
				0.0210	0.0323							
2008	1	5	2	0.000	0	0	0	0.0000	0.0027	0.0054	0.0136	0.0116
				0.0167	0.0303	0.0570	0.0724	0.0560	0.0555	0.0562	0.0575	0.0355
				0.0234	0.0216							
2009	1	5	2	0.000	0	0	0	0.0005	0.0019	0.0050	0.0055	0.0081
				0.0122	0.0206	0.0466	0.0656	0.0866	0.0645	0.0603	0.0523	0.0705
				0.0514	0.0470							
2010	1	5	2	0.000	0	0	0	0.0018	0.0006	0.0037	0.0048	0.0069
				0.0116	0.0213	0.0365	0.0565	0.0927	0.0955	0.0700	0.0509	0.0497
				0.0508	0.0545							
2011	1	5	2	0.000	0	0	0	0.0058	0.0085	0.0092	0.0141	0.0284
				0.0310	0.0384	0.0484	0.0299	0.0530	0.0637	0.0905	0.0635	0.0571
				0.0430	0.0710							
2012	1	5	2	0.000	0	0	0	0.0293	0.0180	0.0191	0.0250	0.0281
				0.0461	0.0351	0.0220	0.0331	0.0355	0.0365	0.0461	0.0663	0.0521
				0.0462	0.0633							
2013	1	5	2	0.000	0	0	0	0.0008	0.0027	0.0093	0.0112	0.0067
				0.0125	0.0202	0.0384	0.0429	0.0450	0.0304	0.0302	0.0455	0.0491
				0.0405	0.0786							
2014	1	5	2	0.000	0	0	0	0.0000	0.0000	0.0012	0.0040	0.0091
				0.0258	0.0219	0.0320	0.0499	0.0770	0.0569	0.0456	0.0307	0.0399
				0.0516	0.0859							
2015	1	5	2	0.000	0	0	0	0.0074	0.0129	0.0110	0.0055	0.0120
				0.0114	0.0107	0.0234	0.0408	0.0461	0.0616	0.0668	0.0531	0.0503
				0.0362	0.0819							
2016	1	5	2	0.000	0	0	0	0.0120	0.0019	0.0036	0.0043	0.0026
				0.0051	0.0143	0.0141	0.0390	0.0714	0.0782	0.1023	0.0737	0.0823
				0.0617	0.1158							
2017	1	5	2	0.000	0	0	0	0.0010	0.0028	0.0030	0.0126	0.0258
				0.0248	0.0167	0.0188	0.0214	0.0511	0.0665	0.0804	0.0885	0.0769
				0.0569	0.0973							
2018	1	5	2	0.000	0	0	0	0.0031	0.0109	0.0172	0.0186	0.0094
				0.0198	0.0516	0.0362	0.0421	0.0296	0.0254	0.0652	0.0462	0.0495
				0.0509	0.0773							
2019	1	5	2	0.000	0	0	0	0.0017	0.0105	0.0018	0.0070	0.0070
				0.0140	0.0143	0.0174	0.0312	0.0355	0.0335	0.0279	0.0515	0.0766
				0.0656	0.1276							

#BSFRF	males											
#Year	Season	Fleet	Sex	Type	Shell	Maturity	Nsamp	Data	Vec			
2007	1	6	1	0	0	0	200	0.0045	0.0074	0.0103	0.0155	0.0198
								0.0321	0.0532	0.0491	0.0443	0.0354
								0.0268	0.0231	0.0236	0.0256	0.0223
								0.032	0.0246	0.0218	0.017	0.0278
2008	1	6	1	0	0	0	200	0.0017	0.001	0.0093	0.0119	0.0175
								0.0279	0.0267	0.0348	0.0428	0.0596
								0.0581	0.0455	0.0371	0.0284	0.0218
								0.0211	0.0156	0.0157	0.0202	0.0294
2013	1	6	1	0	0	0	75.75	0	0.0073	0.0145	0.0291	0.0102
								0.0136	0.0205	0.0341	0.0357	0.0458
								0.0448	0.0383	0.042	0.0348	0.0206
								0.0149	0.0337	0.0426	0.0358	0.0986



**Appendix C. Control File for Models 19.3, 19.3c, 19.3d, and 19.3g**

```

## ----- ##
## LEADING PARAMETER CONTROLS ##
## Controls for leading parameter vector (theta) ##
## LEGEND ##
## prior: 0 = uniform, 1 = normal, 2 = lognormal, 3 = beta, 4 = gamma ##
## ----- ##
## ntheta
91
## ----- ##
## ival lb ub phz prior p1 p2 # parameter ##
## ----- ##
0.18 0.15 0.2 -4 2 0.18 0.04 # M
# 0.18 0.15 0.4 4 2 0.18 0.03 # M
0.0 -0.4 0.4 4 1 0.0 0.03 # M
16.5 -10 18 -2 0 -10.0 20.0 # logR0
19.5 -10 25 3 0 10.0 25.0 # logRini, to estimate if NOT initialized at unfished (n68)
16.5 -10 25 1 0 10.0 20.0 #1 # logRbar, to estimate if NOT initialized at unfished #1
72.5 55 100 -4 1 72.5 7.25 # recruitment expected value (males or combined)
0.726149 0.32 1.64 3 0 0.1 5.0 # recruitment scale (variance component) (males or combined)
0.00 -5 5 -4 0 0.0 20.00 # recruitment expected value (females)
0.00 -1.69 0.40 3 0 0.0 20.0 # recruitment scale (variance component) (females)
-0.10536 -10 0.75 -4 0 -10.0 0.75 # ln(sigma_R)
#-0.10 -5 5.0 4 0 -10.0 10.0 # ln(sigma_R)
0.75 0.20 1.00 -2 3 3.0 2.00 # steepness
0.01 0.00 1.00 -3 3 1.01 1.01 # recruitment autocorrelation
# 0.00 -10 4 2 0 10.0 20.00 # Deviation for size-class 1 (normalization class)
1.107962885630 -10 4 9 0 10.0 20.00 # Deviation for size-class 2
0.563229168219 -10 4 9 0 10.0 20.00 # Deviation for size-class 3
0.681928313426 -10 4 9 0 10.0 20.00 # Deviation for size-class 4
0.491057364532 -10 4 9 0 10.0 20.00 # Deviation for size-class 5
0.407911777560 -10 4 9 0 10.0 20.00 # Deviation for size-class 6
0.436516142684 -10 4 9 0 10.0 20.00 # Deviation for size-class 7
0.40612675395550 -10 4 9 0 10.0 20.00 # Deviation for size-class 8
0.436145974880 -10 4 9 0 10.0 20.00 # Deviation for size-class 9
0.40494522852708 -10 4 9 0 10.0 20.00 # Deviation for size-class 10
0.30401970466854 -10 4 9 0 10.0 20.00 # Deviation for size-class 11
0.2973752673022 -10 4 9 0 10.0 20.00 # Deviation for size-class 12
0.1746800712364 -10 4 9 0 10.0 20.00 # Deviation for size-class 13
0.0845298456942 -10 4 9 0 10.0 20.00 # Deviation for size-class 14
0.0107462399193 -10 4 9 0 10.0 20.00 # Deviation for size-class 15
-0.190468322904 -10 4 9 0 10.0 20.00 # Deviation for size-class 16
-0.376312503735 -10 4 9 0 10.0 20.00 # Deviation for size-class 17
-0.699162895473 -10 4 9 0 10.0 20.00 # Deviation for size-class 18
-1.15881771530 -10 4 9 0 10.0 20.00 # Deviation for size-class 19
-1.17311583316 -10 4 9 0 10.0 20.00 # Deviation for size-class 20
-100.00 -101 5 -2 0 10.0 20.00 # Deviation for size-class 1
-100.00 -101 5 -2 0 10.0 20.00 # Deviation for size-class 2
-100.00 -101 5 -2 0 10.0 20.00 # Deviation for size-class 3
-100.00 -101 5 -2 0 10.0 20.00 # Deviation for size-class 4
-100.00 -101 5 -2 0 10.0 20.00 # Deviation for size-class 5

```

-100.00	-101	5	-2	0	10.0	20.00	# Deviation for size-class 6
-100.00	-101	5	-2	0	10.0	20.00	# Deviation for size-class 7
-100.00	-101	5	-2	0	10.0	20.00	# Deviation for size-class 8
-100.00	-101	5	-2	0	10.0	20.00	# Deviation for size-class 9
-100.00	-101	5	-2	0	10.0	20.00	# Deviation for size-class 10
-100.00	-101	5	-2	0	10.0	20.00	# Deviation for size-class 11
-100.00	-101	5	-2	0	10.0	20.00	# Deviation for size-class 12
-100.00	-101	5	-2	0	10.0	20.00	# Deviation for size-class 13
-100.00	-101	5	-2	0	10.0	20.00	# Deviation for size-class 14
-100.00	-101	5	-2	0	10.0	20.00	# Deviation for size-class 15
-100.00	-101	5	-2	0	10.0	20.00	# Deviation for size-class 16
-100.00	-101	5	-2	0	10.0	20.00	# Deviation for size-class 17
-100.00	-101	5	-2	0	10.0	20.00	# Deviation for size-class 18
-100.00	-101	5	-2	0	10.0	20.00	# Deviation for size-class 19
-100.00	-101	5	-2	0	10.0	20.00	# Deviation for size-class 20
0.425704202053	-10	4	9	0	10.0	20.00	# Deviation for size-class 1
2.268408592660	-10	4	9	0	10.0	20.00	# Deviation for size-class 2
1.810451373080	-10	4	9	0	10.0	20.00	# Deviation for size-class 3
1.370357251111	-10	4	9	0	10.0	20.00	# Deviation for size-class 4
1.158258087990	-10	4	9	0	10.0	20.00	# Deviation for size-class 5
0.596196784439	-10	4	9	0	10.0	20.00	# Deviation for size-class 6
0.225756761257	-10	4	9	0	10.0	20.00	# Deviation for size-class 7
-0.0247857565368	-10	4	9	0	10.0	20.00	# Deviation for size-class 8
-0.214045895269	-10	4	9	0	10.0	20.00	# Deviation for size-class 9
-0.560539577780	-10	4	9	0	10.0	20.00	# Deviation for size-class 10
-0.974218300021	-10	4	9	0	10.0	20.00	# Deviation for size-class 11
-1.24580072031	-10	4	9	0	10.0	20.00	# Deviation for size-class 12
-1.49292897450	-10	4	9	0	10.0	20.00	# Deviation for size-class 13
-1.94135821253	-10	4	9	0	10.0	20.00	# Deviation for size-class 14
-2.05101560679	-10	4	9	0	10.0	20.00	# Deviation for size-class 15
-1.94956606430	-10	4	9	0	10.0	20.00	# Deviation for size-class 16
-100.00	-101	5	-2	0	10.0	20.00	# Deviation for size-class 17
-100.00	-101	5	-2	0	10.0	20.00	# Deviation for size-class 18
-100.00	-101	5	-2	0	10.0	20.00	# Deviation for size-class 19
-100.00	-101	5	-2	0	10.0	20.00	# Deviation for size-class 20
-100.00	-101	5	-2	0	10.0	20.00	# Deviation for size-class 1
-100.00	-101	5	-2	0	10.0	20.00	# Deviation for size-class 2
-100.00	-101	5	-2	0	10.0	20.00	# Deviation for size-class 3
-100.00	-101	5	-2	0	10.0	20.00	# Deviation for size-class 4
-100.00	-101	5	-2	0	10.0	20.00	# Deviation for size-class 5
-100.00	-101	5	-2	0	10.0	20.00	# Deviation for size-class 6
-100.00	-101	5	-2	0	10.0	20.00	# Deviation for size-class 7
-100.00	-101	5	-2	0	10.0	20.00	# Deviation for size-class 8
-100.00	-101	5	-2	0	10.0	20.00	# Deviation for size-class 9
-100.00	-101	5	-2	0	10.0	20.00	# Deviation for size-class 10
-100.00	-101	5	-2	0	10.0	20.00	# Deviation for size-class 11
-100.00	-101	5	-2	0	10.0	20.00	# Deviation for size-class 12
-100.00	-101	5	-2	0	10.0	20.00	# Deviation for size-class 13
-100.00	-101	5	-2	0	10.0	20.00	# Deviation for size-class 14
-100.00	-101	5	-2	0	10.0	20.00	# Deviation for size-class 15
-100.00	-101	5	-2	0	10.0	20.00	# Deviation for size-class 16
-100.00	-101	5	-2	0	10.0	20.00	# Deviation for size-class 17

```

-100.00 -101 5 -2 0 10.0 20.00 # Deviation for size-class 18
-100.00 -101 5 -2 0 10.0 20.00 # Deviation for size-class 19
-100.00 -101 5 -2 0 10.0 20.00 # Deviation for size-class 20

# weight-at-length input method (1 = allometry [w_1 = a*l^b], 2 = vector by sex)
2
## Males
0.000224781 0.000281351 0.000346923 0.000422209 0.000507927 0.000604802
0.000713564 0.00083495 0.0009697 0.00111856 0.00128229 0.00146163
0.00165736 0.00187023 0.00210101 0.00235048 0.00261942 0.00290861
0.00321882 0.0039059
## Females

0.0002151 0.00026898 0.00033137 0.00040294 0.00048437 0.00062711 0.0007216
0.00082452 0.00093615 0.00105678 0.00118669 0.00132613 0.00147539
0.00163473 0.00180441 0.00218315 0.00218315 0.00218315 0.00218315
0.0021831
# Proportion mature by sex
0 0 0 0 0 0 0 0 0 0 0 1 1
1 1 1 1 1 1 1 1 0 0 0 1 1
0 0 0 0 0 1 1 1 1 1 1 1 1
1 1 1 1 1 1 1 1
# Proportion legal by sex
0 0 0 0 0 0 0 0 0 0 0 1 1
1 1 1 1 1 1 1 1
0 0 0 0 0 0 0 0 0 0 0 0 0
0 0 0 0 0 0 0 0
## ----- ##
## ----- ##
## GROWTH PARAMETER CONTROLS ##
## Two lines for each parameter if split sex, one line if not ##
## ----- ##
# Use growth transition matrix option (1=read in growth-increment matrix; 2=read in size-transition; 3=gamma
distribution for size-increment; 4=gamma distribution for size after increment)
3
# growth increment model (1=alpha/beta; 2=estimated by size-class;3=pre-specified/emprical)
3
# molt probability function (0=pre-specified; 1=flat;2=declining logistic)
2
# Maximum size-class for recruitment(males then females)
7 5
## number of size-increment periods
1 3
## Year(s) size-incremnt period changes (blank if no changes)
1983 1994
## number of molt periods
2 2
## Year(s) molt period changes (blank if no changes)
1980 1980

```

## Beta parameters are relative (1=Yes;0=no)

##	ival	lb	ub	phz	prior	p1	p2	# parameter	##
1									
16.5	0	20	-33	0	0	999		# Males	
16.5	0	20	-33	0	0	999		# Males	
16.4	0	20	-33	0	0	999		# Males	
16.3	0	20	-33	0	0	999		# Males	
16.3	0	20	-33	0	0	999		# Males	
16.2	0	20	-33	0	0	999		# Males	
16.2	0	20	-33	0	0	999		# Males	
16.1	0	20	-33	0	0	999		# Males	
16.1	0	20	-33	0	0	999		# Males	
16	0	20	-33	0	0	999		# Males	
16	0	20	-33	0	0	999		# Males	
15.9	0	20	-33	0	0	999		# Males	
15.8	0	20	-33	0	0	999		# Males	
15.8	0	20	-33	0	0	999		# Males	
15.7	0	20	-33	0	0	999		# Males	
15.7	0	20	-33	0	0	999		# Males	
15.6	0	20	-33	0	0	999		# Males	
15.6	0	20	-33	0	0	999		# Males	
15.5	0	20	-33	0	0	999		# Males	
15.5	0	20	-33	0	0	999		# Males	
#1.38403	0.5	3.7	7	0	0	999		# Males (beta)	
1.0	0.5	3.0	6	0	0	999		# Males (beta)	
13.8	0	20	-33	0	0	999		# Females	
12.2	0	20	-33	0	0	999		# Females	
10.5	0	20	-33	0	0	999		# Females	
8.4	0	20	-33	0	0	999		# Females	
7.5	0	20	-33	0	0	999		# Females	
7	0	20	-33	0	0	999		# Females	
6.6	0	20	-33	0	0	999		# Females	
6.1	0	20	-33	0	0	999		# Females	
5.6	0	20	-33	0	0	999		# Females	
5.1	0	20	-33	0	0	999		# Females	
4.6	0	20	-33	0	0	999		# Females	
4.1	0	20	-33	0	0	999		# Females	
3.6	0	20	-33	0	0	999		# Females	
3.2	0	20	-33	0	0	999		# Females	
2.7	0	20	-33	0	0	999		# Females	
2.2	0	20	-33	0	0	999		# Females	
1.7	0	20	-33	0	0	999		# Females	
1.2	0	20	-33	0	0	999		# Females	
0.7	0	20	-33	0	0	999		# Females	
0.4	0	20	-33	0	0	999		# Females	
#1.38403	0.5	3.0	7	0	0	999		# Females (beta)	
1.5	0.5	3.0	6	0	0	999		# Females (beta)	
15.4	0	20	-33	0	0	999		# Females	
13.8	0	20	-33	0	0	999		# Females	
12.2	0	20	-33	0	0	999		# Females	

```

10.5    0    20    -33    0    0    999    # Females
8.9     0    20    -33    0    0    999    # Females
7.9     0    20    -33    0    0    999    # Females
7.2     0    20    -33    0    0    999    # Females
6.6     0    20    -33    0    0    999    # Females
6.1     0    20    -33    0    0    999    # Females
5.6     0    20    -33    0    0    999    # Females
5.1     0    20    -33    0    0    999    # Females
4.6     0    20    -33    0    0    999    # Females
4.1     0    20    -33    0    0    999    # Females
3.6     0    20    -33    0    0    999    # Females
3.2     0    20    -33    0    0    999    # Females
2.7     0    20    -33    0    0    999    # Females
2.2     0    20    -33    0    0    999    # Females
1.7     0    20    -33    0    0    999    # Females
1.2     0    20    -33    0    0    999    # Females
0.7     0    20    -33    0    0    999    # Females
0.0   -1.0 1.0    -7     0     0    999    # Females (beta)
#1.38403 0.5 3.7    -7     0     0    999    # Females (beta)
15.1    0    20    -33    0    0    999    # Females
14      0    20    -33    0    0    999    # Females
12.9    0    20    -33    0    0    999    # Females
11.8    0    20    -33    0    0    999    # Females
10.6    0    20    -33    0    0    999    # Females
8.7     0    20    -33    0    0    999    # Females
7.4     0    20    -33    0    0    999    # Females
6.6     0    20    -33    0    0    999    # Females
6.1     0    20    -33    0    0    999    # Females
5.6     0    20    -33    0    0    999    # Females
5.1     0    20    -33    0    0    999    # Females
4.6     0    20    -33    0    0    999    # Females
4.1     0    20    -33    0    0    999    # Females
3.6     0    20    -33    0    0    999    # Females
3.2     0    20    -33    0    0    999    # Females
2.7     0    20    -33    0    0    999    # Females
2.2     0    20    -33    0    0    999    # Females
1.7     0    20    -33    0    0    999    # Females
1.2     0    20    -33    0    0    999    # Females
0.7     0    20    -33    0    0    999    # Females
0.0   -1.0 1.0    -7     0     0    999    # Females (beta)
#1.38403 0.5 3.7    -7     0     0    999    # Females (beta)
## ----- ##
## ----- ##
## MOLTING PROBABILITY CONTROLS ##
## Two lines for each parameter if split sex, one line if not ##
## ----- ##
## ival  lb  ub  phz  prior  p1  p2  # parameter  ##
## ----- ##
## males and combined
145.0386 100. 500.0 3 0 0.0 999.0 # molt_mu males
0.053036 0.02 2.0 3 0 0.0 999.0 # molt_cv males

```



```

145.0386 100. 500.0 3 0 0.0 999.0 # molt_mu males
0.053036 0.02 2.0 3 0 0.0 999.0 # molt_cv males
## females
300.0000 5. 500.0 -4 0 0.0 999.0 # molt_mu females (molt every year)
0.01 0.001 9.0 -4 0 0.0 999.0 # molt_cv females (molt every year)
300.0000 5. 500.0 -4 0 0.0 999.0 # molt_mu females (molt every year)
0.01 0.001 9.0 -4 0 0.0 999.0 # molt_cv females (molt every year)
## ----- ##
# The custom growth-increment matrix
# custom molt probability matrix

## ----- ##
## SELECTIVITY CONTROLS ##
## Selectivity P(capture of all sizes). Each gear must have a selectivity and a ##
## retention selectivity. If a uniform prior is selected for a parameter then the ##
## lb and ub are used (p1 and p2 are ignored) ##
## LEGEND ##
## sel type: 0 = parametric, 1 = coefficients (NIY), 2 = logistic, 3 = logistic95, ##
## 4 = double normal (NIY) ##
## gear index: use +ve for selectivity, -ve for retention ##
## sex dep: 0 for sex-independent, 1 for sex-dependent ##
## ----- ##
## Gear-1 Gear-2 Gear-3 Gear-4 Gear-5 Gear-6
## PotFshry TrawlByc TCFshry FixedGr NMFS BSFRF
1 1 1 1 2 1 # selectivity periods
1 0 1 0 1 1 # sex specific selectivity
# 9 2 2 2 2 2 # male selectivity type
2 2 2 2 2 2 # male selectivity type
2 2 2 2 2 2 # female selectivity type
0 0 0 0 6 0 #6 # within another gear
# 5 0 0 0 0 0 #-NEW: extra parameters for each pattern by fleet, males
0 0 0 0 0 0 #-NEW: extra parameters for each pattern by fleet, males
0 0 0 0 0 0 #-NEW: extra parameters for each pattern by fleet, females
## Gear-1 Gear-2 Gear-3 Gear-4 Gear-5 Gear-6
2 1 1 1 1 1 # retention periods
1 0 0 0 0 0 # sex specific retention
2 6 6 6 6 6 # male retention type
6 6 6 6 6 6 # female retention type
1 0 0 0 0 0 # male retention flag (0 = no, 1 = yes)
0 0 0 0 0 0 # female retention flag (0 = no, 1 = yes)
0 0 0 0 0 0 #-NEW: extra parameters for each pattern by fleet, males
0 0 0 0 0 0 #-NEW: extra parameters for each pattern by fleet, females
## ----- ##
## gear par sel start end ##
## index index par sex ival lb ub prior p1 p2 phz period period ##
## ----- ##
# Gear-1
1 1 1 1 125.0000 5 190 0 1 999 4 1975 2019 #4
1 2 2 1 8.0 0.1 20 0 1 999 4 1975 2019 #4
# Gear-1
# 1 1 1 1 67.5 0 200 0 1 999 -999 1975 2018 #4 #parameters for cubic spine

```

```

# 1 2 2 1 87.5 0 200 0 1 999 -999 1975 2018 #4
# 1 3 3 1 97.5 0 200 0 1 999 -999 1975 2018 #4
# 1 4 4 1 112.5 0 200 0 1 999 -999 1975 2018 #4
# 1 5 5 1 162.5 0 200 0 1 999 -999 1975 2018 #4
# 1 6 6 1 0.001 0.00001 0.99999 0 1 999 4 1975 2018 #4
# 1 6 7 1 0.1 0.00001 0.99999 0 1 999 4 1975 2018 #4
# 1 6 8 1 0.3 0.00001 0.99999 0 1 999 4 1975 2018 #4
# 1 6 9 1 0.7 0.00001 0.99999 0 1 999 4 1975 2018 #4
# 1 6 10 1 0.99999 0.00001 1.01 0 1 999 -4 1975 2018 #4
1 3 1 2 84.00 5 150 0 1 999 4 1975 2019
1 4 2 2 4.0000 0.1 20 0 1 999 4 1975 2019
# Gear-2
2 5 1 0 165.0 5 190 0 1 999 4 1975 2019
2 6 2 0 15.0000 0.1 25 0 1 999 4 1975 2019
# Gear-3-9
3 7 1 1 115.0 5 190 0 1 999 4 1975 2019
3 8 2 1 15.0 0.1 25 0 1 999 4 1975 2019
3 9 1 2 95.0 5 190 0 1 999 4 1975 2019 # dummy
3 10 2 2 2.5 0.1 25 0 1 999 4 1975 2019
# Gear-4
4 11 1 0 115.0 5 190 0 1 999 4 1975 2019 # dummy
4 12 2 0 9.0 0.1 25 0 1 999 4 1975 2019
# Gear-5
5 13 1 1 75.0 30 190 0 1 999 5 1975 1981 #5
5 14 2 1 5.0 1 50 0 1 999 5 1975 1981 #5
5 15 1 1 80.0 30 190 0 1 999 5 1982 2020 #5
5 16 2 1 10.0 1 50 0 1 999 5 1982 2020 #5
5 17 1 2 70.0 30 180 0 1 999 5 1975 1981 #5
5 18 2 2 9.0 1 50 0 1 999 5 1975 1981 #5
5 19 1 2 70.0 30 180 0 1 999 5 1982 2020 #5
5 20 2 2 4.00 1.0 50 0 1 999 5 1982 2020 #5
# Gear-6
6 21 1 1 75.0 1 180 0 1 999 5 1975 2020 #5
6 22 2 1 8.5 1 50 0 1 999 5 1975 2020 #5
6 23 1 2 85.0 1 180 0 1 999 5 1975 2020 #5
6 24 2 2 10.0 1 50 0 1 999 5 1975 2020 #5

##-----##
## Retained ##
## gear par sel start end ##
## index index par sex ival lb ub prior p1 p2 phz period period ##
##-----##
# Gear-1
-1 25 1 1 135 1 999 0 1 999 4 1975 2004
-1 26 2 1 2.0 1 20 0 1 999 4 1975 2004
-1 27 1 1 140 1 999 0 1 999 4 2005 2019
-1 28 2 1 2.5 1 20 0 1 999 4 2005 2019
-1 29 1 2 591 1 999 0 1 999 -3 1975 2004
-1 30 1 2 591 1 999 0 1 999 -3 2005 2019
# Gear-2
-2 31 1 0 595 1 999 0 1 999 -3 1975 2019
# Gear-3

```

```

-3 32 1 0 595 1 999 0 1 999 -3 1975 2019 #Dummy
# Gear-4
-4 33 1 0 595 1 999 0 1 999 -3 1975 2019
# Gear-5
-5 34 1 0 590 1 999 0 1 999 -3 1975 2020
# Gear-6
-6 35 1 0 580 1 999 0 1 999 -3 1975 2020
##-----##

# Number of asymptotic parameters
1
# Fleet Sex Year ival lb ub phz
1 1 1975 0.000001 0 1 -3
# 1 1 2006 0.044000 0 1 -3
# 1 1 2007 0.019700 0 1 -3
# 1 1 2008 0.019875 0 1 -3
# 1 1 2009 0.032750 0 1 -3
# 1 1 2010 0.015320 0 1 -3
# 1 1 2011 0.011250 0 1 -3
# 1 1 2012 0.024045 0 1 -3
# 1 1 2013 0.063200 0 1 -3
# 1 1 2014 0.160500 0 1 -3
# 1 1 2015 0.070950 0 1 -3
# 1 1 2016 0.082600 0 1 -3
##-----##

## PRIORS FOR CATCHABILITY
## If a uniform prior is selected for a parameter then the lb and ub are used (p1 ##
## and p2 are ignored). ival must be > 0 ##
## LEGEND ##
## prior: 0 = uniform, 1 = normal, 2 = lognormal, 3 = beta, 4 = gamma ##
##-----##

## ival lb ub phz prior p1 p2 Analytic? LAMBDA Emphasis
0.896 0 2 6 1 0.896 0.03 0 1 1
1.0 0 5 -6 0 0.001 5.00 0 1 1 # BSFRF
##-----##

##-----##
## ADDITIONAL CV FOR SURVEYS/INDICES ##
## If a uniform prior is selected for a parameter then the lb and ub are used (p1 ##
## and p2 are ignored). ival must be > 0 ##
## LEGEND ##
## prior type: 0 = uniform, 1 = normal, 2 = lognormal, 3 = beta, 4 = gamma ##
##-----##

## ival lb ub phz prior p1 p2
0.0001 0.00001 10.0 -4 4 1.0 100 # NMFS
0.25 0.00001 10.0 9 0 0.001 1.00 # BSFRF
##-----##

##-----##
## PENALTIES FOR AVERAGE FISHING MORTALITY RATE FOR EACH GEAR
##-----##
## Mean_F Female Offset STD_PHZ1 STD_PHZ2 PHZ_M PHZ_F

```

```

# Upper bound value for male directed fishig mortality deviations
0.22313    0.0505    0.5    45.50    1    1    -12    4    -10    2.95    -10    10 # Pot
0.0183156    1.0    0.5    45.50    1    -1    -12    4    -10    10    -10    10 # Trawl
0.011109    1.0    0.5    45.50    1    1    -12    4    -10    10    -10    10 # Tanner (-1 -5)
0.011109    1.0    0.5    45.50    1    -1    -12    4    -10    10    -10    10 # Fixed
0.00    0.0    2.00    20.00    -1    -1    -12    4    -10    10    -10    10 # NMFS trawl survey (0 catch)
0.00    0.0    2.00    20.00    -1    -1    -12    4    -10    10    -10    10 # BSFRF (0)
## ----- ##

## ----- ##
## OPTIONS FOR SIZE COMPOSTION DATA ##
## One column for each data matrix ##
## LEGEND ##
## Likelihood: 1 = Multinomial with estimated/fixed sample size ##
## 2 = Robust approximation to multinomial ##
## 3 = logistic normal (NIY) ##
## 4 = multivariate-t (NIY) ##
## 5 = Dirichlet ##
## AUTO TAIL COMPRESSION ##
## pmin is the cumulative proportion used in tail compression ##
## ----- ##
# Pot    Trawl Tanner Fixed NMFS  BSFRF
2 2 2 2 2 2 2 2 2 2 2 2 # Type of likelihood
0 0 0 0 0 0 0 0 0 0 0 0 # Auto tail compression (pmin)
1 1 1 1 1 1 1 1 1 1 1 1 # Initial value for effective sample size multiplier
-4 -4 -4 -4 -4 -4 -4 -4 -4 -4 -4 -4 # Phz for estimating effective sample size (if appl.)
1 2 3 4 4 5 5 6 6 7 7 8 8 # Composition aggregator
1 1 1 1 1 1 1 1 1 1 1 1 # LAMBDA
1 1 1 1 1 1 1 1 1 1 1 1 # Emphasis AEP
## ----- ##

## ----- ##
## TIME VARYING NATURAL MORTALIHY RATES ##
## LEGEND ##
## Type: 0 = constant natural mortality ##
## 1 = Random walk (deviates constrained by variance in M) ##
## 2 = Cubic Spline (deviates constrained by nodes & node-placement) ##
## 3 = Blocked changes (deviates constrained by variance at specific knots) ##
## 4 = Time blocks ##
## ----- ##
## Type
6
## M is relative (YES=1; NO=0)
1
## Phase of estimation
3
## STDEV in m_dev for Random walk
0.25
## Number of nodes for cubic spline or number of step-changes for option 3
2
2
## Year position of the knots (vector must be equal to the number of nodes)

```

```

1980 1985
1980 1985
# number of breakpoints in M by size
0
## Specific initial values for the natural mortality devs (0=no, 1=yes)
1
## ----- ##
## ival    lb    ub    phz  extra  prior  p1  p2    # parameter  ##
## ----- ##
1.7342575    0    2    8    0
0.000000    -2    2   -99    0
1.780586    0    2    -8   -1
0.000000    -2    2   -99    0
## ----- ##

## ----- ##
## OTHER CONTROLS
## ----- ##
1975  # First rec_dev
2019  # last rec_dev
2    # Estimated rec_dev phase
2    # Estimated sex_ratio
0.5  # initial sex-ratio
-3   # Estimated rec_ini phase
1    # VERBOSE FLAG (0 = off, 1 = on, 2 = objective func; 3 diagnostics)
3    # Initial conditions (0 = Unfished, 1 = Steady-state fished, 2 = Free parameters, 3 = Free parameters (revised))
1    # Lambda (proportion of mature male biomass for SPR reference points).
0    # Stock-Recruit-Relationship (0 = none, 1 = Beverton-Holt)
10   # Maximum phase (stop the estimation after this phase).
-1   # Maximum number of function calls.
## ----- ##
## EMPHASIS FACTORS (CATCH)
## ----- ##
#Ret_male Disc_male Disc_female Disc_trawl Disc_Tanner_male Disc_Tanner_female Disc_fixed
1 1 1 1 1 1 1
## ----- ##
## EMPHASIS FACTORS (Priors)
## ----- ##
# Log_fdevs meanF Mdevs Rec_devs Initial_devs Fst_dif_dev Mean_sex-Ratio
10000 0 1.0 2 0 0 10 #(10000)
## EOF
9999

```

**Appendix D. Comparison of observer biomass and length composition estimates of crab pot fisheries in 2020 (model 19.3) and updated/standardized estimates in 2021 (models 19.3c-19.6)**

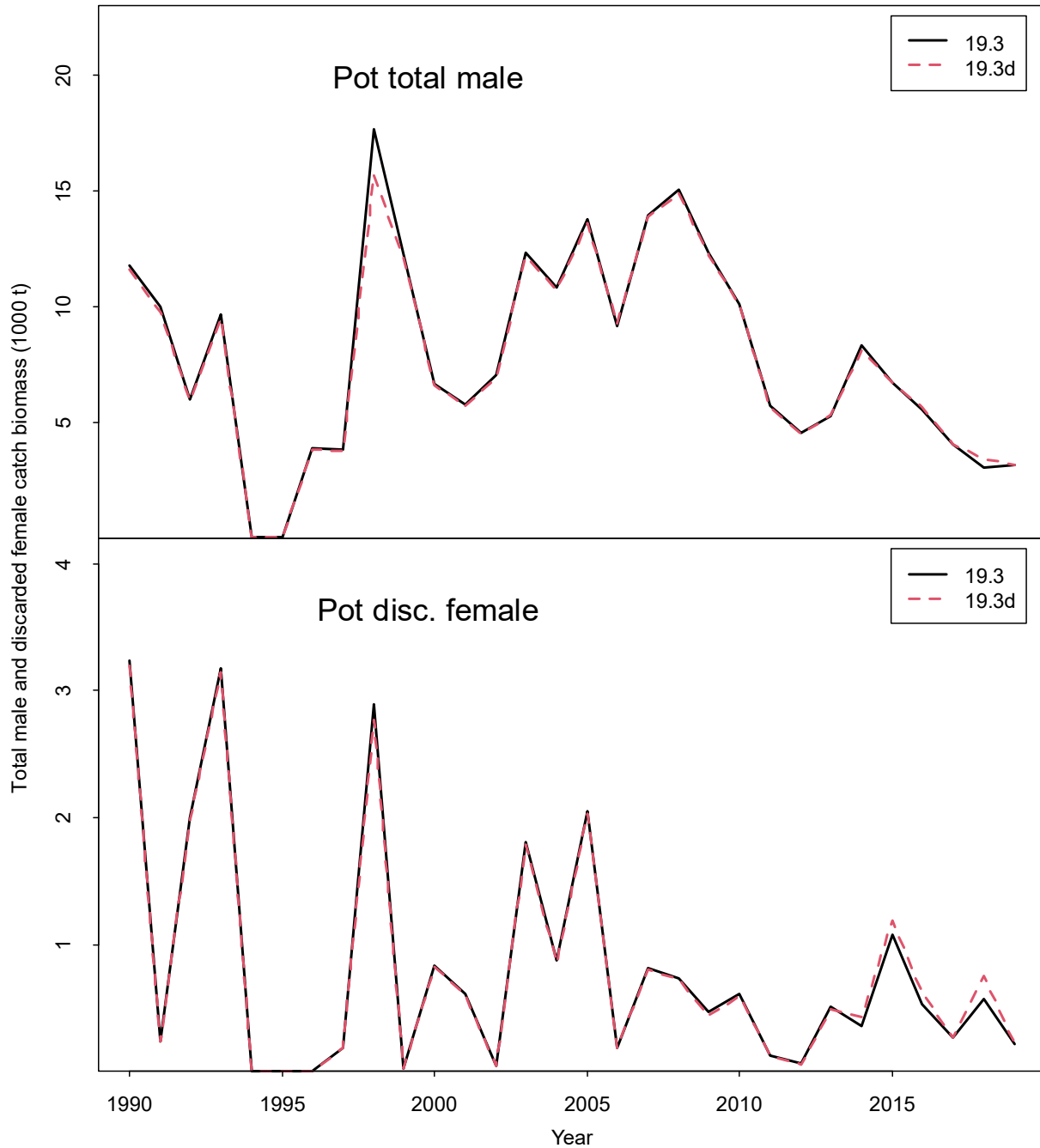


Figure D1. Comparison of observer total male and discarded female biomass in the directed pot fishery used in model 19.3 in 2020 and updated/standardized observer total male and discarded female biomass used in models 19.3c-19.6 in 2021.

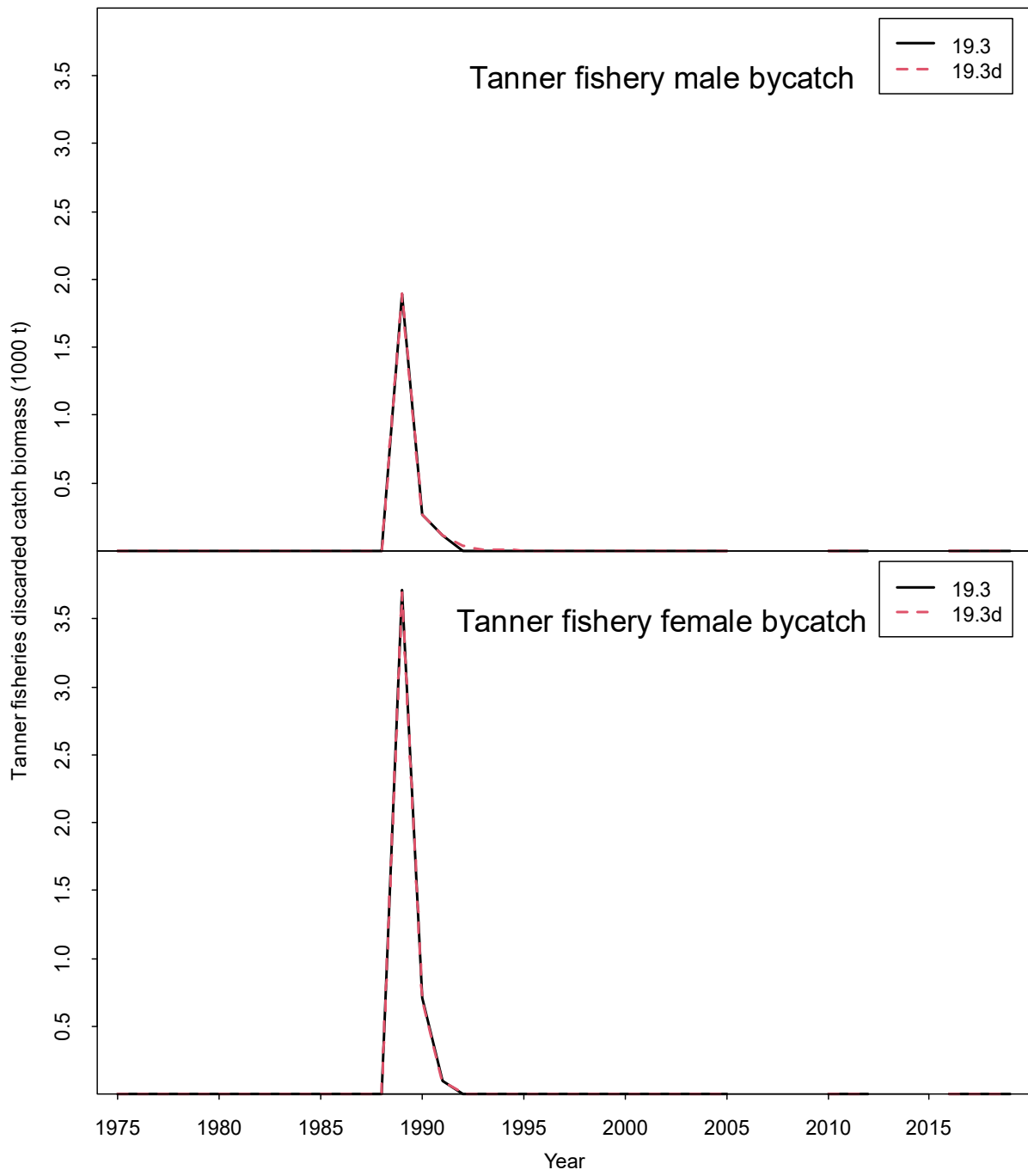


Figure D2. Comparison of observer discarded red king crab biomass in the Tanner crab fishery used in model 19.3 in 2020 and updated/standardized observer discarded red king crab biomass used in models 19.3c-19.6 in 2021.

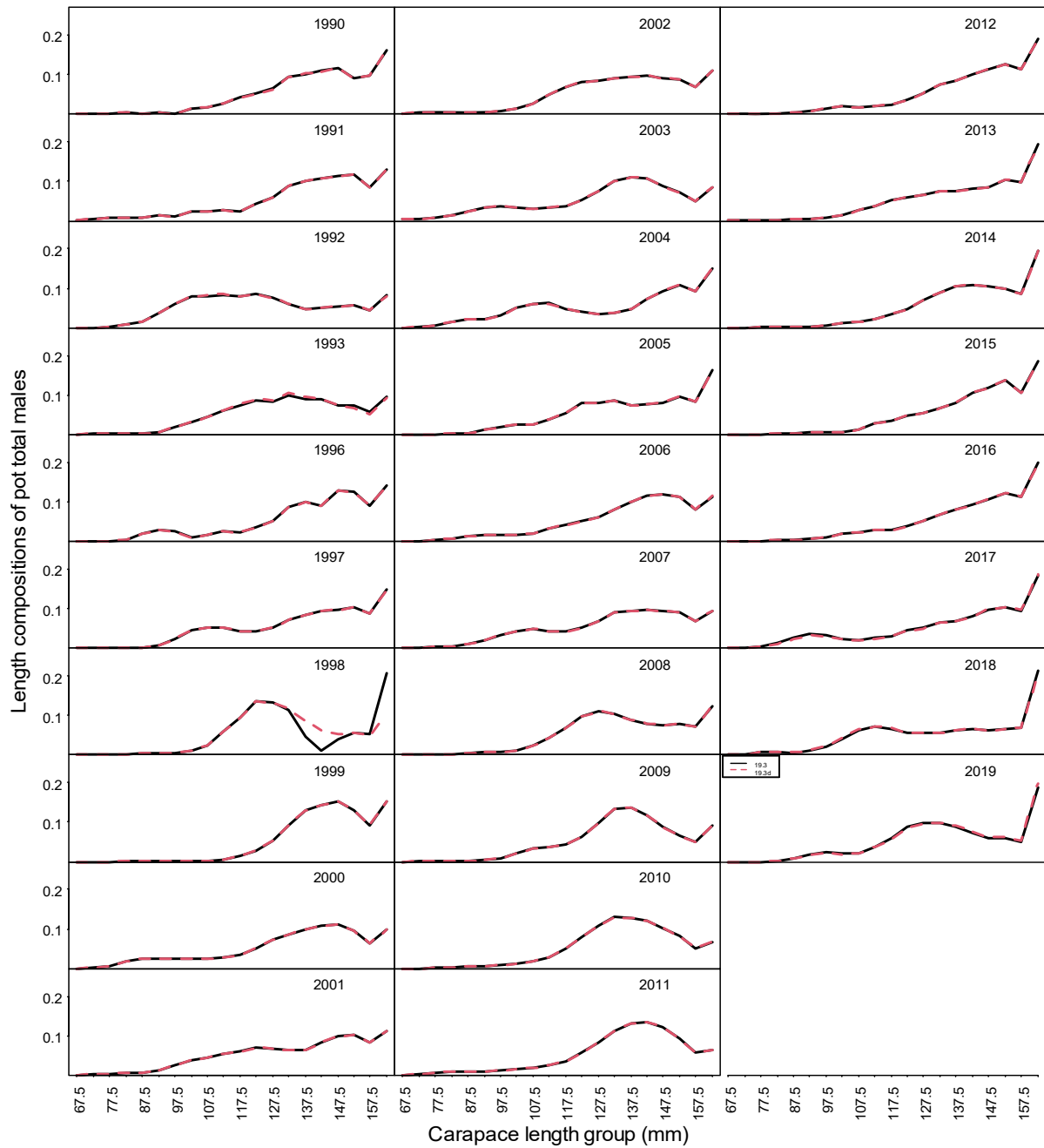


Figure D3. Comparison of observer length compositions of total males in the directed pot fishery used in model 19.3 in 2020 and updated/standardized observer length compositions used in models 19.3c-19.6 in 2021. There are some differences in 1998.



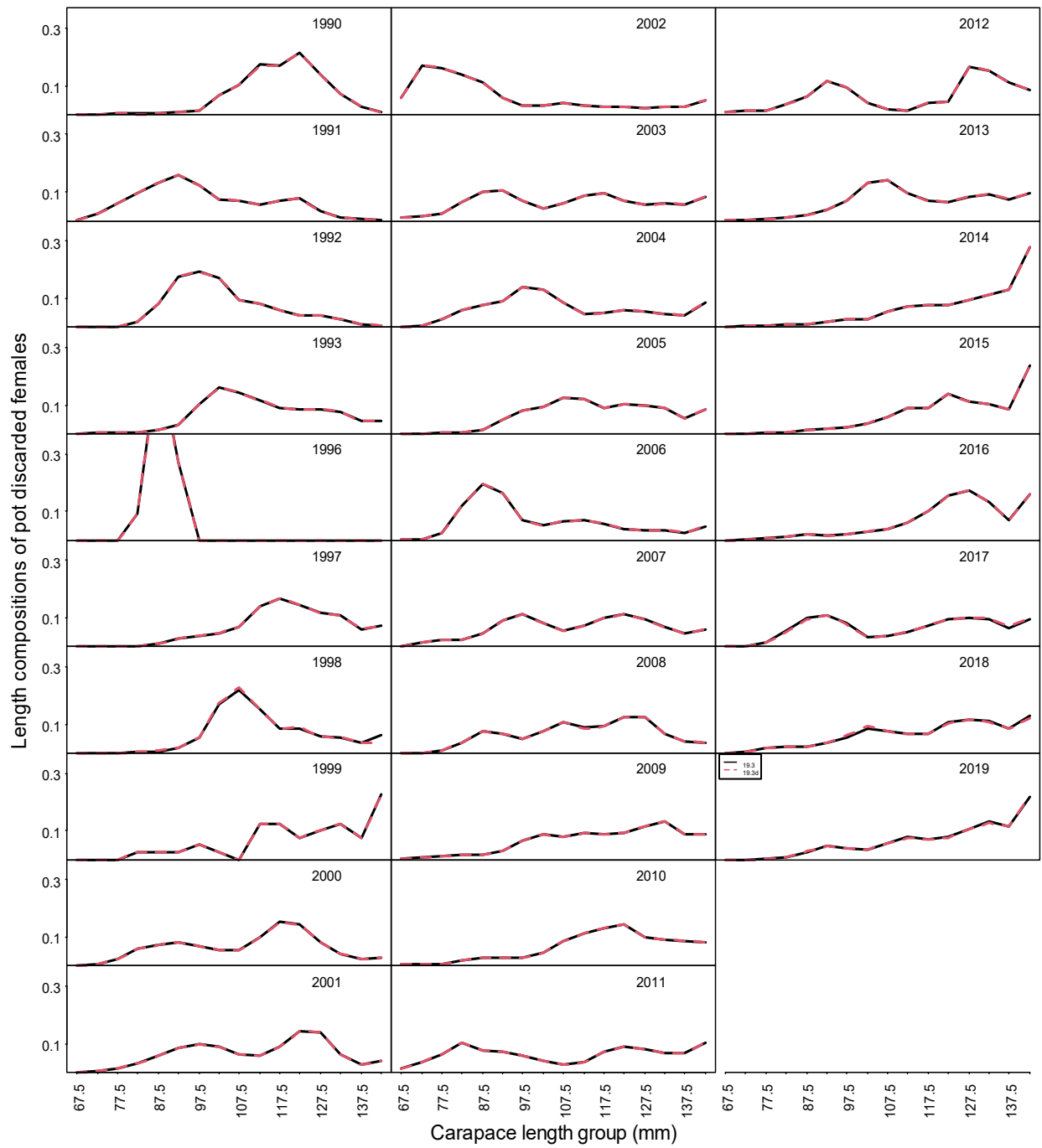


Figure D4. Comparison of observer length compositions of discarded females in the directed pot fishery used in model 19.3 in 2020 and updated/standardized observer length compositions used in models 19.3c-19.6 in 2021.

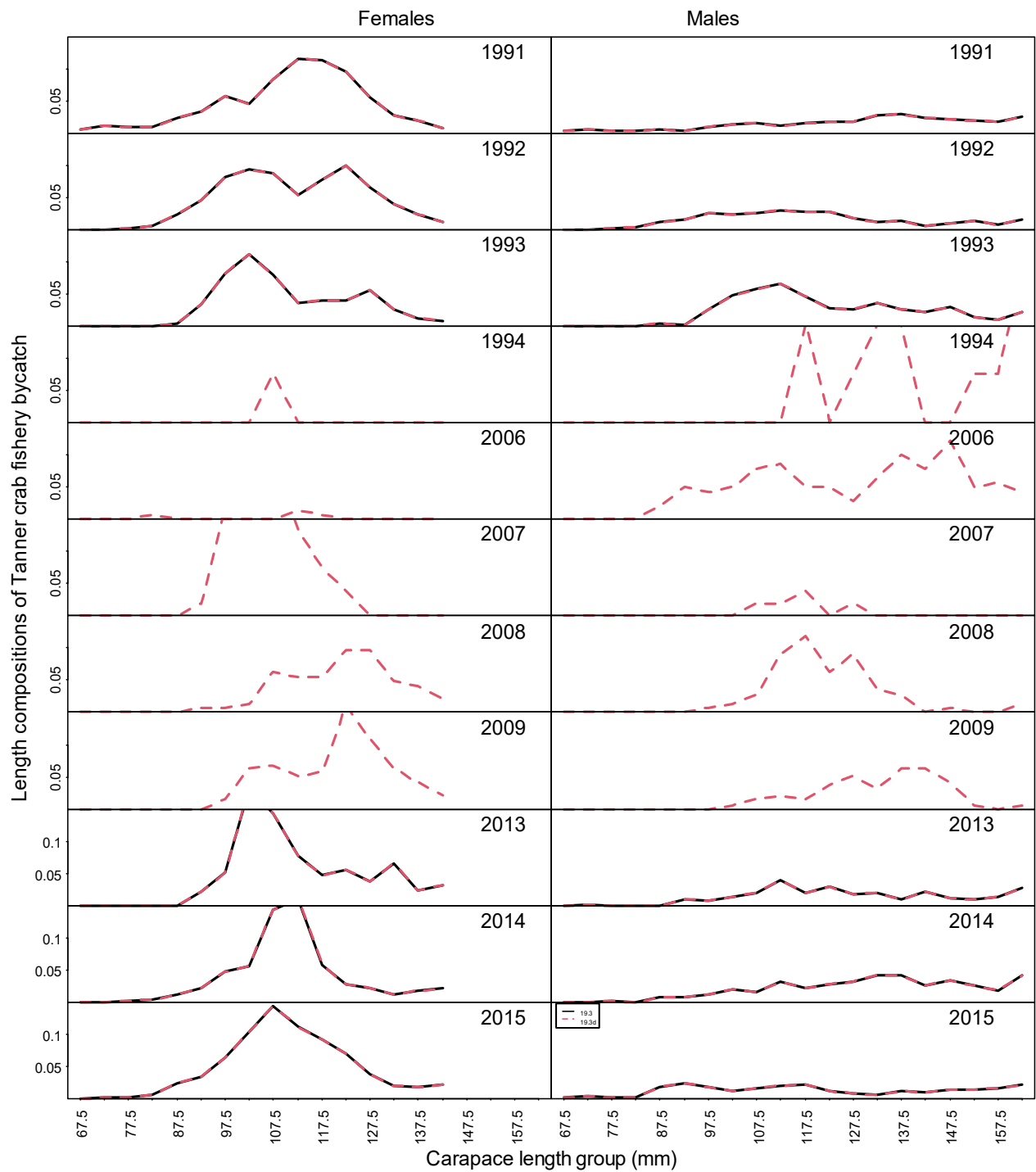


Figure D5. Comparison of observer length compositions of discarded red king crab in the Tanner crab fishery used in model 19.3 in 2020 and updated/standardized observer length compositions used in models 19.3c-19.6 in 2021.

*Appendix E. VAST results and diagnostics by Jon Richar, NMFS*

Overall, VAST performed very well for the Bristol Bay red king crab (BBRKC) stock. Best diagnostic performance was in the total GE65 model, while the female GE65 model was difficult to get to converge and had the most issues of the BBRKC models. Population trends tracked current survey estimates well, with no obvious indications of bias when model estimates are compared against area swept estimates. The one concern here is a divergence in trends between VAST estimates and area swept estimates observed for the female GE65 component from 2006-2009. The cause of this divergence is not currently well-understood, and it is still being investigated. Center of gravity and effective area occupied output suggest that the female GE65 population segment is modestly increasing its range, while shifting somewhat to the northwest, while the male GE65 component is contracting its range while also shifting somewhat to the northwest. The new DHARMA residual-based diagnostics were encouraging, with Q-Q plots indicating that total, male and female biomass models performed well, and were able to adequately capture observed population variability. The plots for rank-transformed model predictions vs. standardized residuals did however indicate that the models had some difficulty with the very highest observations, as positive trends in residuals may be observed these in plots for all models at the larger values/higher ranks. Spatial quantile residual plots do not suggest presence of any concerning trends in residuals in either space or time. In conclusion, I am now much more comfortable supporting use of VAST in the stock assessment process than I was last year.

A. Results and diagnostics for males GE65 mm

Table E1 (extracted from parameter\_estimates.txt). Model parameter estimates for males GE65 mm. Note that L\_epsilon2\_z, the parameter for spatio-temporal variation in biomass only is disabled for this model, due to insufficient variability.

ln_H_input	ln_H_input	beta1_ft	beta1_ft	beta1_ft	beta1_ft
0.4246770	0.1948351	3.2347503	3.1788151	4.3519778	3.5321435
beta1_ft	beta1_ft	beta1_ft	beta1_ft	beta1_ft	beta1_ft
2.7613550	2.9545170	2.8214412	2.1871557	2.4178124	1.4982557
beta1_ft	beta1_ft	beta1_ft	beta1_ft	beta1_ft	beta1_ft
1.3854900	1.3534917	1.8200897	1.6132997	1.9960421	1.4271915
beta1_ft	beta1_ft	beta1_ft	beta1_ft	beta1_ft	beta1_ft
1.0699988	1.1797730	0.5309809	0.3140438	0.5200696	0.6612359
beta1_ft	beta1_ft	beta1_ft	beta1_ft	beta1_ft	beta1_ft
1.0289110	1.2053710	1.0209494	1.6926730	1.2784559	0.8561469
beta1_ft	beta1_ft	beta1_ft	beta1_ft	beta1_ft	beta1_ft
1.6035744	1.1021394	1.0458219	1.7856805	1.6218819	2.2656588
beta1_ft	beta1_ft	beta1_ft	beta1_ft	beta1_ft	beta1_ft
1.3288600	1.1019269	1.0663150	1.2415754	1.1563799	1.5410823
beta1_ft	beta1_ft	beta1_ft	beta1_ft	beta1_ft	L_omega1_z
1.6878197	1.3703371	1.6042728	1.3028284	1.1667935	1.3480184
L_epsilon1_z	logkappa1	beta2_ft	beta2_ft	beta2_ft	beta2_ft
1.0495714	-4.3388659	1.5340859	1.5460340	0.9699323	1.2962917
beta2_ft	beta2_ft	beta2_ft	beta2_ft	beta2_ft	beta2_ft
1.3172601	1.0338801	1.1155396	1.1662783	0.6576483	1.2016665
beta2_ft	beta2_ft	beta2_ft	beta2_ft	beta2_ft	beta2_ft
0.9830857	0.6621450	1.0850035	1.1362303	0.8017747	1.3811555
beta2_ft	beta2_ft	beta2_ft	beta2_ft	beta2_ft	beta2_ft
1.5272657	1.2876578	1.7433139	1.4644895	1.2669372	1.4267320
beta2_ft	beta2_ft	beta2_ft	beta2_ft	beta2_ft	beta2_ft
1.4516482	1.4677695	1.1041354	0.9524132	1.1605098	1.5874892
beta2_ft	beta2_ft	beta2_ft	beta2_ft	beta2_ft	beta2_ft
1.0521914	1.7108559	2.1826606	1.4280594	1.7164466	0.9810639
beta2_ft	beta2_ft	beta2_ft	beta2_ft	beta2_ft	beta2_ft
1.4394978	1.2622499	1.3979922	1.3106862	1.4290399	1.5194971
beta2_ft	beta2_ft	beta2_ft	beta2_ft	beta2_ft	L_omega2_z
1.3870550	1.3922920	1.4670261	1.2116351	1.0194797	0.4350181
L_epsilon2_z	logkappa2	logSigmaM			
0.5071955	-3.4301660	-0.3651867			

Table E2 (extracted from parameter\_estimates.txt). Model parameter estimate diagnostic output for males GE65mm. Starting values for model parameters, upper and lower bounds, MLE, and final gradients for parameters.

	Param	starting_value	Lower	MLE	Upper	final_gradient
1	ln_H_input	0.4246770	-50	0.4246770	50	-5.510007e-09
2	ln_H_input	0.1948362	-50	0.1948351	50	-3.723925e-10
3	beta1_ft	3.2347437	-50	3.2347503	50	-7.993961e-11
4	beta1_ft	3.1788096	-50	3.1788151	50	-1.463718e-11
5	beta1_ft	4.3519711	-50	4.3519778	50	4.783729e-12
6	beta1_ft	3.5321295	-50	3.5321435	50	-1.176659e-11
7	beta1_ft	2.7613474	-50	2.7613550	50	-1.469735e-11
8	beta1_ft	2.9545091	-50	2.9545170	50	2.546408e-12
9	beta1_ft	2.8214352	-50	2.8214412	50	-1.211564e-11
10	beta1_ft	2.1871483	-50	2.1871557	50	-1.672529e-11
11	beta1_ft	2.4178073	-50	2.4178124	50	-2.058931e-11
12	beta1_ft	1.4982417	-50	1.4982557	50	-4.633494e-11
13	beta1_ft	1.3854818	-50	1.3854900	50	-1.053335e-11
14	beta1_ft	1.3534817	-50	1.3534917	50	-1.266365e-11
15	beta1_ft	1.8200752	-50	1.8200897	50	-5.952128e-12
16	beta1_ft	1.6132877	-50	1.6132997	50	-9.197088e-13
17	beta1_ft	1.9960279	-50	1.9960421	50	2.807532e-11
18	beta1_ft	1.4271749	-50	1.4271915	50	2.178813e-11
19	beta1_ft	1.0699726	-50	1.0699988	50	1.539879e-12
20	beta1_ft	1.1797563	-50	1.1797730	50	3.896258e-12
21	beta1_ft	0.5309600	-50	0.5309809	50	-2.937894e-11
22	beta1_ft	0.3140216	-50	0.3140438	50	2.928346e-11
23	beta1_ft	0.5200480	-50	0.5200696	50	3.566680e-11
24	beta1_ft	0.6612178	-50	0.6612359	50	7.180034e-12
25	beta1_ft	1.0288928	-50	1.0289110	50	3.251888e-11
26	beta1_ft	1.2053544	-50	1.2053710	50	1.784350e-11
27	beta1_ft	1.0209341	-50	1.0209494	50	2.645439e-11
28	beta1_ft	1.6926627	-50	1.6926730	50	4.311485e-11
29	beta1_ft	1.2784425	-50	1.2784559	50	1.059597e-12
30	beta1_ft	0.8561281	-50	0.8561469	50	3.409117e-11
31	beta1_ft	1.6035631	-50	1.6035744	50	-5.771561e-11
32	beta1_ft	1.1021215	-50	1.1021394	50	7.861045e-12
33	beta1_ft	1.0457969	-50	1.0458219	50	-1.166498e-10
34	beta1_ft	1.7856635	-50	1.7856805	50	1.387646e-11
35	beta1_ft	1.6218647	-50	1.6218819	50	3.509681e-11
36	beta1_ft	2.2656546	-50	2.2656588	50	-6.827428e-12
37	beta1_ft	1.3288438	-50	1.3288600	50	9.374945e-12
38	beta1_ft	1.1019075	-50	1.1019269	50	7.519851e-11
39	beta1_ft	1.0662975	-50	1.0663150	50	-7.817413e-12

40	beta1_ft	1.2415596	-50	1.2415754	50	4.209832e-11
41	beta1_ft	1.1563633	-50	1.1563799	50	2.644884e-11
42	beta1_ft	1.5410700	-50	1.5410823	50	-1.294875e-11
43	beta1_ft	1.6878073	-50	1.6878197	50	5.561773e-12
44	beta1_ft	1.3703229	-50	1.3703371	50	-8.080314e-12
45	beta1_ft	1.6042561	-50	1.6042728	50	-1.702249e-11
46	beta1_ft	1.3028170	-50	1.3028284	50	-8.401502e-12
47	beta1_ft	1.1667806	-50	1.1667935	50	-2.232214e-12
48	L_omega1_z	1.3480175	-50	1.3480184	50	-3.410241e-08
49	L_epsilon1_z	1.0495711	-50	1.0495714	50	3.175782e-08
50	logkappa1	-4.3388672	-50	-4.3388659	50	-1.977969e-08
51	beta2_ft	1.5341043	-50	1.5340859	50	3.518412e-10
52	beta2_ft	1.5460373	-50	1.5460340	50	-5.931078e-11
53	beta2_ft	0.9699340	-50	0.9699323	50	-1.117861e-10
54	beta2_ft	1.2963031	-50	1.2962917	50	1.640679e-10
55	beta2_ft	1.3172619	-50	1.3172601	50	3.246647e-11
56	beta2_ft	1.0338814	-50	1.0338801	50	-7.904788e-13
57	beta2_ft	1.1155412	-50	1.1155396	50	-1.348788e-11
58	beta2_ft	1.1662823	-50	1.1662783	50	4.388667e-11
59	beta2_ft	0.6576525	-50	0.6576483	50	1.387850e-10
60	beta2_ft	1.2016876	-50	1.2016665	50	1.584439e-10
61	beta2_ft	0.9830891	-50	0.9830857	50	-7.694823e-11
62	beta2_ft	0.6621540	-50	0.6621450	50	6.455103e-11
63	beta2_ft	1.0850082	-50	1.0850035	50	1.288001e-10
64	beta2_ft	1.1362362	-50	1.1362303	50	1.506884e-11
65	beta2_ft	0.8017748	-50	0.8017747	50	-1.619149e-11
66	beta2_ft	1.3811546	-50	1.3811555	50	3.703882e-11
67	beta2_ft	1.5272671	-50	1.5272657	50	2.245697e-10
68	beta2_ft	1.2876523	-50	1.2876578	50	-9.385914e-11
69	beta2_ft	1.7433179	-50	1.7433139	50	8.517631e-13
70	beta2_ft	1.4644727	-50	1.4644895	50	-3.728475e-10
71	beta2_ft	1.2669226	-50	1.2669372	50	-2.471312e-10
72	beta2_ft	1.4267394	-50	1.4267320	50	-2.875122e-11
73	beta2_ft	1.4516377	-50	1.4516482	50	-2.348370e-10
74	beta2_ft	1.4677741	-50	1.4677695	50	-1.358451e-10
75	beta2_ft	1.1041345	-50	1.1041354	50	-3.897860e-11
76	beta2_ft	0.9524004	-50	0.9524132	50	8.397194e-11
77	beta2_ft	1.1605144	-50	1.1605098	50	2.014122e-11
78	beta2_ft	1.5874896	-50	1.5874892	50	-3.676179e-10
79	beta2_ft	1.0522170	-50	1.0521914	50	5.131540e-10
80	beta2_ft	1.7108621	-50	1.7108559	50	5.035528e-11
81	beta2_ft	2.1826575	-50	2.1826606	50	-4.469785e-10
82	beta2_ft	1.4280642	-50	1.4280594	50	-6.323830e-13

83	beta2_ft	1.7164473	-50	1.7164466	50	-1.287486e-10
84	beta2_ft	0.9810607	-50	0.9810639	50	9.283063e-11
85	beta2_ft	1.4395046	-50	1.4394978	50	1.569846e-10
86	beta2_ft	1.2622405	-50	1.2622499	50	-1.901315e-10
87	beta2_ft	1.3979906	-50	1.3979922	50	-2.015126e-10
88	beta2_ft	1.3106768	-50	1.3106862	50	-2.468337e-11
89	beta2_ft	1.4290416	-50	1.4290399	50	-5.722267e-11
90	beta2_ft	1.5195095	-50	1.5194971	50	5.287148e-11
91	beta2_ft	1.3870574	-50	1.3870550	50	-9.208190e-11
92	beta2_ft	1.3922933	-50	1.3922920	50	3.860734e-11
93	beta2_ft	1.4670285	-50	1.4670261	50	5.575718e-11
94	beta2_ft	1.2116363	-50	1.2116351	50	1.250289e-11
95	beta2_ft	1.0194761	-50	1.0194797	50	1.003171e-10
96	L_omega2_z	0.4350162	-50	0.4350181	50	-3.462810e-08
97	L_epsilon2_z	0.5071939	-50	0.5071955	50	-1.597902e-08
98	logkappa2	-3.4301683	-50	-3.4301660	50	5.049451e-09
99	logSigmaM	-0.3651849	-50	-0.3651867	50	1.536594e-08

Table E3 (extracted from parameter\_estimates.txt). Model parameter estimates with parameter standard errors for maleGE65 mm. Note often large sizes of standard errors relative to estimates. Although final model converged and fit observations well, convergence was problematic to achieve, requiring some model simplification via removal of terms. Large standard errors suggest further simplification in future runs may be beneficial. Male and Total GE65 mm models had much larger parameter estimates relative to the standard errors.

sdreport(.) result

	Estimate	Std. Error
ln_H_input	0.4246770	0.07175141
ln_H_input	0.1948351	0.07075685
beta1_ft	3.2347503	0.98598816
beta1_ft	3.1788151	0.96204976
beta1_ft	4.3519778	0.98044580
beta1_ft	3.5321435	0.96715144
beta1_ft	2.7613550	0.95535058
beta1_ft	2.9545170	0.95854249
beta1_ft	2.8214412	0.96092257
beta1_ft	2.1871557	0.97049350
beta1_ft	2.4178124	0.96559370
beta1_ft	1.4982557	0.97948194
beta1_ft	1.3854900	0.98157659

beta1\_ft 1.3534917 0.98313774  
beta1\_ft 1.8200897 0.96971598  
beta1\_ft 1.6132997 0.97605663  
beta1\_ft 1.9960421 0.97095317  
beta1\_ft 1.4271915 0.97695256  
beta1\_ft 1.0699988 0.98285002  
beta1\_ft 1.1797730 0.97948649  
beta1\_ft 0.5309809 0.99510650  
beta1\_ft 0.3140438 1.00611924  
beta1\_ft 0.5200696 0.99842720  
beta1\_ft 0.6612359 0.99068226  
beta1\_ft 1.0289110 0.98695021  
beta1\_ft 1.2053710 0.98129312  
beta1\_ft 1.0209494 0.98241836  
beta1\_ft 1.6926730 0.98578980  
beta1\_ft 1.2784559 0.98379663  
beta1\_ft 0.8561469 0.99119474  
beta1\_ft 1.6035744 0.97978802  
beta1\_ft 1.1021394 0.98098954  
beta1\_ft 1.0458219 0.98341788  
beta1\_ft 1.7856805 0.97135368  
beta1\_ft 1.6218819 0.97184158  
beta1\_ft 2.2656588 0.96722132  
beta1\_ft 1.3288600 0.97521230  
beta1\_ft 1.1019269 0.98542684  
beta1\_ft 1.0663150 0.98647406  
beta1\_ft 1.2415754 0.98209885  
beta1\_ft 1.1563799 0.98193858  
beta1\_ft 1.5410823 0.97442937  
beta1\_ft 1.6878197 0.96955325  
beta1\_ft 1.3703371 0.97086086  
beta1\_ft 1.6042728 0.96785576  
beta1\_ft 1.3028284 0.97409384  
beta1\_ft 1.1667935 0.97335629  
L\_omega1\_z 1.3480184 0.12672906  
L\_epsilon1\_z 1.0495714 0.05392381  
logkappa1 -4.3388659 0.07526483  
beta2\_ft 1.5340859 0.31380413  
beta2\_ft 1.5460340 0.25826648  
beta2\_ft 0.9699323 0.29855034  
beta2\_ft 1.2962917 0.27567715  
beta2\_ft 1.3172601 0.22749975  
beta2\_ft 1.0338801 0.24021826



beta2\_ft 1.1155396 0.24680901  
beta2\_ft 1.1662783 0.25702944  
beta2\_ft 0.6576483 0.23968352  
beta2\_ft 1.2016665 0.25398620  
beta2\_ft 0.9830857 0.27264083  
beta2\_ft 0.6621450 0.27889136  
beta2\_ft 1.0850035 0.24285709  
beta2\_ft 1.1362303 0.27053944  
beta2\_ft 0.8017747 0.25426739  
beta2\_ft 1.3811555 0.23795941  
beta2\_ft 1.5272657 0.27223595  
beta2\_ft 1.2876578 0.25558165  
beta2\_ft 1.7433139 0.27706276  
beta2\_ft 1.4644895 0.28293733  
beta2\_ft 1.2669372 0.27829613  
beta2\_ft 1.4267320 0.26106650  
beta2\_ft 1.4516482 0.28463490  
beta2\_ft 1.4677695 0.28049025  
beta2\_ft 1.1041354 0.27579598  
beta2\_ft 0.9524132 0.30819295  
beta2\_ft 1.1605098 0.27117420  
beta2\_ft 1.5874892 0.26760315  
beta2\_ft 1.0521914 0.28911940  
beta2\_ft 1.7108559 0.27343301  
beta2\_ft 2.1826606 0.26693434  
beta2\_ft 1.4280594 0.26239635  
beta2\_ft 1.7164466 0.25616433  
beta2\_ft 0.9810639 0.25549577  
beta2\_ft 1.4394978 0.25061561  
beta2\_ft 1.2622499 0.28717869  
beta2\_ft 1.3979922 0.26310657  
beta2\_ft 1.3106862 0.25337096  
beta2\_ft 1.4290399 0.26411222  
beta2\_ft 1.5194971 0.25246006  
beta2\_ft 1.3870550 0.24774184  
beta2\_ft 1.3922920 0.24364574  
beta2\_ft 1.4670261 0.23134243  
beta2\_ft 1.2116351 0.23711445  
beta2\_ft 1.0194797 0.23724664  
L\_omega2\_z 0.4350181 0.05009772  
L\_epsilon2\_z 0.5071955 0.05981017  
logkappa2 -3.4301660 0.15228072  
logSigmaM -0.3651867 0.02666241

Table E4 (file = Table\_for\_SS3a.csv). Population index output. Biomass estimates for males GE65 mm, in metric tons, log-SD and SD in metric tons.

Year	Estimate (t)	SD_log	SD_mt
1975	149694.07	0.1250	14893.04
1976	243213.67	0.1125	22185.23
1977	244253.54	0.1012	19842.72
1978	198849.35	0.1135	18384.09
1979	96948.79	0.1265	10137.37
1980	141227.09	0.1244	14389.63
1981	75148.06	0.1023	6280.34
1982	62630.07	0.1227	6305.11
1983	34793.88	0.1114	3170.05
1984	52885.67	0.1388	6064.20
1985	29806.52	0.1123	2755.04
1986	27837.03	0.1186	2726.34
1987	44587.78	0.1197	4394.82
1988	38356.05	0.1174	3679.79
1989	41524.81	0.1193	4036.95
1990	38238.90	0.1239	3912.78
1991	40659.87	0.1286	4375.64
1992	26985.11	0.1217	2728.76
1993	37224.67	0.1290	4028.61
1994	25929.52	0.1306	2806.57
1995	24342.68	0.1233	2497.26
1996	29169.79	0.1298	3134.58
1997	57406.43	0.1262	5970.87
1998	51070.83	0.1161	4868.88
1999	46760.71	0.1382	5303.24
2000	47046.16	0.1276	4843.37
2001	31843.12	0.1099	2880.93
2002	48766.52	0.1278	5142.17
2003	64156.55	0.1313	6922.62
2004	67077.30	0.1325	7343.99
2005	56965.66	0.1177	5692.17
2006	54484.19	0.1106	4947.51
2007	61615.97	0.1136	5740.37
2008	61267.67	0.1288	6415.29
2009	38473.47	0.1406	4423.66
2010	39918.25	0.1259	4100.62
2011	29627.11	0.1303	3195.76
2012	34137.89	0.1475	4111.55
2013	42483.94	0.1351	4692.96
2014	63561.97	0.1233	6403.00
2015	42093.12	0.1170	4038.71
2016	29974.28	0.1120	2799.15
2017	26980.51	0.1050	2385.28
2018	18221.41	0.1117	1714.26
2019	16416.89	0.1162	1597.58

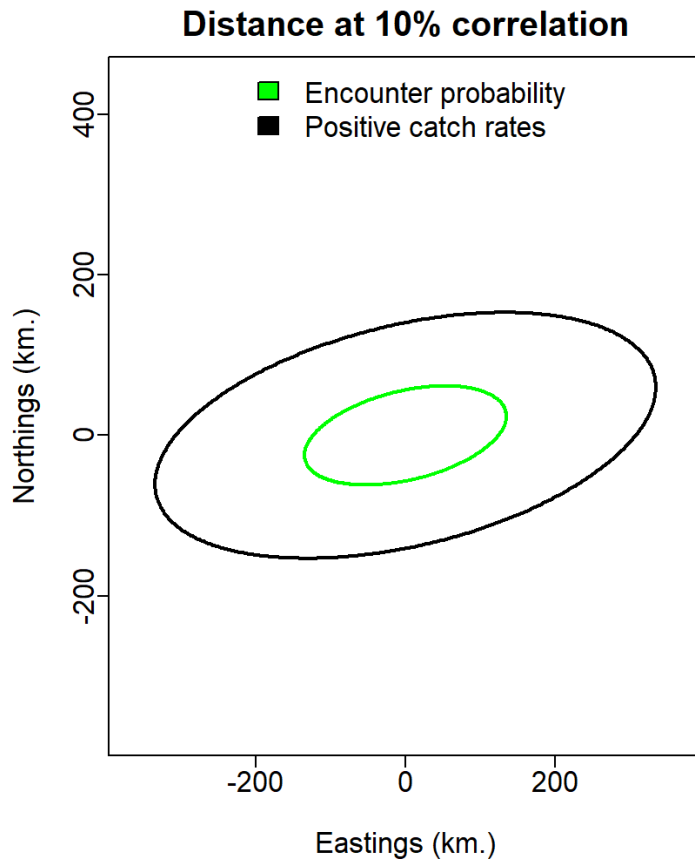


Figure E1 (file = Aniso.png). Geometric anisotropy for males GE65mm. Geometric anisotropy represents the tendency for correlations to decline faster in some directions than others. An ellipse with a major (long) axis pointed northeast-southwest will have correlations that decline slower along this axis than movement northwest-southeast.

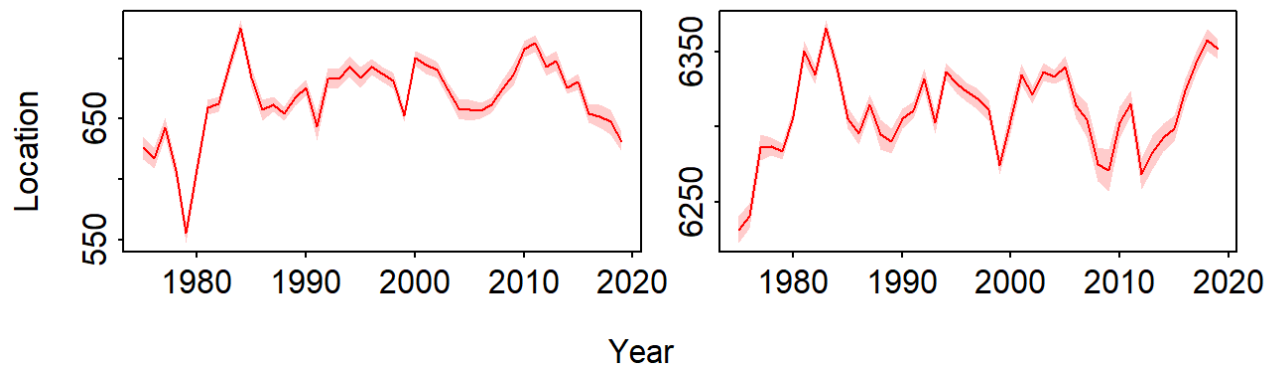


Figure E2 (file = center\_of\_gravity.png). Center of gravity for males GE65 mm: a.) Eastings (km) and b.) Northings (km). These figures indicate that the population has been shifting slightly to the northwest in recent years.

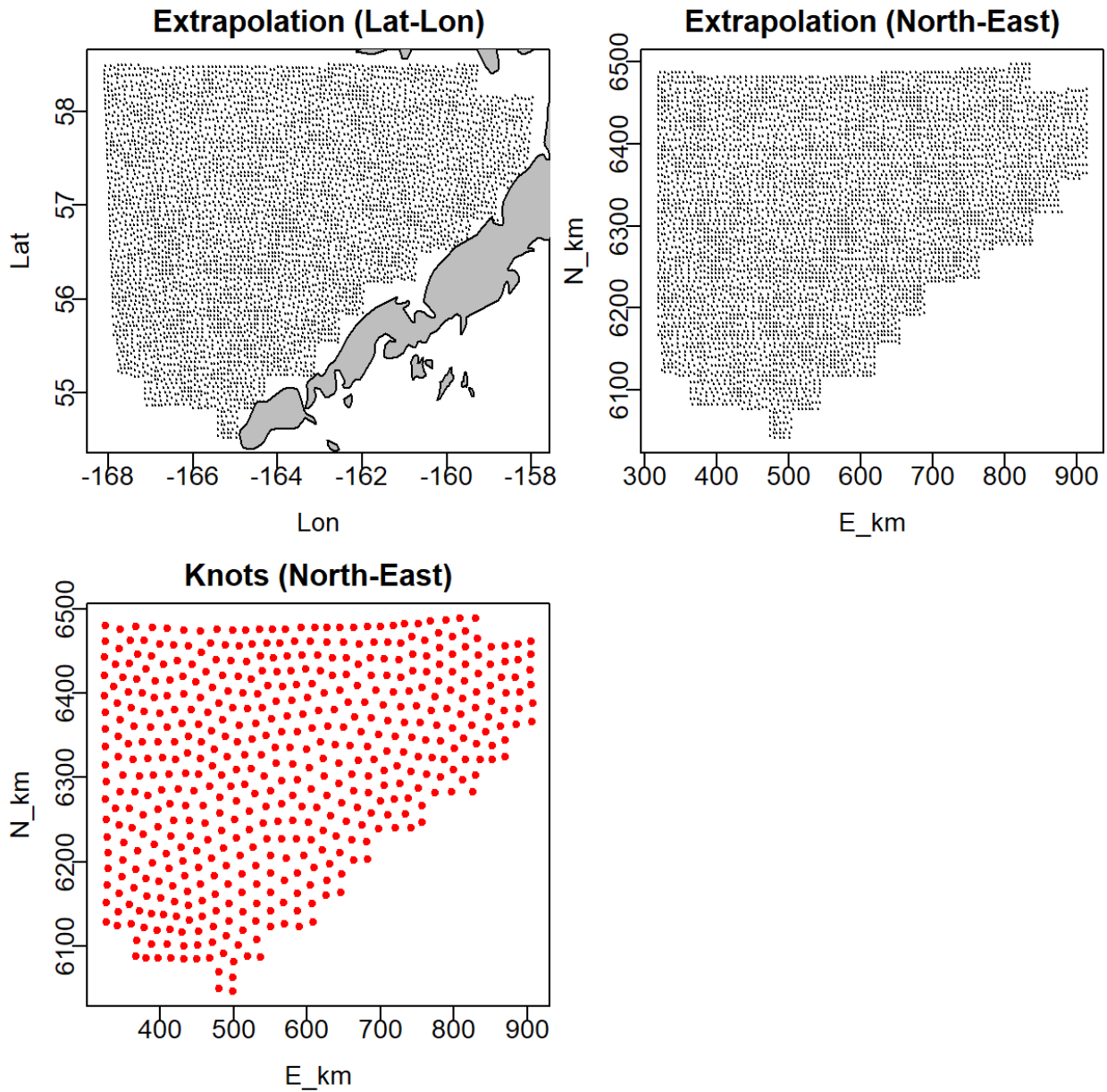


Figure E3 (file = Data\_and\_knots.png). Extrapolation maps for males GE65 mm: a.) Extrapolation grids by latitude and longitude, b.) Extrapolation grid by Northings/Eastings (in km), and c.) Knots used to define spatial mesh for population estimation.

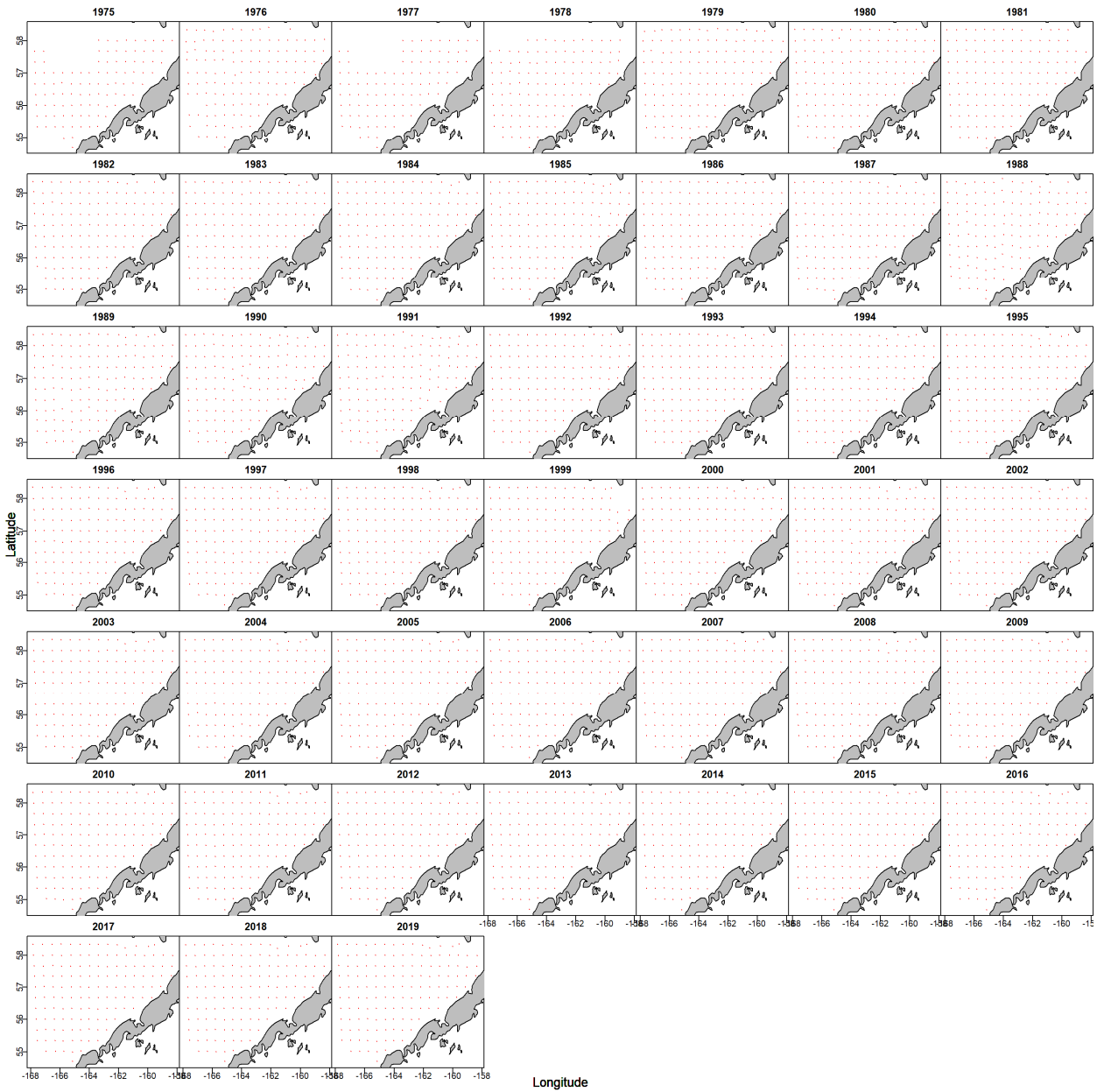


Figure E4 (file = Data\_by\_year.png). Data sampling locations by year for males GE65 mm. Note standardization of survey area in this subregion of the Bering Sea by 1980.

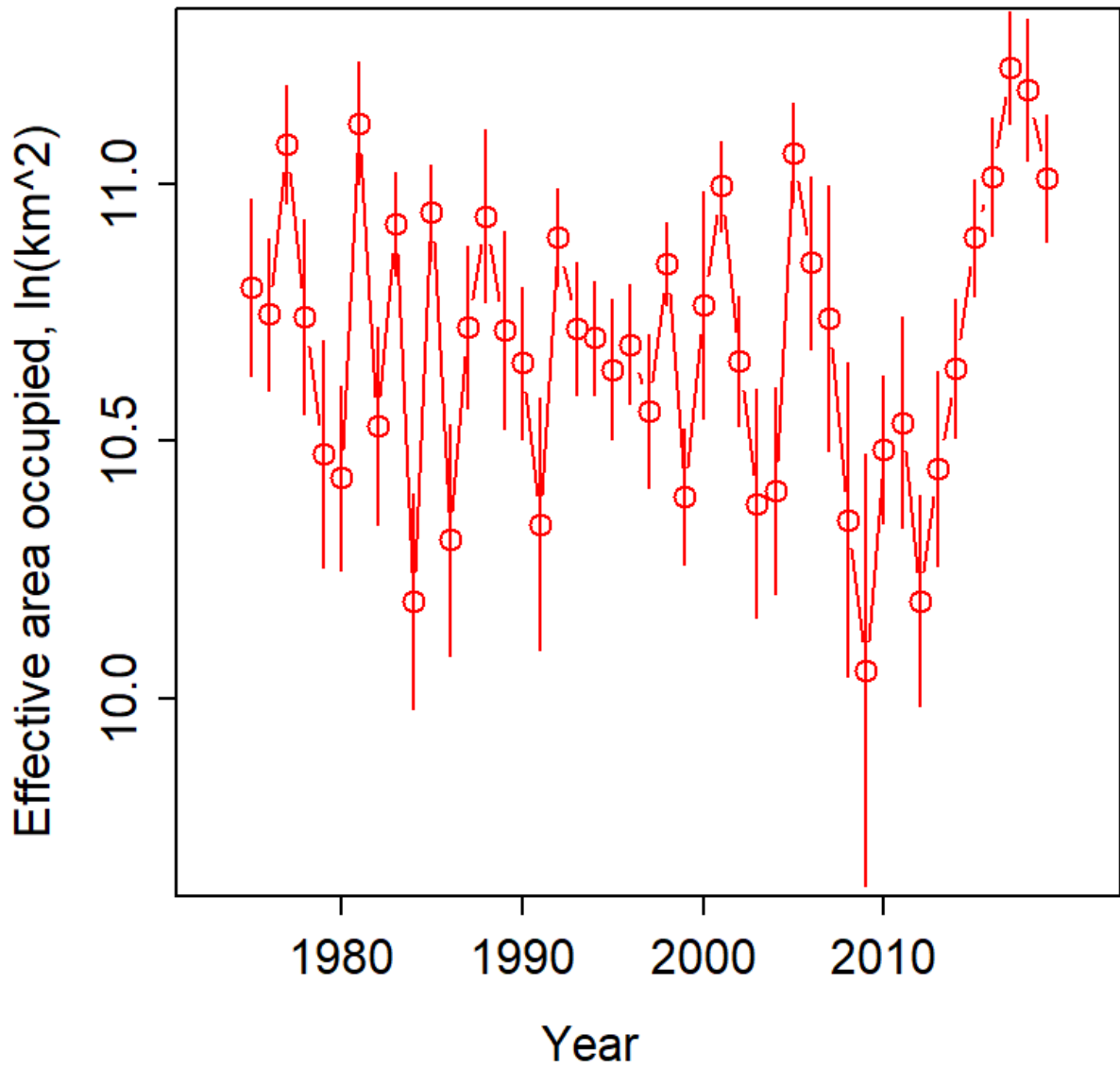


Figure E5 (file = Effective\_Area.png). Effective area occupied, in ln(km<sup>2</sup>) for males GE65 mm. Note modest expansion in area occupied in recent years, despite low population sizes.

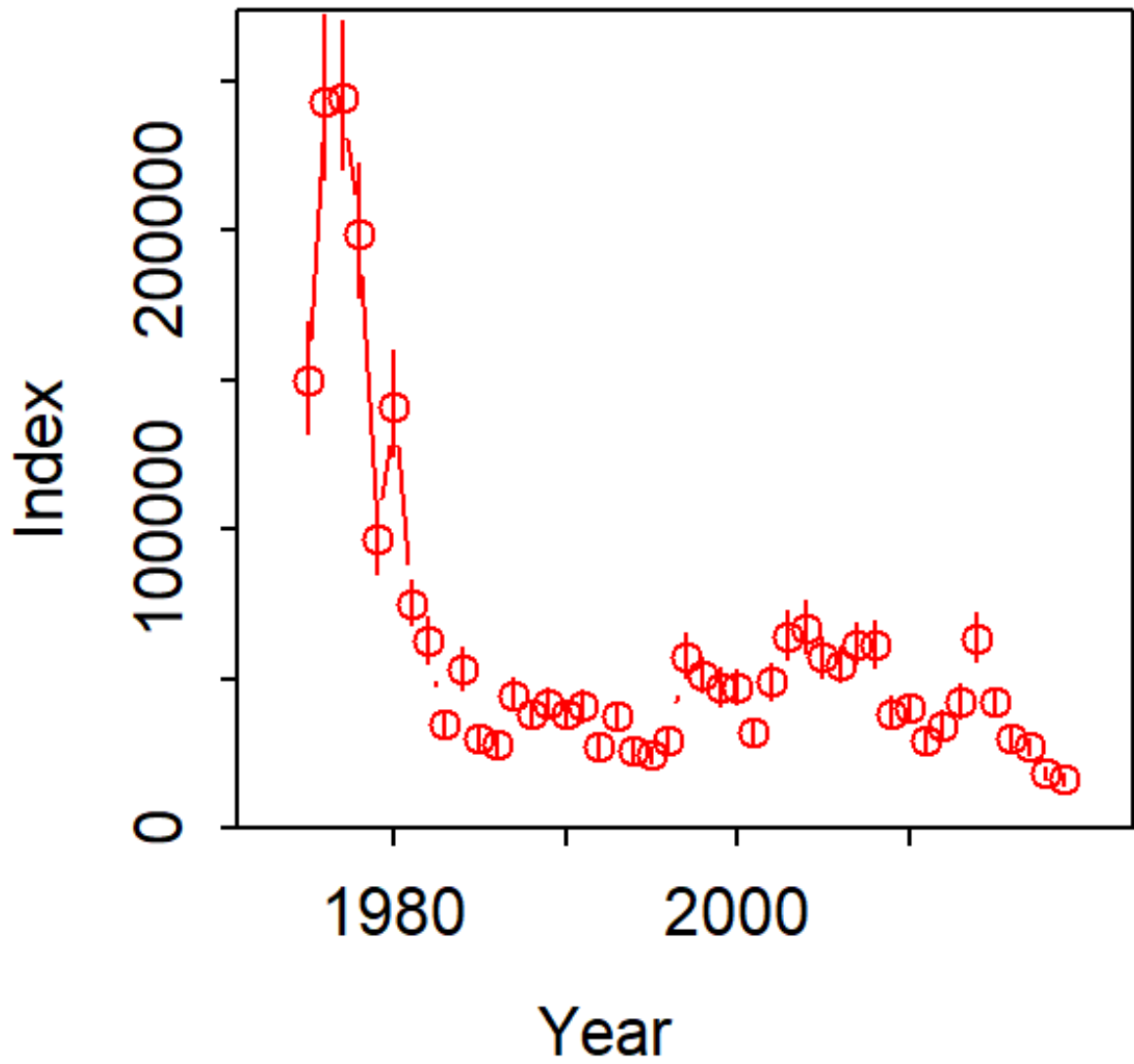


Figure E6 (file = Index-Biomass.png). Biomass estimates by year, with +/- 1SD error bars for males GE65 mm.

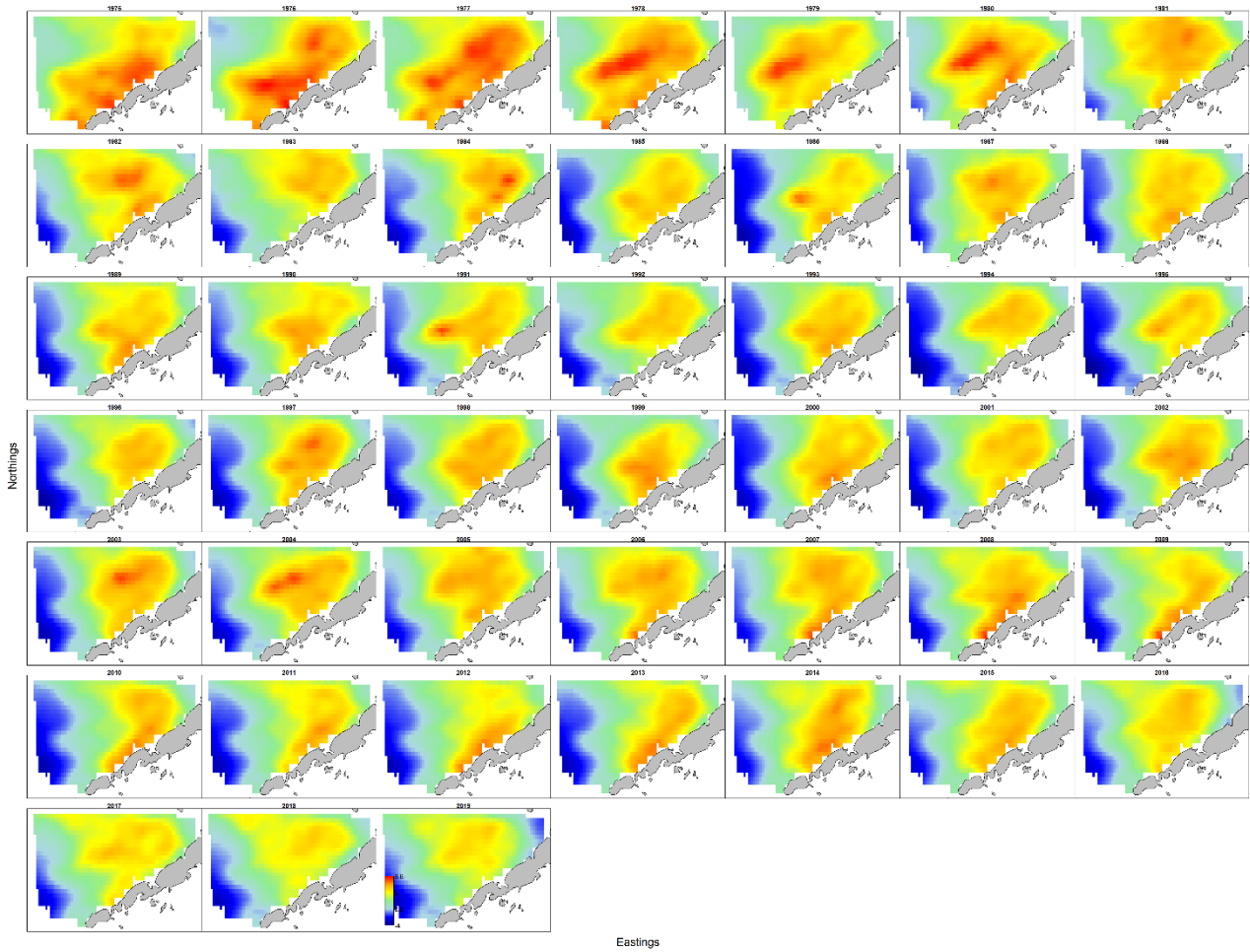


Figure E7 (file = ln\_density-predicted.png). Annual log-density maps for males GE65 mm in  $\ln(\text{kg})/\text{km}^2$ .



DHARMA residual diagnostics

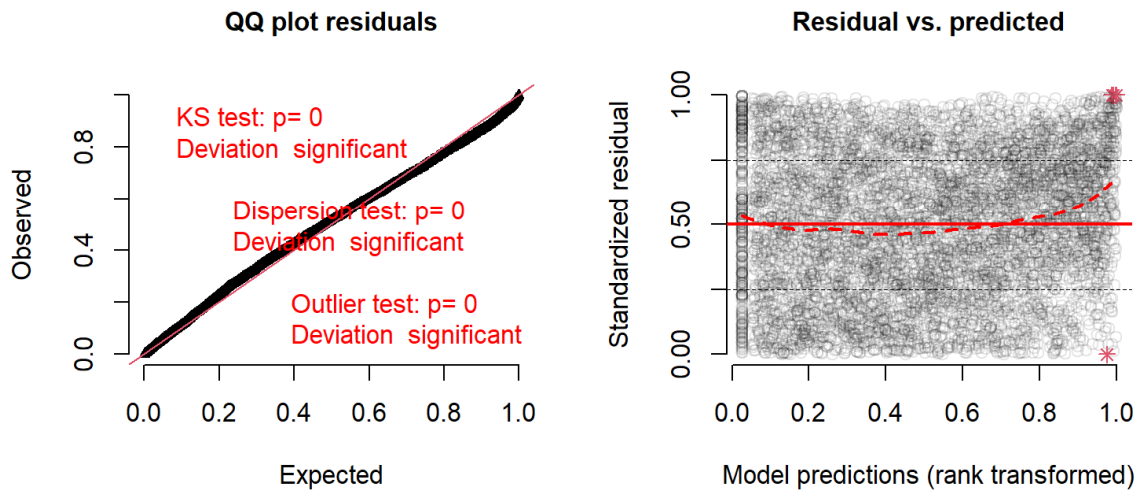


Figure E8 (file = quantile\_residuals). DHARMA residual output for males GE65 mm. a.) Q-Q plot of residuals with p-values for statistical test applied automatically by DHARMA package. These p-values have no meaning here and should be ignored. This figure indicates excellent model performance. b.) Rank-transformed model predictions vs. standardized residuals. Note an indication of a trend in residuals for largest model predictions, as indicated by the dashed red line.

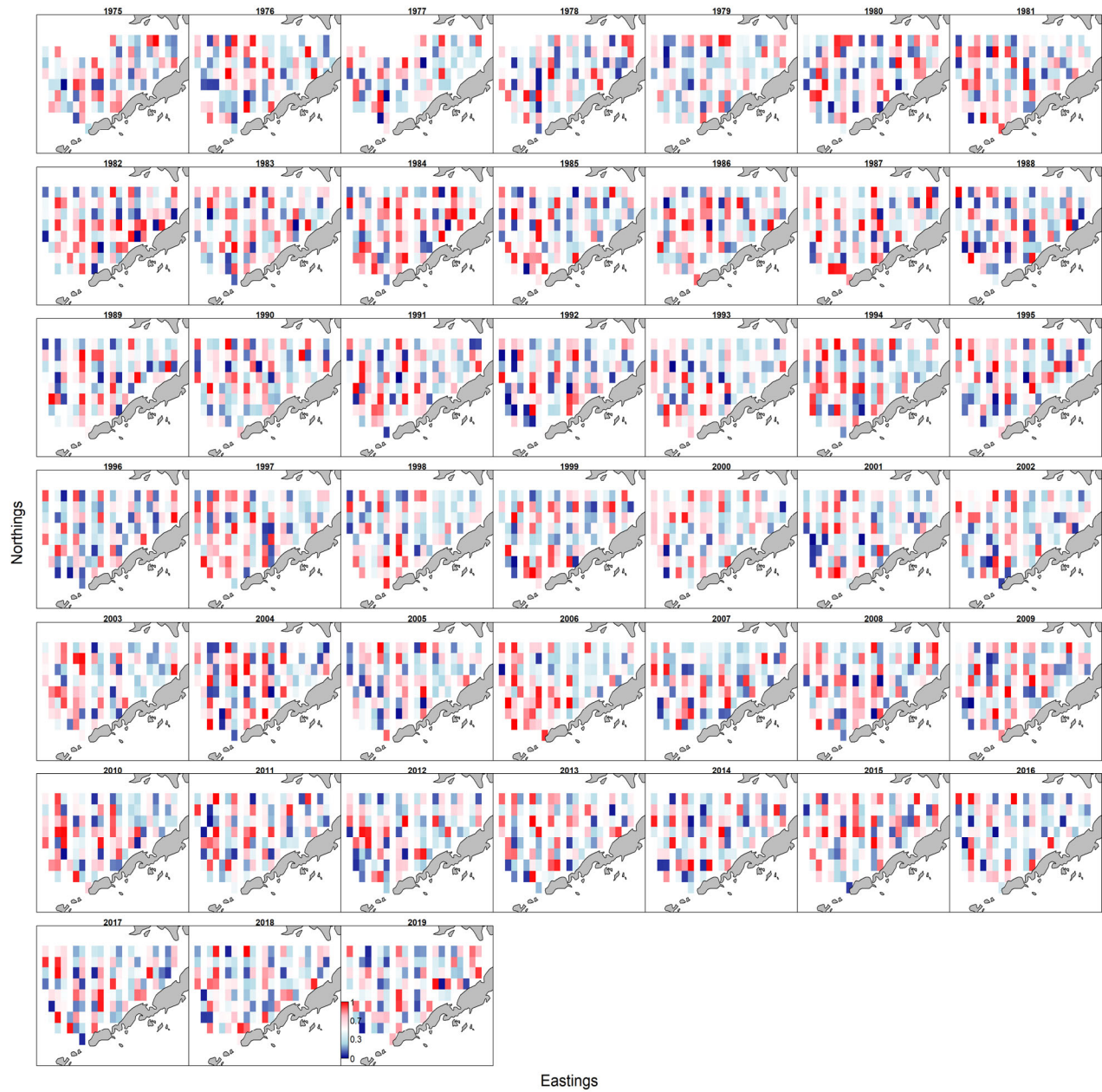


Figure E9 (file = quantile\_residuals\_on\_map.png). Spatial quantile residuals for males GE65 mm with residual “blocks” enlarged for easier viewing. Mapped quantile residuals used to assess for spatial-temporal trends in residuals.

B. Results and diagnostics for females GE65 mm.

Table E5 (extracted from parameter\_estimates.txt). Model parameter estimates for females GE65 mm. Note that L\_epsilon2\_z, the parameter for spatio-temporal variation in biomass only is disabled for this model, due to insufficient variability.

ln_H_input	ln_H_input	beta1_ft	beta1_ft	beta1_ft	beta1_ft
0.146100263	0.335439773	1.394325660	2.341343906	3.146435844	2.240147950
beta1_ft	beta1_ft	beta1_ft	beta1_ft	beta1_ft	beta1_ft
1.544683844	1.420755685	1.768563798	1.885564330	0.884382486	-0.715785757
beta1_ft	beta1_ft	beta1_ft	beta1_ft	beta1_ft	beta1_ft
-0.086417604	-0.516449100	-0.278811523	-0.076622782	0.121955284	-0.227320683
beta1_ft	beta1_ft	beta1_ft	beta1_ft	beta1_ft	beta1_ft
-0.476972715	-0.155499020	-0.706368904	-1.206821249	-0.471400811	-1.182369413
beta1_ft	beta1_ft	beta1_ft	beta1_ft	beta1_ft	beta1_ft
-0.474776530	-0.253160062	-0.136680721	0.022736495	0.080478118	0.004539246
beta1_ft	beta1_ft	beta1_ft	beta1_ft	beta1_ft	beta1_ft
0.841460868	0.545127930	0.813091430	0.717619912	0.792523451	1.384130833
beta1_ft	beta1_ft	beta1_ft	beta1_ft	beta1_ft	beta1_ft
0.886685370	0.232719027	0.306593314	0.078341087	-0.004931193	0.404625040
beta1_ft	beta1_ft	beta1_ft	beta1_ft	beta1_ft	L_omega1_z
0.449762379	1.311637780	0.675056648	0.026527546	0.595523063	1.635392198
L_epsilon1_z	logkappa1	beta2_ft	beta2_ft	beta2_ft	beta2_ft
1.348149780	-3.906944168	0.562883653	-0.174425173	0.054894893	0.674246584
beta2_ft	beta2_ft	beta2_ft	beta2_ft	beta2_ft	beta2_ft
1.026661100	0.466055826	0.174444556	0.509521540	0.084256494	0.174576625
beta2_ft	beta2_ft	beta2_ft	beta2_ft	beta2_ft	beta2_ft
-0.303033573	-0.296763444	0.396748401	0.049622941	0.042630133	0.454973765
beta2_ft	beta2_ft	beta2_ft	beta2_ft	beta2_ft	beta2_ft
0.380748130	0.096104278	0.499816576	0.338381949	0.088470394	0.946669927
beta2_ft	beta2_ft	beta2_ft	beta2_ft	beta2_ft	beta2_ft
0.395813197	0.620655046	0.117854114	0.473391876	0.046214413	0.381860888
beta2_ft	beta2_ft	beta2_ft	beta2_ft	beta2_ft	beta2_ft
0.264985655	0.696545399	0.769477160	0.535063131	0.662513134	0.340985439
beta2_ft	beta2_ft	beta2_ft	beta2_ft	beta2_ft	beta2_ft
0.624217204	0.707408611	1.058690257	0.687316264	0.628149876	0.844395425
beta2_ft	beta2_ft	beta2_ft	beta2_ft	beta2_ft	L_omega2_z
0.738138159	0.486927308	0.492058126	0.876385321	1.047568556	0.383852068
logkappa2	logSigmaM				
-5.028463123	-0.355503492				

Table E6 (extracted from parameter\_estimates.txt). Model parameter estimate diagnostic output for females GE65mm. Starting values for model parameters, upper and lower bounds, MLE, and final gradients for parameters.

	Param	starting_value	Lower	MLE	Upper	final_gradient							
1	ln_H_input	0.146091695	-50	0.146100263	50	-9.653041e-09	41	beta1_ft	-0.004929103	-50	-0.004931193	50	4.020118e-11
2	ln_H_input	0.335439730	-50	0.335439773	50	-7.297056e-09	42	beta1_ft	0.404625596	-50	0.404625040	50	-4.409950e-11
3	beta1_ft	1.394348371	-50	1.394325660	50	-8.267076e-11	43	beta1_ft	0.449764982	-50	0.449762379	50	3.015987e-11
4	beta1_ft	2.341365726	-50	2.341343906	50	-1.209370e-10	44	beta1_ft	1.311646264	-50	1.311637780	50	-3.609779e-11
5	beta1_ft	3.146496395	-50	3.146435844	50	6.357048e-11	45	beta1_ft	0.675066284	-50	0.675056648	50	-5.515499e-11
6	beta1_ft	2.240170799	-50	2.240147950	50	-1.159544e-10	46	beta1_ft	0.026527144	-50	0.026527546	50	-1.961276e-11
7	beta1_ft	1.544692841	-50	1.544683844	50	-8.348544e-11	47	beta1_ft	0.595533673	-50	0.595523063	50	-9.038992e-12
8	beta1_ft	1.420766630	-50	1.420755685	50	-8.020110e-11	48	L_omega1_z	1.635392629	-50	1.635392198	50	9.269360e-08
9	beta1_ft	1.768591432	-50	1.768563798	50	-1.013314e-10	49	L_epsilon1_z	1.348153007	-50	1.348149780	50	2.988806e-08
10	beta1_ft	1.885584713	-50	1.885564330	50	-6.898127e-11	50	logkappa1	-3.906941244	-50	-3.906944168	50	1.784426e-07
11	beta1_ft	0.884396030	-50	0.884382486	50	-2.594280e-11	51	beta2_ft	0.562881952	-50	0.562883653	50	-1.221787e-10
12	beta1_ft	-0.715801814	-50	-0.715785757	50	5.493894e-11	52	beta2_ft	-0.174423669	-50	-0.174425173	50	-4.733387e-10
13	beta1_ft	-0.086421362	-50	-0.086417604	50	-7.910117e-12	53	beta2_ft	0.054919598	-50	0.054894893	50	-1.771829e-09
14	beta1_ft	-0.516459319	-50	-0.516449100	50	2.713141e-11	54	beta2_ft	0.674248705	-50	0.674246584	50	6.082779e-11
15	beta1_ft	-0.278815369	-50	-0.278811523	50	1.251732e-11	55	beta2_ft	1.026674068	-50	1.026661100	50	4.790692e-10
16	beta1_ft	-0.076612194	-50	-0.076622782	50	1.154499e-11	56	beta2_ft	0.466071936	-50	0.466055826	50	-1.897504e-11
17	beta1_ft	0.121953818	-50	0.121955284	50	5.629452e-11	57	beta2_ft	0.174461348	-50	0.174444556	50	-3.870415e-10
18	beta1_ft	-0.227323362	-50	-0.227320683	50	4.072338e-10	58	beta2_ft	0.509528518	-50	0.509521540	50	-1.217323e-09
19	beta1_ft	-0.476983698	-50	-0.476972715	50	7.448564e-11	59	beta2_ft	0.084270318	-50	0.084256494	50	-1.011193e-09
20	beta1_ft	-0.155500315	-50	-0.155499020	50	6.634204e-11	60	beta2_ft	0.174584640	-50	0.174576625	50	1.102949e-09
21	beta1_ft	-0.706384273	-50	-0.706368904	50	5.173756e-11	61	beta2_ft	-0.303015950	-50	-0.303033573	50	1.287417e-09
22	beta1_ft	-1.206851684	-50	-1.206821249	50	9.032272e-11	62	beta2_ft	-0.296757609	-50	-0.296763444	50	7.395640e-10
23	beta1_ft	-0.471410063	-50	-0.471400811	50	9.187281e-11	63	beta2_ft	0.396757683	-50	0.396748401	50	-6.126898e-11
24	beta1_ft	-1.182396055	-50	-1.182369413	50	-2.298062e-11	64	beta2_ft	0.049632222	-50	0.049622941	50	-1.211165e-09
25	beta1_ft	-0.474787880	-50	-0.474776530	50	1.566880e-11	65	beta2_ft	0.042644057	-50	0.042630133	50	-6.045675e-10
26	beta1_ft	-0.253167704	-50	-0.253160062	50	1.378764e-11	66	beta2_ft	0.454988690	-50	0.454973765	50	-1.370531e-09
27	beta1_ft	-0.136685616	-50	-0.136680721	50	4.852241e-11	67	beta2_ft	0.380765824	-50	0.380748130	50	8.237375e-10
28	beta1_ft	0.022735514	-50	0.022736495	50	3.783440e-11	68	beta2_ft	0.096132950	-50	0.096104278	50	-2.664642e-10
29	beta1_ft	0.080476508	-50	0.080478118	50	3.248712e-11	69	beta2_ft	0.499839908	-50	0.499816576	50	1.081549e-09
30	beta1_ft	0.004538283	-50	0.004539246	50	7.227841e-11	70	beta2_ft	0.338385968	-50	0.338381949	50	1.239133e-09
31	beta1_ft	0.841469015	-50	0.841460868	50	-2.181366e-11	71	beta2_ft	0.088484872	-50	0.088470394	50	6.714993e-10
32	beta1_ft	0.545131348	-50	0.545127930	50	-2.765788e-12	72	beta2_ft	0.946659615	-50	0.946669927	50	2.329168e-10
33	beta1_ft	0.813100790	-50	0.813091430	50	-1.728762e-11	73	beta2_ft	0.395836436	-50	0.395813197	50	5.992442e-10
34	beta1_ft	0.717622143	-50	0.717619912	50	-5.641043e-12	74	beta2_ft	0.620663270	-50	0.620655046	50	7.877254e-10
35	beta1_ft	0.792534494	-50	0.792523451	50	-5.767187e-11	75	beta2_ft	0.117869201	-50	0.117854114	50	1.236926e-09
36	beta1_ft	1.384143722	-50	1.384130833	50	-1.862910e-11	76	beta2_ft	0.473402462	-50	0.473391876	50	1.428058e-10
37	beta1_ft	0.886699698	-50	0.886685370	50	-2.849565e-11	77	beta2_ft	0.046229013	-50	0.046214413	50	4.160325e-10
38	beta1_ft	0.232725065	-50	0.232719027	50	-7.110945e-11	78	beta2_ft	0.381867226	-50	0.381860888	50	7.518164e-11
39	beta1_ft	0.306602838	-50	0.306593314	50	-6.003775e-11	79	beta2_ft	0.264998351	-50	0.264985655	50	-3.514602e-10
40	beta1_ft	0.078351311	-50	0.078341087	50	-5.813749e-11	80	beta2_ft	0.696556349	-50	0.696545399	50	7.955636e-10

Table E7 (extracted from parameter\_estimates.txt). Model parameter estimates with parameter standard errors for female GE65 mm. Note often large sizes of standard errors relative to estimates. Although final model converged and fit observations well, convergence was problematic to achieve, requiring some model simplification via removal of terms. Large standard errors suggest further simplification in future runs may be beneficial. Male and Total GE65 mm models had much larger parameter estimates relative to the standard errors.

```

$SD
sdreport(.) result
          Estimate Std. Error
ln_H_input    0.146100263 0.06761266
ln_H_input    0.335439773 0.07906780
beta1_ft      1.394325660 0.86433213
beta1_ft      2.341343906 0.81485185
beta1_ft      3.146435844 0.85263019
beta1_ft      2.240147950 0.83286794
beta1_ft      1.544683844 0.81577316
beta1_ft      1.420755685 0.82529713
beta1_ft      1.768563798 0.82288080
beta1_ft      1.885564330 0.81954828
beta1_ft      0.884382486 0.82259773
beta1_ft     -0.715785757 0.88150391
beta1_ft     -0.086417604 0.85232204
beta1_ft     -0.516449100 0.87245768
beta1_ft     -0.278811523 0.86162127
beta1_ft     -0.076622782 0.85741050
beta1_ft      0.121955284 0.85187886
beta1_ft     -0.227320683 0.85614484
beta1_ft     -0.476972715 0.87145988
beta1_ft     -0.155499020 0.85627857
beta1_ft     -0.706368904 0.87757284
beta1_ft     -1.206821249 0.90185952
beta1_ft     -0.471400811 0.87261906
beta1_ft     -1.182369413 0.88980109
beta1_ft     -0.474776530 0.87016378
beta1_ft     -0.253160062 0.86929090
beta1_ft     -0.136680721 0.85578582
beta1_ft      0.022736495 0.85835301
beta1_ft      0.080478118 0.85903922
beta1_ft      0.004539246 0.85876743
beta1_ft      0.841460868 0.83297103
beta1_ft      0.545127930 0.83437917
beta1_ft      0.813091430 0.82964792
beta1_ft      0.717619912 0.84460846
beta1_ft      0.792523451 0.83628097
beta1_ft      1.384130833 0.81889602

```

Table E8 (file = Table\_for\_SS3a.csv). Population index output. Biomass estimates for females GE65 mm, in metric tons, log-SD and SD in metric tons.

Year	Estimate metric tons	SD log	SD mt
1975	55086.37	0.1718	6972.55
1976	65470.34	0.1420	6845.61
1977	133213.32	0.1311	12833.68
1978	122010.88	0.1338	12174.59
1979	52499.74	0.1174	4772.99
1980	66102.64	0.1614	8010.32
1981	54748.46	0.1400	5805.62
1982	59697.72	0.1354	6146.47
1983	11724.42	0.1363	1232.76
1984	20266.61	0.1945	3059.97
1985	6368.07	0.1520	746.18
1986	5701.93	0.1574	692.78
1987	16230.06	0.1606	2014.92
1988	13305.71	0.1982	2041.40
1989	9390.61	0.1584	1141.87
1990	13575.18	0.1825	1913.29
1991	11706.97	0.1618	1463.03
1992	10441.13	0.1505	1205.25
1993	15263.61	0.1642	1917.22
1994	8029.10	0.1607	1002.58
1995	9336.25	0.1517	1085.74
1996	14420.95	0.1601	1796.38
1997	18651.36	0.1707	2444.73
1998	30921.92	0.1458	3426.45
1999	15125.41	0.1804	2074.79
2000	24137.86	0.1887	3438.51
2001	20193.73	0.1635	2491.42
2002	19796.89	0.1601	2402.94
2003	36323.13	0.1470	4055.72
2004	32061.21	0.1453	3526.41
2005	43414.57	0.1410	4673.71
2006	40916.80	0.1505	4682.61
2007	39118.91	0.1378	4089.42
2008	35149.98	0.1674	4413.73
2009	32823.63	0.1509	3719.33
2010	41408.02	0.1580	4961.70
2011	40670.43	0.1553	4859.63
2012	30134.93	0.1924	4449.19
2013	21878.24	0.1930	3172.41
2014	62170.19	0.1728	8065.27
2015	29304.10	0.1784	3951.88
2016	35323.78	0.1519	4100.59
2017	28353.57	0.1384	2989.85
2018	13977.81	0.1479	1633.91
2019	15791.49	0.1251	1582.16

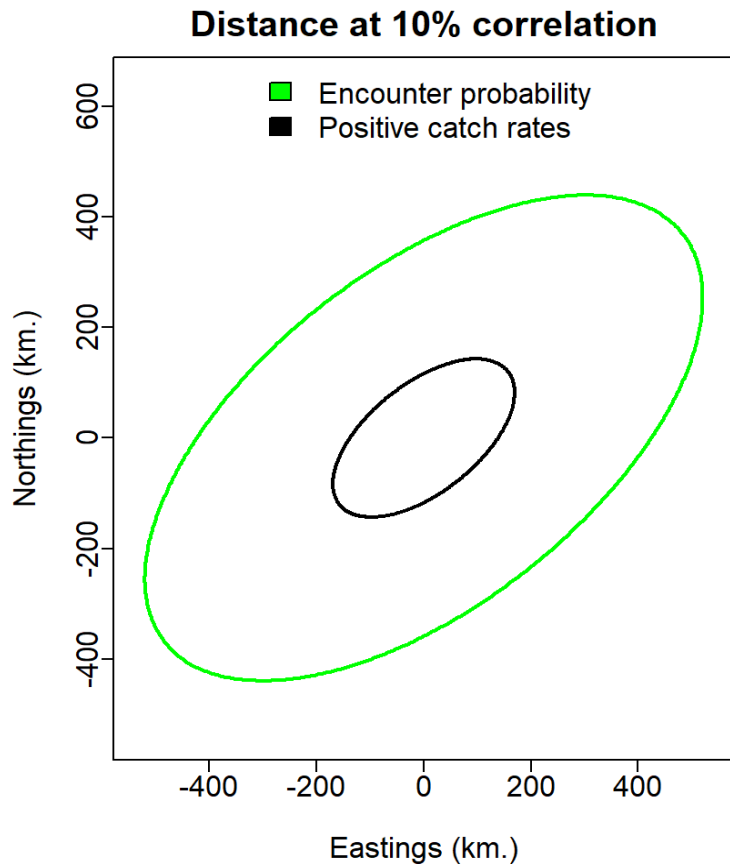


Figure E10 (file = Aniso.png). Geometric anisotropy for females GE65mm. Geometric anisotropy represents the tendency for correlations to decline faster in some directions than others. An ellipse with a major (long) axis pointed northeast-southwest will have correlations that decline slower along this axis than movement northwest-southeast.

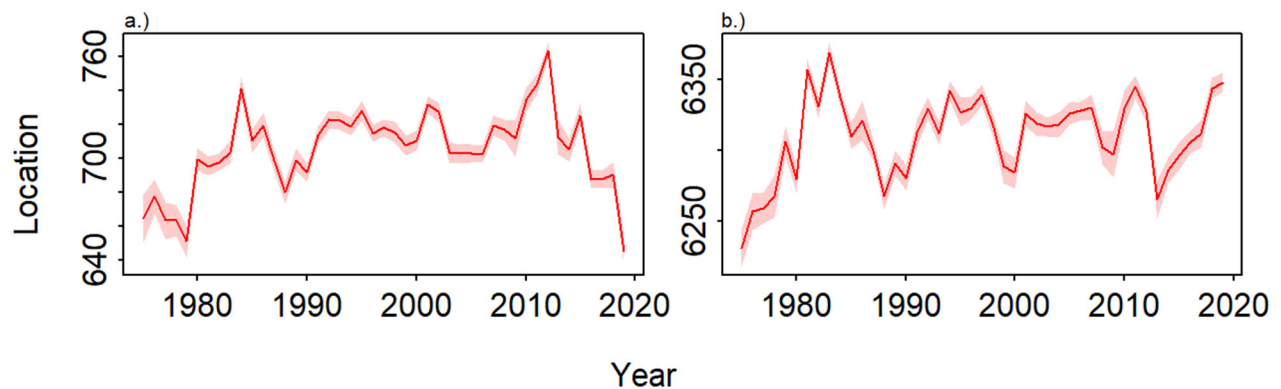


Figure E11 (file = center\_of\_gravity.png). Center of gravity for females GE65 mm: a.) Eastings (km) and b.) Northings (km). These figures indicate that the population has been shifting slightly to the northwest in recent years.

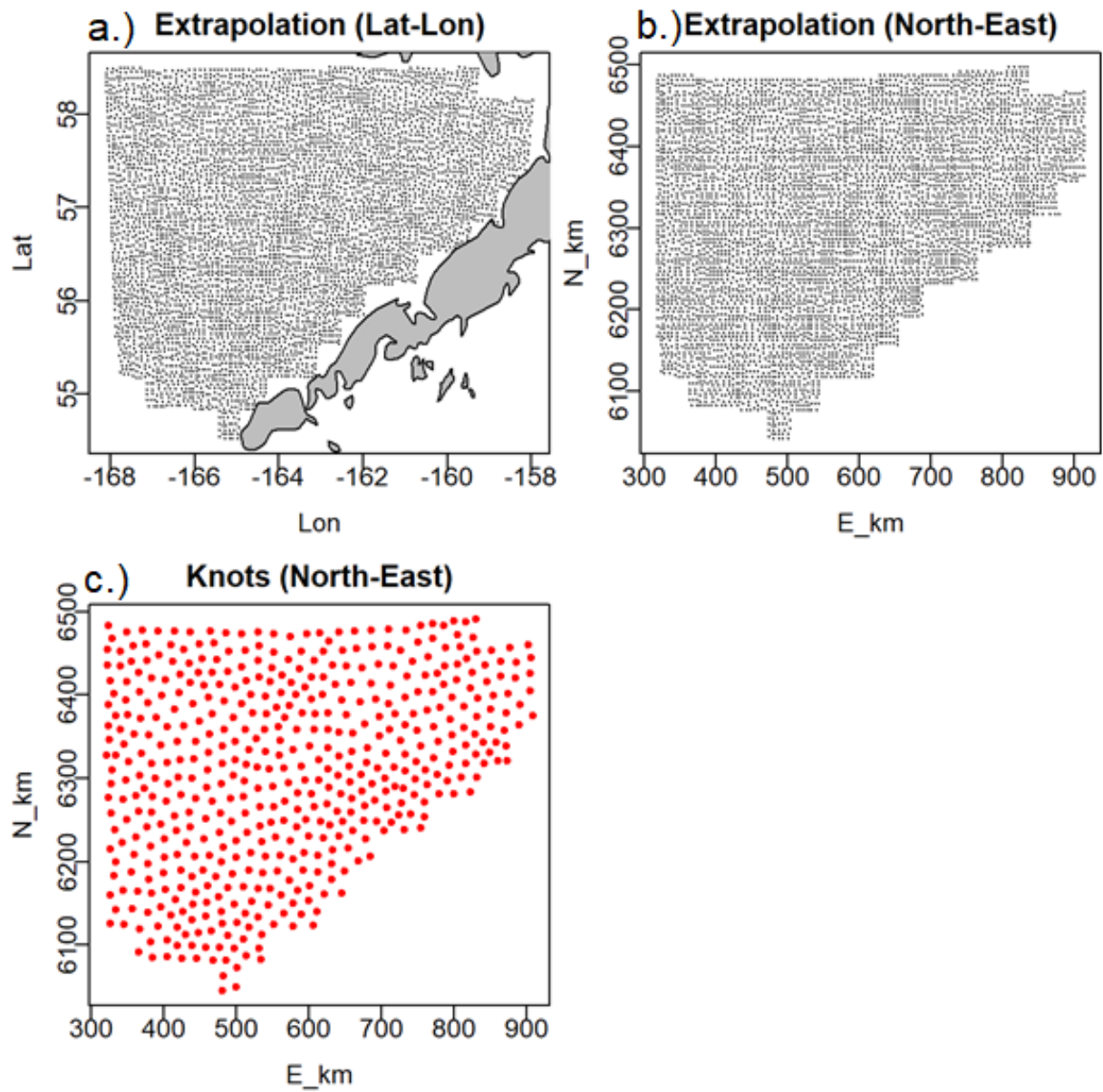


Figure E12 (file = Data\_and\_knots.png). Extrapolation maps for females GE65 mm: a.) Extrapolation grids by latitude and longitude, b.) Extrapolation grid by Northings/Eastings (in km), and c.) Knots used to define spatial mesh for population estimation.



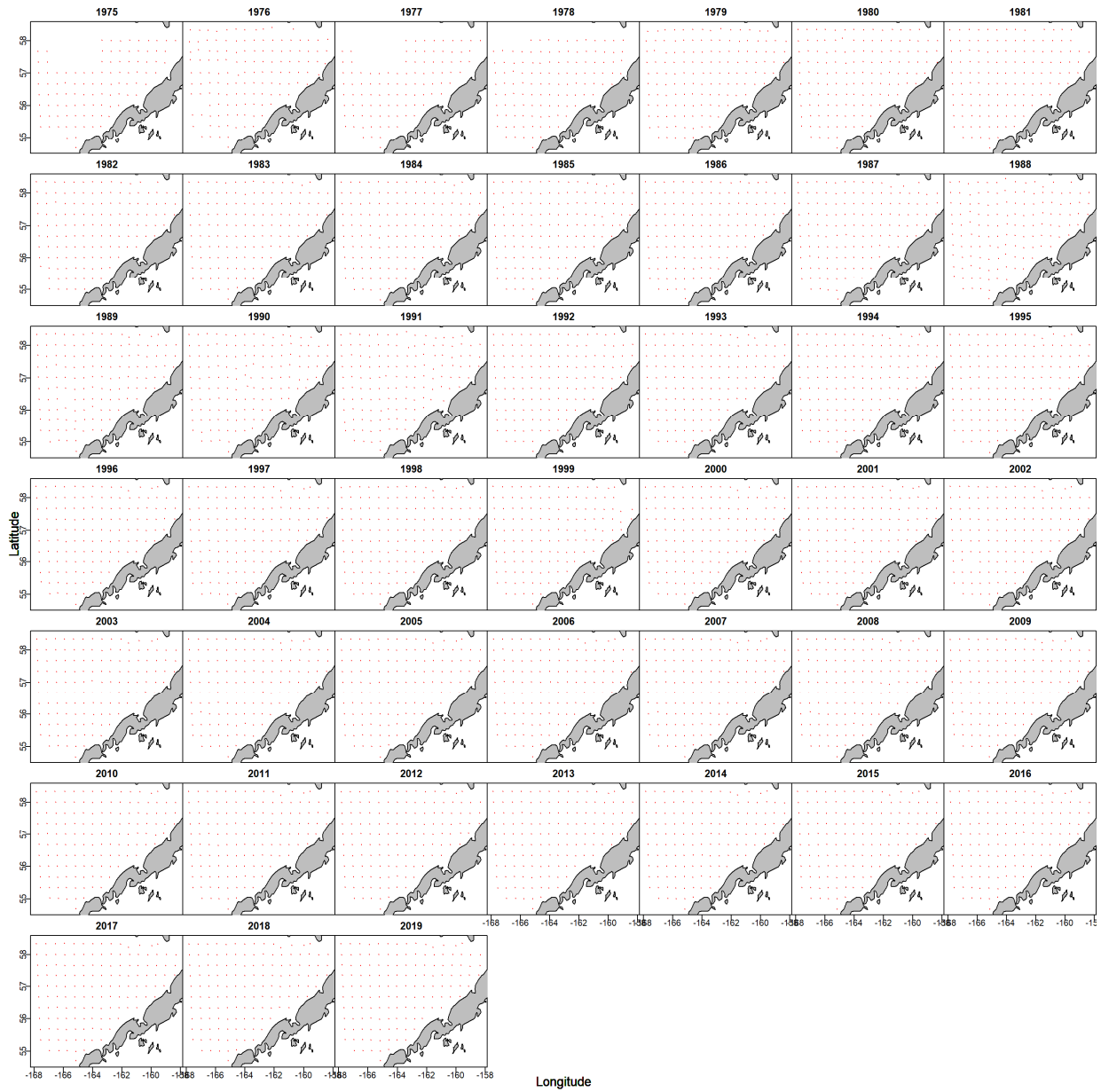


Figure E13 (file = Data\_by\_year.png). Data sampling locations by year for females GE65 mm. Note standardization of survey area in this subregion of the Bering Sea by 1980.

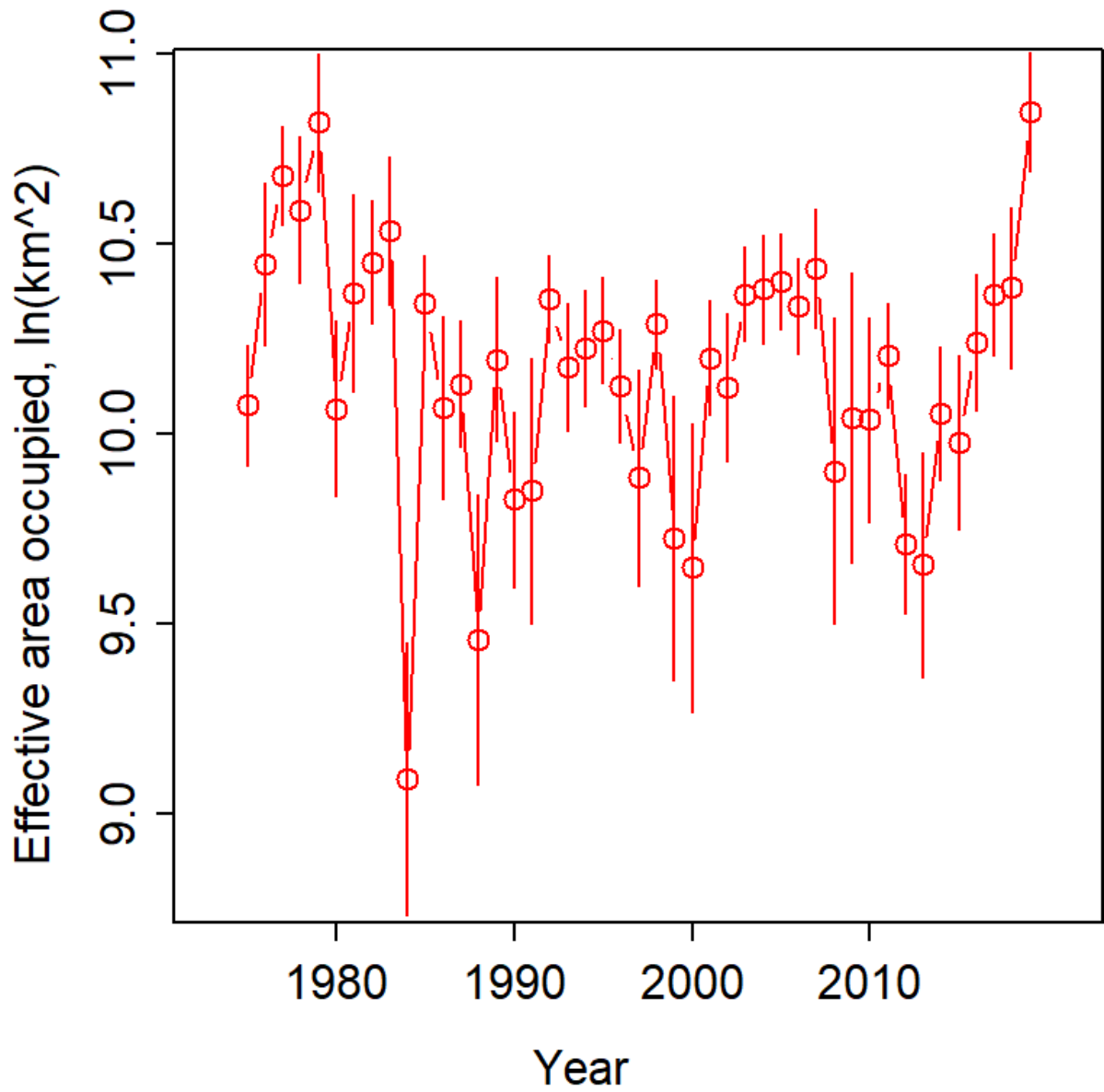


Figure E14 (file = Effective\_Area.png). Effective area occupied, in ln(km<sup>2</sup>) for females GE65 mm. Note modest expansion in area occupied in recent years, despite low population sizes.

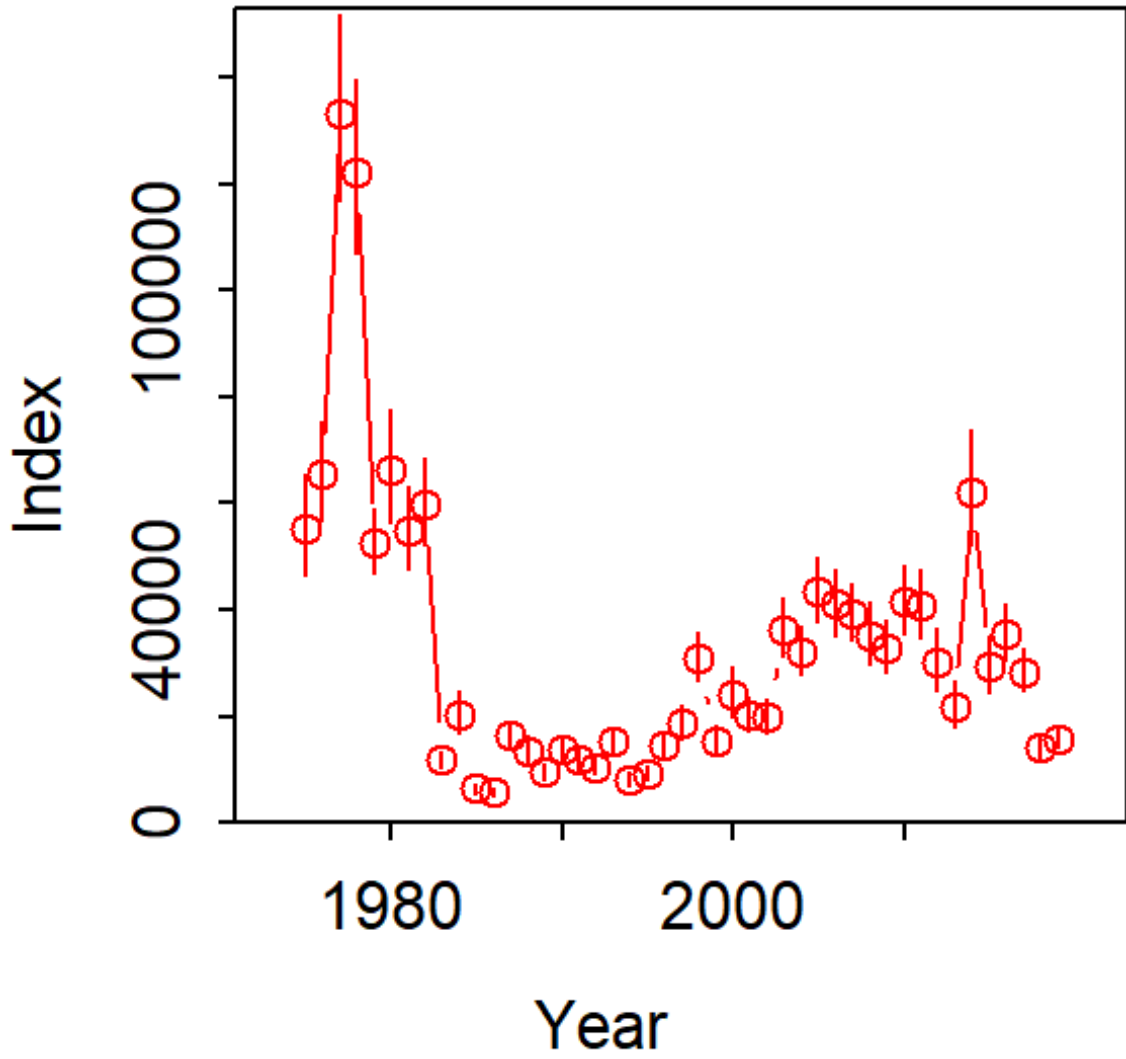


Figure E15 (file = Index-Biomass.png). Biomass estimates by year, with +/- 1SD error bars for females GE65 mm.

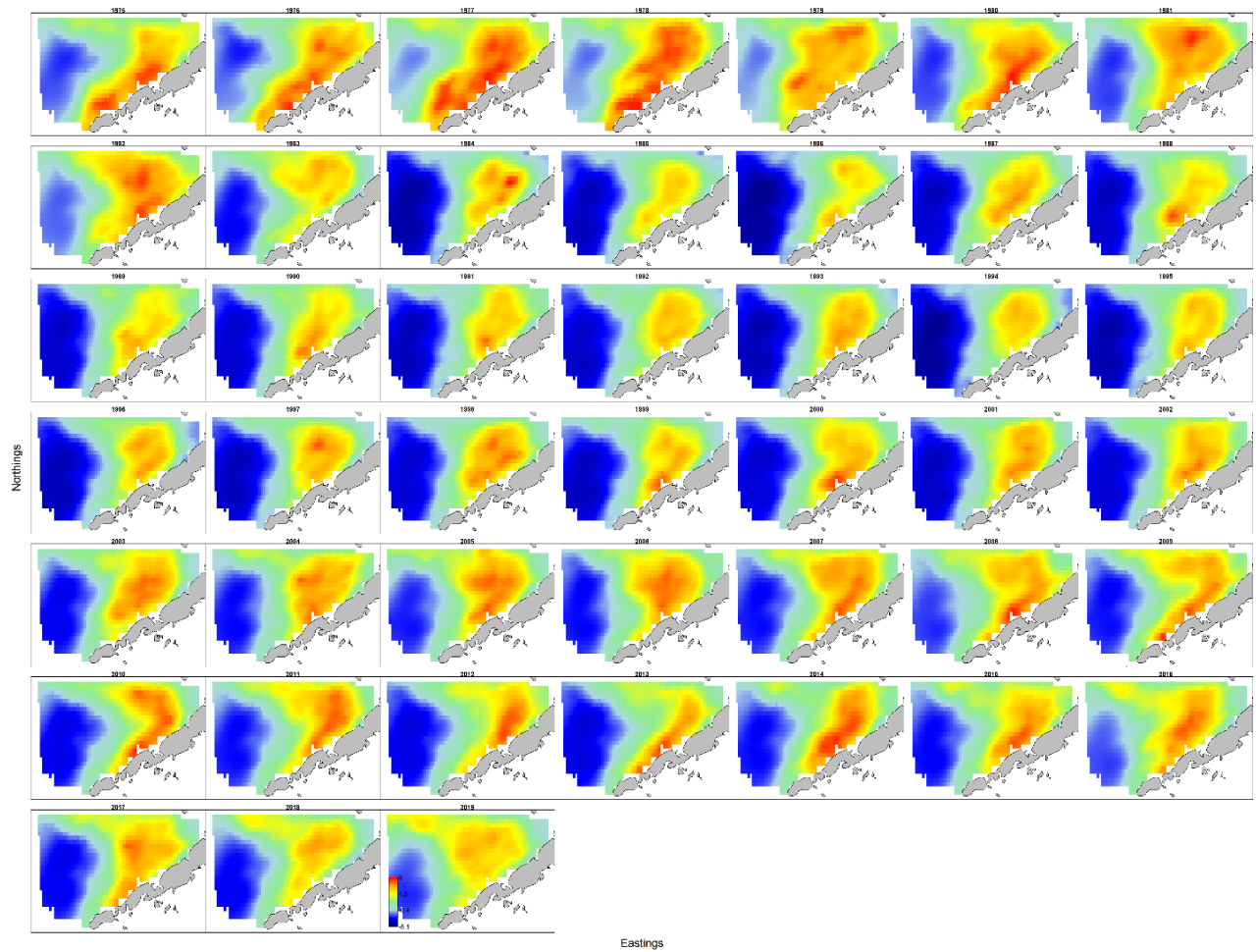


Figure E16 (file = ln\_density-predicted.png). Annual log-density maps for females GE65 mm in  $\ln(\text{kg})/\text{km}^2$ .

DHARMA residual diagnostics

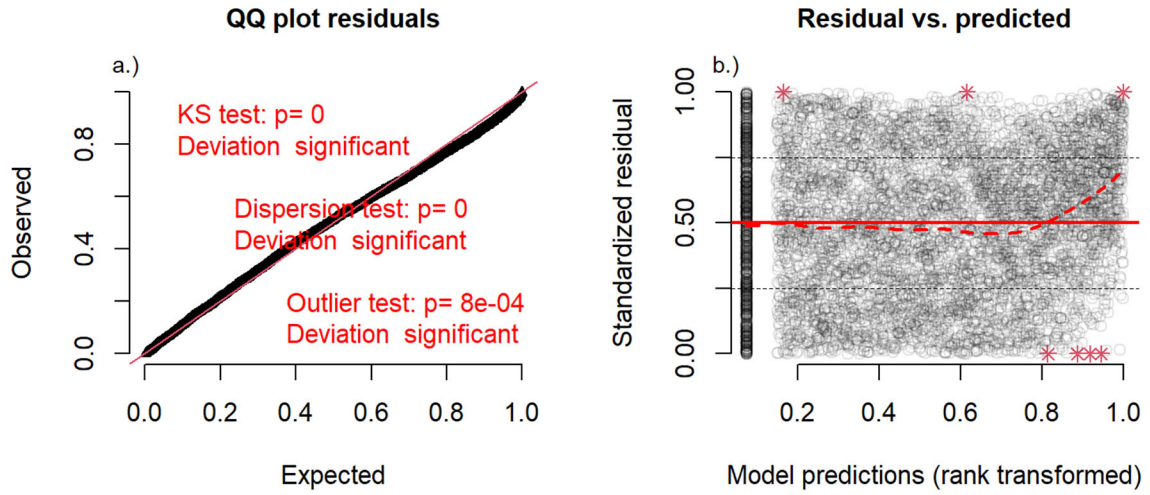


Figure E17 (file = quantile\_residuals). DHARMA residual output for females GE65 mm. a.) Q-Q plot of residuals with p-values for statistical test applied automatically by DHARMA package. These p-values have no meaning here and should be ignored. This figure indicates excellent model performance. b.) Rank-transformed model predictions vs. standardized residuals. Note an indication of a trend in residuals for largest model predictions, as indicated by the dashed red line.

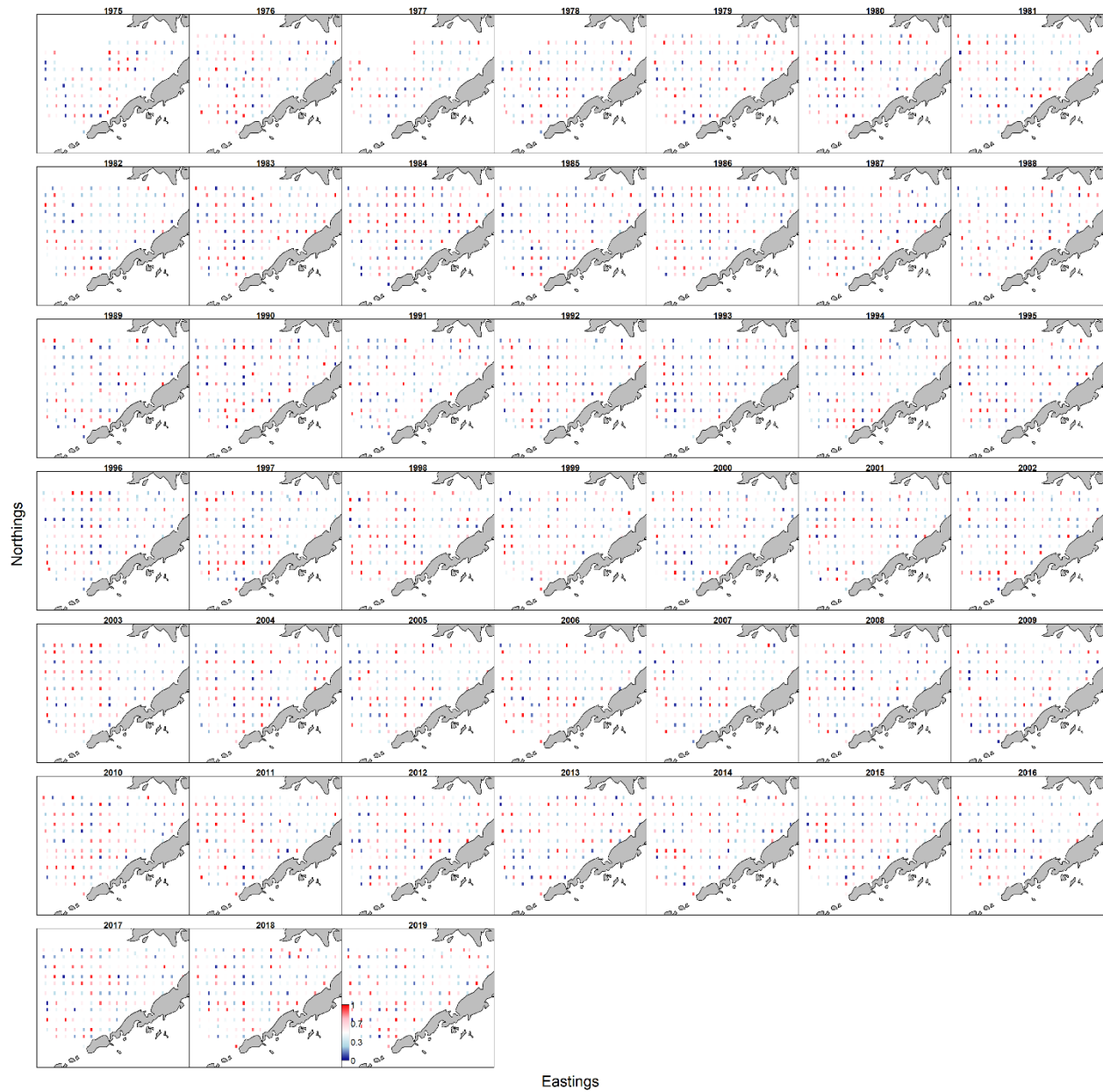


Figure E18 (file = quantile\_residuals\_on\_map.png). Spatial quantile residuals for females GE65 mm. Mapped quantile residuals used to assess for spatial-temporal trends in residuals. While difficult to read here, an absence of notable trends may be observed at higher resolutions. Work is underway to implement code to improve this graphic, please see Figure E9 for an example.

C. Results and diagnostics for total GE65 mm.

Table E9 (extracted from parameter\_estimates.txt). Model parameter estimates for total GE65 mm. Note that L\_epsilon2\_z, the parameter for spatio-temporal variation in biomass only is disabled for this model, due to insufficient variability.

```

ln_H_input ln_H_input beta1_ft beta1_ft beta1_ft beta1_ft
0.43677326 0.18710855 0.54829502 0.17235147 0.81963778 0.49439896
beta1_ft beta1_ft beta1_ft beta1_ft beta1_ft beta1_ft
-0.35855266 -0.31378110 -0.09204112 -0.40514673 -0.57961487 -1.66383056
beta1_ft beta1_ft beta1_ft beta1_ft beta1_ft beta1_ft
-1.40809570 -1.78732456 -1.11514733 -1.38214691 -1.15824325 -1.62861091
beta1_ft beta1_ft beta1_ft beta1_ft beta1_ft beta1_ft
-1.95352757 -1.92301581 -2.52265776 -2.75288338 -2.60297078 -2.53035173
beta1_ft beta1_ft beta1_ft beta1_ft beta1_ft beta1_ft
-2.03379254 -1.77293178 -1.65093602 -1.39177173 -1.81340767 -2.02357425
beta1_ft beta1_ft beta1_ft beta1_ft beta1_ft beta1_ft
-1.23998442 -1.67968188 -1.66557964 -1.10576545 -1.26197607 -0.57466328
beta1_ft beta1_ft beta1_ft beta1_ft beta1_ft beta1_ft
-1.27234779 -1.71376421 -1.89174877 -1.66812545 -1.84010008 -1.25664492
beta1_ft beta1_ft beta1_ft beta1_ft beta1_ft L_omega1_z
-1.12845900 -1.00601021 -0.85921221 -1.58563801 -1.70131456 1.48548278
L_epsilon1_z logkappa1 beta2_ft beta2_ft beta2_ft beta2_ft
1.00890543 -4.36665192 4.40098145 4.73447557 4.71953602 4.68484113
beta2_ft beta2_ft beta2_ft beta2_ft beta2_ft beta2_ft
4.93175596 4.60228232 4.21029499 4.23650101 3.82847586 4.37133805
beta2_ft beta2_ft beta2_ft beta2_ft beta2_ft beta2_ft
3.88486157 3.85357327 4.05994729 4.23289071 4.04901240 4.57177581
beta2_ft beta2_ft beta2_ft beta2_ft beta2_ft beta2_ft
4.56746814 4.52603655 4.82937481 4.46522311 4.53038262 4.73236058
beta2_ft beta2_ft beta2_ft beta2_ft beta2_ft beta2_ft
4.66785284 4.54915366 4.10873769 4.23099596 4.39171184 4.63861582
beta2_ft beta2_ft beta2_ft beta2_ft beta2_ft beta2_ft
4.14067821 4.76548445 5.08584612 4.61494732 5.00157015 4.21858149
beta2_ft beta2_ft beta2_ft beta2_ft beta2_ft beta2_ft
4.50883301 4.38801650 4.73283330 4.51147788 4.60625284 4.56997068
beta2_ft beta2_ft beta2_ft beta2_ft beta2_ft L_omega2_z
4.49295268 4.16123810 4.20421448 4.35722994 4.38539795 0.48545691
L_epsilon2_z logkappa2 logSigmaM
-0.61385674 -3.45977764 -0.35926265

```

Table E10 (extracted from parameter\_estimates.txt). Model parameter estimate diagnostic output for total GE65mm. Starting values for model parameters, upper and lower bounds, MLE, and final gradients for parameters.

	Param	starting_value	Lower	MLE	Upper	final_gradient
1	ln_H_input	0.43677362	-50	0.43677326	50	-8.597976e-09
2	ln_H_input	0.18710868	-50	0.18710855	50	8.751693e-10
3	beta1_ft	0.54827929	-50	0.54829502	50	4.760459e-11
4	beta1_ft	0.17235040	-50	0.17235147	50	4.587708e-11
5	beta1_ft	0.81965636	-50	0.81963778	50	3.157297e-11
6	beta1_ft	0.49442039	-50	0.49439896	50	-1.043610e-11
7	beta1_ft	-0.35854516	-50	-0.35855266	50	1.415312e-11
8	beta1_ft	-0.31377513	-50	-0.31378110	50	2.559020e-11
9	beta1_ft	-0.09203195	-50	-0.09204112	50	1.211653e-11
10	beta1_ft	-0.40515826	-50	-0.40514673	50	6.784440e-11
11	beta1_ft	-0.57961383	-50	-0.57961487	50	2.593126e-11
12	beta1_ft	-1.66381926	-50	-1.66383056	50	-1.858336e-11
13	beta1_ft	-1.40808852	-50	-1.40809570	50	1.607336e-11
14	beta1_ft	-1.78731605	-50	-1.78732456	50	1.014122e-11
15	beta1_ft	-1.11514222	-50	-1.11514733	50	1.630784e-11
16	beta1_ft	-1.38214218	-50	-1.38214691	50	2.212186e-11
17	beta1_ft	-1.15823712	-50	-1.15824325	50	1.974598e-11
18	beta1_ft	-1.62860092	-50	-1.62861091	50	-7.965184e-12
19	beta1_ft	-1.95353975	-50	-1.95352757	50	-3.131204e-10
20	beta1_ft	-1.92300256	-50	-1.92301581	50	-2.252110e-11
21	beta1_ft	-2.52264980	-50	-2.52265776	50	-1.291367e-11
22	beta1_ft	-2.75287439	-50	-2.75288338	50	-6.977086e-12
23	beta1_ft	-2.60295438	-50	-2.60297078	50	-4.847767e-11
24	beta1_ft	-2.53034889	-50	-2.53035173	50	1.385647e-11
25	beta1_ft	-2.03378203	-50	-2.03379254	50	-6.157741e-12
26	beta1_ft	-1.77292701	-50	-1.77293178	50	2.082690e-11
27	beta1_ft	-1.65092759	-50	-1.65093602	50	8.383516e-12
28	beta1_ft	-1.39176733	-50	-1.39177173	50	1.655742e-11
29	beta1_ft	-1.81340264	-50	-1.81340767	50	1.128697e-11
30	beta1_ft	-2.02356864	-50	-2.02357425	50	9.623857e-12
31	beta1_ft	-1.23998109	-50	-1.23998442	50	1.440981e-11
32	beta1_ft	-1.67967173	-50	-1.67968188	50	2.169820e-11
33	beta1_ft	-1.66556065	-50	-1.66557964	50	-5.094236e-11
34	beta1_ft	-1.10575082	-50	-1.10576545	50	-3.726797e-12
35	beta1_ft	-1.26196297	-50	-1.26197607	50	7.757350e-12
36	beta1_ft	-0.57464545	-50	-0.57466328	50	-2.125589e-11
37	beta1_ft	-1.27233402	-50	-1.27234779	50	-1.682121e-11
38	beta1_ft	-1.71375878	-50	-1.71376421	50	1.617195e-11



39	beta1_ft	-1.89174362	-50	-1.89174877	50	-1.006306e-11
40	beta1_ft	-1.66812166	-50	-1.66812545	50	1.604361e-11
41	beta1_ft	-1.84009411	-50	-1.84010008	50	9.063861e-12
42	beta1_ft	-1.25664125	-50	-1.25664492	50	2.242828e-11
43	beta1_ft	-1.12845190	-50	-1.12845900	50	6.816769e-12
44	beta1_ft	-1.00598923	-50	-1.00601021	50	-3.577316e-11
45	beta1_ft	-0.85919644	-50	-0.85921221	50	-3.243628e-12
46	beta1_ft	-1.58563531	-50	-1.58563801	50	2.014167e-11
47	beta1_ft	-1.70130979	-50	-1.70131456	50	9.175549e-12
48	L_omega1_z	1.48548338	-50	1.48548278	50	-3.586138e-08
49	L_epsilon1_z	1.00890614	-50	1.00890543	50	2.264434e-08
50	logkappa1	-4.36665182	-50	-4.36665192	50	-3.174932e-08
51	beta2_ft	4.40099140	-50	4.40098145	50	1.024958e-11
52	beta2_ft	4.73446957	-50	4.73447557	50	3.198810e-10
53	beta2_ft	4.71955493	-50	4.71953602	50	-5.154011e-10
54	beta2_ft	4.68484262	-50	4.68484113	50	-5.099388e-11
55	beta2_ft	4.93176019	-50	4.93175596	50	-7.858070e-11
56	beta2_ft	4.60227752	-50	4.60228232	50	3.593552e-10
57	beta2_ft	4.21030675	-50	4.21029499	50	-3.163372e-10
58	beta2_ft	4.23650814	-50	4.23650101	50	-9.772450e-11
59	beta2_ft	3.82848617	-50	3.82847586	50	-1.657128e-10
60	beta2_ft	4.37134647	-50	4.37133805	50	-2.012817e-10
61	beta2_ft	3.88487619	-50	3.88486157	50	-3.851337e-10
62	beta2_ft	3.85356909	-50	3.85357327	50	2.813163e-10
63	beta2_ft	4.05995021	-50	4.05994729	50	9.005774e-11
64	beta2_ft	4.23289750	-50	4.23289071	50	-1.068905e-10
65	beta2_ft	4.04901962	-50	4.04901240	50	-1.221689e-10
66	beta2_ft	4.57178778	-50	4.57177581	50	-2.042349e-10
67	beta2_ft	4.56747924	-50	4.56746814	50	5.318981e-10
68	beta2_ft	4.52604992	-50	4.52603655	50	-4.666472e-10
69	beta2_ft	4.82938703	-50	4.82937481	50	-2.455396e-10
70	beta2_ft	4.46522435	-50	4.46522311	50	1.497131e-10
71	beta2_ft	4.53040008	-50	4.53038262	50	-4.307434e-10
72	beta2_ft	4.73234464	-50	4.73236058	50	6.076863e-10
73	beta2_ft	4.66785179	-50	4.66785284	50	1.187459e-10
74	beta2_ft	4.54915140	-50	4.54915366	50	1.475424e-10
75	beta2_ft	4.10875137	-50	4.10873769	50	-1.811760e-10
76	beta2_ft	4.23100237	-50	4.23099596	50	-7.147882e-11
77	beta2_ft	4.39172456	-50	4.39171184	50	-1.877449e-10
78	beta2_ft	4.63862457	-50	4.63861582	50	-1.289262e-10
79	beta2_ft	4.14068722	-50	4.14067821	50	-1.351488e-10
80	beta2_ft	4.76546374	-50	4.76548445	50	7.154171e-10
81	beta2_ft	5.08588056	-50	5.08584612	50	-2.155015e-09

82	beta2_ft	4.61496308	-50	4.61494732	50	-6.560317e-10
83	beta2_ft	5.00157949	-50	5.00157015	50	-3.255085e-10
84	beta2_ft	4.21857503	-50	4.21858149	50	4.254943e-10
85	beta2_ft	4.50883858	-50	4.50883301	50	1.476863e-11
86	beta2_ft	4.38801013	-50	4.38801650	50	2.635065e-10
87	beta2_ft	4.73285249	-50	4.73283330	50	-3.754783e-10
88	beta2_ft	4.51147972	-50	4.51147788	50	1.122276e-10
89	beta2_ft	4.60625707	-50	4.60625284	50	2.724576e-11
90	beta2_ft	4.56997268	-50	4.56997068	50	1.009717e-10
91	beta2_ft	4.49296250	-50	4.49295268	50	-2.314771e-10
92	beta2_ft	4.16123304	-50	4.16123810	50	3.339302e-10
93	beta2_ft	4.20421884	-50	4.20421448	50	-1.521059e-10
94	beta2_ft	4.35722779	-50	4.35722994	50	3.761693e-10
95	beta2_ft	4.38540426	-50	4.38539795	50	1.736034e-11
96	L_omega2_z	0.48545775	-50	0.48545691	50	-2.974647e-08
97	L_epsilon2_z	-0.61385663	-50	-0.61385674	50	9.579793e-09
98	logkappa2	-3.45977963	-50	-3.45977764	50	2.510703e-08
99	logSigmaM	-0.35926334	-50	-0.35926265	50	5.326118e-08

Table E11 (extracted from parameter\_estimates.txt). Model parameter estimates with parameter standard errors for total GE65 mm. Note often large sizes of standard errors relative to estimates. Although final model converged and fit observations well, convergence was problematic to achieve, requiring some model simplification via removal of terms. Large standard errors suggest further simplification in future runs may be beneficial. Male and Total GE65 mm models had much larger parameter estimates relative to the standard errors.

	Estimate	Std. Error
ln_H_input	0.43677326	0.07125020
ln_H_input	0.18710855	0.06922590
beta1_ft	0.54829502	1.06521177
beta1_ft	0.17235147	1.03844497
beta1_ft	0.81963778	1.05499567
beta1_ft	0.49439896	1.04101278
beta1_ft	-0.35855266	1.02911677
beta1_ft	-0.31378110	1.03316275
beta1_ft	-0.09204112	1.03492621
beta1_ft	-0.40514673	1.04348530
beta1_ft	-0.57961487	1.03701595
beta1_ft	-1.66383056	1.05150023
beta1_ft	-1.40809570	1.04970992
beta1_ft	-1.78732456	1.05496242
beta1_ft	-1.11514733	1.04197956

beta1\_ft -1.38214691 1.04995339  
beta1\_ft -1.15824325 1.04378422  
beta1\_ft -1.62861091 1.04885595  
beta1\_ft -1.95352757 1.04978080  
beta1\_ft -1.92301581 1.04899738  
beta1\_ft -2.52265776 1.06603223  
beta1\_ft -2.75288338 1.07208896  
beta1\_ft -2.60297078 1.06748590  
beta1\_ft -2.53035173 1.06195500  
beta1\_ft -2.03379254 1.05593722  
beta1\_ft -1.77293178 1.05178834  
beta1\_ft -1.65093602 1.04937519  
beta1\_ft -1.39177173 1.05685647  
beta1\_ft -1.81340767 1.05474695  
beta1\_ft -2.02357425 1.05961781  
beta1\_ft -1.23998442 1.04981667  
beta1\_ft -1.67968188 1.04968985  
beta1\_ft -1.66557964 1.04971036  
beta1\_ft -1.10576545 1.04521784  
beta1\_ft -1.26197607 1.04266897  
beta1\_ft -0.57466328 1.04123777  
beta1\_ft -1.27234779 1.04645734  
beta1\_ft -1.71376421 1.05315626  
beta1\_ft -1.89174877 1.05686101  
beta1\_ft -1.66812545 1.05363784  
beta1\_ft -1.84010008 1.05262940  
beta1\_ft -1.25664492 1.04380201  
beta1\_ft -1.12845900 1.04114926  
beta1\_ft -1.00601021 1.03886168  
beta1\_ft -0.85921221 1.04181062  
beta1\_ft -1.58563801 1.04356583  
beta1\_ft -1.70131456 1.04450095  
L\_omega1\_z 1.48548278 0.14028785  
L\_epsilon1\_z 1.00890543 0.04759218  
logkappa1 -4.36665192 0.06735172  
beta2\_ft 4.40098145 0.36978347  
beta2\_ft 4.73447557 0.30669709  
beta2\_ft 4.71953602 0.33380265  
beta2\_ft 4.68484113 0.31147675  
beta2\_ft 4.93175596 0.26781540  
beta2\_ft 4.60228232 0.27236877  
beta2\_ft 4.21029499 0.28164418  
beta2\_ft 4.23650101 0.29997743

beta2\_ft 3.82847586 0.26230395  
beta2\_ft 4.37133805 0.27831774  
beta2\_ft 3.88486157 0.30528643  
beta2\_ft 3.85357327 0.31035370  
beta2\_ft 4.05994729 0.27289567  
beta2\_ft 4.23289071 0.31367090  
beta2\_ft 4.04901240 0.27860818  
beta2\_ft 4.57177581 0.27073996  
beta2\_ft 4.56746814 0.28284895  
beta2\_ft 4.52603655 0.28098105  
beta2\_ft 4.82937481 0.30868137  
beta2\_ft 4.46522311 0.31411678  
beta2\_ft 4.53038262 0.30774403  
beta2\_ft 4.73236058 0.29177052  
beta2\_ft 4.66785284 0.30782346  
beta2\_ft 4.54915366 0.30490868  
beta2\_ft 4.10873769 0.30264172  
beta2\_ft 4.23099596 0.33309644  
beta2\_ft 4.39171184 0.30104200  
beta2\_ft 4.63861582 0.30324470  
beta2\_ft 4.14067821 0.32568258  
beta2\_ft 4.76548445 0.29983933  
beta2\_ft 5.08584612 0.29810572  
beta2\_ft 4.61494732 0.29704305  
beta2\_ft 5.00157015 0.28203320  
beta2\_ft 4.21858149 0.29024624  
beta2\_ft 4.50883301 0.29251135  
beta2\_ft 4.38801650 0.30436965  
beta2\_ft 4.73283330 0.29987888  
beta2\_ft 4.51147788 0.28929694  
beta2\_ft 4.60625284 0.28942773  
beta2\_ft 4.56997068 0.28169234  
beta2\_ft 4.49295268 0.27141072  
beta2\_ft 4.16123810 0.28022206  
beta2\_ft 4.20421448 0.28669247  
beta2\_ft 4.35722994 0.27178128  
beta2\_ft 4.38539795 0.27256156  
L\_omega2\_z 0.48545691 0.05618568  
L\_epsilon2\_z -0.61385674 0.04007210  
logkappa2 -3.45977764 0.11475488  
logSigmaM -0.35926265 0.02454837  
Maximum gradient component: 5.326118e-08

Table E12 (file = Table\_for\_SS3a.csv). Population index output. Biomass estimates for total GE65 mm, in metric tons, log-SD and SD in metric tons.

Year	Estimate metric tons	SD log	SD mt
1975	218754.39	0.1358	23271.11
1976	314111.29	0.1152	28865.53
1977	386129.11	0.1045	31892.88
1978	329878.97	0.1055	27723.90
1979	156699.97	0.1073	13620.71
1980	207590.45	0.1113	18672.42
1981	128568.60	0.1119	11563.99
1982	122784.92	0.1212	12023.72
1983	47552.03	0.1156	4445.68
1984	77405.14	0.1497	9475.66
1985	37762.70	0.1156	3521.08
1986	33697.07	0.1186	3269.71
1987	64163.18	0.1206	6275.24
1988	55005.67	0.1277	5634.27
1989	51208.71	0.1249	5137.01
1990	54859.72	0.1305	5797.54
1991	48655.55	0.1466	6050.62
1992	38260.57	0.1232	3852.48
1993	52436.90	0.1324	5693.26
1994	34174.79	0.1316	3661.82
1995	33740.79	0.1223	3394.04
1996	43935.26	0.1339	4801.85
1997	76636.25	0.1303	8126.61
1998	83045.51	0.1205	8106.89
1999	71976.39	0.1365	7922.89
2000	81405.59	0.1479	9571.48
2001	53605.54	0.1195	5176.42
2002	71258.23	0.1292	7461.41
2003	104859.91	0.1263	10741.51
2004	101626.77	0.1241	10273.17
2005	99198.04	0.1144	9302.95
2006	108514.28	0.1183	10348.81
2007	115708.62	0.1109	10352.46
2008	112276.57	0.1277	11469.91
2009	81248.65	0.1296	8415.20
2010	81562.84	0.1270	8332.95
2011	69928.79	0.1293	7323.51
2012	67635.30	0.1422	7715.65
2013	65508.40	0.1421	7464.34
2014	124659.81	0.1322	13263.56
2015	72566.87	0.1270	7419.99
2016	63672.67	0.1153	5944.26
2017	57769.97	0.1111	5184.21
2018	32083.29	0.1157	3049.00
2019	32470.03	0.1134	3019.25

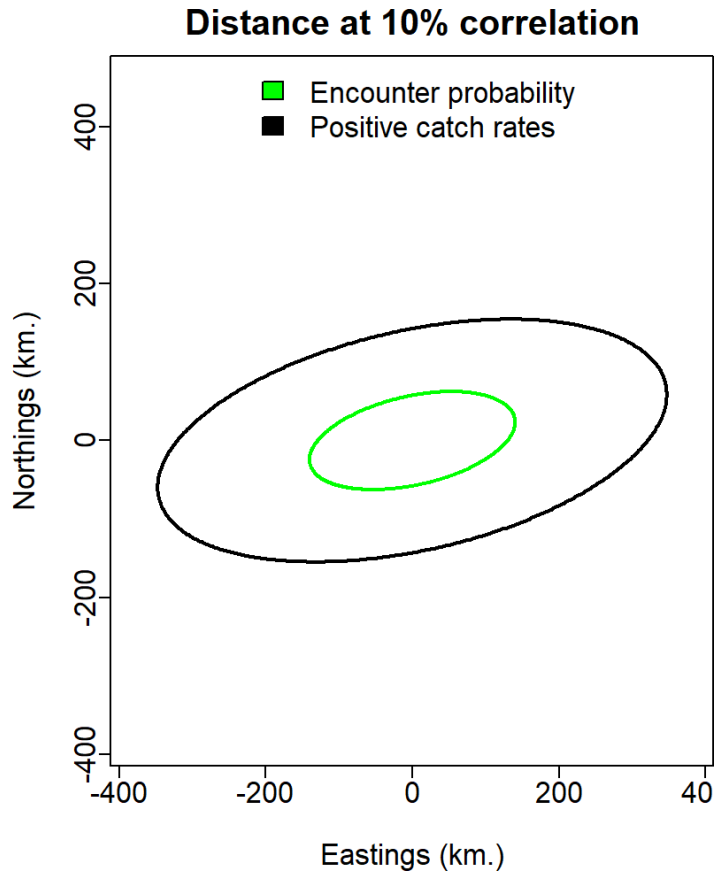


Figure E19 (file = Aniso.png). Geometric anisotropy for total GE65mm. Geometric anisotropy represents the tendency for correlations to decline faster in some directions than others. An ellipse with a major (long) axis pointed northeast-southwest will have correlations that decline slower along this axis than movement northwest-southeast.

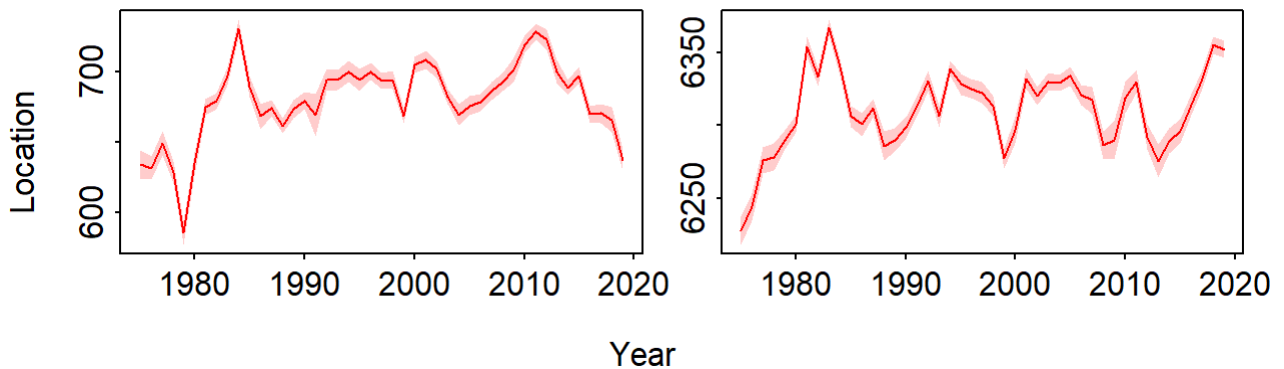


Figure E20 (file = center\_of\_gravity.png). Center of gravity for total GE65 mm: a.) Eastings (km) and b.) Northings (km). These figures indicate that the population has been shifting slightly to the northwest in recent years.

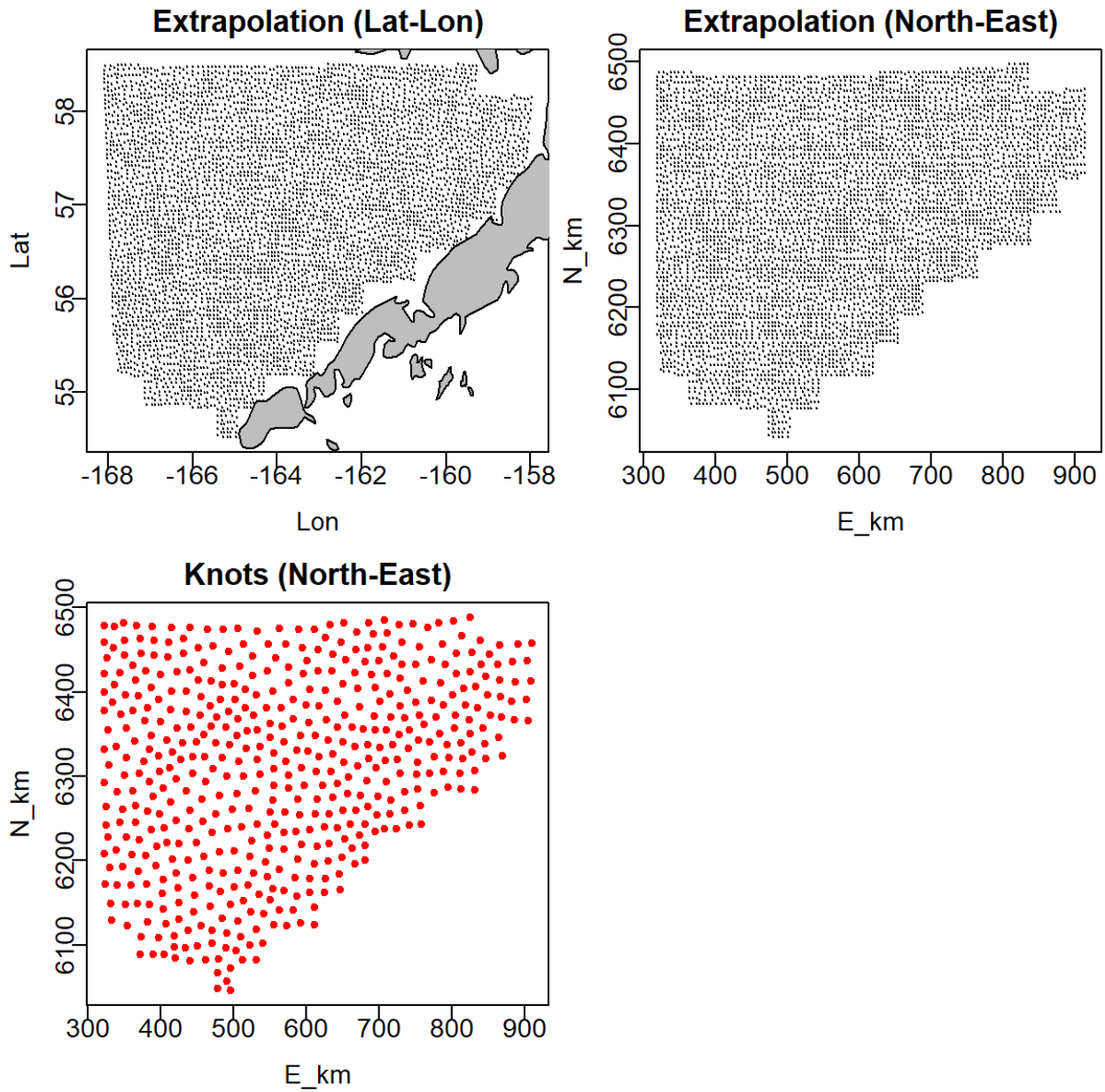


Figure E21 (file = Data\_and\_knots.png). Extrapolation maps for total GE65 mm: a.) Extrapolation grids by latitude and longitude, b.) Extrapolation grid by Northings/Eastings (in km), and c.) Knots used to define spatial mesh for population estimation.

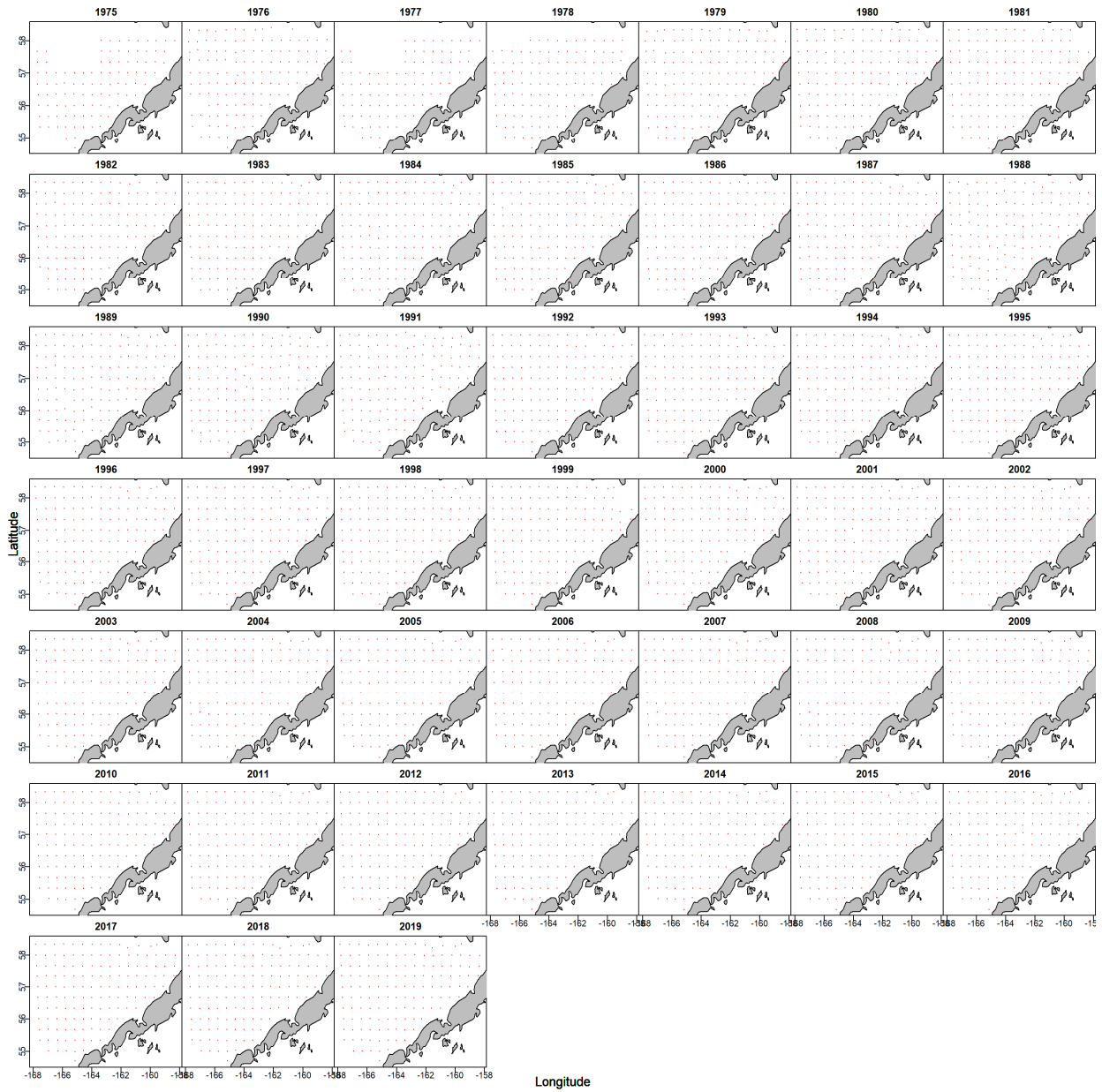


Figure E22 (file = Data\_by\_year.png). Data sampling locations by year for total GE65 mm. Note standardization of survey area in this subregion of the Bering Sea by 1980.



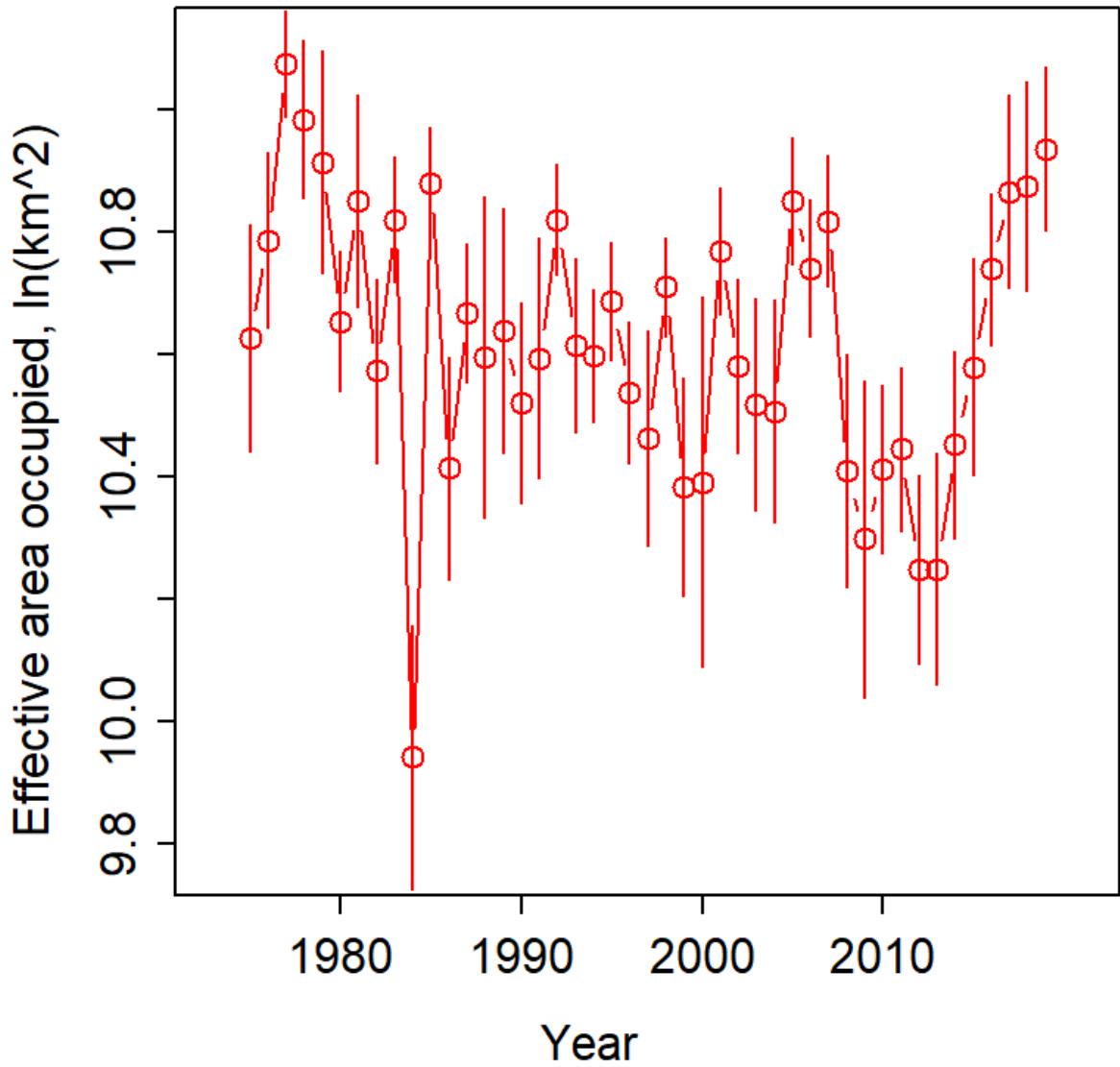


Figure E23 (file = Effective\_Area.png). Effective area occupied, in ln(km<sup>2</sup>) for total GE65 mm. Note modest expansion in area occupied in recent years, despite low population sizes.

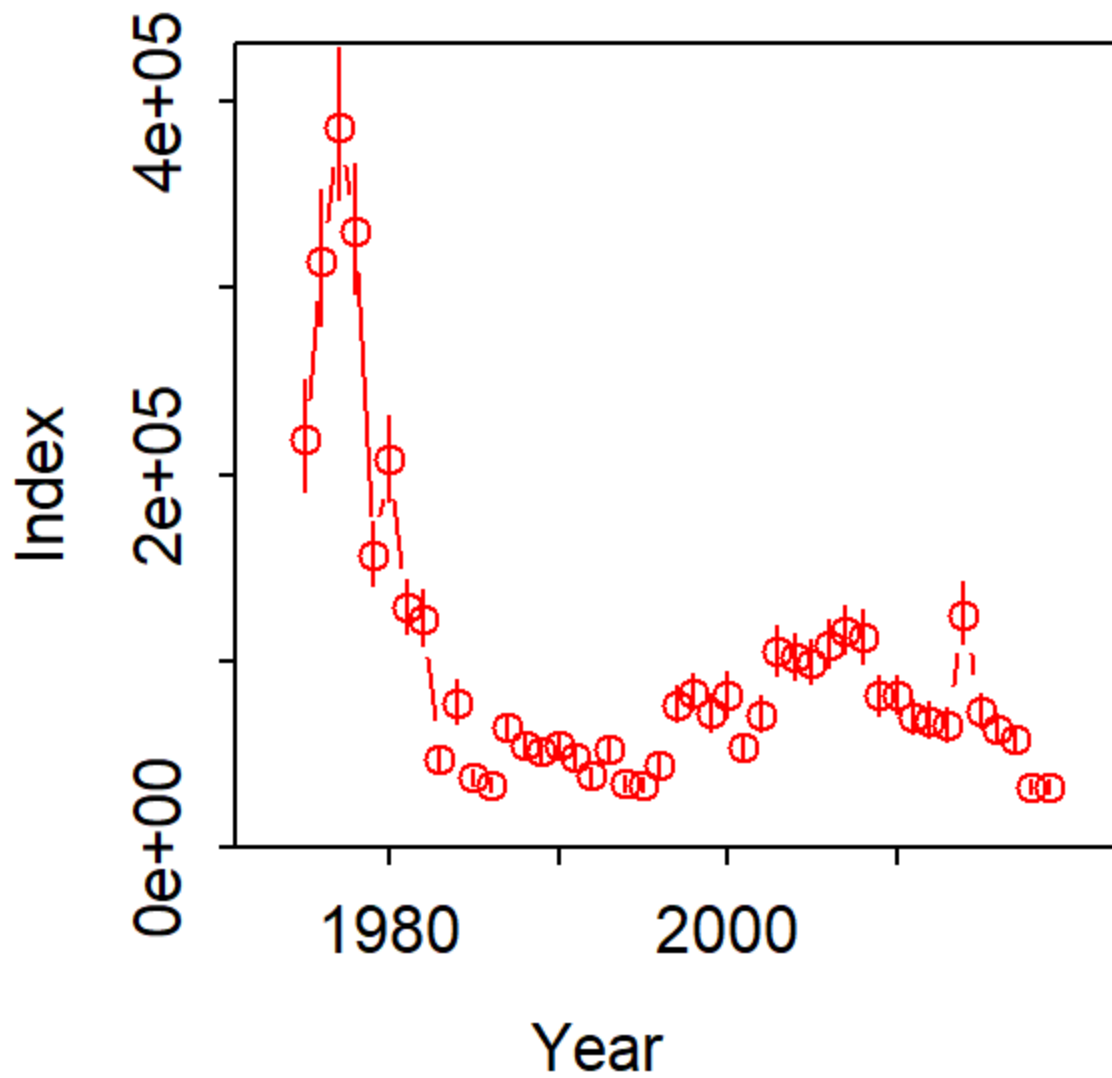


Figure E24 (file = Index-Biomass.png). Biomass estimates by year, with +/- 1SD error bars for total GE65 mm.

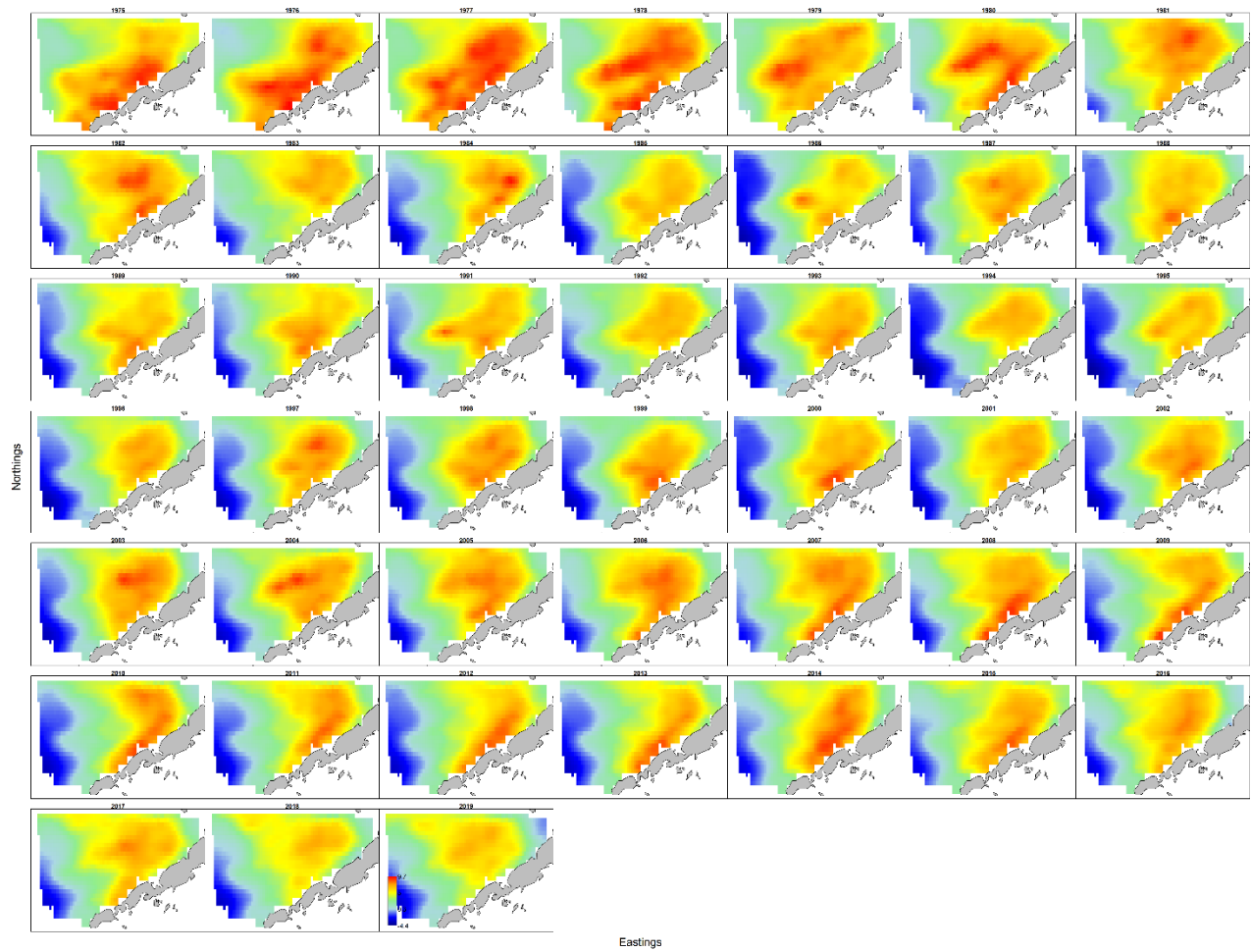


Figure E25 (file = ln\_density-predicted.png). Annual log-density maps for total GE65 mm in  $\ln(\text{kg})/\text{km}^2$ .

DHARMA residual diagnostics

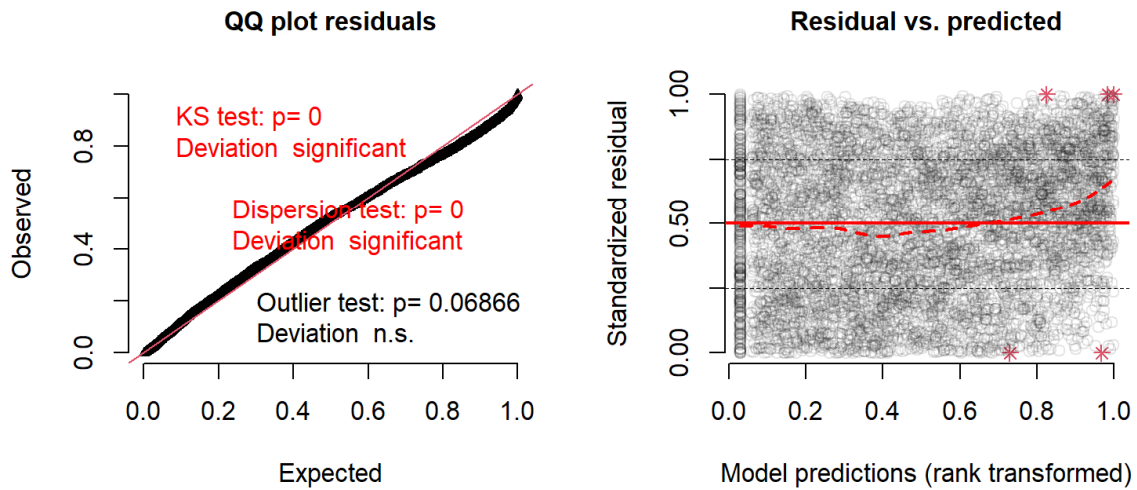


Figure E26 (file = quantile\_residuals). DHARMA residual output for total GE65 mm. a.) Q-Q plot of residuals with p-values for statistical test applied automatically by DHARMA package. These p-values have no meaning here and should be ignored. This figure indicates excellent model performance. b.) Rank-transformed model predictions vs. standardized residuals. Note an indication of a trend in residuals for largest model predictions, as indicated by the dashed red line.

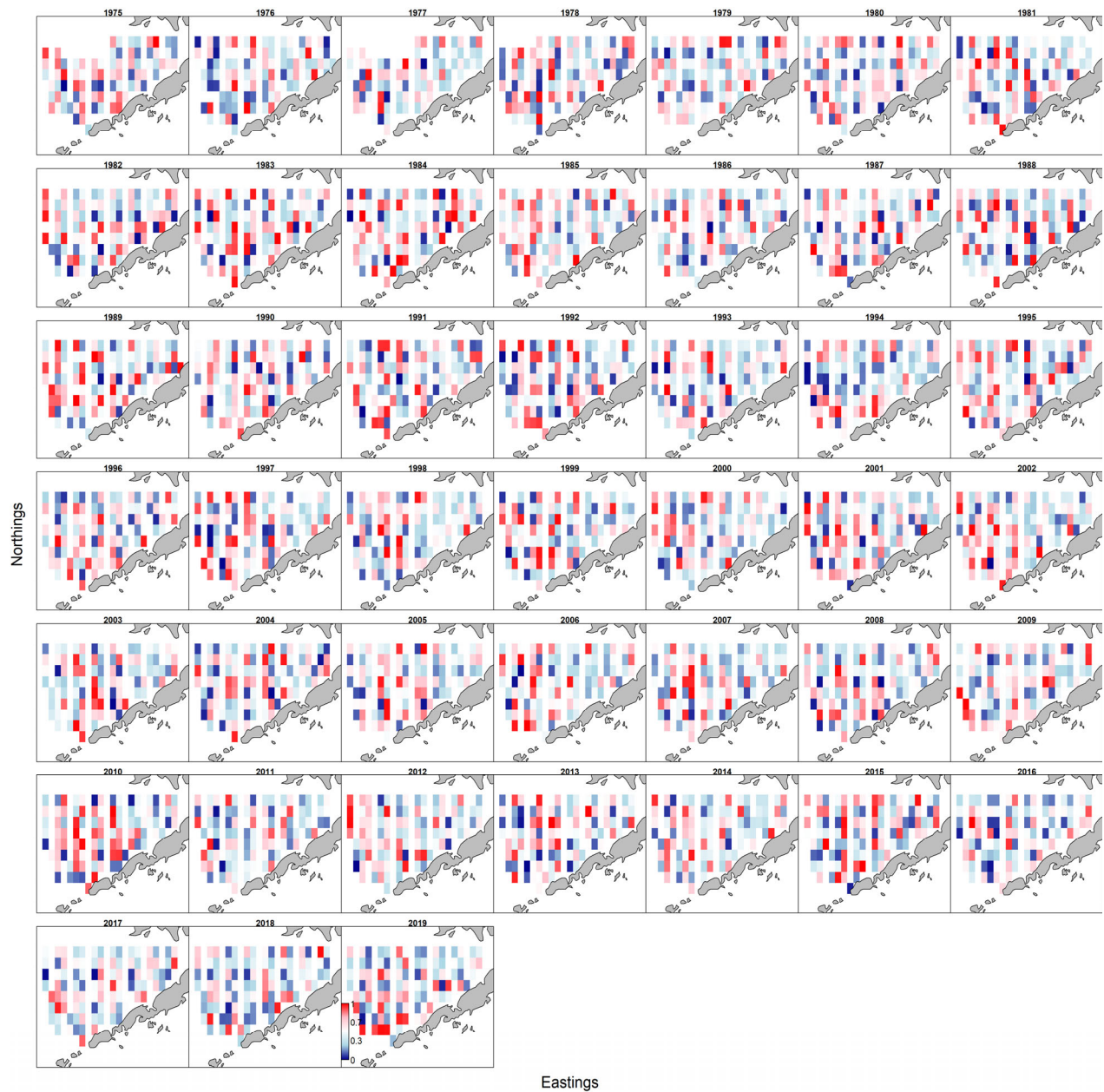


Figure E27 (file = quantile\_residuals\_on\_map.png). Spatial quantile residuals for total GE65 mm with residual “blocks” enlarged for easier viewing. Mapped quantile residuals used to assess for spatial-temporal trends in residuals.